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STORMWATER CHARACTERISTICS AS DESCRIBED IN THE

NATIONAL STORMWATER QUALITY DATABASE

by

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A DISSERTATION

Submitted in partial fulfillment of the requirements
for the degree of Doctor of Philosophy in the
Department of Civil and Environmental
Engineering in the Graduate School
of The University of Alabama

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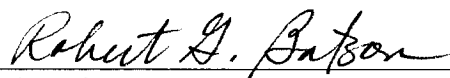Submitted by Alexander Alberto Maestre Martinez in partial fulfillment of the requirements for the degree of Doctor of Philosophy specializing in Civil / Environmental Engineering.
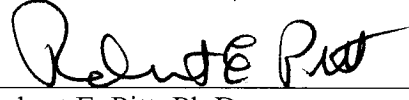
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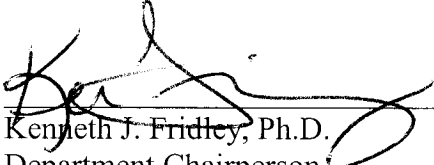
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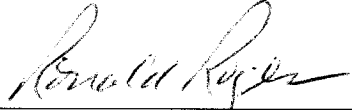
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ii

DEDICATION

To God, that always will guide the steps of human beings.

LIST OF ABBREVIATIONS AND SYMBOLS

µg/L	Microgram per liter
ANOVA	Analysis of Variance
ASCE	American Society of Civil Engineers
BMP	Best Management Practice
BOD_5	Biochemical Oxygen Demand
Cd	Cadmium
CDM	Camp Dresser and McKee
CO	Commercial Land Use
COMIX	Mixed Commercial Land Use
COV	Coefficient of Variation
CFR	Code of Federal Regulation
Cr	Chromium
Cu	Copper
CWA	Clean Water Act
DCIA	Directly Connected Impervious Area
DO	Dissolved Oxygen
EIA	Effective Impervious Area
EMC	Event Mean Concentration

iv

EPA	Environmental Protection Agency
FA	Fall
FHWA	Federal Highway Administration
FW	Freeway Land Use
FWMIX	Mixed Freeway Land Use
ID	Industrial
IDMIX	Mixed Industrial Land Use
IS	Institutional
mg/L	Milligrams per Liter
MLE	Maximum Likelihood Estimator
MS4	Municipal Separate Storm Sewer System
NCASI	National Council of the Paper Industry for Air and Stream Improvement
NCHRP	National Cooperative Highway Research Program
Ni	Nickel
NPDES	National Pollutant Discharge Elimination System
NSQD	National Stormwater Quality Database
NURP	Nationwide Urban Runoff Program
OP	Open Space
OPMIX	Mixed Open Space Land Use
Pb	Lead
pH	pH
QA/QC	Quality Assurance and Quality Control
RE	Residential

REMIX	Mixed Residential Land Use
Rv	Volumetric Runoff Coefficient
SCS	Soil Conservation Service (Natural Resources Conservation Service)
SP	Spring
SS	Suspended Solids
SU	Summer
TAWMS	Toronto Area Watershed Management Strategy Study
TDS	Total Dissolved Solids
TKN	Total Kjeldahl Nitrogen
TPH	Total Petroleum Hydrocarbon
TSS	Total Suspended Solids
TR-55	Technical Report Number 55
USGS	Unite States Geological Survey
WI	Winter
Zn	Zinc

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CONTENTS

Chapter

1. INTRODUCTION

2. THE NATIONAL STORMWATER QUALITY DATABASE

LIST OF TABLES

LIST OF FIGURES

ABSTRACT

The National Stormwater Quality Database v. 1.1 (NSQD) contain water quality characteristics from the monitoring required by the NPDES Phase 1 stormwater permit applications and subsequent permits, during the period of 1992 to 2002. This database contains about 3,765 events from 360 sites in 65 communities from throughout the U.S. For each site, more additional data, including the percentage of each land use in the catchment, the total area, the percentage of imperviousness, the geographical location and the season, has been included in the database. Information about the characteristics of each event is also included. Total precipitation, precipitation intensity, total runoff and antecedent dry period are also included, if collected. The database only contains data collected at the outfall; in-stream samples were not included in the database.

The first phase of the federal stormwater permit program was first published in the Federal Register by the EPA in 1987 and was initially applied to large cities (>100,000 in population), while Phase II of the stormwater permit program was applied to all urban areas as of early 2003. This program requires significant changes in how stormwater is to be managed. Historical approaches only examined drainage issues, while the new regulations also required consideration of water quality issues. Unfortunately, there are a number of commonly accepted notions that are used by stormwater managers and regulators that can have major impacts on local costs and program effectiveness. This dissertation will examine a number of these potential misconceptions to see how well

they hold up under a comprehensive set of actual monitoring data collected throughout

the U.S. as part of the Phase I stormwater permit program.

Chapter 1
Introduction

1.1 Introduction

The first phase of the federal stormwater permit program was first published in the Federal Register by the EPA in 1987 and was initially applied to large cities (>100,000 in population), while Phase II of the stormwater permit program was applied to all urban areas as of early 2003. This program requires significant changes in how stormwater is to be managed. Historical approaches only examined drainage issues, while the new regulations also required consideration of water quality issues. Unfortunately, some professionals involved with the stormwater permit program may not have an adequate understanding of stormwater characteristics, including its effects, and treatability. As an example, there are a number of commonly accepted notions that are used by stormwater managers and regulators that can have major impacts on local costs and program effectiveness. This dissertation will examine a number of these potential misconceptions to see how well they hold up under a comprehensive set of actual monitoring data collected throughout the U.S. as part of the Phase I stormwater permit program. This dissertation also includes a predictive tool that can assist stormwater managers in predicting expected stormwater conditions for local areas.

Researchers from the University of Alabama and the Center for Watershed Protection assembled a large database of stormwater characteristics, the National Stormwater Quality Database (NSQD), as part of a recent EPA-funded 104(b)3 project from the Office of Water. This is the largest collection of information on stormwater characteristics ever assembled. The research described in this dissertation used this information to test the validity of several commonly accepted notions, and will produce a statistical tool to enable managers and regulators to develop more appropriate regulations and strategies for the control of stormwater. In addition, many suggestions concerning monitoring strategies for stormwater are summarized, based on the experiences of many of the Phase I permitted communities.

The University of Alabama and the Center for Watershed Protection were awarded an EPA Office of Water 104(b)3 grant in 2001 to collect and evaluate stormwater data from a representative number of NPDES (National Pollutant Discharge Elimination System) MS4 (Municipal Separate Storm Sewer System) stormwater permit holders. These data were collected and reviewed to provide guidance for future sampling needs, and to enhance local stormwater management activities in areas having limited data. The cumulative value of the monitoring data collected over nearly a ten-year period from more than 200 municipalities throughout the country has a great potential in characterizing the quality of stormwater runoff and comparing it against historical benchmarks.

The data set received a comprehensive quality assurance/quality control review, based on reasonableness of data, extreme values, relationships among parameters, sampling methods, and a review of the analytical methods. The statistical analyses were

conducted at several levels. Probability plots were used to identify range, randomness and normality. Multivariate analyses were also being utilized to characterize significant factors affecting the data patterns. The master data set was also evaluated to develop descriptive statistics, such as measures of central tendency and standard errors. Testing was done for regional and climatic differences, the influences of land use, and the effects of storm size, drainage area and season, among other factors.

This National Stormwater Quality Database (NSQD), in its first version, is not intended for comprehensive characterization purposes for all conceivable situations. Some communities may have obviously unusual conditions, or adequate data may not be available in the database for their region. In these conditions, local outfall monitoring may be needed. In addition, stormwater monitoring will continue to be needed for other purposes in many areas having, or anticipating, active stormwater management programs (especially when supplemented with other biological, physical, and hydrologic monitoring components). These new monitoring programs should be designed specifically for additional objectives, beyond characterization. These may include receiving water assessments to understand local problems, source area monitoring to identify critical sources, treatability tests to verify performance of stormwater controls for local conditions, and assessment monitoring to verify the success of local stormwater management approaches (including model calibration and verification). In many cases, however, the resources being spent for conventional outfall monitoring could be more effectively spent to better understand many of these other aspects of an effective stormwater management program.

1.2 Research Hypotheses

This dissertation will address the following hypothesis, using the National Stormwater Quality Database:

Commonly accepted notions used by the stormwater regulatory and management community are correct for a wide range of conditions, including land use, precipitation depth, season, watershed area and geographic location throughout the United States.

Examples of commonly accepted notions tested in this dissertation are:

1. *Lognormal distributions are robust descriptions of stormwater quality data and a few unusually high values have little effect on data set descriptions.*

2. *Censored data can be adequately adjusted by substituting half of the detection limit so there is little resulting effect on the mean and variance of stormwater data sets.*

3. *Different levels of imperviousness are more important than differences in land use categories on stormwater constituent concentrations.*

4. *Antecedent dry periods have a significant effect on stormwater constituent concentrations.*

5. *Outfall samples collected during the "first flush" periods of storms have significantly greater concentrations than whole storm composite samples.*

Total suspended solids, COD, total nitrogen, total phosphorus, total copper, total lead, and total zinc, amongst other constituents will be analyzed in this dissertation.

1.3 Dissertation Organization

This dissertation is divided into eight chapters and five appendices. Chapter 2 describes the National Stormwater Quality Database (NSQD). Chapter 3 describes the QA/QC procedures used during the collection of data and creation of the database, including an evaluation of alternative methods to address the presence of non-detected values. Chapter 4 addresses the hypothesis concerning the probability distributions most appropriate for the stormwater constituents. Chapter 5 describes the results of the investigations relating constituent concentrations to main factors and interactions of parameters described in the site description and hydrologic information sections of the database. Chapter 6 presents the results from the "first flush" analysis. Chapter 7 presents an example of how the data in the NSQD can be used to estimate the concentration of stormwater constituents for Maryland and Virginia (the region best represented in the database). Chapter 8 presents the conclusions and recommendations of this dissertation.

Chapter 2
The National Stormwater Quality Database

2.1 Introduction

The National Stormwater Quality Database (NSQD) was prepared by the
University of Alabama and the Center for Watershed Protection under 104(b)3 funding
from the U.S. Environmental Protection Agency (EPA). The NSQD is a spreadsheet and
supporting documents describing the monitoring efforts of 65 communities from
throughout the U.S. that are larger than 100,000. The monitoring period covered by the
NSQD is from 1992 to 2002.

Several efforts have been performed in the past to describe the water quality
characteristics of stormwater constituents at different locations. The importance of this
EPA-sponsored project is based on the scarcity of nationally summarized and accessible
data from the existing U.S. EPA's NPDES (National Pollutant Discharge Elimination
System) stormwater permit program. There have been some local and regional data
summaries, but little has been done with nationwide data. A notable exception is the
Camp, Dresser, and McGee (CDM) national stormwater database (Smullen and Cave
2002) that combined historical Nationwide Urban Runoff Program (NURP) (EPA 1983),
available urban U.S. Geological survey (USGS), and selected NPDES data. Their main
effort had been to describe the probability distributions of these data (and corresponding
EMCs, the event mean concentrations). They concluded that concentrations for different

6

land uses were not significantly different, so all their data were pooled into a single category.

Between 1978 and 1983, the EPA conducted the Nationwide Urban Runoff Program that examined stormwater quality from separate storm sewers in different land uses (EPA 1983). This program studied 81 outfalls in 28 communities throughout the U.S. and included the monitoring of approximately 2,300 storm events. The data was presented for several land use categories, although most of the information was obtained from residential lands. Since NURP, other important studies have been conducted that characterize stormwater. The USGS created a database with more than 1,100 storms from 98 monitoring sites in 20 metropolitan areas. The Federal Highway Administration (FHWA) analyzed stormwater runoff from 31 highways in 11 states during the 1970s and 1980s. Strecker (personal communication) is also collecting information from highway monitoring as part of a current NCHRP (National Cooperative Highway Research Program) funded project. The city of Austin also developed a database having more than 1,200 events.

Other regional databases also exist for U.S. data, mostly using local NPDES data. These include the Los Angeles area database, the Santa Clara and Alameda County (California) databases, the Oregon Association of Clean Water Agencies Database, and the Dallas, Texas, area stormwater database. These regional data are included in the NSQD. However, the USGS and historical NURP data are not included in the NSQD due to lack of consistent descriptive information for the older drainage areas and because of the age of the data from those prior studies. Much of the NURP data is available in

electronic form at the University of Alabama student American Water Resources Association web page at: http://www.eng.ua.edu/~awra/download.htm.

Outside the U.S., there have been important efforts to characterize stormwater. In Toronto, Canada, the Toronto Area Watershed Management Strategy Study (TAWMS) was conducted during 1983 and 1984 and extensively monitored industrial stormwater, along with snowmelt in the urban area, for example. Numerous other investigations in South Africa, the South Pacific, Europe and Latin America have also been conducted over the past 30 years, but no large-scale summaries of that data have been prepared. About 4,000 international references on stormwater have been reviewed and compiled since 1996 by the Urban Wet Weather Flows literature review team for publication in *Water Environment Research* (most recently by Clark, *et al.* 2001, 2002, 2003, 2004). An overall compilation of these literature reviews is available at: http://www.eng.ua.edu/~rpitt/Publications/Publications.shtml. These reviews include short summaries of the papers and are organized by major topics. Besides journal articles, many published conference proceedings are also represented (including the extensive conference proceedings from the 8th International Conference on Urban Storm Drainage held in Sydney, Australia, in 1999, the 9th International Conference on Urban Storm Drainage held in Portland, OR, in 2002, and the Toronto Stormwater and Urban Water Systems Modeling conference series, amongst many other specialty conferences).

The Clean Water Act (CWA) of 1972 was the first major national regulation in the U.S. requiring control of conventional point source discharges of water pollutants (affecting municipal and industrial discharges). Section 208 also provided the capability to implement stormwater management plans at the regional level. In 1976, the EPA

enlarged the planning initiative through the "Section 208: Areawide Assessment

Procedures Manual". However, in the late 1970s, some problems arose with the 208

planning projects due to inadequate data and lack of technological development

(Whipple, as quoted by Pitt et al. 1999).

NURP is still an important reference for water quality characteristics of urban

stormwater; however, the collected data poorly represented the southern part of the

country and was focused mainly in residential and mixed land use areas. There are other

important studies that characterized stormwater quality. The USGS created a database

containing more than 1100 storms at 98 sites in 20 metropolitan areas summarizing their

monitoring projects for Phase I communities (Smullen and Cave 2003).

In 1987, the amendments to the CWA established a two-phase program to

regulate 13 classes of stormwater discharges. Two of these classifications were

discharges from large and medium-sized Municipal Separate Storm Sewer Systems. A

large MS4 serves an urban population of 250,000 or more, while a medium MS4 serves

communities between 100,000 and 250,000. EPA set up a permit strategy for

communities complying with NPDES requirements. Monitoring data from this program

have been included in some databases. The CDM National Stormwater Runoff Pollution

database included 816 NPDES storm events in a database that totals approximately 3,100

events. The Rouge River National Wet Weather Demonstration Program office in Detroit

included their NPDES data in their database (Smullen and Cave 2003).

Another important effort has been the developing of the National Stormwater Best

Management Practices Database. This database was created with the purpose to evaluate

the performance and effectiveness of stormwater control practices, frequently labeled

"best management practices," or BMP's. Detention ponds, street cleaning, and hydrodynamic devices are examples of BMPs (ASCE/EPA 2000).

In 2001, the EPA granted the University of Alabama and the Center for Watershed Protection to collect and evaluate NPDES MS4 municipal stormwater permits. This project collected 3,765 storm events from 65 agencies and municipalities in 17 states. This database (National Stormwater Quality Database) includes geographic and seasonal information that can be useful for various analyses (Pitt et al. 2003).

2.2 Data Collection

A total of 3,765 storm events at 360 monitoring sites were collected and stored in version 1.1 of the NSQD. This version contains the results of approximately one third of the total number of communities that participated in the Phase I NPDES stormwater permit monitoring activities.

According to the published sampling guidance (40 CFR 122.21) for the permit application, each community was required to sample at least a residential, a commercial and an industrial watershed. At least three samples should be collected every year at each location. Each storm should be at least one month apart and have at least a 3 days antecedent dry period. Only samples from rain events greater than 0.1 inches, and close to the annual mean conditions, were considered valid for the analysis. It was required to collect a composite sample with subsamples collected during the first three hours of the event. An additional grab sample was required during the first 30 minutes of the event to evaluate the "first flush" effect. "First flush" refers to the hypothesis that the concentration of the stormwater constituents is higher at the beginning of the discharge

than during the complete event. Designated states were able to modify some of these sampling requirements to better address local concerns.

Most communities were required to submit annual reports describing the sampling locations and procedures, the equipment, and the quality control and quality assurance (QA/QC) procedures used during the sampling and analysis of the samples, the analytical methods used in the laboratory, and problems encountered during the sample collection. The reports also included the results of the chemical analysis performed by the laboratories.

Figure 1 is a map showing the 65 communities and 17 states included in the first version of the NSQD. The EPA-funded project was intended to focus on the Chesapeake Bay area and parts of the southern U.S. (specifically Birmingham, AL, and Atlanta, GA) as a demonstration of the usefulness of the data. However, it was possible to obtain some data from other parts of the country during the project period and these data were incorporated in the database, allowing some regional analyses. States representing most of the samples included Virginia (24%) and Maryland (13%). The states with low numbers of observations included Pennsylvania, Massachusetts, and Indiana.

Figure 1 also shows the EPA Rain Zones. Each zone corresponds to a geographical region with similar climatic conditions (EPA 1986). There is at least one community per rainfall zone indicating some geographical representation for the entire country. However, Table 1 indicates that most of the samples were collected west, south and east of the continental part of the country, with few of the large amounts of data from EPA Rain Zone 1 included in the database. EPA Rain Zones 8 and 9 have sparse available data from the Phase I monitoring program, due to few large cities in these areas.
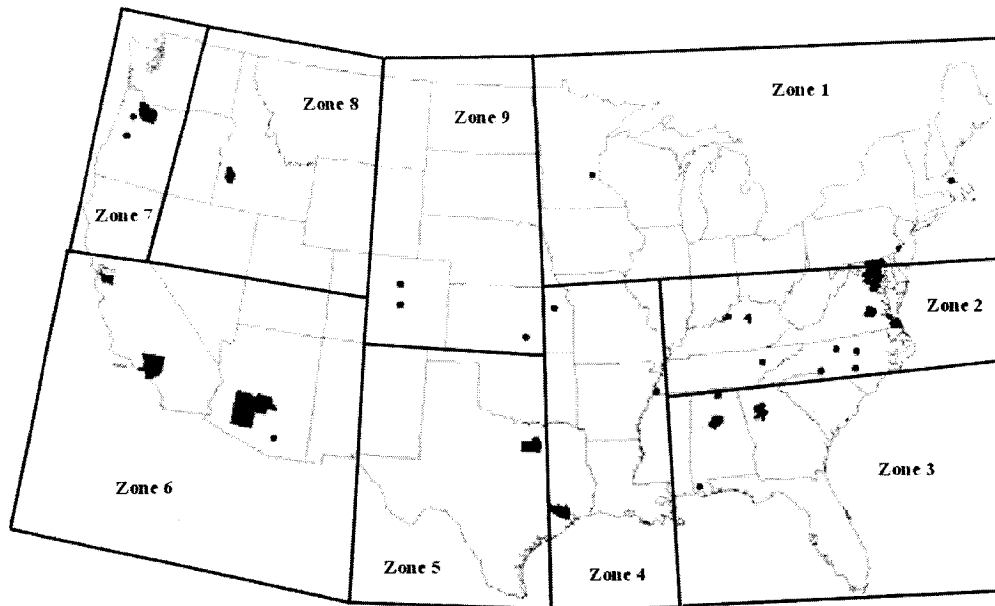
Fig. 1. Communities included in the NSQD version 1.1 by rainfall zones

Table 1. Total Samples and Sites by EPA Rain Zone

EPA Rain Zone	Total Samples	Percentage of Samples	Number of Communities	Number of Sites
1	69	1.83	2	12
2	2000	53.12	28	185
3	266	7.07	8	30
4	212	5.63	4	21
5	485	12.88	9	33
6	356	9.46	4	30
7	229	6.08	6	28
8	24	0.63	1	4
9	124	3.29	3	17

Each site in the database corresponds to an outfall where the runoff produced in the watershed is discharged. During the monitored events, samples were collected to identify the characteristics of the stormwater discharge. According to the activity performed in the watershed, each site was classified as residential, commercial,

industrial, open space, freeway, or mixed. When a single activity was not identified for

the watershed, then the site was considered mixed, with a predominant land use. Table 2

indicates the total number of sites included in the database, separated by land use.

Table 2. Total Samples and Sites by Land Use

Land use	Number of Sites	Percentage	Number of Events	Percentage
Residential	111	30.8	1042	27.7
Mixed Residential	44	12.2	611	16.2
Commercial	51	14.2	526	14
Mixed Commercial	29	8.1	325	8.6
Industrial	54	15.0	566	15.0
Mixed Industrial	22	6.1	249	6.6
Institutional	1	0.3	18	0.5
Open Space	10	2.8	49	1.3
Mixed Open Space	13	3.6	168	4.5
Freeways	22	6.1	185	4.9
Mixed Freeways	3	0.8	26	0.7

About one third of the sites included in the database correspond to residential

areas, another third is shared by commercial and industrial land uses. The remaining third

correspond to freeways, open space, institutional and all the mixed land uses. Several

schools were identified in the sites, however only one site was considered 100%

institutional.

2.3 Summary of U.S. NPDES Phase I Stormwater Data in the NSQD

Table 3 is a summary of selected data collected and entered into the database. The

data are separated into 11 land use categories: residential, commercial, industrial,

institutional, freeways, and open space, plus mixtures of these land uses. Summaries are

shown for the major land use areas and for the total data set combined. The full database

includes all of the data. The total number of observations and the percentage of observations above the detection limits are also shown on this summary table. In general, the coefficient of variation (COV) range from 1.0 to 2.0 for the majority of pollutants across all major land uses.

The following sections describe the structure of the full database and present some findings. The findings presented are focused on specific issues and are illustrated using small portions of the complete database to minimize the effects of other interacting factors (such as using data from a single region and land use to show the effects of sampling methods, for example). Later sections of this dissertation present more comprehensive discussions of the data that do consider interactions of the many factors available in the database.

2.4 Database Structure

The database has five major sections: General Information, Items Description, Constituents and Parameters, and the Database itself. In addition, detailed site information along with aerial photographs and topographic maps is provided for each municipality and monitoring location. Each of the sections is a tab in the bottom part of the spreadsheet.

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1

	Area (acres)	% Impervious	Precipitation Depth (in)	Runoff Depth (in)	Conductivity (μS/cm @25°C)	Hardness (mg/L CaCO3)	Oil and Grease (mg/L)	pH	Temperature (C)	TDS (mg/L)	TSS (mg/L)	BOD$_5$ (mg/L)	COD (mg/L)
Overall Summary (3765)													
Number of observations	3765	2209	3316	1495	685	1082	1834	1665	861	2956	3493	3105	2750
% of samples above detection	100	100	100	100	100	98.7	66.1	100	100	99.0	97.9	96.2	98.4
Median	57.3	50.0	0.48	0.15	121	38.0	4.3	7.5	16.5	80	59	8.6	53
Coefficient of variation	3.7	0.4	1.0	1.9	1.6	1.4	9.7	0.1	0.4	3.4	1.8	7.4	1.1
Residential (1042)													
Number of observations	1042	614	919	372	104	215	483	286	181	814	978	908	748
% of samples above detection	100	100	100	100	100	100	54.9	100	100	99.1	98.3	97.1	98.7
Median	57.3	37.0	0.48	0.10	102	32.0	4.0	7.2	17.0	72.0	49	9.0	54.5
Coefficient of variation	4.8	0.4	1.0	1.5	1.6	1.1	7.8	0.1	0.4	1.1	1.8	1.5	0.93
Mixed Residential (611)													
Number of observations	611	278	491	262	105	168	283	333	137	491	582	549	465
% of samples above detection	100	100	100	100	100	98.2	70.3	100	100	99.2	98.3	94.2	99.6
Median	150.8	44.9	0.53	0.12	112	40.0	4.0	7.50	15.5	86	66	7.8	43
Coefficient of variation	2.1	0.3	0.8	1.3	1.2	1.1	2.6	0.1	0.3	5.2	1.6	1.3	1.2
Commercial (527)													
Number of observations	527	284	462	146	78	156	331	191	98	418	503	452	393
% of samples above detection	100	100	100	100	100	100	71.9	100	100	99.5	95.2	97.6	98.5
Median	38.8	84.5	0.42	0.29	107	36.5	4.6	7.4	16.0	72	43	11.0	58
Coefficient of variation	1.2	0.1	1.0	1.0	1.0	1.1	3.0	0.1	0.4	1.9	2.0	1.1	1.0
Mixed Commercial (324)													
Number of observations	324	237	305	118	59	98	134	156	98	265	297	277	267
% of samples above detection	100	100	100	100	100	99.0	79.9	100	100	99.6	99.7	98.9	99.6
Median	75.0	60.0	0.47	0.28	100	36.0	5.0	7.60	14.5	69.5	54.5	9.0	60
Coefficient of variation	1.4	0.3	1.0	0.9	0.8	1.8	2.9	0.1	0.4	1.9	1.3	1.7	1.0
Industrial (566)													
Number of observations	566	292	482	215	102	132	315	248	140	431	521	455	386
% of samples above detection	100	100	100	100	100	96.2	64.8	100	100	99.5	97.7	95.4	99.0
Median	39.5	75.0	0.50	0.16	139	39.0	4.8	7.50	17.9	86	81	9.0	58.6
Coefficient of variation	1.1	0.3	0.9	1.2	1.3	1.5	11.8	0.1	0.3	3.6	1.6	10.0	1.2

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 – *Continued*

	Area (acres)	% Impervious	Precipitation Depth (in)	Runoff Depth (in)	Conductivity (µS/cm @25°C)	Hardness (mg/L CaCO3)	Oil and Grease (mg/L)	pH	Temperature (C)	TDS (mg/L)	TSS (mg/L)	BOD5 (mg/L)	COD (mg/L)
Mixed Industrial (218)													
Number of observations	218	118	193	117	56	75	72	152	57	186	207	178	175
% of samples above detection	100	100	100	100	100	93.3	80.6	100	100	99.5	100	95.5	98.9
Median	168.0	44.0	0.45	0.29	126	29.3	9.0	7.70	18.0	90	82	7.5	39.9
Coefficient of variation	1.8	0.3	0.9	1.2	0.8	0.6	1.8	0.1	0.3	0.8	1.4	1.8	1.2
Institutional (18)													
Number of observations	18	18	17	14						18	18	18	18
% of samples above detection	100	100	100	100						100	94.4	88.9	88.9
Median	36.0	45.0	0.18	0.00						52.5	17	8.5	50
Coefficient of variation	0	0	0.9	2.1						0.7	0.83	0.7	0.9
Freeways (185)													
Number of observations	185	154	182	144	86	127	60	111	31	97	134	26	67
% of samples above detection	100	100	100	100	100	100	71.7	100	100	99.0	99.3	84.6	98.5
Median	1.6	80.0	0.54	0.41	99	34.0	8.0	7.10	14.0	77.5	99	8	100
Coefficient of variation	1.4	0.13	1.1	1.7	1.0	1.9	0.6	0.1	0.4	0.8	2.6	1.3	1.1
Mixed Freeways (26)													
Number of observations	26		26		21	12	20	17	17	15	23	23	15
% of samples above detection	100		100		100	100	100	100	100	100	100	100.0	100.0
Median	63.1		0.47		353	83	4.5	7.7	16.0	177	88	8.2	47
Coefficient of variation	0.7		0.8		0.6	0.3	1.8	0.1	0.3	0.4	1.1	1.2	0.5
Open Space (49)													
Number of observations	49	37	41	11	2	8	19	19	2	45	44	44	43
% of samples above detection	100	100	100	100	100	100	36.8	100	100	97.8	95.5	86.4	76.74
Median	85	2.0	0.52	0.05	113	150	1.3	7.70	14.6	125	48.5	5.4	42.1
Coefficient of variation	1.5	1.0	1.2	1.4	0.5	0.6	0.7	0.08	0.7	0.7	1.5	0.7	1.5
Mixed Open Space (168)													
Number of observations	168	131	167	93	65	70	90	128	76	148	153	145	145
% of samples above detection	100	100	100	100	100	100	60.0	100	100	99.3	97.4	96.6	96.6
Median	115.4	33.0	0.51	0.10	215	64.2	8.5	7.9	16.0	109	78.0	6.0	34
Coefficient of variation	0.8	0.4	0.8	1.2	1.7	1.3	1.5	0.1	0.3	2.2	1.6	2.7	1.6

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 – Continued

	Fecal Coliform (mpn/100 mL)	Fecal Streptococcus (mpn/100 mL)	Total Coliform (mpn/100 mL)	Total E. Coli (mpn/100 mL)	NH3 (mg/L)	NO2+NO3 (mg/L)	Nitrogen, Total Kjeldahl (mg/L)	Phosphorus, filtered (mg/L)	Phosphorus, total (mg/L)	Sb, total (µg/L)	As, total (µg/L)	As, filtered (µg/L)	Be, total (µg/L)
Overall Summary (3765)													
Number of observations	1704	1141	83	67	1908	3075	3191	2477	3285	874	1507	210	947
% of samples above detection	91.2	94.0	90.4	95.5	71.3	97.3	95.6	85.1	96.5	7.2	49.9	27.1	7.7
Median	5091	17000	12000	1750	0.44	0.60	1.4	0.13	0.27	3.0	3.0	1.5	0.4
Coefficient of variation	4.6	3.8	2.4	2.3	1.4	0.97	1.2	1.6	1.5	1.7	2.6	1.0	2.5
Residential (1042)													
Number of observations	402	257		14	572	889	922	690	926		395		282
% of samples above detection	87.8	87.9		100	82.2	97.6	96.5	83.5	96.8		40.8		7.8
Median	7000	24300		700	0.31	0.60	1.5	0.18	0.31		3.0		0.5
Coefficient of variation	5.2	1.7		1.6	1.1	1.1	1.1	0.9	1.1		2.2		2.5
Mixed Residential (611)													
Number of observations	336	178	26	11	282	531	517	430	552		158		97
% of samples above detection	94.3	97.8	84.6	90.9	58.5	97.9	95.0	83.3	96.2		65.9		11.3
Median	11210	27500	5667	1050	0.39	0.57	1.4	0.13	0.28		3.0		0.3
Coefficient of variation	3.2	2.1	1.3	2.1	1.6	0.78	1.7	1.1	1.7		3.9		2.7
Commercial (527)													
Number of observations	253	201			300	445	469	343	466		235		
% of samples above detection	88.9	92.5			83.3	98.0	97.4	81.0	95.9		33.6		
Median	4600	12000			0.50	0.6	1.5	0.11	0.22		2.3		
Coefficient of variation	3.0	2.7			1.2	1.1	0.9	1.3	1.2		2.9		
Mixed Commercial (324)													
Number of observations	116	95			173	284	276	221	290	89	139		
% of samples above detection	94.8	98.9			67.1	96.8	96.0	93.7	98.6	11.9	45.5		
Median	5400	11900			0.60	0.58	1.4	0.12	0.26	15.0	2.0		
Coefficient of variation	3.0	2.6			1.0	0.7	0.9	2.1	1.5	1.0	1.0		
Industrial (566)													
Number of observations	315	189			272	461	483	344	478	152	255		197
% of samples above detection	87.3	93.7			78.3	96.3	96.3	88.1	96.2	14.5	52.9		10.7
Median	2400	12000			0.42	0.69	1.4	0.10	0.25	3.7	4.0		0.38
Coefficient of variation	5.7	7.0			1.3	0.92	1.1	1.2	1.4	1.4	1.4		2.5

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 – *Continued*

	Fecal Coliform (mpn/100 mL)	Fecal Streptococcus (mpn/100 mL)	Total Coliform (mpn/100 mL)	Total E. Coli (mpn/100 mL)	NH3 (mg/L)	N02+NO3 (mg/L)	Nitrogen, Total Kjeldahl (mg/L)	Phosphorus, filtered (mg/L)	Phosphorus, total (mg/L)	Sb, total (µg/L)	As, total (µg/L)	As, filtered (µg/L)	Be, total (µg/L)
Mixed Industrial (218)													
Number of observations	79	59	14		99	173	160	179	177		93		
% of samples above detection	98.7	96.9	71.4		30.3	98.8	92.5	84.4	95.5		88.2		
Median	3033	11000	2467		0.58	0.59	1.1	0.08	0.20		3.5		
Coefficient of variation	2.5	2.5	1.5		0.8	0.7	1.5	2.3	1.6		0.9		
Institutional (18)													
Number of observations					18	18	18	17	17				
% of samples above detection					88.9	100	100	82.4	94.1				
Median					0.31	0.6	1.35	0.13	0.18				
Coefficient of variation					0.5	0.6	0.5	0.5	1.0				
Freeways (185)													
Number of observations	49	25	16	13	79	25	125	22	128		61	72	
% of samples above detection	100	100	100	100	87.3	96.0	96.8	95.5	99.2		55.7	50.0	
Median	1700	17000	50000	1900	1.07	0.28	2.0	0.20	0.25		2.4	1.4	
Coefficient of variation	2.0	1.2	1.5	2.2	1.3	1.2	1.4	2.1	1.8		0.7	2.0	
Mixed Freeways (26)													
Number of observations	20	16				22	22	11	22		15		
% of samples above detection	85.0	93.8				100	100	100	100		80		
Median	2600	19000				0.9	2.3	0.03	0.34		3.0		
Coefficient of variation	2.3	1.1				0.7	1.3	0.9	0.7		0.7		
Open Space (68)													
Number of observations	23	22			32	44	45	44	46		19		
% of samples above detection	91.3	90.9			18.8	84.1	71.1	79.6	84.8		31.6		
Median	7200	24900			0.18	0.59	0.74	0.13	0.31		4.0		
Coefficient of variation	1.1	1.0			1.24	0.9	0.9	0.9	3.5		0.4		
Mixed Open Space (168)													
Number of observations	86	75			71	152	123	148	152		88		
% of samples above detection	97.7	100			22.5	97.4	90.2	85.8	96.1		44.3		
Median	3000	21000			0.51	0.7	1.1	0.09	0.25		3.0		
Coefficient of variation	2.3	2.4			1.2	0.8	0.9	1.1	1.1		0.9		

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 – *Continued*

	Cd, total (µg/L)	Cd, filtered (µg/L)	Cr, total (µg/L)	Cr, filtered (µg/L)	Cu, total (µg/L)	Cu, filtered (µg/L)	Pb, total (µg/L)	Pb, filtered (µg/L)	Hg, total (µg/L)	Ni, total (µg/l)	Ni, filtered (µg/L)	Zn, total (µg/L)	Zn, filtererd (µg/L)
Overall Summary (3765)													
Number of observations	2574	389	1598	261	2722	411	2949	446	1014	1430	246	3007	381
% of samples above detection	40.6	30.3	70.2	60.5	87.4	83	77.7	49.8	10.2	59.8	64.2	96.6	96.3
Median	1.0	0.50	7.0	2.1	16	8.0	17.0	3.0	0.20	8.0	4.0	116	52
Coefficient of variation	3.7	1.1	1.5	0.7	2.2	1.6	1.8	2.0	2.5	1.2	1.5	3.3	3.9
Residential (1042)													
Number of observations	695		404		771	90	762	108	275	392	25	784	87
% of samples above detection	31.1		53.2		83.1	63.3	69.4	33.3	6.9	44.1	44.0	96.2	89.7
Median	0.5		4.5		12	7.0	12.0	3.0	0.20	5.6	2.0	73	31.5
Coefficient of variation	3.4		1.2		1.8	2.0	1.9	1.9	0.9	1.2	0.5	1.3	0.8
Mixed Residential (611)													
Number of observations	420	30	193	21	432	29	500	30	115	150	25	515	28
% of samples above detection	34.5	40.0	81.3	52.4	83.8	72.4	78.4	46.7	15.7	60	72.0	92.6	100
Median	0.9	0.30	7.0	2.0	16	5.5	16	3.0	0.20	7.8	5.5	95	48
Coefficient of variation	3.6	0.6	1.5	0.8	1.2	0.9	1.4	0.7	0.8	0.8	0.9	0.9	0.9
Commercial (527)													
Number of observations	379	47	257	27	408	48	399	59	170	242	23	414	49
% of samples above detection	41.7	23.4	60.7	40.7	92.9	79.2	85.5	52.5	6.5	60.3	47.8	99.0	100
Median	0.96	0.30	6.0	2.0	17	7.57	18.0	5.0	0.20	7.0	3.0	150	59
Coefficient of variation	2.7	1.3	1.3	0.6	1.5	0.8	1.6	1.6	0.8	1.2	0.8	1.2	1.4
Mixed Commercial (324)													
Number of observations	188	41	128	27	191	41	244	41		102	26	243	39
% of samples above detection	49.5	34.1	88.3	66.7	93.2	80.5	88.1	63.4		78.4	69.2	98.8	100
Median	0.9	0.35	5.0	2.5	17.5	10	17.0	3.5		5.1	3.5	131.4	73
Coefficient of variation	1.1	0.8	1.1	0.7	3.0	0.6	1.4	0.8		1.3	0.6	1.7	0.8
Industrial (566)													
Number of observations	435	42	250	36	455	42	452	51	199	237	36	473	42
% of samples above detection	49.0	54.8	72.0	55.6	88.6	90.5	75.0	52.9	13.9	61.6	58.3	98.9	95.2
Median	2.0	0.60	12.0	3.0	20.8	8.0	24.9	5.0	0.20	14.0	5.0	199	112
Coefficient of variation	2.2	1.1	1.2	0.7	2.0	0.7	1.9	1.6	2.7	1.0	1.4	1.5	3.6

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 – *Continued*

	Cd, total (µg/L)	Cd, filtered (µg/L)	Cr, total (µg/L)	Cr, filtered (µg/L)	Cu, total (µg/L)	Cu, filtered (µg/L)	Pb, total (µg/L)	Pb, filtered (µg/L)	Hg, total (µg/L)	Ni, total (µg/l)	Ni, filtered (µg/L)	Zn, total (µg/L)	Zn, filtered (µg/L)
Mixed Industrial (218)													
Number of observations	145	25	109	15	150	24	213	25	58	74	15	212	24
% of samples above detection	60.7	92.0	92.7	66.7	90.0	100.0	82.6	92.0	22.4	83.8	100.0	98.6	95.8
Median	1.6	0.60	8.0	2.0	23	6.0	20.0	5.0	0.3	12	5.0	172	2100
Coefficient of variation	1.9	0.6	1.7	0.7	0.8	0.6	1.4	1.0	0.6	0.8	0.6	3.1	1.2
Institutional (18)													
Number of observations							18					18	
% of samples above detection							77.8					100	
Median							5.75					305	
Coefficient of variation							0.8					0.8	
Freeways (185)													
Number of observations	95	114	76	101	97	130	107	126		99	95	93	105
% of samples above detection	71.6	26.3	98.7	78.2	99.0	99.2	100	50.0		89.9	67.4	96.8	99.1
Median	1.0	0.68	8.3	2.3	34.7	10.9	25	1.8		9.0	4.0	200	51
Coefficient of variation	0.9	1.0	0.7	0.7	1.0	1.5	1.5	1.7		0.9	1.4	1.0	1.9
Mixed Freeways (26)													
Number of observations	23		15		23		23					23	
% of samples above detection	56.5		100		100		56.5					100	
Median	0.5		6.0		14		10.0					130	
Coefficient of variation	2.2		1.0		1.0		1.3					0.9	
Open Space (68)													
Number of observations	38		36		39		45					45	
% of samples above detection	55.3		36.1		74.4		42.2					71.1	
Median	0.38		5.4		10		10.0					40	
Coefficient of variation	1.9		1.7		2.0		1.7					1.3	
Mixed Open Space (168)													
Number of observations	107		88		108		155		27	51		156	
% of samples above detection	18.7		81.8		89.8		74.2		14.8	72.5		98.1	
Median	2.0		6.0		9.0		10		0.15	8.0		80	
Coefficient of variation	1.4		1.3		1.0		2.3		0.4	1.1		1.1	

In the General Information tab, the spreadsheet lists the states and municipalities included in the current version of the database. The second tab describes the two main sections of the database: site descriptions and event descriptions. In the items description section, each column in the database is described. The last column in this table shows an example of the value expected in each column. The third tab describes the constituents and parameters included in the database, the number of observations, and the percentage of samples having detected observations. This table is useful to identify those constituents with high percentage of detected values.

The last tab in the database contains the data itself; a matrix of 232 columns by 3,765 rows containing all the data collected and reviewed. Each row represents a storm event for each monitoring location. This part of the table is divided in seven subsections describing the site location, the hydrology of the event and equipment used, and the constituent classifications. Each section of the database is described in the following discussion, with detailed analyses presented in Chapters 4, 5, 6, and 7 of this dissertation.

The following discussion will require a copy of the database for reference. This is available at: http://unix.eng.ua.edu/~rpitt/Research/ms4/mainms4.shtml. Each of the sections and columns included in the spreadsheet will be explained in detail. Summary statistics, probability plots and box and whiskers plots will be used to describe the most important parameters.

2.4.1 Site Description [Columns A through Y]

Column A is an identifier of each storm event stored in the database. It is the table key. Column B describes the site main land use or activity: residential (RE), commercial

(CO), industrial (ID), institutional (IS), open space (OS), and freeways (FW). In the case when more than one land use is present, a combination code is used beginning with the land use with most area in the watershed. For example, if a site was 70% residential and 30% commercial, the site was coded as RE_CO. The percentage of each land use is indicated in the columns J through O.

Column C describes the month of the year when the sample was collected as follows: winter (WI) if the sample was collected in November, December or January; spring (SP) if the sample was collected in February, March, or April; summer (SU) if the sample was collected in May, June or July; and fall (FA) if the sample was collected in August, September, or October. A reasonably uniform number of samples were collected during each of the four periods: about 28.5% of the samples were collected in the winter, 30% in the spring, 18.5% in the summer, and 23% in the fall.

Columns D through F indicate the location of the site. LOCATION_ID is the key for sorting the sites, and is a code of eight characters: the first two letters indicate the name of the state, the next four letters is a code for the community, and the last two letters represent the site name. Columns E and F are the name of the community and the name of the site. Column G is the contact information of the person in that community that supplied the database information. Columns H through M are the percentages of the land uses described in column B

Column N indicates watershed drainage area in acres. Figure 2 shows the distribution of the area by land use. The distribution of the watersheds areas can be considered approximately lognormal. Commercial, industrial, open space, and residential land uses have approximately the same distribution of drainage areas, with a range

between ten and one thousand acres. The median monitored watershed area for commercial and industrial sites was about 43 acres, while the median watershed area in residential and open space areas was about 65 acres. Freeways had smaller areas than the other land uses, with median areas being about 2 acres, with a range varying between one and one hundred acres.

Columns O and P list the approximate latitude and longitude of the outfall location in degrees, minutes, and seconds. Most of these coordinates were obtained using the Terraserver website. Column S indicates the EPA Rain Zone location of each site (Figure 1 and Table 1). About 52% of the sites are located in the EPA Rain Zone 2, which contains the Chesapeake Bay region, the main targeted area for this database. Each of the Rain Zones 3 through 7 has about 8% of the total sites. Rain Zones 1 and 9 have each about 3% of the sites. Rain Zone 8 has only one community with four locations, or about 1% of the total number of sites.
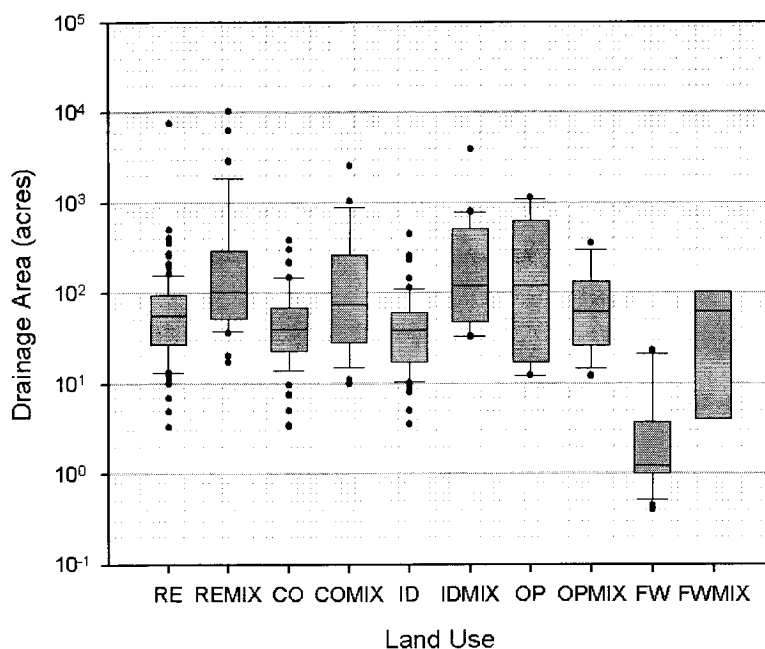
Fig. 2. Drainage areas by land use

Column R indicates the total percentage of impervious reported for each site. Only Newport News, Virginia, contained information describing how the impervious areas were hydraulically connected to the drainage systems. It is expected that a watershed with high levels of impervious (a parking lot for example), is mostly directly connected due to little opportunity for draining to pervious areas. Less water is therefore infiltrated and the stormwater rapidly moves to the connected outfall. About 169 sites (about 47% of the total number of sites) included percentage of impervious in their annual reports or permit applications. Of this response, about 69 sites were for single or mixed residential areas, 34 sites were single or mixed industrial areas, 34 sites were single or mixed commercial area, 17 sites were single or mixed freeway areas, and 15 sites were single or mixed open space areas.

Figure 3 shows a box and whiskers plot of the reported impervious values for the predominant land uses. As expected, the open space sites have the lowest percentage of impervious areas (mean about 3.3%), while the mean impervious value for the freeway sites is 92%. Industrial and commercial area impervious values are higher, with means of 67% and 81% respectively. Residential areas cover almost the complete range, from about 7 to 89%. The impervious covers for residential areas are intermediate between the values for open space and the industrial/commercial values, as expected. The mean percentage of impervious areas in residential areas is approximately 41%.
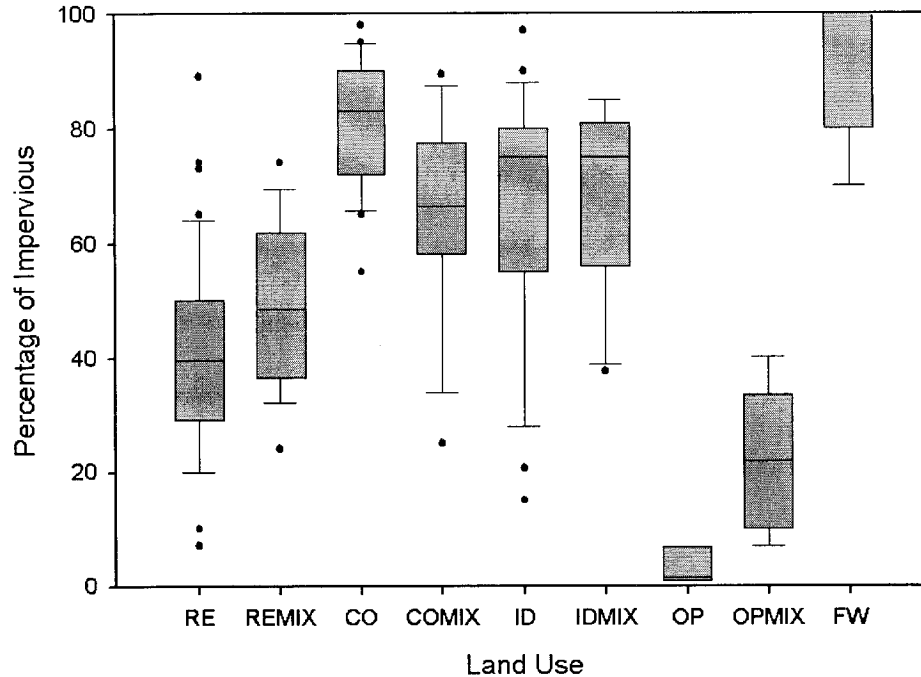
Fig. 3. Percentage of impervious area by land use

Column S is a qualifier for the total percentage of impervious area in the test

watershed, indicating if there was an apparent increase the percentage of imperviousness

during the monitoring period, based on examinations of aerial photographs. Only one site

(Pylon Street in Forth Worth, TX) had an apparent increase in the percentage of

impervious area during the monitoring period. Column T indicates the volumetric runoff

coefficient (Rv), or the ratio between the total runoff depth divided by the precipitation

depth for each event. Figure 4 is a scatter plot of the reported percentage of impervious

areas and reported Rv. As expected, higher volumetric runoff coefficients are reported for

heavily paved areas, such as parking lots or freeways, compared to areas having much

more landscaped areas, such as residential areas or parks. However, this plot is suspect

and it is possible that the reported Rv values are simply calculated from the percent

impervious values, and not from monitored rainfall and monitored runoff values.
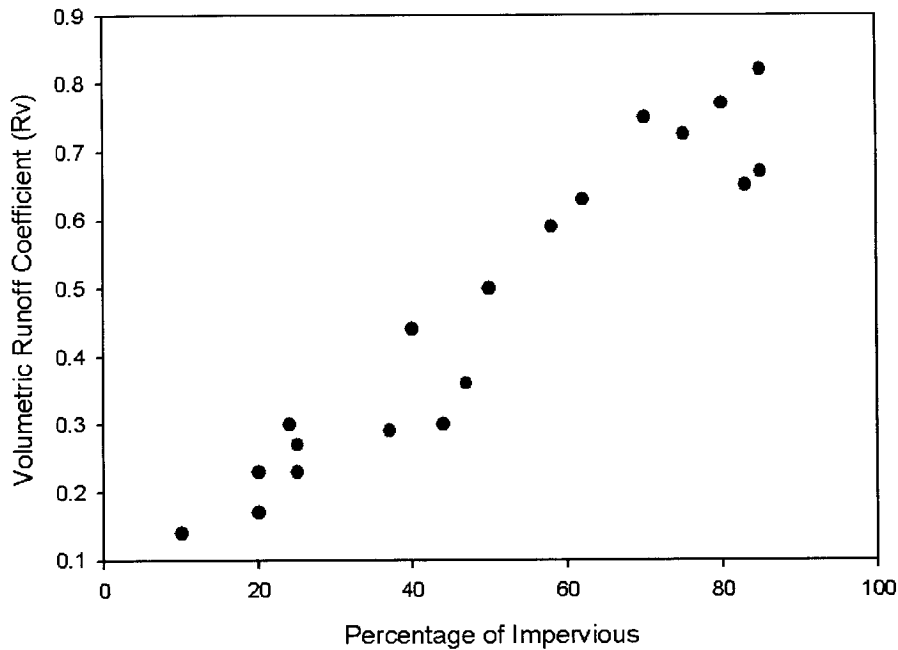
Fig. 4. Scatter plot of percentage of impervious and Rv

None of the monitoring agencies reported the TR-55 curve number for the sites.

This value is used to estimate the runoff volume using the Soil Conservation Service,

SCS (now Natural Resources Conservation Service, NRCS) TR-55 method. Curve

numbers (column U) were therefore not examined during this analysis. Only eight sites

indicated the period when the land was developed, and are shown in column V. Because

of the low number of observations, this factor also could not be used in the data analyses.

Column W indicates the type of stormwater conveyance reported for the

monitored area. This parameter indicates if the site is drained with "curb and gutter"

systems typical of areas with high percentages of impervious, or if the water is

transported beside the road through a small grass-lined channel, more common in lower

density areas. About 26% of the sites did not report the type of conveyance or it was not

possible to identify them using the aerial photographs. Curb and gutter systems were

reported for 65% of the sites, while grass swales were reported for 9% of the sites. Grass

swales are usually considered a stormwater control, or "BMP," due to their ability to

infiltrate large fractions of the runoff before discharge. They may provide some limited

concentration reduction of particulate pollutants, but only for the shallowest flows.

Detailed analyses are presented in Chapter 5 of this dissertation.

The next column indicates if the site has wet detention ponds. About seven sites

(out of the 360 total sites) have a wet pond at the outfall, nine sites have ponds in the

watershed, and three sites have ponds in series, all above the monitoring location. Other

reported stormwater controls included: dry detention ponds (4 sites), small underground

detention storage tanks (2 sites), besides the 32 sites having grass swales as noted above.

A wet detention pond is usually installed to reduce the discharge of stormwater

pollutant to receiving water. A wet pond is a sedimentation pond where the particles

settle during the storm event, plus reducing the energy of the discharged water to enhance

channel stability. Discharges from wet detention ponds usually have much reduced

concentrations of particulate pollutants, compared to influent conditions. Detailed

analyses about controls are presented in Chapter 5 of this dissertation

The final column in this section (Y) includes important comments that were not

assigned to any of the other columns. Typical information in this column is the size of the

pipe; if the outlet is circular (pipe), or square (box culvert); the number of pipes

discharging from the watershed; or if there is a USGS monitoring station at the outfall

that reported the data in the NSQD.

2.4.2 Hydrologic Information [Columns Z through AN]

Column Z is the identifier of each storm event stored in the database. It is used as a table sorting "key." Generally, it contains information about the location and the sampling date. Column AA indicates the precipitation depth recorded during the event, in inches. About 3,300 events included this parameter. Precipitation depth, flow volume and similar hydrologic parameters were included in the annual reports or permit applications usually as appendices. During the data collection process, some of these appendices were not copied or located. The highest percentage of events with precipitation by land use was observed in single and mixed freeways (about 99%). The lowest percentage of events with precipitation data was observed in single and mixed residential areas with 85%. The percentage for the other land uses were: 87% for single and mixed industrial, 90% in single and mixed commercial, 96% in single and mixed open space.

Figure 5 shows the distribution of the available precipitation depth data by land use. The range of precipitation depth varies between 0.01 and 6 inches, indicating that some of the reported events were outside of the range specified by the general monitoring guidance (minimum of 0.1 inches and close to annual average characteristics). The distribution of the rainfall depth data is approximately lognormal, with a median between 0.4 to 0.6 inches. All the land uses have a similar pattern, with approximately the same variance. The mixed freeway category seems to have a narrower range, but they only represent 0.5% of the total events that have precipitation data. Column AB is a qualifier for the precipitation depth data. Some communities collected the data on site, while others used rain gauge data collected from a local airport. Rain gauges located on site are preferred as they are expected to better represent the rainfall conditions that occurred on

the monitored site for the monitored event. Twelve percent of the total database events did not include precipitation depth data, 42% of the events were associated with rain data collected on site, 23% of the events did not indicate how the reported rain data was obtained, 7% of the events are associated with rain data from the local airport rain gauge, and the remaining 16% used other methods to determine the event rainfall data, such as regional rain gauges associated with flood monitoring systems.

Columns AC through AF indicate the starting and ending date and time of the event. Column AG indicates the maximum reported 15-minute rain intensity for each event. Events having high rain intensities have high kinetic energies, and it is hypothesized that these events will have increased washoff or erosion of particulate pollutants from watershed surfaces. However, only 1% of the database events reported this parameter. Column AG information was therefore not included in any of the data analyses.

Runoff depth (column AH) is the total volume of stormwater that leaves the monitored watershed during the rain event. For a directly connected paved parking lot, the runoff depth (expressed in inches of runoff for the complete drainage area) is only slightly smaller than the precipitation depth. In contrast, a park having mostly pervious surfaces would record total runoff volumes much smaller than the rain depth because most of the rainwater is infiltrated before it drains from the site. About 36% of the events included runoff data.

Figure 5 also shows the probability plots of runoff depth for each land use. As expected, smaller runoff values were observed in open space and residential areas, while freeways, mixed commercial, and mixed industrial land uses have runoff distributions

similar to the rain distributions observed in the precipitation panel. A different pattern

was observed for runoff at freeways which are characterized by their small area and high

percentage of imperviousness.

Probability Plot of Precipitation and Runoff Depth
Lognormal
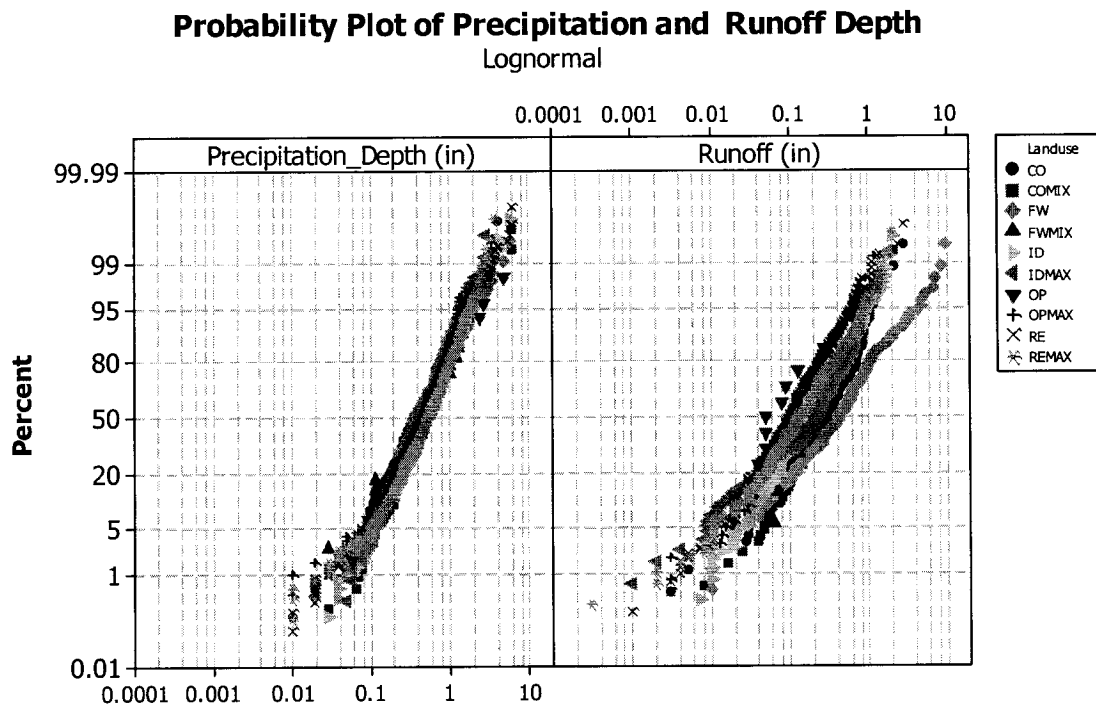
Fig. 5. Precipitation and runoff depth by land use

Column AJ indicates if the runoff and precipitation were measured during the

complete event or only the first three hours of the storm. The basic NPDES stormwater

monitoring guidelines indicates that samples must be collected at least during the first

three hours of the event. If the runoff and precipitation were not monitored for the

complete event, then further site hydrology confusion would occur. Most of the

communities recorded the runoff for the complete event, even if monitoring only

occurred for three hours. Only Greensboro, Topeka, Chesterfield County, and

Fayetteville recorded runoff only for the first three hours of the events.

Column AK indicates that all the events included in this version of the database are composite. First flush events were included in the first version of the database, version 1.0. After the paired first flush statistical analyses (see Chapter 6), these first-flush data were removed from the main database to eliminate confusion, leaving only the composite samples in the main database.

Column AL indicates if the composite sample was collected using automatic equipment, or if manual sampling was used. This column can be used to evaluate possible differences in the recorded concentrations due to the sampling method. About 81% of the events were collected using automatic samplers, 10.5% used manual sampling, and about 8.5% of the events did not have any reported sampling method. Detailed analysis about the effect of manual versus automatic sampling is discussed in Chapter 5 of this dissertation.

Column AM describes if the collected sample was flow-weighted or time-weighted. A flow-weighted composite sample is comprised of several equal volume subsamples that were collected according to the flow rate of the runoff water. The sampler is programmed to collect a subsample for a specified constant flow increment. The total volume in the single composite bottle is therefore proportionate to the total runoff volume associated with the monitored event. A time-weighted composite sample is made up of several equal volume subsamples that were collected at constant periods of time and collected into a single large composite sample bottle. At the end of the event, the total volume of sample in the composite sample bottle is proportionate to the duration of the event. About 73% of the events in the database were collected using flow-weighted composite sampling methods, while only 8% of the events were collected using time-

weighted composite sampling methods. No composite sampling method information was available for the remaining 19% of the events.

The last column in this database section describes the number of days without rain prior to the event sampling. It is usually hypothesized that an increase in the number of dry days prior to an event would cause an increase in the constituent concentration. About 38% of the events had this information available. Detailed analyses are presented in Chapter 5 of this dissertation.

2.4.3 Conventional Constituents [Columns AO through BS]

This section of the database contains conventional stormwater constituents (conductivity, DO, hardness, oil and grease, pH, temperature, TDS, TSS, BOD_5, COD, fecal coliforms, and fecal streptococcus).

Table 3, presented earlier, contains a summary showing the total number of samples included in the database classified by land use, the percentage of samples detected, the median, and the coefficient of variation. In general, the lowest concentrations were usually found at open space land uses, followed by residential areas. The highest concentrations were observed at freeway land use sites. Table 4 is a summary contrasting the land uses having the lowest and the highest concentrations of these constituents.

The Mann-Whitney test was used to determine if there is a significant difference between the land uses having the lowest and highest concentrations. As a complement, one-way ANOVA analyses were used to identify if a significant difference existed among any of the land uses. As the number of samples increases, the power of the test also

increases. P-values close to zero will indicate that the concentration of at least one land

use is statistically different than the other land uses (true for all constituents in Table 4,

except for Dissolved Oxygen).

Table 4. Conventional Constituents Summary

Constituent	Land use having the lowest median concentration			Land use having the highest median concentration			Mann-Whitney Test	1-Way ANOVA by Land Use
	n	Land Use	Median	n	Land Use	Median	p-value	p-value
Conductivity (µS/cm)	106	RE	96.5	108	ID	135.5	0	0
Dissolved Oxygen (mg/L)	39	ID	7.3	30	RE	7.8	0.064	0.325
Hardness (mg/L CaCO3)	350	RE	32	139	CO	38.9	0.009	0
Oil and Grease Total (mg/L)	308	RE	3.85	43	FW	8.0	0	0.001
pH (s.u.)	111	FW	7.1	234	ID	7.5	0	0
Temperature (°C)	31	FW	14	140	ID	17.8	0	0
Total Dissolved Solids (mg/L)	854	RE	72	411	ID	92	0	0
Total Suspended Solids (mg/L)	977	RE	49	133	FW	99	0	0
Biological Oxygen Demand (mg/L)	38	OP	5.4	421	CO	11	0	0
Chemical Oxygen Demand (mg/L)	33	OP	42.1	66	FW	100	0	0
Fecal Coliform (colonies/100 mL)	261	ID	2500	21	OP	7200	0.014	0
Fecal Streptococcus (colonies/100 mL)	166	CO	10285	273	RE	24600	0	0.003

Figure 6 contains examples of grouped box and whiskers plots for several

constituents for different major land use categories. The freeways sites had the highest

reported TSS, COD and oil and grease concentrations. Statistical ANOVA analyses for

all land use categories found significant differences for land use categories for all

constituents except for dissolved oxygen. Turbidity, total solids, total coliform and total

E-coli have not enough samples in each group to evaluate if there is a difference among

all land uses. Chapter 5 presents more comprehensive analyses for specific site conditions

(considering interactions of land use, geographical location, etc.).
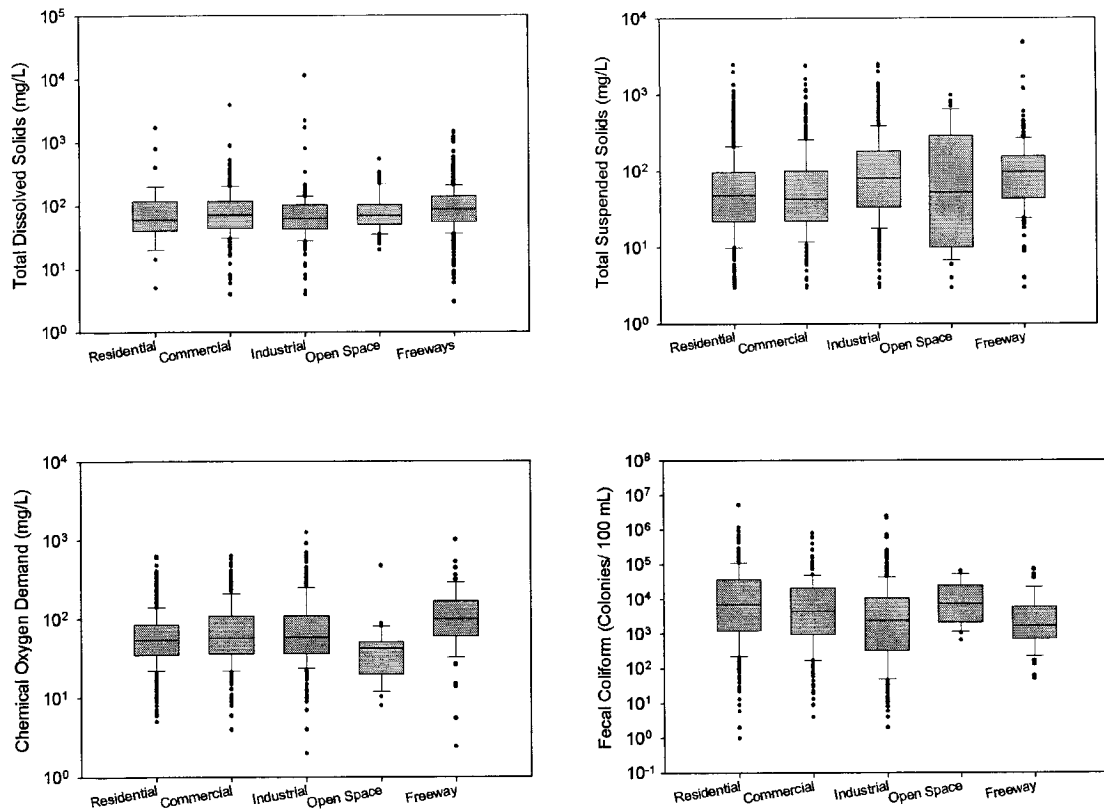
Fig. 6. Box and whiskers plots for conventional constituents by single land use

Stormwater temperature depends of many factors, including season, the time of the day, and the types of surfaces in a land use. Column C shows the season of the year when each sample was obtained, the most obvious factor affecting runoff temperature.

Figure 7 shows the water temperatures for each month for the samples collected in the EPA Rain Zones 5 and 6 combined. Similar patterns were observed in the other EPA Rain Zones. Two main periods can be identified in this plot: from February to July the water temperature rises and from August to January the water temperature decreases. Table 4 shows that for almost all conventional constituents, residential and open space

land uses have the lowest concentrations, except for pathogen indicators. Industrial and freeway land uses generally have the highest concentrations.
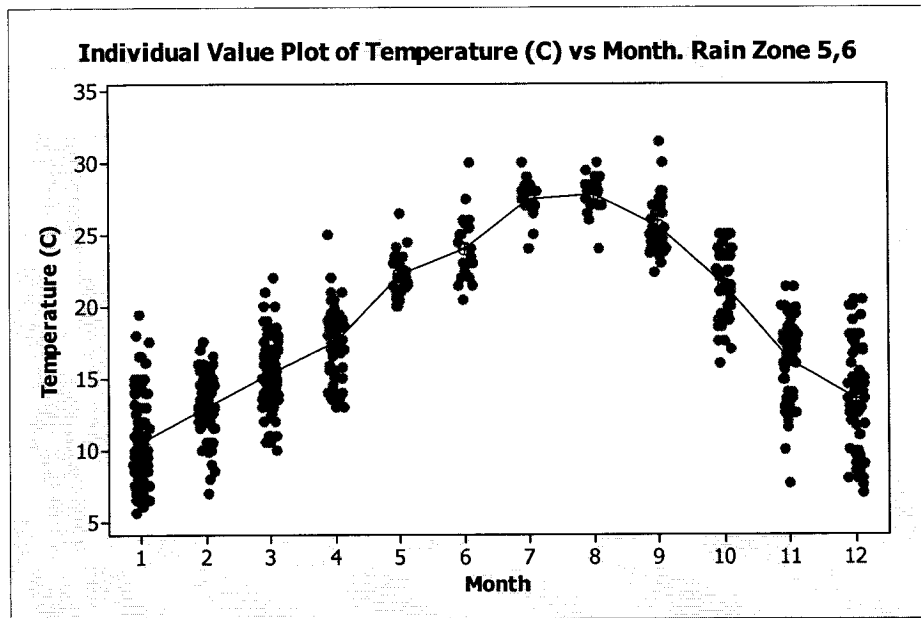
Fig. 7. Water temperature in EPA Rain Zones 5 and 6 (line links median values for each month)

2.4.4 Nutrients [Columns BU through CG]

This section in the database contains the compounds associated with nitrogen and phosphorus compounds. Table 5 shows a summary of the land uses having the lowest and highest concentrations for each constituent. Again, the Mann Whitney and ANOVA tests were used to evaluate if there was a significant difference between land uses.

In contrast to conventional constituents, dissolved and total phosphorus have the highest concentrations in residential land uses. There was no significant difference land uses for total nitrogen. Another observation is that the median ammonia concentration in freeway stormwater is almost three times the median concentration observed in residential and open space land uses. Freeways have the lowest orthophosphate and

nitrite-nitrate concentrations; almost half of the concentrations were observed in industrial land uses. Figure 8 shows box plots for TKN, total phosphorus, and nitrite-nitrate for several land uses. It shows that even if there are differences in the median concentrations by a factor of two or three between the land uses, the extreme range of the concentrations within the same land uses can vary by two or three orders of magnitude. Again, Chapter 5 analyses many factors affecting these concentrations, in addition to land use.

Table 5. Nutrients Summary

Constituent	Land use having the smallest median concentration			Land use having the largest median concentration			Mann-Whitney Test	1-Way ANOVA by Land Use
	n	Land Use	Median	n	Land Use	Median	p-value	p-value
Ammonia (mg/L)	485	RE	0.31	69	FW	1.07	0	0
Nitrogen Nitrite-Nitrate (NO_2+NO_3) (mg/L)	24	FW	0.28	429	ID	0.71	0	0.001
Nitrogen Total (mg/L)	63	ID	2.03	81	RE	2.30	0.25	0.698
Nitrogen Kjeldahl Total (TKN) (mg/L)	32	OP	0.74	121	FW	2.00	0	0
Phosphate Ortho (mg/L)	103	FW	0.09	66	ID	0.23	0	0
Phosphorous Dissolved (mg/L)	283	ID	0.11	621	RE	0.17	0	0
Phosphorous Total (mg/L)	427	CO	0.22	933	RE	0.30	0	0

2.4.5 Metals [Columns CK through EK]

This section in the database contains the descriptions of the metals concentrations. Industrial land uses have higher median concentrations of heavy metals than any of the other land uses, followed by freeways. Table 6 shows the ANOVA results for metals. As expected, open space and residential land uses have the lowest median concentrations
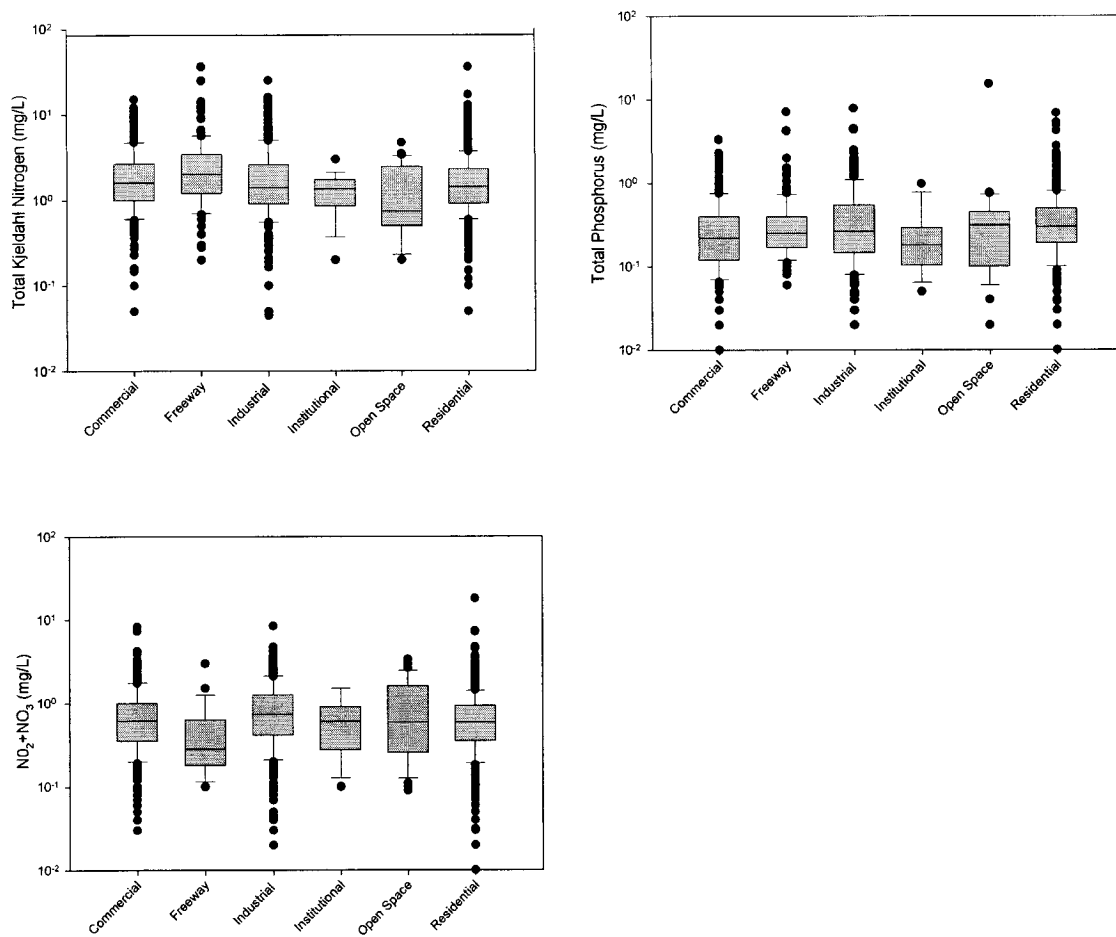
Fig. 8. Box and whiskers plots for nutrients by single land use

In almost all cases, the median metal concentrations in industrial sites were about three times the median concentrations observed in open space and residential land uses. Arsenic, cadmium, chromium, copper, lead, nickel, and zinc showed significant differences between the extreme land uses at the 1% level of confidence, or less. Other constituents are also included in the database (antimony, beryllium, cyanide, mercury, selenium, silver, thallium) in addition with all the dissolved forms of the metals. Few observations and large fractions of undetected observations hindered statistical analyses of these other metals.

Table 6. Summary of Metals Concentration

Constituent	Land use having the smallest concentration			Land use having the largest concentration			Mann-Whitney Test	1-Way ANOVA by Land Use
	n	Land Use	Median	p-value	Land Use	Median	p-value	p-value
Arsenic Total (µg/L)	70	CO	2.4	145	ID	4.0	0	0
Cadmium Total (µg/L)	219	RE	0.5	223	ID	1.9	0	0
Chromium Total (µg/L)	241	RE	4.6	186	ID	14.0	0	0
Copper Total (µg/L)	29	OP	10	96	FW	34.7	0	0
Lead Total (µg/L)	19	OP	10	343	ID	26	0	0
Nickel Total (µg/L)	190	RE	5.4	156	ID	16	0	0
Zinc Total (µg/L)	32	OP	40	455	ID	200	0	0

Figure 9 contains examples of grouped box and whiskers plots for lead, copper, and zinc constituents for different major land use categories. The highest lead and zinc concentrations were found in industrial land uses, while the highest copper concentrations were observed at freeways sites.

2.4.6 Additional Constituents [Columns EM through HW]

These columns contain additional constituents that were sampled only during the permit application (first year of sampling). Some constituents having more than a 30% detection level included: methylenechloride, total petroleum hydrocarbon, total organic carbon, chloride, nitrate nitrogen, nitrite nitrogen, total organic nitrogen, and iron.
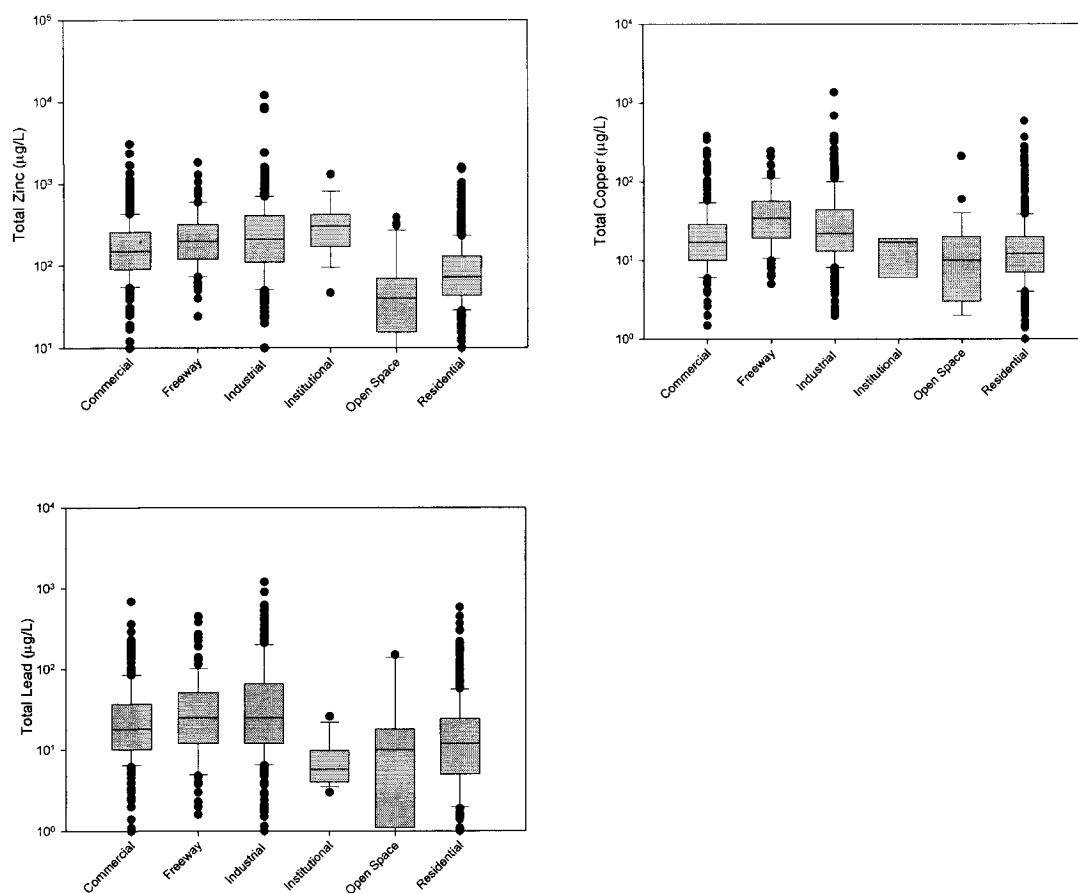
Fig. 9. Box and whiskers plots for metals by single land use

Table 7 shows summaries for these additional constituents that have enough samples to identify significant differences between land uses. Only total petroleum hydrocarbon (TPH) and nitrite nitrogen showed significant differences (at the 5% significance level) between the land uses having lowest and highest median concentrations. The median stormwater TPH concentration in residential areas is almost half the median TPH stormwater concentration at freeway sites.

Table 7. Summary of Additional Constituents

Constituent	Land use having the smallest concentration			Land use having the largest concentration			Mann-Whitney Test	
	n	Land Use	Median	n	Land Use	Median	p-value	Significant α =5%
Total Petroleum Hydrocarbon (mg/L)	36	RE	0.38	20	FW	0.78	0	Yes
Chloride (mg/L) (FW and OP not included)	42	ID	7.1	38	CO	9.5	0.25	No
Nitrogen Nitrate (mg/L) (CO and OP not included)	13	RE	0.69	98	FW	0.84	0.58	No
Nitrogen Nitrite (mg/L) (CO and OP not included)	7	ID	0.07	42	FW	0.17	0.01	Yes
Nitrogen Total Organic (mg/L) (FW not included)	12	RE	0.96	5	CO	1.97	0.19	No
Iron (mg/L)	6	RE	2.99	27	FW	3.60	0.27	No

2.5 Site Description and Additional Supporting Information

Supplemental reports were created containing additional information for each

community. These site descriptions include (depending on available information) the land

use and impervious surfaces for the monitored site, aerial photographs and a topographic

map of the area, and descriptions of the sampling procedures and quality control

(QA/QC) used during sample collection and analysis. The QA/QC description indicates if

blank samples were used during the analysis to check the equipment, the protocols used

during the sample collection, and in some cases, the chain of custody of the samples.

These supplemental reports also contain descriptions of the sampled parameters,

analytical methods, and field instrumentation used by the community.

About 38% of the aerial photographs have below 1-meter resolution and the

remaining photos have 1-meter resolution. The locations of most of the outfalls were

included in the database in the Q and R columns (Latitude and Longitude). Table 8 shows

the total number of sites with high-resolution aerial photos and with watershed

delineations.

Table 8. Additional Site Information

State	Number of Communities	Number of Sites	Sites with high-resolution aerial photos (resolution 0.25 m)	Sites with watershed delineation
1	2	12	0	2
2	28	185	38	18
3	8	30	15	20
4	4	21	15	17
5	9	33	18	0
6	4	30	20	9
7	6	28	19	0
8	1	4	0	0
9	3	17	13	8

Watershed delineations are an important component of the site descriptions by identifying the extent of the contributing area, the different land uses located in the watershed and the sampling location. Only 20% of the sites included their watershed delineations.

Most communities followed the sampling recommendations presented in the Code of Federal Register (40 CFR 122.21), although delegated NPDES state agencies were able to modify the specific requirements to better address local concerns. Almost all communities collected samples at least during 3 hours of the event (or the complete event if the duration was shorter). For about 66% of the events, the communities calculated the total runoff for the duration of the total event discharge, but used the concentrations from the shorter monitoring period. Chapter 6 includes a detailed analysis of first-flush concentrations that may indicate the maximum errors that may occur with truncated sampling periods. Seven percent of the events included runoff for only the first three hours of the event. The remaining 25% of the events did not include runoff volume data,

or it was not clear if the runoff volume data was obtained during the first three hours, or for the whole event.

Another important monitoring aspect described in the site descriptions is how the composite sample was created. There are two compositing options: flow-weighted and time-weighted. During the time-weighted compositing scheme, subsamples of equal volume were obtained a specific time intervals during the three hour sampling period. All the subsamples were collected in a single bottle, creating the composite sample. In the flow-weighted compositing case, the subsamples were collected for every set flow increment. About 71% of the events were collected using flow-weighted sampling, 5% of the events were collected using time-weighted sampling, and it was not clear how the remaining 24% of the samples were collected. Roa-Espinosa and Bannerman (1995) found that time-weighted composite sampling could be representative of the sampling period, if many subsamples are collected throughout the storm period. Time-weighted compositing is much simpler and less expensive than flow-weighted composite sampling, but may have a slight error in the measured concentrations, compared with the flow-weighted method.

About 62% of the 65 communities represented in the NSQD indicated that they used automatic samplers during their monitoring activities, about 34% did not indicate how they collected their samples, and 4% collected their samples manually. ISCO samplers were the most commonly used automatic sampler, with about 24% of the sites used ISCO 2700, 3700 or 6700 samplers. American Sigma samplers were used in about 12% of the 65 communities. The most common American Sigma sampler models included 800SL, 900AV and 900 MAX. About 69% of the communities did not indicate

how, or if, they measured flow, and did not report any flow data. About 20% of the sites used of ISCO 3230 or 4230 flow meters. The remaining 11% used other methods to estimate the stormwater discharge volumes.

About 58% of communities also described problems found during the monitoring process and these are summarized in the site summary reports. Some communities reported more than one problem. One of the basic sampling requirements was to collect three samples every year for each of the land use stations. These samples were to be collected at least one month apart during rains having at least 0.1 inch rains, and with at least 72 hours from the previous 0.1-inch storm event. It was also required (when feasible), that the variance in the duration of the event and the total rainfall not exceeded the median rainfall for the area. About 47% of the communities reported problems meeting these requirements. In many areas of the country, it was difficult to have three storm events per year having these characteristics. The second most frequent problem, reported by 26% of the communities, concerned backwater tidal influences during sampling, or the outfall became submerged during the event. In other cases, it was observed that there was flow under the pipe (flowing outside of the pipe, in the backfill material, likely groundwater), or sometimes there was not flow at all. About 12% of the communities described errors related to malfunctions of the sampling equipment. Most of the communities with equipment failures did not report the reasons of the failure. When reported, the equipment failures were due to incompatibility between the software and the equipment, clogging of the rain gauges, and obstruction in the sampling or bubbler lines. Memory losses in the equipment recording data were also periodically reported. Other reported problems were associated with lighting, false starts of the automatic sampler

before the runoff started, and operator error due to misinterpretation of the equipment
configuration manual.

2.6 Problems Encountered during NPDES Stormwater Monitoring

The annual reports prepared by the monitoring agencies and submitted to the
states periodically described problems that they encountered during the monitoring
activities. Some of these are summarized below:

The NPDES monitoring guidance required the sampled events to have at least a
72 hours antecedent dry period since the previous 0.1 inch or larger rain. At some sites,
72 hours still did not prevent the occurrence of base flows, and in others, rains commonly
occur at high frequency. Previous events with large precipitation depths, or very large
watersheds, were more commonly associated with base flows. Sites located on the East
coast (Hampton, VA for example) where the hurricane season produces frequent large
storms, especially having a high water table, were especially susceptible. Base flows can
commonly occur in separate storm drainage systems for a variety of reasons and they
may be more important during some seasons than during others. In many cases, they
cannot be avoided and should be included in the monitoring program, and their effects
need to be recognized as an important flow phase. As an example, Pitt and McLean
(1986) found dry weather base flows to be significant sources of many pollutants, even
during a comprehensive research project that spent much time surveying the test
watersheds to ensure they did not have any inappropriate discharges entering the storm
drainage system.

Capturing runoff events within the acceptable range of rain depth was difficult for some monitoring agencies. Rain depth cannot be precisely predicted in many areas of the country. Also, if using rain gauge data from a location distant from the monitoring location, the reported rain depth may not have been representative of the depth that occurred at the site.

The rain gauges need to be placed close to the monitored watersheds. This was likely one of the reasons why the runoff depths periodically exceeded the reported rain depths. Rain in urban areas can vary greatly over small distances. The ASCE/EPA (2002) recommended that rainfall gauges be located as close as possible to the monitoring station. In the NSQD, about 7% of the events had site precipitation estimated using rain gauge located at the city airport. About 16% of the events had precipitation depth estimated using their own monitoring network (Hampton Road Sanitation District, for example). Some communities had precipitation networks that were used for flood control purposes for the surrounding area. These networks can be considered better than the single airport rain gauge, but should at least be supplemented with a rain gauge located in the monitored watershed. Another factor that needs to be considered is the size of the watershed. Large watersheds cannot be represented with a single rain gauge at the monitoring station; in those cases the monitoring networks will be a better approach. Large watersheds are more difficult to represent with a single rain depth value.

Many of the monitoring stations lacked flow monitoring instrumentation, or did not properly evaluate the flow data. Accurate flow monitoring can be difficult, but it greatly adds to the value of the expensive water quality data. As noted previously, base flows also need to be properly removed from the event measurements so only direct

runoff quantities are reported. It is probably unreasonable to expect to have a permanent flow monitoring station installed at a location where only manual grab samples are being obtained. However, manual flow monitoring can be conducted during manual sampling by carefully noting the flow stage in previously surveyed locations. These observations will need to be obtained during the complete duration of the event.

The three hour monitoring period may have resulted in some bias in the reported water quality data. This limit was likely used to minimize the length of time personnel at a monitoring location during manual sampling. Also, it is unlikely that manual samplers were able to initiate sampling near the beginning of the events, unless they were deployed in anticipation of an event later in the day. A more cost-effective and reliable option would be to have semi-permanent monitoring stations located at the monitoring locations and sampling equipment installed in anticipation of a monitored event. Most monitoring agencies operated three to five land use stations at one time. This number of samplers, and flow equipment, could have been deployed in anticipation of an acceptable event and would not need to be installed in the field continuously.

Some of the site descriptions lacked important information and local personnel sometimes did not have the needed information. This was especially critical for watershed delineations on maps of the area. Also, few of the watershed descriptions adequately described how the impervious areas were connected to the drainage system, one of the most important factors affecting urban hydrologic analyses. In most cases, information concerning local stormwater controls was able to be determined from a variety of sources, but it was not clearly described in the annual reports.

2.7 Discussion

2.7.1 Comparison of NSQD with Existing Stormwater Databases

The NSQD, with 3,765 events (from the 1992-2002 period) representing sites

throughout the US for most land uses, and for many constituents, is the most

comprehensive stormwater quality database currently available. The historical NURP

database (sampling period in the late 1970s and early 1980s) contains the results from

2,300 national stormwater events, while the CDM National Urban Stormwater Quality

Database includes the results of approximately 3,100 events (including the NURP data,

plus additional data collected by the USGS and about 30 NPDES permits; Smullen and

Cave, 2002). Table 9 compares the results of the pooled EMC's from the NURP

(calculated by Smullen and Cave 2002), CDM, and NSQD databases.

The NURP means and medians were computed by Smullen and Cave (2002)

using the EPA (1983) data. The CDM and the NSQD results are similar for all

constituents, except for lead and zinc. All three databases have similar reported median

and mean concentrations for COD and BOD and the nutrients, but are apparently

different for TSS and the heavy metals. The pooled mean event mean concentration

(EMC) for TSS was 2.3 times larger in the NURP database compared to the NSQD. The

largest reduction in mean EMCs was found for lead (7.9 times larger for NURP) followed

by copper (7.9 times larger for NURP) and zinc (1.6 times large for NURP).

Table 9. Comparison of Stormwater Databases

Constituent	Units	Source	Event Mean Concentrations		Number of events
			Mean	Median	
Total Suspended Solids	mg/L	NURP	174	113	2000
		CDM	78.4	54.5	3047
		NSQD	79.1	49.8	3404
Biochemical Oxygen Demand	mg/L	NURP	10.4	8.4	474
		CDM[a]	14.1	11.5	1035
		NSQD	10.9	8.6	2973
Chemical Oxygen Demand	mg/L	NURP	66.1	55.0	1538
		CDM	52.8	44.7	2639
		NSQD	71.2	55.6	2699
Total Phosphorus	mg/L	NURP	0.337	0.266	1902
		CDM	0.315	0.259	3094
		NSQD	0.373	0.289	3162
Dissolved Phosphorus	mg/L	NURP	0.100	0.078	767
		CDM[b]	0.129	0.103	1091
		NSQD	0.107	0.078	2093
Total Kjeldahl Nitrogen	mg/L	NURP	1.67	1.41	1601
		CDM	1.73	1.47	2693
		NSQD	1.74	1.37	3034
Nitrite and Nitrate	mg/L	NURP	0.837	0.666	1234
		CDM	0.658	0.533	2016
		NSQD	0.767	0.606	2983
Copper	µg/L	NURP	66.6	54.8	849
		CDM	13.5	11.1	1657
		NSQD	17.8	14.2	2356
Lead	µg/L	NURP	175	131	1579
		CDM	67.5	50.7	2713
		NSQD	24.4	16.5	2250
Zinc	µg/L	NURP	176	140	1281
		CDM	162	129	2234
		NSQD	110	88	2888

Note: a. No BOD_5 for USGS dataset. b. No DP for CDM portion of NPDES dataset

The differences found in both the NURP and the NSQD databases are most likely due to differences in geographical areas emphasized by each database. Figure 10 is a national map showing the percentage of events collected in each state as contained in the NSQD database, while Figure 11 shows the percentage of events contained in the NURP

database. Half of the events included in the NSQD database were collected in EPA Rain

Zone 2 (Maryland, Virginia, North Carolina, Kentucky and Tennessee), while half of the

events contained in the NURP database were collected in EPA Rain Zone 1 (Minnesota,

Wisconsin, Michigan, Illinois, New York, Massachusetts and New Hampshire).

Only 3% of the events in the NSQD are located in EPA Rain Zone 1, while 50%

of the NURP data is from this area. Twenty four percent of the NURP data is located in

the Mid-Atlantic and southeast states, while 60% of the NSQD data is from this area (the

area that was to be emphasized during the EPA-funded project). The NSQD is slightly

better representative of other parts of the country compared to NURP. As an example, the

percentage of the total event data from the west coast is similar for both databases, but

the NSQD represents 10 communities with almost 60 different sites, while NURP has

only 3 communities and only 7 sites. The total number of sites, communities and events

collected in the NURP study are shown in Table 10.

Table 10. Total Events Monitored During NURP by EPA Rain Zones

Rain Zone	Total Events	Percentage of Events	Number of Communities	Number of Sites
1	804	50.6	12	42
2	324	20.4	3	10
3	65	4.1	1	5
4	0	0	0	0
5	24	1.5	1	2
6	45	2.8	2	5
7	136	8.6	1	2
8	0	0	0	0
9	188	11.8	3	12
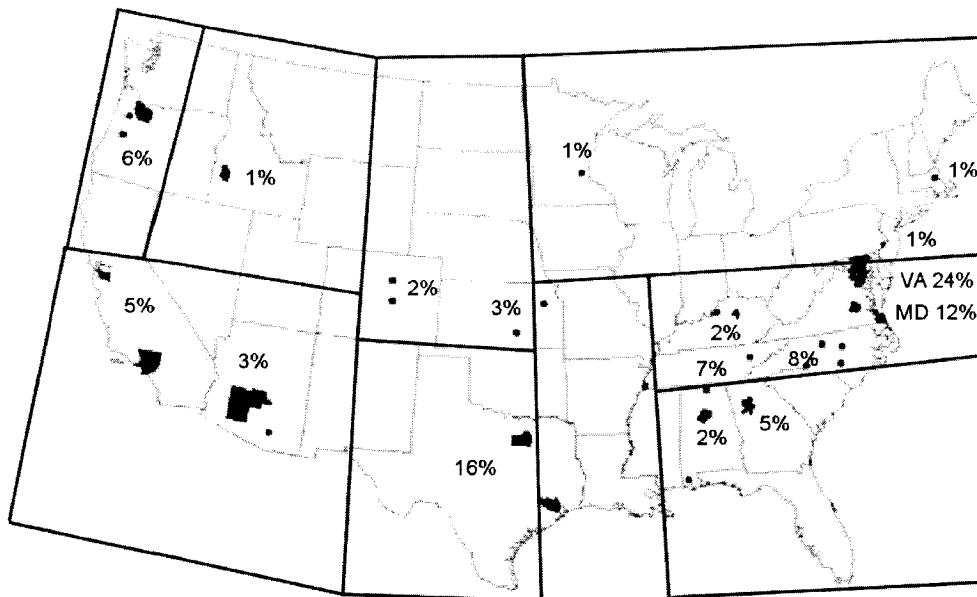
Fig. 10. Distribution of collected events using the NSQD database
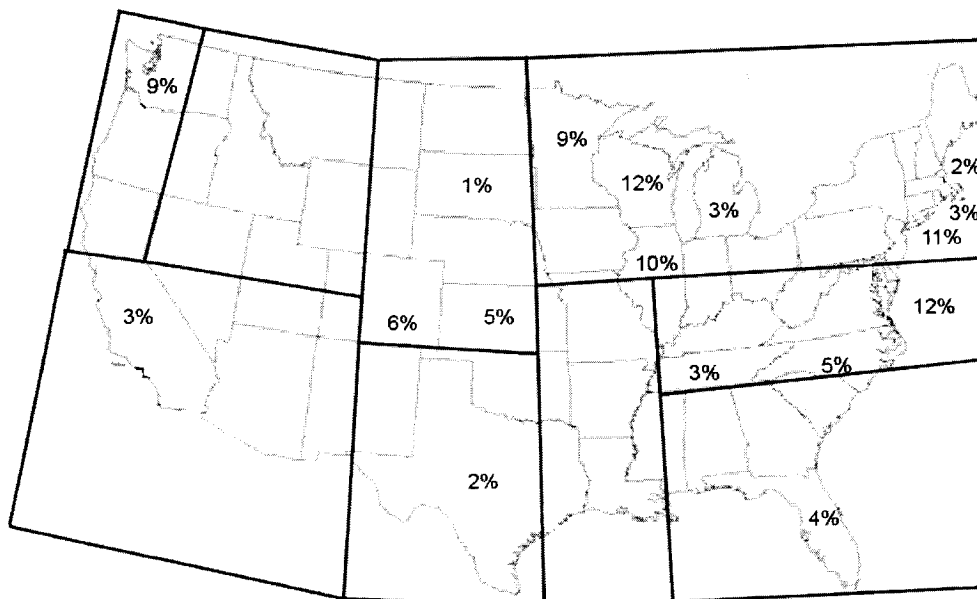
Fig. 11. Distribution of collected events using the NURP database

Figure 12 presents example plots for selected residential area data for different

EPA Rain Zones for the country as contained in the NSQD. Rain Zones 3 and 7 (the

wettest areas of the country) had the lowest concentrations for most of the constituents, while Rain Zone 1 has some of the highest concentrations.
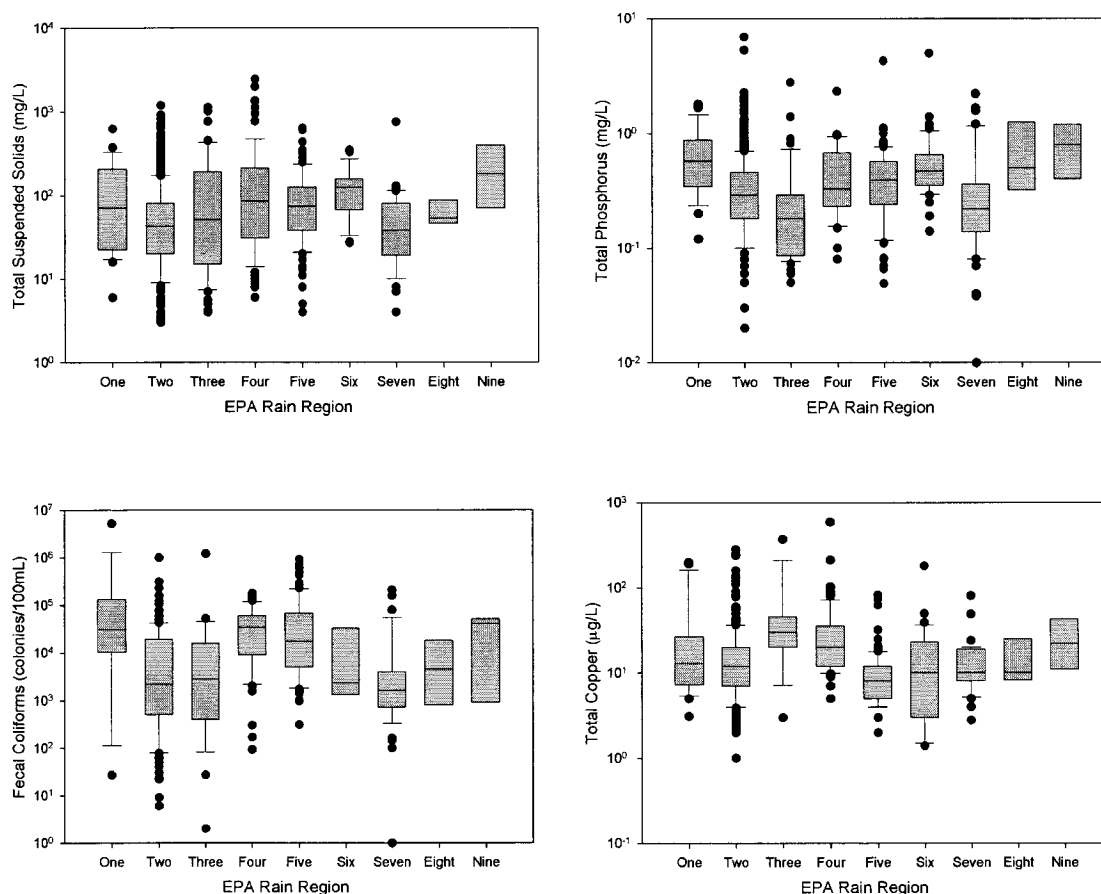
Fig. 12. Example of constituents collected in residential land use by EPA Rain Zone

It is likely that the few data from EPA Rain Zone 1 (having relatively high concentrations) in the NSQD and the few data in EPA Rain Zones 2 and 3 (having relatively low concentrations) in NURP are the main reason for the differences in the database summary values.

2.7.2 Land use effects

Another factor that may affect the difference in reported concentrations between the NURP and NSQD databases is the percentage of samples collected for each different land use category. Although each database summarized observed concentrations by land use, having few data from few sites in a land use category reduces the reliability of the estimate. Almost 45% of the NURP database represents residential sites, while residential sites comprise about 30% of the NSQD. The percentage of industrial sites in the NSQD is 15%, while industrial sites in the NURP database represent only 6% of the total. The NSQD contains samples for freeways sites, which are not included in the NURP database. The percentages of mixed land uses and commercial areas are similar for both databases. However, a better representation of open space land uses was observed in the NURP database (10% of the total) compared with the NSQD (3% of the total).

2.7.3 Other Factors

Other factors may influence the differences in reported EMCs in the different databases. Figure 13 shows the probability plot for drainage areas for sites included in the NSQD and NURP databases.

This plot shows that the NURP watersheds are larger than those observed in the NSQD. The median NSQD drainage area was about 50 acres, while it was about twice as large during NURP. The NSQD also has about 10% of the watersheds smaller than 10 acres, representing freeways sites. No literature was found that indicates that there is a relationship between the drainage area and the concentration of stormwater constituents.
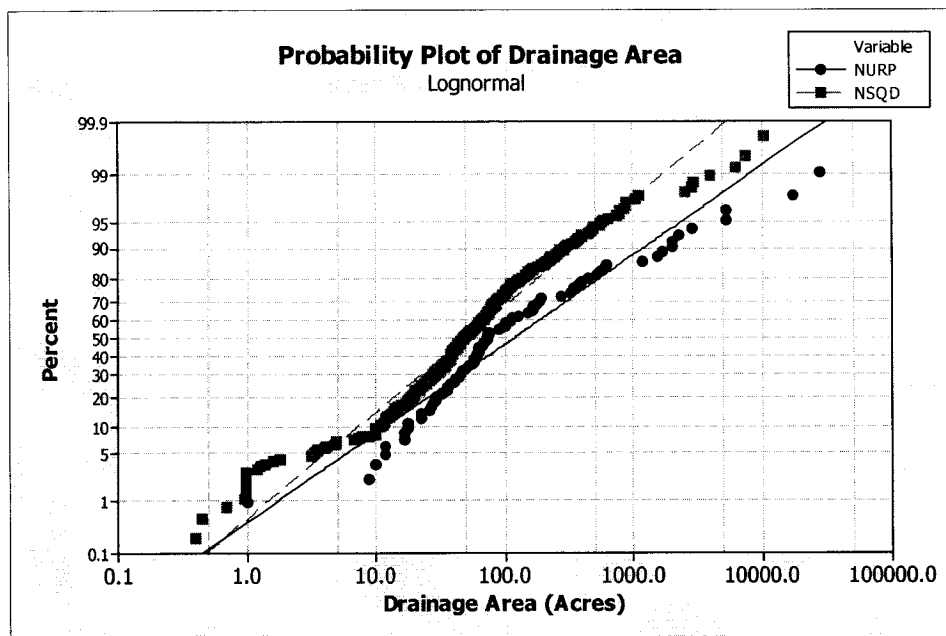
Fig. 13. Distribution of collected events using the NURP database

2.8 Chapter Summary

This chapter describes the National Stormwater Quality Database. The information collected from the NPDES Phase I stormwater monitoring program was stored in a spreadsheet containing more than 3,700 rows and 250 columns. Each row represents a single monitored event. The main structure of the database is divided into six sections: site descriptions, hydrologic information, conventional constituents, nutrients, metals, and additional constituents. The collected data is grouped into 11 land use categories: residential, commercial, industrial, open space, freeways, mixtures of these land uses, and institutional. Support documents were also created for each community. These documents include aerial photos of the watershed and outfall area (when available), narrative descriptions about the main activities and land uses in the watersheds, sampling and quality control procedures, analytical methods, and equipment used during the collection and analysis of the samples. The last part of the support

documents describe the problems that occurred during the collection and analyses of the samples, and meeting discharge permit requirements that specified sampling requirements. This information is useful for interpreting the reported monitoring data and as guidance for future stormwater programs in other communities around the country.

The data from the NSQD was compared with information from the most commonly used stormwater database, the EPA's Nationwide Urban Runoff Program (NURP) conducted more than 20 years ago. It was observed the concentrations in the NSQD were in general lower than those found during the NURP program. The analysis indicates that the main reason of these differences is the geographical differences represented by the monitoring locations represented in the databases. Most of the samples during the NURP program were collected in the upper Midwest and northern east coast areas of the country, while most of the samples represented in the NSQD were collected in the mid-east coast and southeast areas of the country. The preliminary regional analyses shown in this chapter indicate that southeast areas have lower stormwater concentrations than northeast areas.

Chapter 3
QA/QC Procedures

3.1 Introduction

This chapter presents the quality assurance and quality control procedures followed during the creation of the database. These tasks relied on two basic activities: identification of unusual observations and monitoring locations, and the examination of alternative methods to address non-detected pollutant concentration observations (left-censored data).

3.2 Quality Control/ Quality Assurance

More than 70 communities were contacted to request information concerning their NPDES Phase I monitoring activities. Most communities submitted their reports in electronic media or on paper. In cases where the data were in electronic form, the data were manipulated with macros and stored in the main Excel spreadsheet. For those communities with data only in paper, the information was typed into the spreadsheet.

Once the database was completed, the main table was first reviewed by rows (corresponding to individual runoff events) and then by columns (corresponding to measured constituents). Each row and column in the database was reviewed at least once and compared to information contained in the original reports (when available). For each constituent, probability plots, box and whisker plots, and time series plots were used to identify possible errors (likely associated with the transcription of the information, or as

55

typographical errors in the original reports). Most of the identified errors were attributed

to the transcription process and, in some cases, errors during unit conversions (such as

metal results reported as mg/L when they were really as µg/L).

Additional "logical" plots were used to identify possible errors in the database. A

plot of the dissolved (filtered) concentrations against the total concentrations for metals

should indicate that the dissolved concentrations are lower than the total forms, for

example (Figure 14). Other plots included TKN versus NH_3, COD versus BOD_5, SS

versus turbidity and TDS versus conductivity.
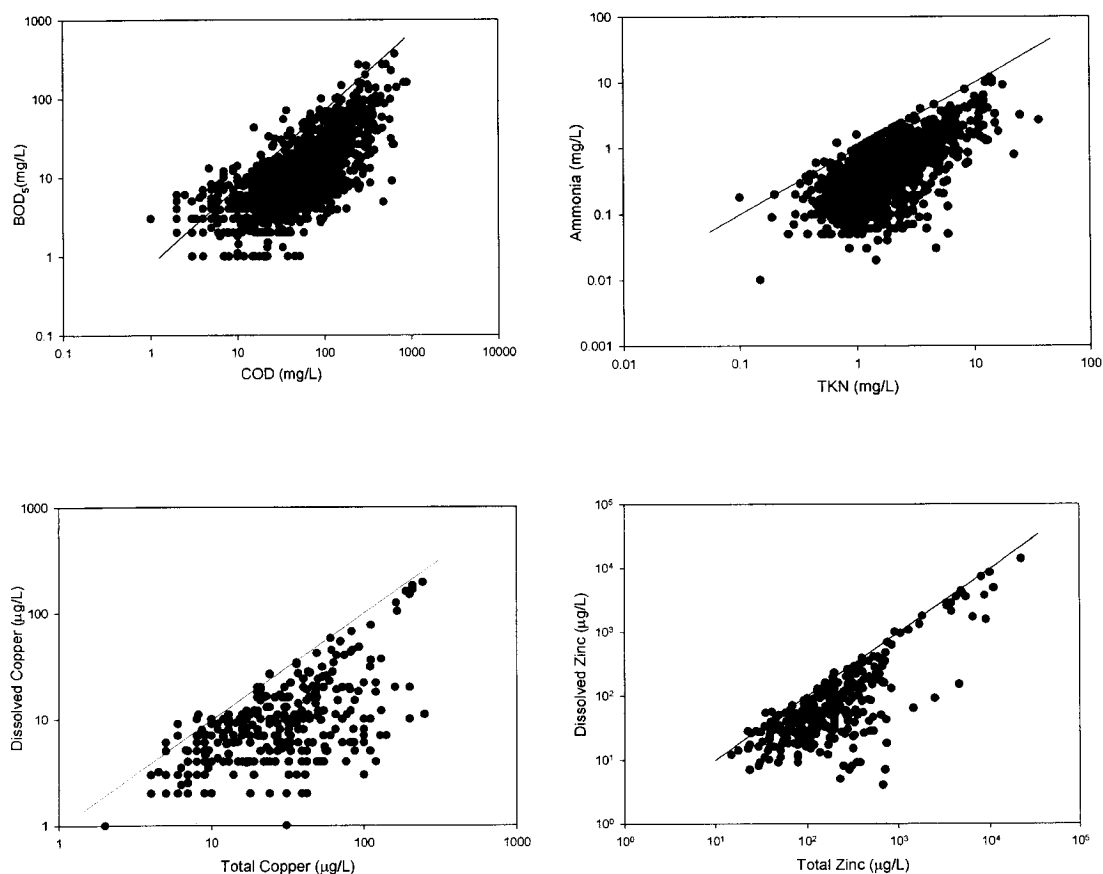
Fig. 14. Example scatter plots of stormwater data (line of equivalent concentration shown)

In all cases, suspect values were carefully reviewed and many were found to be associated with simple transcription errors, or obviously improper units, which could be corrected. However, about 300 suspect values were removed from the database as they could not be verified. None of the data were deleted without sufficient evidence of a highly probable error. For example, if a set of samples from the same community had extremely high concentrations (in one case, 20 times larger than the typical concentrations reported for other events for the same community) at different sites, but for the same event, this will indicate a very likely error during the collection or analysis of the sample. If just a single site had high concentrations (especially if other related constituents were also high), it would not normally be targeted for deletion, but certainly subject to further scrutiny. If a value was deleted from the database, or otherwise modified, a question mark notation was assigned to the respective constituent in the qualifier column. Appendix B includes all the modifications performed in the database.

In order to calculate the standard deviations for the site quality control tests, each location must have at least two observations. Nine sites were not included in that analysis because they had only one observation. These sites were: ALHUDRAV, KYLXEHL4, KYLXEHL5, KYLXNEL1, MABOA007, ORCCA001, ORODA001, ORODA002, and ORODA004

3.3 Non-Detected Analyses

To estimate the censored values, it is important to identify the distribution of the dataset and the level of censoring. It was observed that most of the constituents followed a lognormal distribution (See Chapter 4). Appendix C shows several approaches to

analyze censoring observations with single and multiple detection limits. Different comparisons substantiated the conclusion that the non-detected values in the NSQD can be best estimated using the Cohen's maximum likelihood method, compared to other traditional methods.

The values of the detection limits and their frequencies varied among the different constituents and monitoring locations. The first step in evaluating the different methods to address censored data was to identify the distribution of the censoring dataset. The identification of the detection limits assists our understanding of the methods used during the laboratory analysis and monitoring activities. The second step was applying and evaluating the different estimation methods.

3.3.1 Censored Data Distribution

The level of censoring for each constituent was calculated for each land use and site, for 18 selected constituents. These constituents contained low level of censored values. The National Council of the Paper Industry for Air and Stream Improvement found that for levels of censoring above 60%, the use of estimation methods is not appropriate (NCASI 1995). Table 11 shows the maximum, minimum, and percentage of detected values by constituent for each main land use for the complete dataset. In general, freeway sites have the largest percentage of detected observations, while open space sites have the highest percentage of non-detected observations.

Table 11. Percentages of Detected Values by Land Use Category and for the Complete Database

Constituent	Land use*	Total Events	Minimum Detected Concentration	Maximum Detected Concentration	Percentage with detected values
Conductivity (μS/cm)	RE	106	27.3	2020	100
	CO	66	17	894	100
	ID	108	42	1958	100
	OP	2	75	150	100
	FW	86	20	870	100
	TOTAL	685	16.8	5955	100
Hardness (mg/L)	RE	250	3	401	100
	CO	139	1.9	356	100
	ID	138	5.5	888	96.4
	OP	8	11	270	100
	FW	127	5	1000	100
	TOTAL	1082	1.9	1100	98.71
Oil and Grease (mg/L)	RE	533	0.2	2980	57.79
	CO	308	0.8	359	70.78
	ID	327	0.5	11000	65.14
	OP	19	0.5	4	36.84
	FW	60	3	30	71.67
	TOTAL	1834	0.2	11000	66.09
Total Dissolved Solids (mg/L)	RE	861	3	1700	99.19
	CO	399	4	3860	99.5
	ID	412	4.5	11200	99.51
	OP	45	32	542	97.78
	FW	97	12	470	98.97
	TOTAL	2956	3	17900	99.32
Total Suspended Solids (mg/L)	RE	991	3	2426	98.59
	CO	458	3	2385	98.25
	ID	427	3	2490	99.06
	OP	44	3	980	95.45
	FW	134	3	4800	99.25
	TOTAL	3389	3	4800	98.76
BOD (mg/L)	RE	941	1	350	97.56
	CO	432	2	150	97.45
	ID	406	1	6920	95.32
	OP	44	1	20	86.36
	FW	26	2	89	84.62
	TOTAL	3105	1	6920	96.2
COD (mg/L)	RE	796	5	620	98.87
	CO	373	4	635	98.39
	ID	361	2	1,260	98.89
	OP	43	8	476	76.74
	FW	67	2.44	1,013	98.51
	TOTAL	2,750	1	1,260	98.4

RE = residential; CO=commercial; ID=industrial; OP=open space; FW=freeways
Total=total database, all land uses combined, including mixed land uses

Table 11. Percentages of Detected Values by Land Use Category and for the Complete Database - *Continuation*

Constituent	Land use	Total Events	Minimum Detected Concentration	Maximum Detected Concentration	Percentage of detected values
Fecal Coliform (Colonies/100mL)	RE	446	1	5,230,000	88.34
	CO	233	4	610,000	87.98
	ID	297	2	2,500,000	87.88
	OP	23	650	63,000	91.3
	FW	49	50	70,000	100
	TOTAL	1704	1	5,230,000	91.2
Fecal Streptococcus (Colonies/100mL)	RE	305	20	840,000	89.51
	CO	181	20	1,100,000	91.71
	ID	195	22	6,000,000	93.85
	OP	22	160	101,000	90.91
	FW	25	560	130,000	100
	TOTAL	1141	20	6,000,000	94.04
Ammonia (mg/L)	RE	595	0.01	6	81.51
	CO	299	0.02	8	83.28
	ID	253	0.03	10	83.4
	OP	32	0.07	2	18.75
	FW	79	0.08	12	87.34
	TOTAL	1908	0.01	12	71.33
NO2 + NO3 (mg/L)	RE	927	0.01	18	97.41
	CO	425	0.03	8.21	98.12
	ID	417	0.02	8.4	96.16
	OP	44	0.09	3.33	84.09
	FW	25	0.1	3	96
	TOTAL	3075	0.01	18	97.33
TKN (mg/L)	RE	957	0.05	36	95.61
	CO	449	0.05	15	96.76
	ID	439	0.05	25	97.33
	OP	45	0.2	5	95.9
	FW	125	0.2	36	71.11
	TOTAL	3191	0.05	66	96.8
Dissolved Phosphorus (mg/L)	RE	738	0.01	2	84.15
	CO	323	0.01	2	81.11
	ID	325	0.02	2	87.38
	OP	44	0.01	1	79.55
	FW	22	0.06	7	95.45
	TOTAL	2477	0.01	7	85.14
Total Phosphorus (mg/L)	RE	963	0.01	7	96.88
	CO	446	0.02	3	95.74
	ID	434	0.02	8	95.85
	OP	46	0.02	15	84.78
	FW	128	0.06	7	99.22
	TOTAL	3285	0.01	15	96.5

Table 11. Percentages of Detected Values by Land Use Category and for the Complete Database - *Continuation*

Constituent	Land use	Total Events	Minimum Detected Concentration	Maximum Detected Concentration	Percentage of detected values
Total Copper (µg/L)	RE	799	1	590	83.6
	CO	387	1.5	384	92.76
	ID	415	1.97	1360	89.64
	OP	39	2	210	74.36
	FW	97	5	244	98.97
	TOTAL	2723	0.6	1360	87.37
Total Lead (µg/L)	RE	788	0.5	585	71.32
	CO	377	1	689	85.41
	ID	411	1	1200	76.4
	OP	45	0.2	150	42.22
	FW	107	1.6	450	100
	TOTAL	2949	0.2	1200	77.65
Total Nickel (µg/L)	RE	419	1	100	45.35
	CO	232	2	110	59.48
	ID	249	1	110	62.65
	OP	38	12	120	18.42
	FW	99	2.8	100	89.9
	TOTAL	1430	1	120	59.79
Total Zinc (µg/L)	RE	810	3	1580	96.42
	CO	392	5	3050	98.98
	ID	432	5.77	8100	98.61
	OP	45	5	390	71.11
	FW	93	6	1829	96.77
	TOTAL	3007	2	22500	96.57

The constituents having greater than 95% detected observations (of these 18) are conductivity, pH, hardness, TSS, TDS, and COD (except for open space). Most of the non-detected observations of these 18 constituents were for copper, lead, nickel and zinc analyses. The percentage of detected observations for these four constituents in open space areas varied between 18% and 75%, while freeways recorded valid vales for 89% to 100% of the analyses for these metals.

Residential, commercial and industrial land uses have similar percentages of detected observations for each constituent shown in Table 11. The most frequent detection limit for each constituent was also identified. Because of the duration of the

monitoring activities reported in the NSQD, the large number of municipalities involved,

and the large number of analytical methods used, each constituent usually had several

reported detection limits. The number and percentage of non-detected observations at

each detection limit was calculated with respect to the total number of non-detected

observations. For example, there are a total of 60 oil and grease observations at freeway

sites: 43 detected and 17 non-detected. There were three separate detection limits

reported for the non-detected observations: < 0.5, < 1 and < 3 mg/L with 1, 2 and 14

observations reported for each, respectively. The frequency distribution of non-detected

oil and grease observations at freeways sites was therefore 5.8%, 11.8% and 82.3%,

respectively. The results for the remaining land uses and constituents are shown in Table

12. A discussion about the percentage of the detected values and their distributions for

each constituent is presented in Appendix D.

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database

Constituent	Land use*	<**	<0.5	<0.6	<1	<1.2	<1.4	<1.9	<2	<2.24	<2.47	<2.5	<2.9	<3	<3.97	<5	<5.2	<6	<6.5	<7	<8.3	<10	<14	<17
Oil and Grease (mg/L)	RE	0.44	1.78		36.89		2.22		5.78	0.44	0.44	3.56		1.78		39.56		2.67	0.89		0.44	3.11		
	CO		2.22	1.11	34.44		2.22		5.56			3.33	1.11	4.44		40.00		4.44		1.11				
	ID		2.63		51.75		1.75		6.14			3.51				28.95	1.75	2.63				0.88		
	OP				100.00																			
	FW		5.88		11.76									82.35										
	TOTAL	0.16	1.93	0.16	44.69	0.16	2.09	0.16	5.14	0.16	0.16	2.57	0.16	4.02	0.16	32.64	0.32	2.09	0.32	0.16	0.16	2.25	0.16	0.16

*see footnote for Table 11 for definitions of land use categories
** the < sign without a value implies a non-detected value that was not identified

Constituent	Land use	<1	<5	<6	<10
Total Dissolved Solids (mg/L)	RE	14.29	71.43		14.29
	CO				100.00
	ID		50.00	50.00	
	OP		100.00		
	FW	100.00			
	TOTAL	15.00	55.00	5.00	25.00

Constituent	Land use	<0.5	<1	<2	<3	<4	<5	<10
Total Suspended Solids (mg/L)	RE	7.14	7.14				85.71	
	CO	12.50		12.50		62.50	12.50	
	ID		25.00			75.00		
	OP		50.00			50.00		
	FW							
	TOTAL	4.76	9.52	4.76		78.57	2.38	

Constituent	Land use	<1	<2	<3	<4	<5	<6	<10	<15	<20	<100
BOD (mg/L)	RE	8.70	13.04	8.70	4.35	34.78	8.70	4.35		8.70	
	CO	9.09	18.18	9.09		18.18	9.09	9.09	5.26		
	ID	5.26	31.58	5.26		47.37			5.26		
	OP	16.67	50.00	16.67						16.67	
	FW			50.00		50.00					
	TOTAL	33.05	20.34	6.78	3.39	18.64	2.54	1.69	0.85	3.39	1.69

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database - *Continued*

Constituent	Land use	<1	<2	<5	<10	<20	<25
COD (mg/L)	RE	33.33		55.56	11.11		
	CO	50.00		33.33	16.67		
	ID				50.00	25.00	25.00
	OP				20.00	80.00	
	FW				100.00		
	TOTAL	15.91	2.27	15.91	40.91	22.73	2.27

Constituent	Land use	<	<1	<2	<3	<10	<20	<30	<100	<200	<2000	>	>1.6K	>2.4K	>3K	>6K	>12K	>16K	>24K	>30K	>35K	>40K	>60K	>80K	>160K
Fecal Coliform (Colonies/100mL)	RE		7.84	3.92		1.96				3.92			25.49	3.92		1.96		13.73	1.96			1.96	21.57	3.92	7.84
	CO		17.86		3.57	3.57			3.57	3.57			10.71	3.57	3.57			10.71			3.57	3.57	28.57		3.57
	ID		44.44	11.11			2.78	2.78	2.78	5.56			2.78			2.78		8.33					11.11	2.78	2.78
	OP																								100.00
	FW																								
	TOTAL	0.67	22.00	4.00	0.67	2.67	0.67	0.67	1.33	3.33	2.67	0.67	11.33	2.00	2.00	2.00	2.67	2.67	0.67	0.67	0.67	1.33	15.33	2.67	5.33

Constituent	Land use	<	<1	<2	<4	<200	>16K	>35K	>60K	>80K	>100K	>160K	>240K
Fecal Streptococcus (Colonies/100mL)	RE	21.88	6.25			9.38	15.63	3.13	21.88	6.25	6.25	6.25	3.13
	CO	20.00	20.00	6.67	6.67		6.67		46.67	6.25			
	ID	50.00	8.33			8.33	25.00		8.33				
	OP					50.00	50.00				50.00		
	FW												
	TOTAL	27.94	13.24	1.47		5.88	14.71	1.47	22.06	5.88	2.94	2.94	1.47

Constituent	Land use	<	<0.01	<0.02	<0.04	<0.05	<0.1	<0.2	<0.27	<0.3	<0.5	<1
Ammonia (mg/L)	RE	30.00		0.91	0.91	13.64	18.18	14.55		1.82	20.00	
	CO	18.00				2.00	22.00	30.00	2.00		26.00	
	ID	23.81				2.38	7.14	28.57			35.71	2.38
	OP	19.23			3.85	7.69					69.23	
	FW							100.00				
	TOTAL	14.63	0.37	0.18	0.55	4.94	9.87	53.75	0.18	0.37	14.81	0.37

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database - *Continued*

NO2 + NO3 (mg/L)

Land use	<	<0.01	<0.02	<0.03	<0.05	<0.06	<0.1	<0.2	<0.3	<0.5	<1
RE	16.67		8.33	8.33			37.50	4.17			25.00
CO	25.00		12.50				50.00		12.50		
ID	6.25		6.25		6.25		43.75	6.25	6.25		25.00
OP							100.00				
FW										100.00	
TOTAL	15.85	1.22	8.54	4.88	4.88	1.22	35.37	4.88	6.10	3.66	13.41

TKN (mg/L)

Land use	<	<0.01	<0.05	<0.1	<0.2	<0.28	<0.3	<0.5	<1	<1.5
RE	25.81	3.23		3.23	19.35			25.81	19.35	3.23
CO	25.00	8.33			16.67			25.00	8.33	16.67
ID	16.67	5.56			11.11			22.22	38.89	5.56
OP			11.11					100.00		
FW				50.00					50.00	
TOTAL	15.71	2.86	0.71	2.86	30.71	0.71	7.86	22.14	13.57	2.86

Dissolved Phosphorus (mg/L)

Land use	<	<0.001	<0.01	<0.016	<0.02	<0.03	<0.04	<0.05	<0.06	<0.1	<0.12	<0.15	<0.2	<0.5
RE	41.03		0.85	1.71	11.97		1.71	19.66		12.82	0.85			8.55
CO	37.70		6.56	1.64	18.03	1.64		11.48	1.64		3.28			18.03
ID	36.59				4.88			29.27		12.20			2.44	14.63
OP			11.11					22.22		11.11				55.56
FW		100.00												
TOTAL	34.24	0.27	3.80	2.17	13.04	1.90	0.82	14.40	0.82	14.95	0.82		1.09	10.60

Total Phosphorus (mg/L)

Land use	<	<0.01	<0.02	<0.03	<0.04	<0.05	<0.06	<0.1	<0.12	<0.15	<0.5
RE	10.00	6.67	3.33	3.33		10.00		30.00	3.33		33.33
CO	15.79	5.26	10.53							10.53	57.89
ID	38.89		5.56			11.11	16.67	5.56			22.22
OP		14.29				28.57		14.29			42.86
FW				100.00							
TOTAL	13.91	4.35	9.57	3.48	0.87	7.83	2.61	20.87	2.61	6.09	27.83

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database – Continued

Total Copper (µg/L)

Constituent	Land use	<	<0.45	<1	<1.5	<2	<4	<5	<7	<8	<10	<15	<20	<25	<40	<41	<60
	RE		0.76	1.53	1.53	3.82	0.76	19.85	0.76	0.76	41.98	2.29	25.95				
	CO	3.57			7.14	7.14		14.29			57.14		17.86				
	ID					9.30		9.30		4.65	18.60		48.84	2.33	2.33	2.33	2.33
	OP	20.00									70.00		10.00				
	FW										100.00						
	TOTAL	2.62	0.29	0.87	0.58	4.94	0.87	14.24	0.29	1.16	39.53	0.87	32.56	0.29	0.29	0.29	0.29

Total Lead (µg/L)

Constituent	Land use	<	<0.2	<0.65	<0.7	<1	<1.5	<1.6	<2	<2.5	<3	<4	<5	<6	<7	<9	<10	<15	<20	<25	<30	<40	<42	<50	<53	<55	<60	<100	<200	<250	<500
	RE	10.6	0.4	0.4		0.4			2.2	4.0	9.3	1.3	21.7	0.4		7.5	5.8	1.3	4.4	5.8	0.9	3.5	0.4	8.9		0.9	1.3	5.3	3.1		
	CO	38.2			1.8	1.8		3.6	3.6	1.8		1.8	9.1			1.8	1.8				3.6	14.6	3.6	1.8				14.6			
	ID	17.5				3.1	1.0	1.0	3.1		3.1		13.4			8.3	8.3	2.1	8.3	5.2		3.1	1.0	9.3	1.0	3.1	2.1	10.3		1.0	1.0
	OP	11.5	3.9						3.9		7.7		11.5									19.2		19.2				23.1			
	FW																														
	TOTAL	14.7	0.3	0.2	0.2	6.2	0.2	0.2	2.6	1.7	4.9	3.3	12.4	0.2		2.06	14.3	1.1	2.9	4.3	0.9	3.6	0.6	6.1	0.3	1.1	0.9	6.4	2.0	0.2	0.2

Total Nickel (µg/L)

Constituent	Land use	<	<0.5	<1	<2	<2.5	<3	<4	<5	<7	<8	<10	<14	<15	<20	<25	<30	<40	<50	<60	<100
	RE	12.66		1.75	7.86	0.87	1.31	1.31	1.31	1.31		6.99		15.72	29.26	2.18	0.44	2.62	3.93	0.87	
	CO	22.34		1.06	6.38		1.06		14.89			6.38		7.45	36.17			2.13	1.06	1.06	
	ID	20.43			3.23			2.15	9.68	3.23	3.23	10.75	1.08	16.13	17.20	4.30	4.30	4.30	2.15		1.08
	OP	16.13		32.26				3.23	6.45						29.03			3.23	3.23		
	FW			10.00					60.00						30.00						
	TOTAL	15.30	0.17	1.22	8.17	0.52	1.74	0.87	12.00	1.57		7.65	0.35	14.26	27.30	1.57	1.04	2.26	2.96	0.87	0.17

Total Zinc (µg/L)

Constituent	Land use	<	<0.0057	<0.6	<1	<2	<4	<5	<8	<10	<20	<25	<30	<80	<100	<110	<200
	RE	13.79		3.45		10.34		3.45		31.03	13.79			3.45	13.79		6.90
	CO	25.00													50.00		25.00
	ID	16.67					16.67		16.67	16.67					16.67	16.67	16.67
	OP	38.46						15.38	15.38	15.38					15.38	15.38	
	FW							33.33							66.67		
	TOTAL	15.53	0.97	0.97	0.97	3.88	0.97	3.88	0.97	13.59	17.48	2.91		0.97	15.53	10.68	4.85

Total lead had the largest number of different detection limits (31 in total) with <10 μg/L as the most frequent censored observation at 14.3%. The constituent with the lowest number of detection limits was TDS, with four levels: < 1, < 5, < 6 and <10 mg/L. Less than 5 mg/L was the most common reported censored TDS observation occurring 55% of the time.

3.3.2 Expected Percentage of Observations at Different Levels of Detection

There are different approved methods to calculate the concentration of a specific constituent in a water sample. *Standard Methods for the Examination of Water and Waste Water* (APHA 1995) lists several approved methods for the detection of many of these constituents. The choice of methods presents a problem as these methods have varying features and costs. The objective is usually to select a method with a detection limit that results in useable data for most samples.

The distribution of the data, including the non-detected values, can be used to estimate the percentage of observations that will be detected using different analytical methods. Table 13 shows the expected percentage of observations below a specific detection limit for each of these constituents using the cumulative density function for each constituent and land use. For example, if a stormwater sample is collected in a freeway site and the detection limit of the method is 100 μS/cm, about 51% of the observations will be not detected.

Table 13. Percentage of Observations below Specific Concentrations

Constituent	Land use	Percentage of observations smaller than			
		20 µS/cm	100 µS/cm	200 µS/cm	2000 µS/cm
Conductivity (µS/cm)	RE	0	54	84	99
	CO	0	39	82	100
	ID	0	26	72	100
	OP	-	-	-	-
	FW	0	51	85	100
	TOTAL	0	39	73	99

Constituent	Land use	Percentage of observations smaller than				
		1 mg/L	4 mg/L	10 mg/L	160 mg/L	2500 mg/L
Hardness (mg/L)	RE	0	0	5	98	100
	CO	0	4	7	91	100
	ID	0	0	3	95	100
	OP	-	-	-	-	-
	FW	0	0	2	96	100
	TOTAL	0	0.1	3	94	100

Constituent	Land use	Percentage of observations smaller than				
		0.5 mg/L	1 mg/L	2 mg/L	5 mg/L	10 mg/L
Oil and Grease (mg/L)	RE	2	19	31	75	91
	CO	1	11	23	64	87
	ID	1	20	31	66	86
	OP	-	74	-	-	-
	FW	2	5	5	55	75
	TOTAL	0.3	17	29	67	84

Constituent	Land use	Percentage of observations smaller than		
		1 mg/L	5 mg/L	10 mg/L
Total Dissolved Solids (mg/L)	RE	0	0.8	1.5
	CO	0	0	0.5
	ID	0	0	2
	OP	0	0	0
	FW	0	0	0
	TOTAL	0.1	0.7	1.5

Constituent	Land use	Percentage of observations smaller than		
		1 mg/L	5 mg/L	10 mg/L
Total Suspended Solids (mg/L)	RE	0.2	4	11
	CO	0.2	3	9
	ID	0.2	3	5
	OP	0	11	23
	FW	0	2	2
	TOTAL	0.2	3	7

Constituent	Land use	Percentage of observations smaller than		
		1 mg/L	2 mg/L	5 mg/L
BOD (mg/L)	RE	0.2	2	18
	CO	0.2	1	16
	ID	0.2	3	18
	OP	2	11	55
	FW	0	0	31
	TOTAL	1	3	22

Table 13. Percentage of Observations below Specific Concentrations - *Continued*

Constituent	Land use	Percentage of observations smaller than				
		0.7 mg/L	1 mg/L	5 mg/L	10 mg/L	20 mg/L
COD (mg/L)	RE	0	0.4	1	3	9
	CO	0	1	2	3	7
	ID	0	0	0.5	2	7
	OP	0	0	0	7	37
	FW	0	0	1	4	7
	TOTAL	0	0.2	2	5	13

Constituent	Land use	Percentage of observations smaller than				
		0.01 mg/L	0.05 mg/L	0.1 mg/L	0.2 mg/L	0.5 mg/L
Ammonia (mg/L)	RE	0	3	12	36	71
	CO	0	2	9	28	53
	ID	0	1	7	21	57
	OP	0	11	15	22	93
	FW	0	0	5	20	27
	TOTAL	0.1	2	10	37	65

Constituent	Land use	Percentage of observations smaller than				
		0.01 mg/L	0.05 mg/L	0.1 mg/L	0.2 mg/L	0.5 mg/L
NO2 + NO3 (mg/L)	RE	0	2	5	11	40
	CO	0	1	4	11	40
	ID	0	2	6	11	31
	OP	0	0	18	21	50
	FW	0	0	0	28	72
	TOTAL	0	2	4	10	40

Constituent	Land use	Percentage of observations smaller than				
		0.01 mg/L	0.05 mg/L	0.1 mg/L	0.2 mg/L	0.5 mg/L
TKN (mg/L)	RE	0.1	0.1	0.5	2	6
	CO	0.2	0.2	0.7	2	6
	ID	0.2	0.5	0.7	2	8
	OP	0	0	0	0	44
	FW	0	0	2	2	6
	TOTAL	0.1	0.2	0.6	2	10

Constituent	Land use	Percentage of observations smaller than				
		0.01 mg/L	0.02 mg/L	0.05 mg/L	0.1 mg/L	0.5 mg/L
Dissolved Phosphorus (mg/L)	RE	0.3	3.5	11	32	93
	CO	1	6	21	48	91
	ID	0.3	2.2	16	46	95
	OP	2	7	23	45	93
	FW	5	5	5	14	82
	TOTAL	0.7	4.5	17.5	44.5	94

Constituent	Land use	Percentage of observations smaller than				
		0.01 mg/L	0.02 mg/L	0.05 mg/L	0.1 mg/L	0.5 mg/L
Total Phosphorus (mg/L)	RE	0.2	0.4	1.5	10	28
	CO	0.2	0.6	3	16	82
	ID	0	0.2	3	14	74
	OP	2	2	11	24	80
	FW	0	0	0	3	83
	TOTAL	0.1	0.5	3	12	78

Table 13. Percentage of Observations below Specific Concentrations - *Continued*

Constituent	Land use	Percentage of observations smaller than				
		2 µg/L	5 µg/L	10 µg/L	20 µg/L	40 µg/L
Total Copper (µg/L)	RE	2.3	14	44	76	92
	CO	0.7	6	26	58	84
	ID	1.2	6	16	46	75
	OP	0	32	54	73	92
	FW	0	0	8	26	58
	TOTAL	1.4	9	31	63	85

Constituent	Land use	Percentage of observations smaller than				
		1 µg/L	3 µg/L	5 µg/L	10 µg/L	50 µg/L
Total Lead (µg/L)	RE	2	14	28	47	88
	CO	0.6	3	8	23	80
	ID	0.7	7	12	24	72
	OP	12	21	33	38	76
	FW	0	3	9	22	72
	TOTAL	2	9	17	36	82

Constituent	Land use	Percentage of observations smaller than				
		1 µg/L	2 µg/L	5 µg/L	10 µg/L	20 µg/L
Total Nickel (µg/L)	RE	1	6	33	55	91
	CO	0.5	3	29	56	92
	ID	0	2	12	33	64
	OP	0	30	39	39	73
	FW	0	1	19	55	84
	TOTAL	0.6	5	26	52	84

Constituent	Land use	Percentage of observations smaller than				
		5 µg/L	10 µg/L	20 µg/L	100 µg/L	200 µg/L
Total Zinc (µg/L)	RE	1	3	7	65	87
	CO	0	0.2	1	28	51
	ID	0.2	0.7	1	24	48
	OP	5	25	35	85	92
	FW	1	2	3	20	51
	TOTAL	0.6	2	4	44	73

Appendix D describes the methods used to analyze censored observations for each constituent. Based on the results presented in Table 13 and these methods, it is possible to estimate the percentage of non-detected observations that can be obtained by constituent and land use. For example, the most frequent non-detected ammonia detection limit was 0.2 mg/L. About 37% of the detected and non-detected observations were located below this detection limit. One of the EPA approved methods to measure ammonia uses the

Nessler's reaction. The detection limit of this method is close to 0.02 mg/L. If this method was commonly used, the number of non-detected ammonia observations would have been significantly reduced. Of course, other factors must be considered when selecting analytical methods besides detection limits, including sample preparation, sample storage limits, sample volume needed, safety, cost, disposal problems associated with wastes, interferences, comparisons with other methods, etc. Burton and Pitt (2002) present a review of many alternative analytical methods that may be suitable for stormwater samples.

3.3.3 Effects of Non-detected Observations on Calculating Mean and Standard Deviation Values

The selection of the proper procedure to deal with non-detected values is not an easy task. One option is to ignore the non-detected values and make a statement indicating the percentage of non-detected values found in the dataset. The problem arrives when it is desired to calculate the mean and standard deviation values of a dataset. The presence of non-detected values can strongly bias these parameters, depending on their prevalence. Three methods for dealing with non-detected values were explored during this research: 1) Ignore them; 2) Estimate them with the Cohen's multi level MLE method for left censored data (NCASI 1995); and 3) replace them with half of the detection limit. In cases were Cohen's method could not be used (i.e. when only two values were detected), half of the detection limit was used as the estimated value to replace the non-detected observations. It was expected that for low percentages of non-detected data, their effects on the mean, standard deviation, and coefficient of variation

would be minimal. There effects were expected to increase as the level of censored observations increases.

Appendix D shows the results for each constituent and land use using the three substitution methods. In general, it was observed that if the censored data was deleted, the mean of the constituent was larger than in the case where the non-detected values were replaced by half of the detection limit. The same pattern was observed for the standard deviation calculations. The behavior for the coefficient of variation was opposite: the coefficient of variation was reduced when the censored observation was deleted.

When the frequencies of the censored observations were lower than 5%, the means, standard deviations and coefficients of variation were almost identical when the censored observations were replaced by half of the detection limit, or estimated using Cohen's Method. As the percentage of non-detected values increases, replacing the censored observation by half of the detection limit instead of estimating them using the Cohen's maximum likelihood method produces lower means and larger standard deviations.

3.3.4 Effects on Mean, Median and Coefficient of Variation Values at Different Percentages of Censored Observations

As noted above, when the percentage of detected values is high, there are minimal changes in the calculated means, standard deviations, and coefficients of variation for any of the replacement methods.

In this discussion, the ratios of the calculated values using the different methods for different frequencies of detection are examined. This analysis will identify the sensitivity of the detection frequencies for each substitution method.

The first task was to evaluate the effect of the substitutions and detection frequencies on the calculated means. When the percentage of detected values is close to 100%, all of the substitution methods produce the same mean, as expected. As the percentage of non-detected values increases, the Cohen's estimated values and half of detection limit methods produces smaller means than if ignored.

Figure 15 is a scatter plot of both ratios (Cohen estimated/ignore and half of the detection limit/ignore) of the calculated mean values. If the scatter plot values formed a line near the 1.0 ratio value, then the "ignore" and the other option would be accurate. If the scatter plot values formed the same line for both of the sets of ratios, then either substitution method would be accurate.

The regression equation 3.1 for the Cohen estimated/ignore ratio of calculated mean values has a coefficient of determination of almost 93%. The coefficients in the equation are significant, with a probability that the coefficients are equal to zero smaller than 0.0001.

(3.1)   *Ratio Mean (Estimated/Ignore) = 0.316 + 0.0068*D*

Where D is the percentage of detected values (0 to 100).

Effect on the Mean

| ● Estimated / Ignore |
| O Half Detection / Ignore |
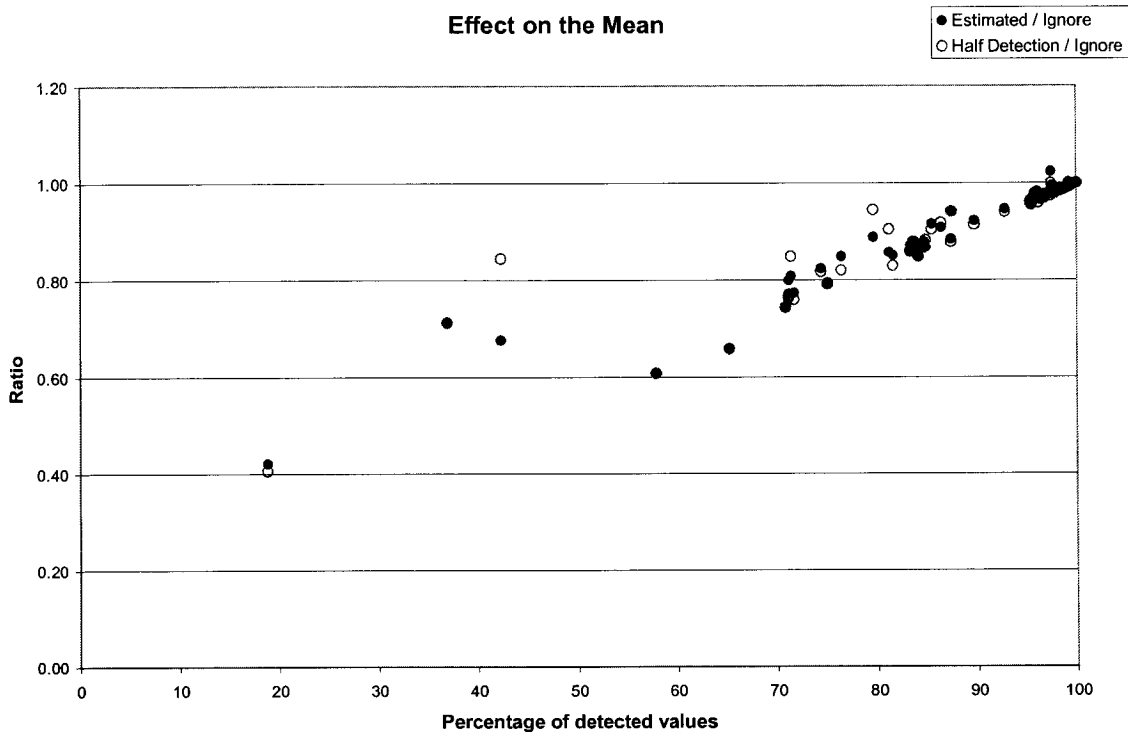
Fig. 15. Effects on the mean when using random estimated values versus ignoring the non-detected observations, at different percentage of detected values

For percentages of detected values smaller than 60%, the ratios are located away from the line formed by the other observations. The residual plot of the regression indicates those observations that are most affecting the departure from the regression line. Six observations are considered influential in this plot: oil and grease in open space (most influential), residential and industrial land uses, plus ammonia and lead in open space land uses. The Cook's distance procedure was used to remove the overly influential points in the regression. After removing the influential observations the final regression is therefore:

(3.2) $Ratio\ Mean\ (Estimated/Ignore) = 0.248 + 0.0075*D$

Equation 3.2 indicates that a dataset having 30% non-detectable observations would have an expected reduction in the calculated mean of 23% when the censored data is appropriately estimated instead of being ignored. The standard deviation of the residuals is 0.014. The coefficient of determination in this case was higher than 96% with no potential or influential points. This equation can be used to estimate the mean of the distribution for data sets with percentages of detected values higher than 60%. When the non-detected observations are replaced by half of the detection limit, the coefficient of determination was reduced to 92% of the actual value. Equation 3.3 describes the relationship between the ratio of the means and the percentage of detected observations.

(3.3)   *Ratio Mean (Half Detection/Ignore) = 0.250 + 0.0075*D*

From the regression of the ratios "estimated/ignore" and "half detection/ignore," replacing by half of the detection limit, or estimating the censored observations using Cohen's method, will produce the same results when the percentage of detected observations is larger than 80%.

The effects on the median are similar to those observed in the mean. When the non-detected values are estimated with Cohen's method instead of ignoring the non-detected values, the regression of the coefficient of determination reduces to 86%.

Effect on the Median
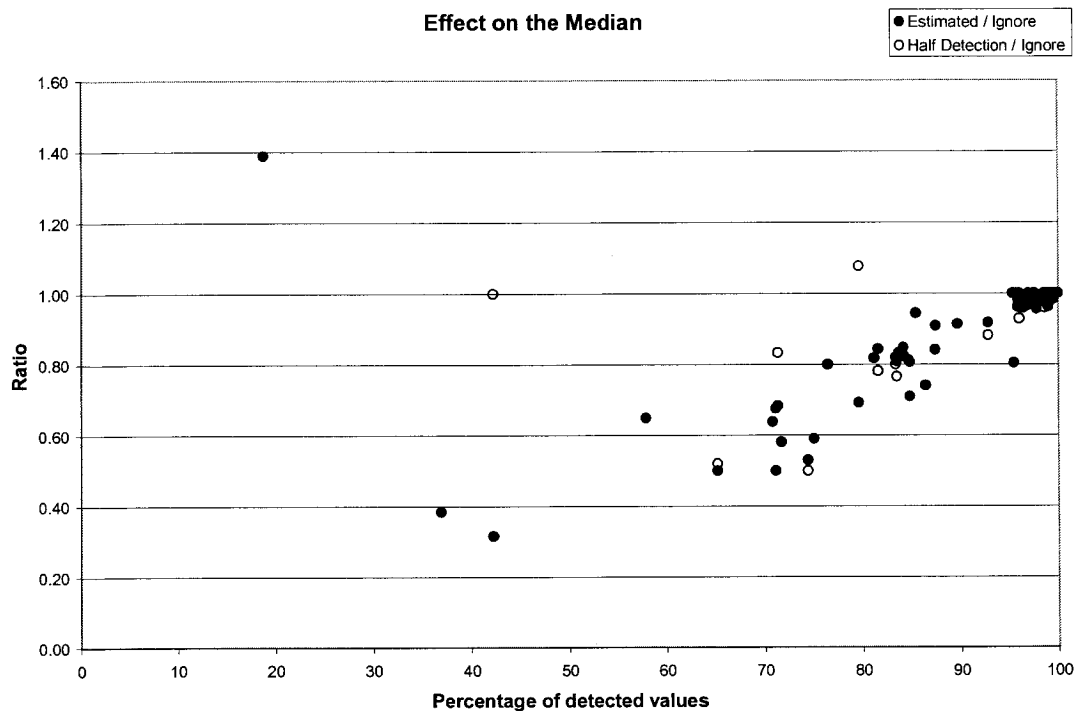
Fig. 16. Effect of ignoring the non-detected observations on the median

Equation 3.4 shows the estimated regression line for the median case.

(3.4)   *Ratio Median (Estimate /Ignore) = -0.326 + 0.0134*D*

This equation is valid for percentage of detected observations higher than 70%. A reduction of 40% in the median value is expected in a 30% censored dataset when the non-detected observations are estimated using Cohen's method instead of being ignored. The standard deviation of the residuals for this equation is 0.05.

When the censored observations are replaced by half of the detection limit, the coefficient of determination is about 73%. The regression equation for the ratio of the median is therefore not as good in explaining the variability as it was for the mean.

Equation 3.5 shows the calculated regression line for the median when the non-detected values are replace by half of the detection limit.

(3.5)   *Ratio Median (Half Detection/Ignore) = -0.195 + 0.012*D*

This equation is valid when the percentage of detected observations is higher than 70%. Replacing the censored observations by half of the detection limit has the same effect on the median as estimating them using Cohen's method, except for dissolved and total phosphorus in open space and lead in residential land uses.

The effects on the calculated standard deviation values also indicate a good correlation between the level of detected observations and the ratio between the "estimate the non-detected or ignore them" values. Figure 17 shows the scatter plot of the median values as a function of the percentage of detected observations. Equation 3.6 presents the estimated regression line of these data.

(3.6)   *Ratio Standard Deviation (Estimate/Ignore) = 0.68 + 0.003226*D*

Effect on the Standard Deviation

| ● Estimated / Ignored |
| O Half Detection / Ignored |
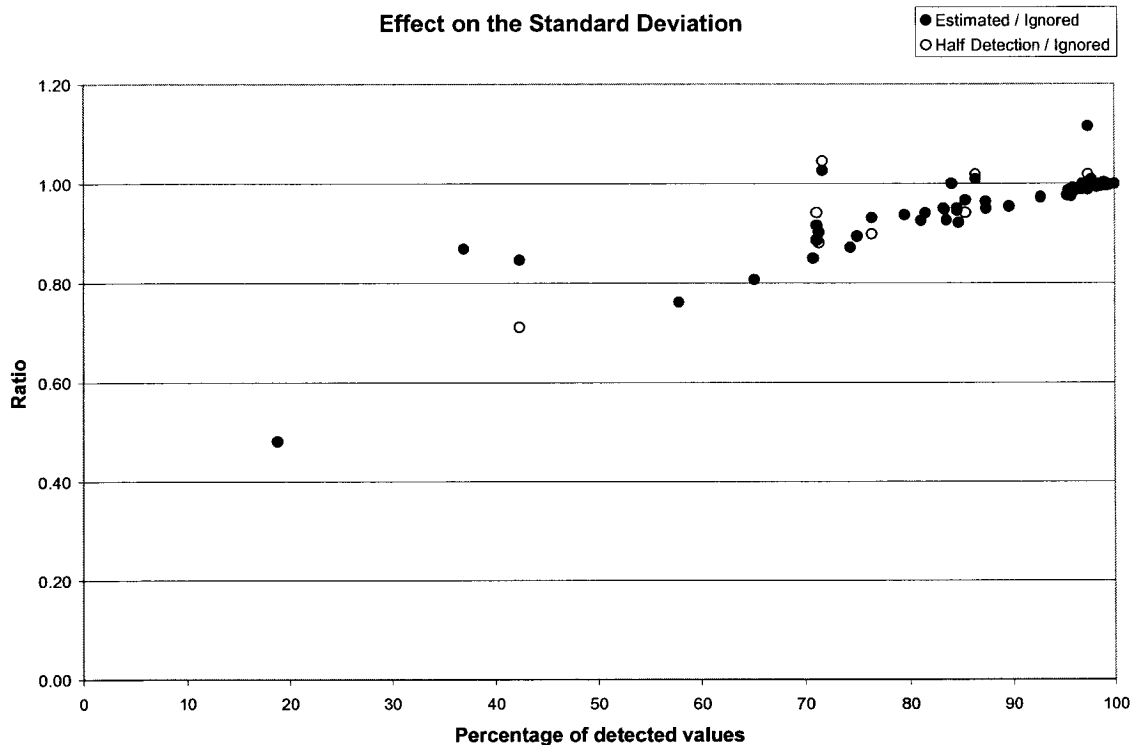
Fig. 17. Effect of ignoring the non-detected observations in the median

The regression has a low coefficient of determination (56%) compared to the prior

regressions. Oil and grease at freeway sites was considered unusual according to its

Cook's distance. The data for this case was examined and no reason was found to

eliminate it from the analysis. It was also observed that BOD in commercial land use

areas had 3 right-censored observations. Because the Cohen method must be used with

left censored observations, these data were eliminated from this analysis. Observations

where the percentage of detection was smaller than 70% were not included. Equation 3.7

shows the estimated regression line for those constituents with more than 60% detected

observations.

(3.7)   *Ratio Standard Deviation (Estimate/Ignore) = 0.68 + 0.003226*D*

This equation indicates that for a dataset with 30% censored observations, the standard deviation will be reduced by 9.5% when the non-detected observations are estimated instead of ignored. The standard deviation of the residuals is 0.023. When the censored observations are replaced by half of the detection limits, the coefficient of determination and the equation coefficients were almost the same. Equation 3.8 presents the estimated regression equation for the standard deviation when the censored observations are replaced by half of the detection limits.

(3.8)   *Ratio Standard Deviation (Half Detection/Ignore) = 0.6778 + 0.00325*D*

The last parameter examined was the coefficient of variation. The coefficient of determination (69%) for the fitted regression equation was better than for the standard deviation regression, but not as high as for the median and mean regressions. The calculated regression equation is presented as equation 3.9

(3.9)   *Ratio Coefficient of Variation (Estimate/Ignore) = 1.53 -0.0053*D*

The standard deviation of the residuals is 0.033. As the number of non-detected observations increases, the coefficient of variation also increases. The regression equation is valid for percentages of detected values higher than 70%. For a data set with 30% censored observations, the expected coefficients of variation using Cohen's method will be 16% higher than if the non-detected values are ignored.

In the case that the censored observations are replaced by half of the detection limits, the coefficient of determination of the resulting equation (equation 3.10) is reduced to 58%. Figure 18 shows the scatter plot for the ratios "estimated/ignore" and "half detection/ignore" for the coefficient of variation.

(3.10)   *Ratio Coefficient of Variation (Half Detection/Ignore) = 1.543 -0.0054*D*
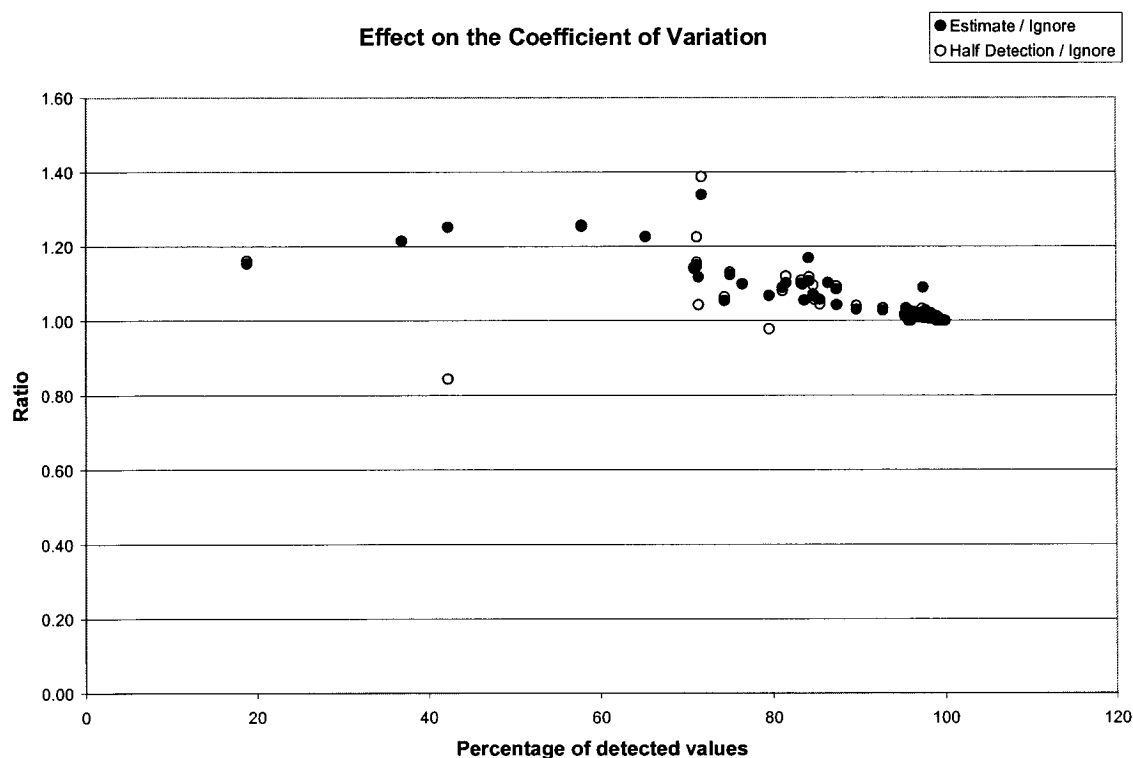
Fig. 18. Effect of ignoring the non-detected observations on the coefficient of variation

3.3.5 Total Suspended Solids Analyses at Different Levels of Censoring

To evaluate the effect of the non-detected values in the mean and standard deviation observations at different levels of censoring, one of the constituents with low percentages of non-detected observation (TSS) was trimmed in the lower tail until

reduced to 50% of the original distribution. All TSS observations were used during this

analysis.

Table 14. Descriptive Statistics for TSS Truncated at Different Levels

Total number of samples	% of original samples	Minimum concentration in set (mg/L)	RATIO			
			Average	Median	Standard Deviation	Coefficient of Variation
2025	100.00	3	1.00	1.00	1.00	1.00
2015	99.51	3	1.00	1.00	1.00	1.00
1995	98.52	4	0.99	0.98	0.99	1.01
1974	97.48	5	0.98	0.96	0.99	1.02
1954	96.49	6	0.97	0.95	0.99	1.02
1934	95.51	7	0.96	0.95	0.98	1.03
1914	94.52	8	0.95	0.93	0.98	1.03
1873	92.49	10	0.93	0.90	0.97	1.05
1833	90.52	11	0.91	0.87	0.96	1.06
1792	88.49	13	0.89	0.84	0.96	1.07
1752	86.52	15	0.87	0.81	0.95	1.08
1711	84.49	17	0.86	0.78	0.94	1.10
1671	82.52	18	0.84	0.74	0.93	1.11
1630	80.49	20	0.82	0.73	0.92	1.12
1589	78.47	22	0.80	0.71	0.91	1.14
1545	76.30	24	0.78	0.68	0.90	1.15
1496	73.88	26	0.76	0.65	0.89	1.17
1468	72.49	27	0.75	0.64	0.89	1.18
1428	70.52	29	0.73	0.63	0.88	1.20
1387	68.49	31	0.72	0.60	0.87	1.21
1347	66.52	33	0.70	0.58	0.86	1.23
1306	64.49	35	0.68	0.57	0.85	1.24
1266	62.52	37	0.67	0.55	0.84	1.26
1225	60.49	40	0.65	0.53	0.83	1.28
1185	58.52	42	0.63	0.52	0.82	1.29
1144	56.49	44	0.62	0.50	0.81	1.31
1104	54.52	47	0.60	0.47	0.80	1.33
1063	52.49	50	0.58	0.46	0.79	1.35
1023	50.52	52	0.57	0.44	0.78	1.37

The results are similar to those observed during the analysis of the censoring

observations within multiple constituents and land uses. Real mean, median and standard

deviation are smaller than the calculated values when censored observations are ignored

(Figure 19). The true coefficients of variation are larger than those calculated when the level of trimming is increased.
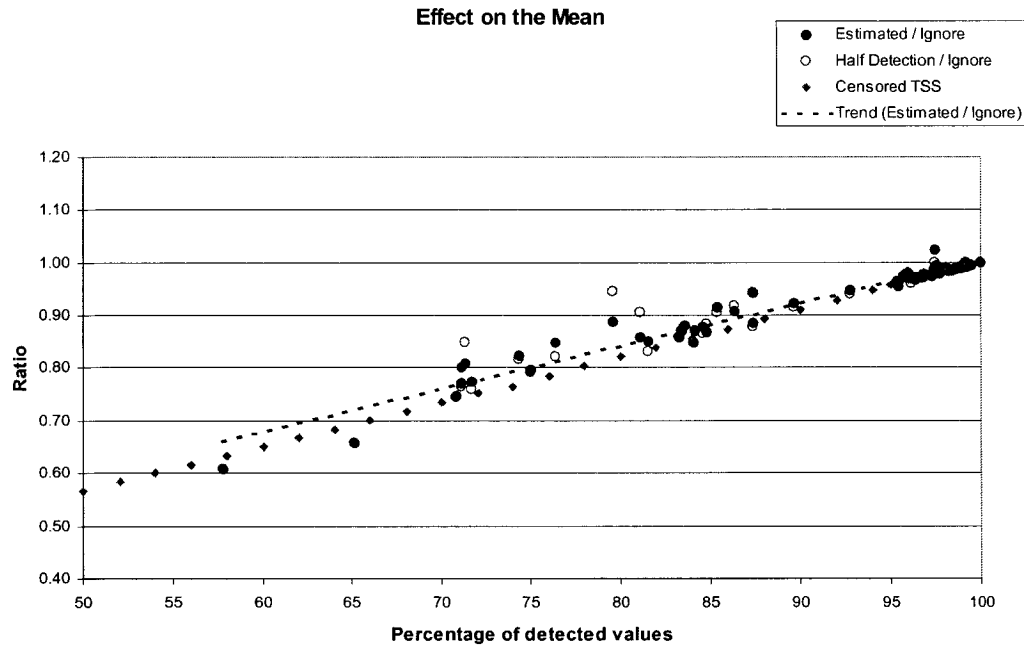
Effect on the Mean

Fig. 19. Effect on the mean when TSS observations are truncated

The effect on the mean indicates that only when about 5% of the data is censored or is trimmed, the ratios "replace/ignore," "estimated/ignore," or "trimmed/total" observations produced the same results in the mean of the distribution. When the percentage of non-detected observations is increased, the ratios "estimate/ignore" and "half detection/ignore" are higher than the ratio "trimmed/complete" in the TSS distribution. This means that trimming the data set has a larger effect than when the observations are censored. This is explained because for the trimmed/complete ratios, all the censored observations were at one value. In the other case, several detection limits were used during the analysis.

In the previous discussion, it was observed that censored levels less than 30% can

be used for predicting simple statistics about the distribution. The previous figure

indicates that levels of censoring close to 45% followed the trend indicated by the ratio

"trimmed/complete." This indicates that even if the regression analysis was

recommended for levels of non-detected values smaller than 30%, they can be used for

levels of censoring up to 45%.

The effects on the medians are stronger than on the means. When the level of

censored observations is close to 30%, the ratio "trimmed/complete" is close to 0.6,

compared with 0.75 in the case of the mean (Figure 20). Levels of censoring around 5%

do not show the straight-line pattern that it was observed with the mean. The trend for

censoring levels between 5 and 45% is similar for the "estimated/ignore" ratio; however

the dispersion around the trend line is higher.
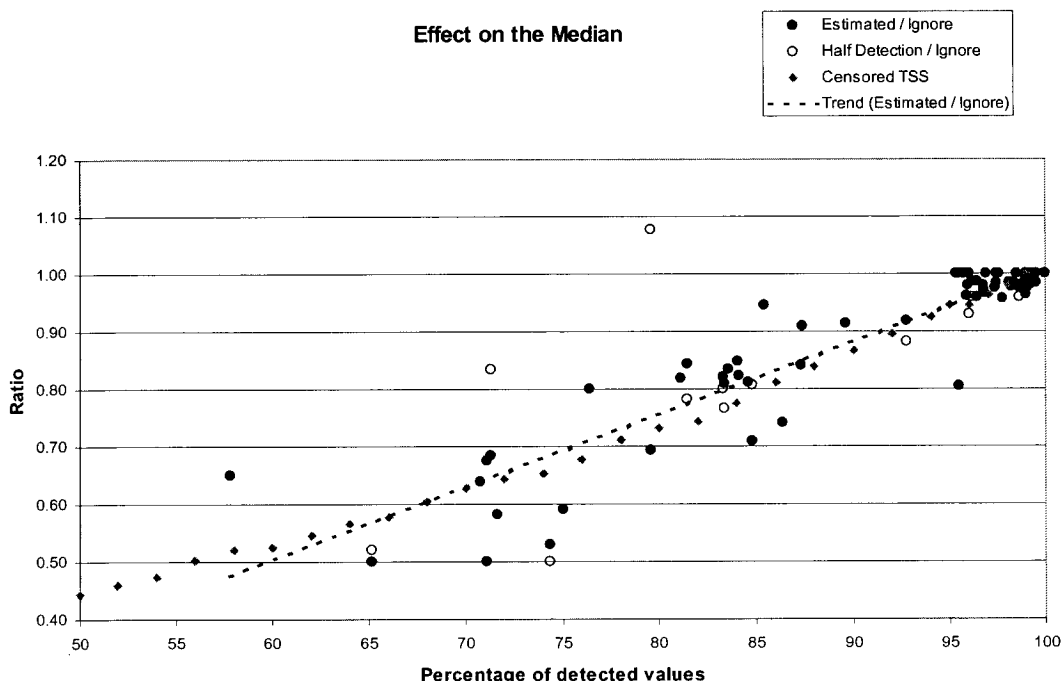
Fig. 20. Effect on the median when the TSS dataset is truncated

The effect on the standard deviation of the trimming the TSS is similar to the

effect in the mean (Figure 21). When the level of censoring is close to 30%, the ratio

"trimmed/complete" is close to 0.85. The dispersion around the trend line is lower than in

the median case. When the percentage of non-detected values is lower than 5%, the ratios

"estimated/ignore," "half detection/ignore," and "trimmed/complete" are almost the

same. For levels of censored observations larger than 15%, the differences among the

ratios increase.
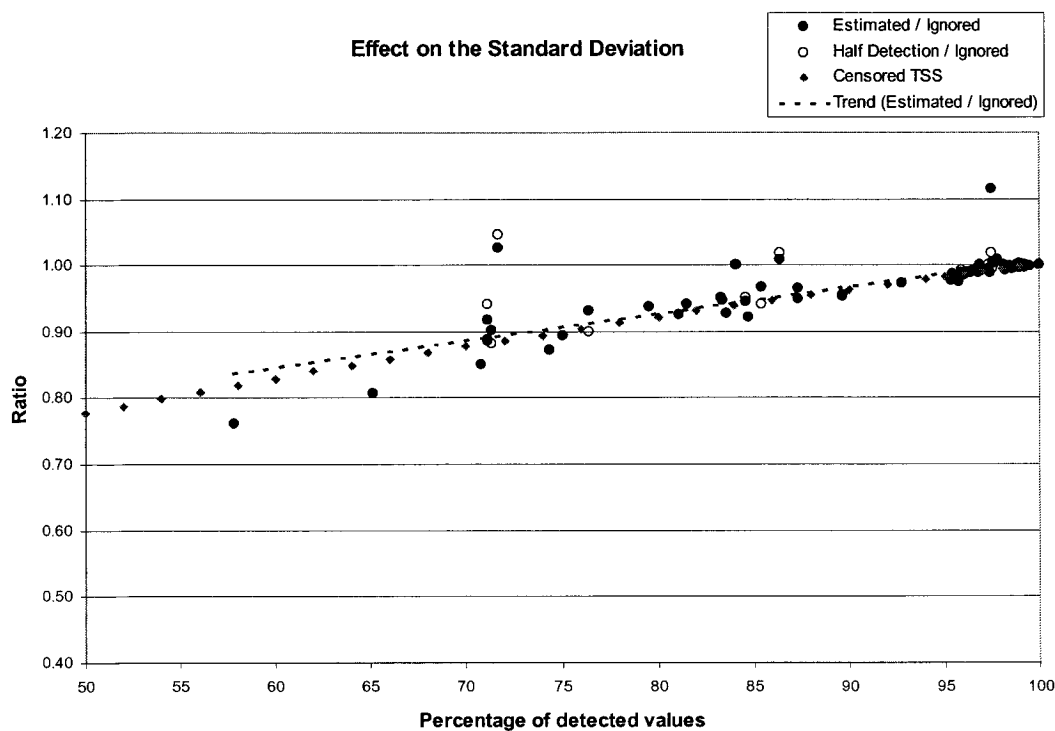
Fig. 21. Effect on the standard deviation when the TSS dataset is truncated

The ratio of the effects on the calculated coefficients of variation has a different

slope than the previous statistics. As in the mean case when the level of censoring is

smaller than 5%, a linear trend between the percentage of detected and the ratio was

observed (Figure 22). When the percentage of censored observations is larger than 15%, the differences among the three ratios increase.
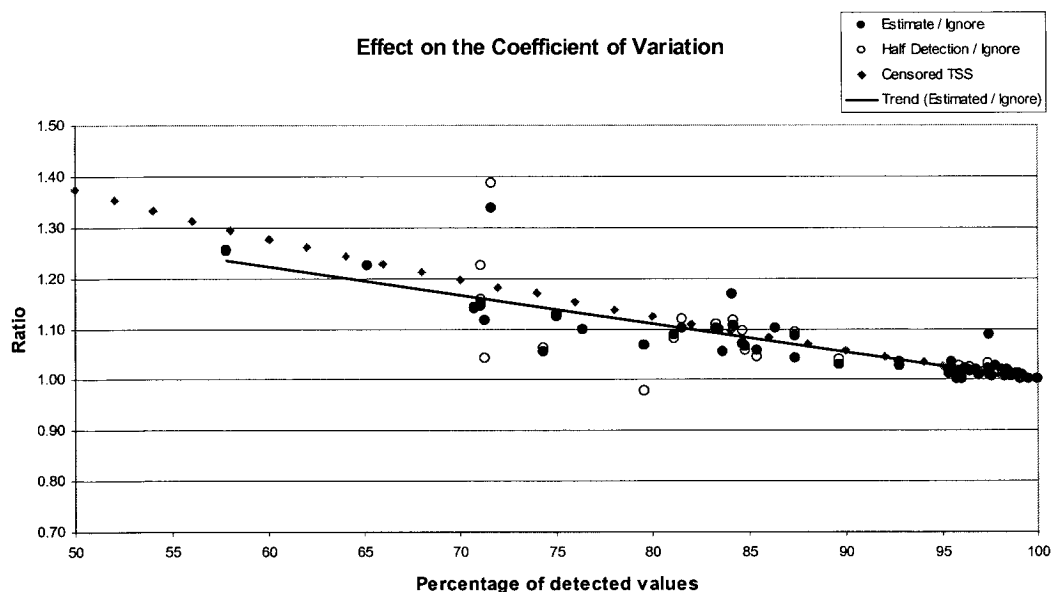
Fig. 22. Effect on the coefficient of variation when the TSS dataset is truncated

3.4 Unusual Monitoring Locations

Box and whisker plots can be used as a preliminary examination of the principal factors and interactions between EPA Rain Zones and land use for any constituent. These plots can also be used to identify sites that do not fit within an established pattern shown by other land use locations from other regions of the country. Figure 23 shows box and whisker plots for residential, commercial, and industrial land uses for EPA Rain Zones 1 through 9. These plots indicate that there are significant differences between EPA Rain Zones and between land uses. Statistical tests also found that the interaction of these two factors was also significant. The median observations by land use have patterns similar to those found during NURP (EPA 1983), and other studies. Residential and open space areas have lower concentrations than commercial and industrial land use areas.
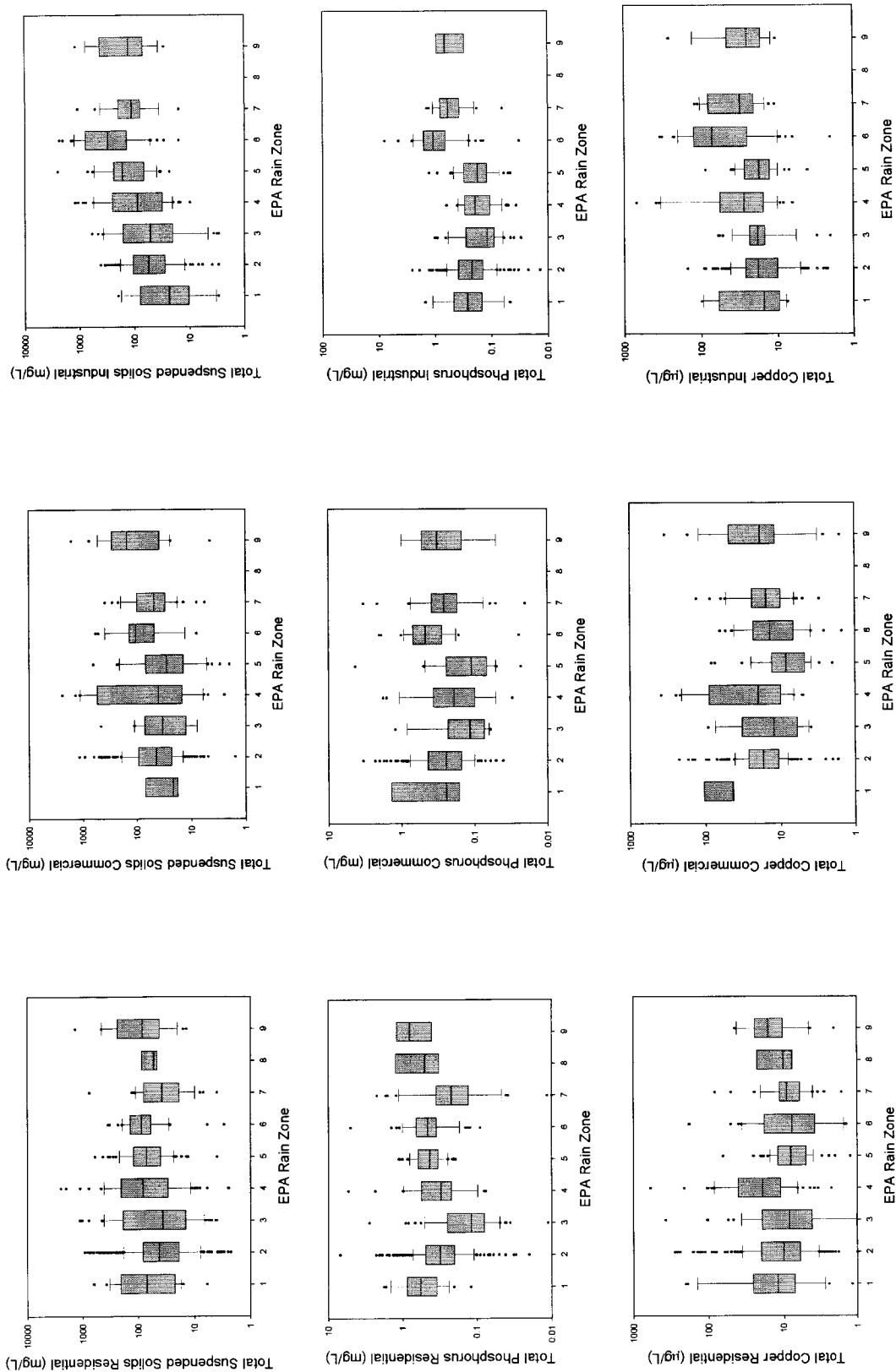
Fig. 23. Box and whisker plots for TSS, total phosphorus and total copper by EPA Rain Zone and land use (Two-way ANOVA analyses indicate that both land use and geographical main factors and the interaction are significant at the 0.1% level)

Residential, commercial and industrial areas are the single land uses having the most observations in the database. These three land uses were analyzed separately to identify those sites with different characteristics than the remaining sites in the same land use and EPA Rain Zone. The following is an example using TSS at residential land use sites to demonstrate the method used to detect unusual monitoring sites in the database. Summaries of additional constituents in residential, commercial and industrial land uses are given in Appendix D.

3.4.1 Example Using Single Residential Land Use

The following example explains the steps used to identify unusual locations in the database. This analysis was performed in three steps. First, box and whisker plots we used to identify any site with concentrations unusually high or low compared with the other residential locations. The plot was used to identify preliminary differences between and within EPA Rain Zones. Figure 24 shows that there are some sites in EPA Rain Zone 2 having lower TSS concentrations than the remaining residential sites included in the database. On the other hand, it seems that sites located in EPA Rain Zone 4 have higher concentrations than other groups. The second step was to identify those single residential sites that failed the Xbar and S chart tests for all the observations and by EPA Rain Zone.

A total of 10 Xbar and S charts were created for each EPA Rain Zone and for all the zones combined. An indication of geographical differences is if the Xbar chart using all observations shows clusters close or outside the control limits. The effect will be confirmed if none of the sites failed the Xbar test within EPA Rain Zones. The S chart identifies those sites that have a larger or smaller variation than the overall sites in the set.

Figure 25 shows the Xbar and S chart for the residential land use sites. Six sites have mean TSS values different from the remaining sites in the same group. One important characteristic of this plot is that the control limits change with the number of samples collected at each site. The S chart identifies those sites with standard deviations different than the pooled deviation of the data set. In this case, two sites are outside the control limits. Table 15 shows the sites that failed the Xbar and S chart for all residential sites and for each EPA Rain Zone. Table 15 shows that most of the sites located below the lower control limit were located in North Carolina, Virginia (EPA Rain Zone 2) or Oregon (EPA Rain Zone 7). Sites above the upper control limit were located in Arizona (EPA Rain Zone 6), Kansas (EPA Rain Zone 4), and Colorado (EPA Rain Zone 9).

Xbar plots by EPA Rain Zones also indicate differences within groups. EPA Rain Zones 2, 3, and 4 showed nine sites failing the Xbar test. Six sites out of 54 failed the Xbar chart test in residential land use EPA Rain Zone 2. Each of these sites will be described individually.

The first site was located in Kentland Village (Flagstaff Street), in Prince George County, Maryland (Location_ID = MDPGCOS2, median TSS = 132 mg/L). This site with 63 events has the largest number of observations in the database. An industrial park and a commercial area surrounded this high-density residential site. A special characteristic of this site is the construction of a stadium close of the watershed during the monitoring period.

Boxplot of TSS in Residential Land Uses

Total Suspended Solids mg/L

1.000 100 10

EPA Rain Zone and Location_ID

1-2   2-3   3-4   4-5   5-6 6-7   78 9
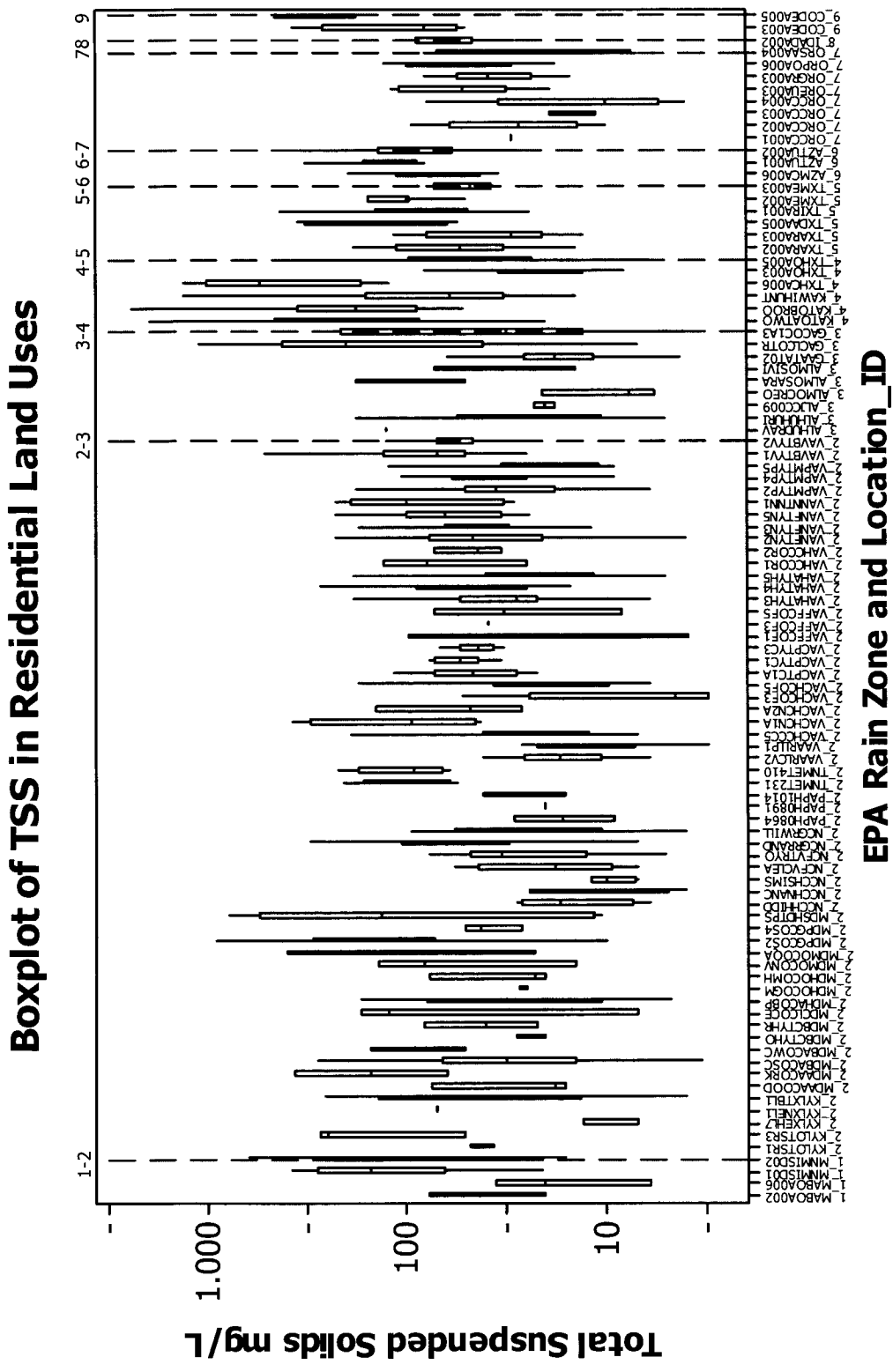
Fig. 24. TSS box and whisker plots in residential land use by EPA Rain Zone and location

Xbar-S Chart of TSS in Residential Land Use - EPA Rain Zone
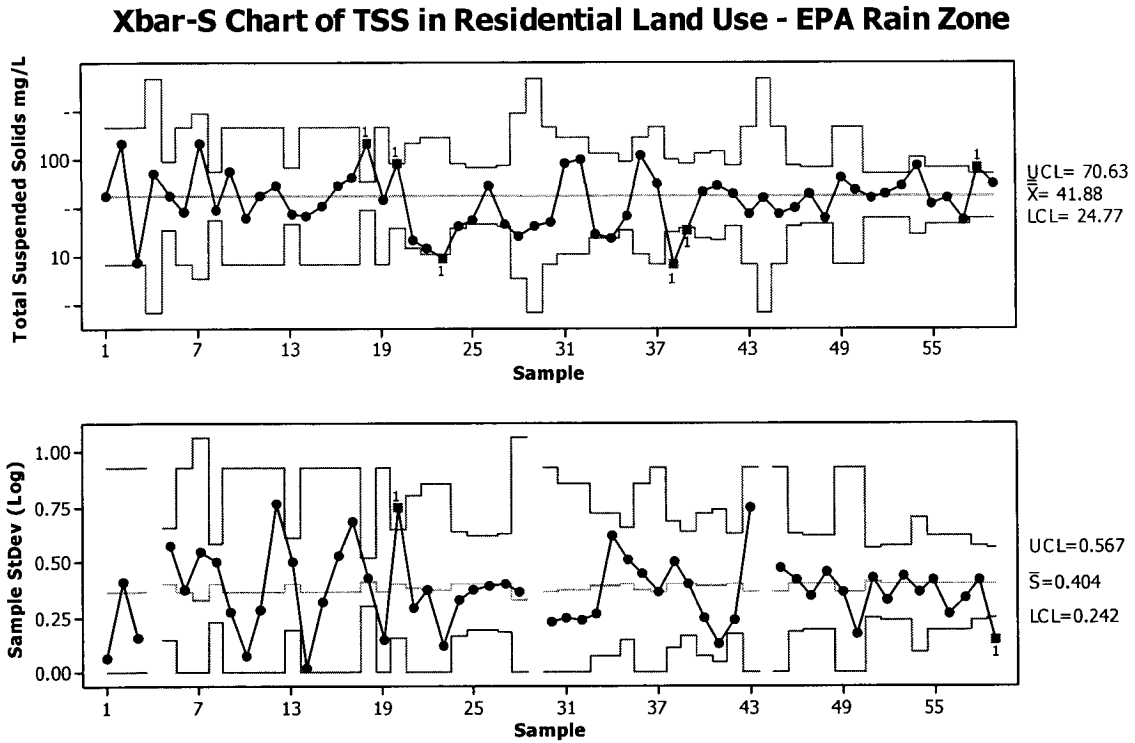
Fig. 25. Xbar and S chart for residential land use in EPA Rain Zone 2

Table 15. Sites Failing Xbar and R Chart in Residential Land Uses

Rain Zone	Sites Failing Xbar chart				Sites Failing S Chart
ALL	AZTUA001(H) KATOATWO(H) MNMISD01(H) NCGRWILL(L) VAARLCV2(L) VAHATYH5(L)	CODEA005(H) KATOBROO(H) NCCHSIMS(L) ORCCA004(L) VAARLLP1(L) VAPMTYP5(L)	GAATAT02(L) KYLXEHL7(L) NCFVCLEA(L) TXHCA006(H) VACHCOF3(L)	GACLCOTR(H) MDPGCOS2(H) NCFVTRYO(L) TXHOA003(L) VACHCOF5(L)	VAVBTYV2(L)
1	None				None
2	MDPGCOS2(H) VACHCOF5(L)	MDSHDTPS(H) VAVBTYV1(H)	NCCHSIMS(L)	VACHCOF3(L)	MDSHDTPS(H) VAVBTYV2(L)
3	GACLCOTR(H)				None
4	TXHCA006(H)	TXHOA003(L)			None
5	None				None
6	None				None
7	None				None
8	None				None
9	None				None

The second site has 13 observations and was operated by the Maryland State Highway Department (MDSHDTPS, median TSS = 135 mg/L). This 51-acre site is considered 96% single family residential, with 4% agricultural land use. The site is located close to the intersection of two highways. Observed concentrations ranged from 10 mg/L and up to 750 mg/L. The highest concentrations were observed in summer and the lowest in spring. The last site in EPA Rain Zone 2 with elevated values has 26 observations and is located close to Bow Creek in Virginia Beach, VA (VAVBTYV1, median TSS = 69). This site is located close to a golf course and is drained by a natural channel.

The site with a standard deviation below the lower control limit (VAVBTYV2) is located next to VAVBTYV1. It has also a high TSS concentration but inside the control limits. A total of 30 samples were collected at VAVBTYV2. The aerial photograph did not indicate any unusual conditions at this site.

In EPA Rain Zone 4, only one site had high concentrations compared with the remaining residential sites. This site (TXHCA006) is located in Harris County, TX. Six samples were collected, having a median TSS of 550 mg/L. This site is also analyzed in Chapter 5 and seems to be affected by flooding or erosion activity. In EPA Rain Zone 3, site GACLCOTR is a new development in Tara Road, Clayton County, and Georgia. There were collected 22 samples at this location. The median TSS was 200 mg/L. No unusual conditions were identified when analyzing the aerial photographs.

Site mean concentrations below the lower control limit in the Xbar chart were located in Virginia, North Carolina and Texas. The two sites located in Virginia are located in Chesterfield County. The first site is located in King Mills Road (VACHCOF3,

10 observations, median TSS = 4 mg/L) and is located in a forested area with less than

20% impervious. The second site (VACHCOF5, 14 observations, median TSS = 15

mg/L) is 50% impervious, but surrounded by a forested area. Only four events were

collected at the site between March and August 1993, in Silo Lane, Charlotte, North

Carolina (NCCHSIMS, median TSS = 10 mg/L), no unusual characteristics were

observed from the aerial photographs. The unusual low concentration site in Houston,

Texas is located on Lazybrook Street (TXHOA003, median TSS = 21 mg/L). Freeways

(I-610) are located in the north and west part of the watershed. Tall trees surrounding the

houses were also observed inside the watershed.

The final step was using ANOVA to evaluate if any EPA Rain Zone was different

than the others. The ANOVA table indicated a p-value close to zero, indicating that there

are significant differences in the TSS concentration among at least two of the different

EPA Rain Zones. The Dunnett's comparison test with a family error of 5% indicate that

concentrations in EPA Rain Zones 4 (median TSS= 91 mg/L), 5 (median TSS = 83

mg/L), 6 (median TSS=118 mg/L), 7 (median TSS = 69 mg/L), and 9 (median TSS = 166

mg/L) are significantly higher than the concentrations observed in EPA Rain Zone 2

(median TSS = 49 mg/L).

This same procedure was performed for the following 13 additional constituents

in residential, commercial and industrial land use areas: hardness, TSS, TDS, oil and

grease, BOD, COD, NO2 + NO3, ammonia, TKN, dissolved phosphorus, total

phosphorus, copper, lead and zinc.

3.4.2 Identification of Unusual Sites

The Xbar charts were created for residential, commercial and industrial land uses. In residential areas, 54 sites were identified with at least one constituent out of control. These sites failed when compared with sites in the same EPA Rain Zone. Table 16 shows the sites with more than 4 constituents outside the control limits.

These eight sites were located in EPA Rain Zone 2. Three sites show elevated concentrations, one in all constituents, and another in metals and the third in nutrients. The site located near a golf course in Virginia Beach (VAVBTYV1) shows elevated concentrations in TSS, phosphorus and COD. The site located in Prince George close to an industrial park (MDPGCOS2), indicated elevated concentrations of total phosphorus, lead and zinc.

The site with the highest number of constituents outside the control limits (10 out of 14 constituents evaluated) was located in Mt. Vernon, Lexington, Kentucky (KYLXTBL1). This site was monitored between 1992 an 1997; it is located close to two high schools and the University of Kentucky. It is interesting that one of the sites having elevated concentrations is located next to one of the sites with a large number of constituents below the lower control limit (VAVBTYV2 is located close to VAVBTYV1). VAVBTYV1 has low concentrations for 6 out of 14 constituents. This indicates that not only can geographical differences be expected; there are also differences between locations in the same EPA Rain Zone. Lead was most frequently found with high concentrations within the same EPA Rain Zone. Eight sites had elevated lead concentrations, while 11 sites had lower concentrations in the same group. The least frequent out-of-bound constituent was oil and grease: none of the sites indicated elevated

concentrations of oil and grease when compared with other locations in the same EPA

Rain Zone.

Table 16. Sites Failing Xbar Chart in Residential Land Uses

SITE	HA	TSS	TDS	OG	BOD	COD	NO2	NH3	TKN	DP	TP	Cu	Pb	Zn
1MABOA006											L	H		
2KYLOTSR3			H											
2KYLXTBL1	H		H		H	H	H		L	H	H	H	H	H
2MDAACORK													L	
2MDBACOSC							H						L	
2MDBCTYHR									H			H		
2MDCLCOCE														H
2MDHACOBP					L				L		L	L	L	L
2MDHOCOGM					H									
2MDPGCOS2		H									H	L	H	H
2MDSHDTPS		H										L		
2NCCHHIDD							H	H						
2NCCHNANC					L									
2NCCHSIMS		L						H						
2NCFVCLEA			L								L	L		
2NCFVTRYO												L	L	
2NCGRWILL								L		H				
2VAARLCV2					L	L					L		L	
2VAARLLP1			H										L	L
2VACHCN2A												L		L
2VACHCOF3	L	L	L	L		L				L	L	L		
2VACHCOF5		L	L			L					L			
2VACPTSF2												L		L
2VAFFCOF1			L											
2VAHATYH3									L	L				
2VAHATYH5										H				L
2VANFTYN2										H	H			
2VANFTYN3						H							H	
2VANFTYN5													H	
2VAPMTYP2										L			H	
2VAPMTYP4								L		L				
2VAPMTYP5														L
2VAVBTYV1		H				H				H	H			
2VAVBTYV2			H		L		L	L		L		L		L
3GAATAT02												H		
3GACOC1A3					L									
3GACLCOTR		H												
4KATOATWO				L										
4KATOBROO				L										
4TXHOA003		L	H									L		
5TXARA002						L							L	
5TXARA003											H			
5TXDAA005													H	
5TXIRA001													H	
5TXMEA002	H													
5TXMEA003													L	
6AZMCA006												H		
6AZTUA001			H			H								
6AZTUA002													L	
7ORCCA004													L	
7OREUA003	H												H	
7ORGRA003				L										
7ORPOA006	L													
7ORSAA004										L			L	L

Note: H: Site with mean concentrations larger than UCL. L: Site with mean
concentrations lower than LCL

In commercial land use areas, six out of 25 locations indicated more than three

constituents outside of the control limits (Table 20). Five sites have more than one

constituent above the upper detection limit. The site with the largest number was located

in Wilhite Drive behind a K-Mart in Lexington, Kentucky (KYLXWHL1). This site was

monitored between 1992 and 1996. The site indicates elevated nutrients, BOD, hardness

and TDS concentrations. The second site was also located in Kentucky. East Land is

located in an old commercial area in Lexington (KYLXNEL3). This site has elevated

total and dissolved phosphorus concentrations.

Table 17. Sites Failing Xbar Chart in Commercial Land Uses

SITE	HA	TSS	TDS	OG	BOD	COD	NO2	NH3	TKN	DP	TP	Cu	Pb	Zn
2KYLXNEL3								L		H	H			
2KYLXWHL1	H		H		H				H	H	H			
2MDAACOPP							L						L	
2MDHOCODC					L						L			
2MDHOCODC														
2MDPGCOS1		H										H	H	H
2NCGRATHE											H			
2NCGRMERR														L
2VAARLRS3					L	L							L	L
2VACHCCC4	H	L		L	L	L								
2VAHATYH1										L				
2VAHCCOC2														H
2VAPMTYP1										L				
3ALHUMASM													H	
3ALHUWERP													H	
3ALMODAPH					H									
4KATOJACK		H		H										H
4TXHOA004				L										
6AZTUA003													H	
7OREUA001	H													
7ORPOA001	L													
9CODEA001												H		H
9CODEA002				H										
9KAWITOWN					L									

Note: H: Site with mean concentrations larger than UCL. L: Site with mean
concentrations lower than LCL

A third site having elevated stormwater concentrations was found in Brightseat Road adjacent to Landover Mall in Prince George County, Maryland (MDPGCOS1). This site was monitored between 1992 and 1996. It has elevated TSS, copper, lead and zinc concentrations. A fourth site with elevated stormwater concentrations is located in Topeka, Kansas (KATOJACK). This site is located close to a sand quarry. Median TSS concentrations at this location were close to 600 mg/L. Elevated oil and grease, total lead and total zinc were also found at this location. The last elevated concentration site is located in Denver, Colorado. Cherry Creek at Colfax Avenue (CODEA001) has elevated copper and zinc concentrations. The site is 87% commercial and contains a convention center, hotels and restaurants on 16th Street Mall, the State Capital and other government buildings.

Four out of 25 industrial land use locations indicated more than three constituents with median concentrations outside the upper control limit (Table 18). One site is located in Boston, Massachusetts. The Brighton (MABOA004) watershed drains runoff from warehouses and manufacturing operations associated with mechanical, roofing and electrical activities. According to the site description, there is a large potential for storage of rainfall on rooftops and poorly maintained parking lots and roadways. Extremely high ammonia and TKN concentrations were observed at this location. Another industrial site having high concentrations is located in Greensboro, North Carolina. The site is located at Husband Street (NCGRHUST). Zinc and especially copper concentrations were elevated (median copper = 29 µg/L).

A site located at Santa Fe Shops in Topeka, Kansas (KATOSTFE) had elevated metal concentrations. Railroad activity was present in the watershed. Another industrial

site of interest is located on 27th Avenue at the Salt River in Maricopa, Arizona

(AZMCA003). It had a median TSS of 668 mg/L. Copper, lead and zinc had extremely

high concentrations at this location compared with many other single land uses sites in

the database.

Table 18. Sites Failing Xbar Chart in Industrial Land Uses

SITE	HA	TSS	TDS	OG	BOD	COD	NO2	NH3	TKN	DP	TP	Cu	Pb	Zn
1MABOA004								H	H			H		
1MNMISD03												L		
2KYLOTSR2					H									
2KYLXTBL2								L		H	H			L
2MDBACOTC													H	
2MDCHCOIP												L		
2MDPGCOS6		H											H	
2NCCHBREV							H	H						
2NCCHHOSK								H						H
2NCFVWINS														H
2NCGRHUST					H							H		H
2VAARLTC4			H									L	L	L
2VACHCOF1												L		
2VACPTYC5		L	L		L	L	L	L	L	L	L	L		
2VAFFOF10									H				L	
2VAFFOF11					H				H					
2VAHATYH2												H		
2VAVBTYV4		L	L									L		
3ALHUCHIP													H	
3GAATAT01		L										L		
3GACLCOSI										H				
4KATOSTFE												H	H	H
4TNMET211							H							
4TXHCA004							L							
5TXDAA001										H		L		
5TXDAA002		L			L									L
5TXFWA004														H
6AZMCA001							L						L	
6AZMCA003		H					H					H	H	
6AZTUA004												L		
6CAALAL09			L						L					
7ORSAA003					L						L			L
9CODEA007												H		

Note: H: Site with mean concentrations larger than UCL. L: Site with mean
concentrations lower than LCL

3.5 Discussion

The level of censoring observations in a dataset affects the mean, median, standard deviation and coefficient of variation. As the level of non-detected observations increase, the mean, median and standard deviation are larger than if the censored observations are detected. The opposite behavior is expected for the coefficient of variation.

Different laboratories report different detection limits for the same constituents. In many cases, the detection limits are calculated by each laboratory based on their measured repeatability (precision) for a specific laboratory test. Using methods with low precision increases the percentage of non-detected values and the uncertainty of the real mean and standard deviation values.

Open space has the largest number of non-detected observations among land uses. The largest percentages of detected observations were observed in freeways and industrial land uses.

Estimating or replacing by half of the detection limit for levels of censoring smaller than 5% does not have a significant effect on the mean, standard deviation and coefficient of variation values.

Substituting the censored observations by half of the detection limit produces smaller values than when using Cohen's maximum likelihood method. Replacing the censored observations by half of the detection limit is not recommended for levels of censoring larger than 15%.

The censored observations in the database for the dissertation analyses were replaced using estimated values using Cohen's maximum likelihood method for each site. Because this method uses the detected observations to estimate the non-detected values, it is not very accurate, and therefore not recommended, when the percentage of censored observations is larger than 40%. Table 14 shows those constituents having percentages of non-detected observations smaller than 40% for the three main land uses.

All the methods used in this chapter are approximations to calculate the EMC when censored observations are present. These problems would not exist if appropriate analytical methods were used to analyze the samples. It is very important to select analytical methods capable of detecting the desired range of concentrations in the samples in order to reduce the numbers of censored observations to acceptable levels.

Chapter 4
Stormwater Quality Descriptions Using
the Three Parameter Lognormal Distribution

4.1 Introduction

The National Stormwater Quality Database v. 1.1 (NSQD) contains water quality

characteristics from the monitoring required by the NPDES Phase I stormwater permit

applications and subsequent permits, during the period of 1992 to 2002. This database

contains about 3,765 events from 360 sites in 65 communities from throughout the U.S.

For each site, much additional data, including the percentage of each land use in the

catchment, the total area, the percentage of imperviousness, the geographical location and

the season, has been included in the database. Information about the characteristics of

each event is also included. Total precipitation, precipitation intensity, total runoff and

antecedent dry period are also included, if collected. The database contains data collected

at the outfall. In-stream samples were not included in the database. Water quality

characteristics where divided in four main groups: common parameters, nutrients, metals

and others (pesticides and organic compounds for example). Much time and effort was

spent in reviewing these data for QA/QC problems and correcting the information.

The Nationwide Urban Runoff Program (NURP) evaluated the characteristics of

stormwater discharges at 81 outfalls in 28 communities throughout the U.S. (EPA 1983).

One of the conclusions was that stormwater constituent concentrations could be described

using a lognormal distribution. Recently, Van Buren (1997) found that stormwater

100

concentrations were described using a lognormal distribution for almost all constituents, with the exception of dissolved constituents that were better described with a normal distribution. Beherra (2000) also found that some stormwater constituent concentrations were better described using a lognormal distribution, while others were better described with gamma or exponential distributions. The constituents that were best described with a gamma distribution were: total solids, total Kjeldahl nitrogen (TKN), total phosphorous, chemical oxygen demand (COD), barium and copper. The constituents that were best described with an exponential distribution were: suspended solids, nitrates and aluminum. In both of these studies, fewer than 50 samples (collected at the same site) were available for evaluation.

During the research reported in this chapter, statistical tests were used to evaluate the log-normality of a selection of the constituents in the NSQD database. Statistical descriptions were obtained of each set of data including box and whiskers and probability plots for each land use category and for the pooled dataset. It was found in almost all cases that the log-transformed data followed a straight line between the 5th and 95th percentile, as illustrated in Figure 26 for total dissolved solids (TDS) in residential areas.

For many statistical tests focusing on the central tendency (such as for determining the average concentration that is used for mass balance calculations), this may be a suitable fit. As an example, WinSLAMM, the Source Loading and Management Model (Pitt and Voorhees 1995), uses a Monte Carlo component to describe the likely variability of stormwater source flow pollutant concentrations using either lognormal or normal probability distributions for each constituent. However, if the extreme values are of importance (such as when dealing with the influence of many non-detectable values on

the predicted concentrations, or determining the frequency of observations exceeding a

numerical standard), a better description of the extreme values may be important.
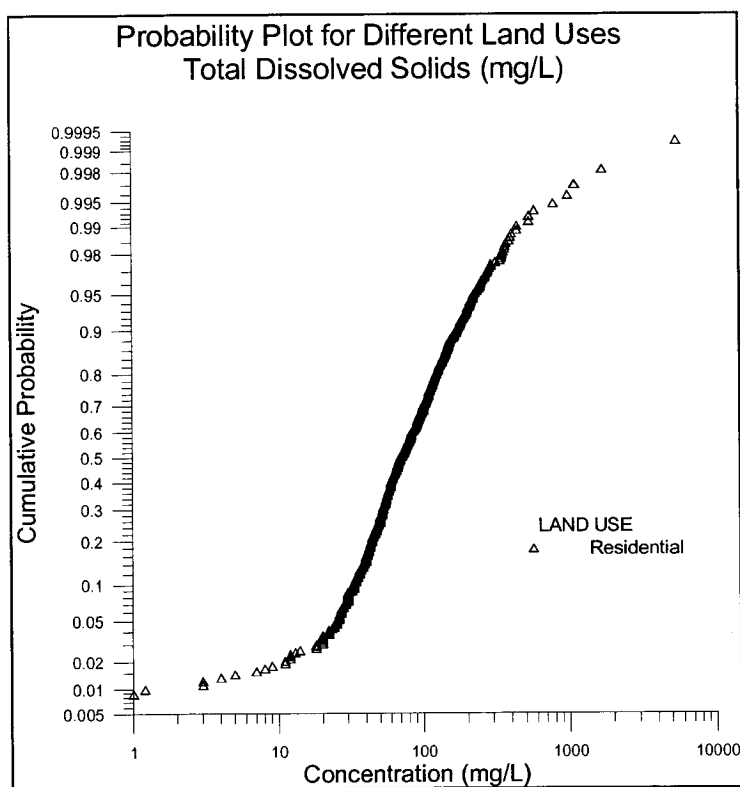
Fig. 26. Probability plot of total dissolved solids in residential land use

The NSQD underwent an extensive data evaluation process, including multiple

comparisons of the all data value in the database to original documents. In some cases,

data was available from the local agency in electronic form. These spreadsheets were

reformatted to be consistent to the NSQD format. However, it was found that all of the

submitted electronic data needed to be verified against original data sheets and reports.

When reviewing the NSQD, it was assumed that some of the events in the upper and

lower tails of the distributions were caused by errors, most likely due to faulty

transcription of the data (such as mislabeling the units for heavy metals or nutrients as

mg/L instead of μg/L, for example). Unusual values were verified with the original reports and datasets. While some values (less than 5% of the complete dataset) were found to be in error and were corrected, most of the suspected values were found to be stormwater observations. Besides the targeted extreme values, all reported values were also examined in relationship to other related constituents (COD vs. BOD; total metal concentrations vs. dissolved metal concentrations; TKN vs. NH3; TDS vs. specific conductivity; SS vs. turbidity; etc) and unusual behavior was further checked and corrected, as necessary. In some cases, unusual values could not be verified and were therefore eliminated from the dataset, although this was very unusual.

After the extensive QA/QC activities and corrections were made to the NSQD, the next step was to conduct a sensitivity analysis to determine the effects of the remaining unusual high and low values on the probability distribution parameters.

4.2 The Effects of Unusual High and Low Values on Probability Distribution Parameters

For this evaluation, 10,000 sets of 200 samples each were randomly generated following a lognormal distribution (1, 1), but having differing amounts of extreme values in each data set. For each set, the mean, variance and coefficient of variation were calculated. Two main factors were analyzed using these data: the extreme value factor and percentage of extreme values in each sample. The following percentages of extreme values were selected for evaluation: 0.5, 1, 5, 10, 25 and 50%. For each percentage of extreme values, the following factors were analyzed: 0.001, 0.01, 0.1, 10, 100, 1.000, 10,000, 100,000 and 1,000,000. For example (5%, 100) indicates that in each set, 5% of the data were increased by a factor of 100. The coefficient of variation was then

calculated for each set of data. The medians of the coefficients of variation for the 10,000

runs are shown in Figure 27 for each level of extreme values.
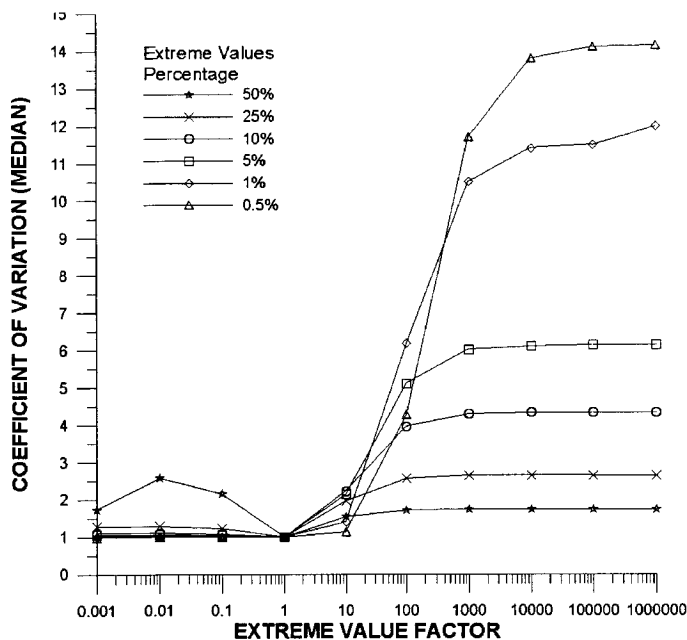
Fig. 27. Effect of unusual values on the coefficient of variation (based on LN(1,1))

For a lognormal distribution (1, 1) the coefficient of variation is equal to one.

Figure 27 shows how this original value is changed for different amounts of extreme

values in the data sets, and for different factors in these extreme values. The horizontal

axis represents the factor used in the extreme values. As an example, many of the

incorrect extreme values observed in the NSQD for heavy metals were because the units

were originally incorrectly reported as mg/L in the submitted information, while the

correct units were actually μg/L. This would be an extreme value factor of 1,000.

Extreme value factors of 10 were also fairly common and were associated with simple

misplacements of decimal points in the data.

Figure 27 also shows that for small error factors (0.1, 0.01 and 0.001) there is not a large effect in the coefficient of variation for percentages smaller than 10%. For larger percentages the effect in the coefficient of variation is important. When 50% of the data are affected by an error factor of 0.01, the coefficient of variation was increased by almost three times.

High extreme value factors can have an important effect on the coefficient of variation. When 10% of the data were increased by a factor of 10, the coefficient of variation was increased almost three times. Notice that affecting 10% of the data by a factor of ten have almost the same effect as affecting 50% of the data by a factor of a hundredth. This effect is reduced when the percentage of elevated values in the dataset is smaller than 10%.

For factors larger than a hundred, the effect on the coefficient of variation is much greater. Very low percentages of elevated values can increase the coefficient of variation by up to 15 times. For example, when only 0.5% of the sample is affected by a factor of a thousand, the coefficient of variation increases almost 12 times more than the correct value. As noted earlier this is important because it is not unusual to find reported values affected by a factor larger than a hundred (See Figure 26). Some of these values can be due to incorrect reporting units, but in many cases they were considered as valid observations because they were supported by similarly high values of other closely related constituents. For factors greater than 10,000 the multiplying value of the coefficient of variation remains stable at the maximum value obtained.

The above analyses indicate that in lognormal distributions, the presence of just a few unusual elevated values is important and can dramatically affect the reported

coefficient of variation for the distribution of concentration. This observation is critical in the relatively common case where one or a very few observations are affected by a factor larger than a hundred. In the other extreme, factors smaller than one do not have a large impact on the reported coefficient of variation, except when the percentage of extreme values is greater than 50%.

The effect of extreme values on the mean and standard deviation was also analyzed. Figure 28 shows the effect of the extreme values on calculated standard deviation. For large extreme value factors (larger than one) the standard deviation increases as the percentage of extreme values increases.
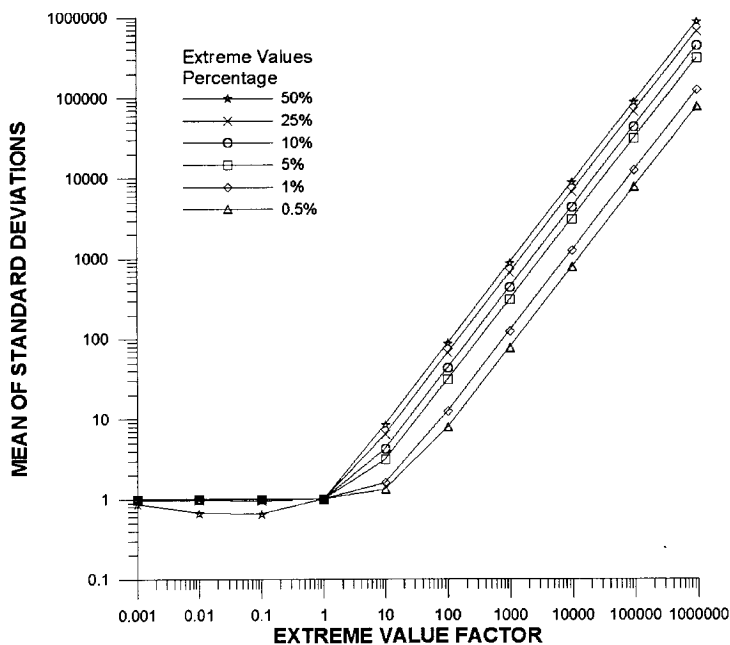
Fig. 28. Effect of unusual values on the standard deviation (based on LN(1,1))

Percentages smaller than 25% do not have an important effect on the standard deviation for small extreme value factors. For a specific extreme value factor, changing

the extreme value percentages from 0.5% to 50% increases the standard deviation close to 10 times.

The effect of the presence of extreme value on the distribution mean is shown in Figure 29. For small extreme value factors, the mean is reduced almost 80% when the extreme value percentage is close to 50%. This is expected because in a lognormal (1, 1) most of the values are located in the lower tail of the distribution. For extreme value occurrences less than 25%, the mean value is reduced by less than 20%.

Large extreme values factor have much larger effects on the distribution means. As the extreme value percentage increases, the calculated means also increase. If 0.5% of the values are affected by a factor of a hundred, the mean value is doubled. If 50% of the values are affected by the same factor, the mean values are increased by almost 50 times. For factors larger than a thousand, increasing the percentages of extreme values from 0.5% to 50% increases the mean values by up to two orders of magnitude.

These evaluations are important because it points out that for a lognormal distribution the effects of few elevated values in the upper tail have much greater effect on common statistics than unusual values in the lower tail. Many stormwater researchers have focused on the lower tail, especially when determining how to handle the detection limits and unreported data. Stormwater constituents usually have unusual values in both tails of the probability distribution. It is a common to delete elevated values from the observations assuming they are expendable "outliers". This practice is not recommended unless is there sufficient evidence that the observed values are a mistake. Actual elevated values can have a large effect on the calculated distribution parameters. If these are arbitrarily removed, the data analyses will likely be flawed.
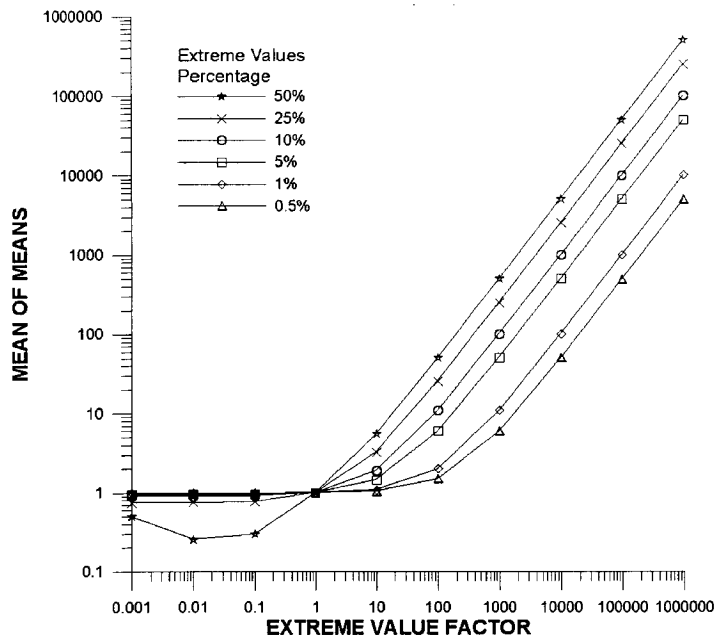
Fig. 29. Effect of unusual values on the mean (based on LN(1,1))

4.3 Analysis of Lognormality of Stormwater Constituents Parameters

The goodness of fitness of twenty nine stormwater constituent probability

distributions was evaluated using the Kolmogorov-Smirnov test. Figure 30 shows how

the test accepts or rejects the null hypothesis that the empirical and the estimated

distributions are the same. If the null hypothesis is valid, then the constituent can be

adequately represented by the lognormal distribution. The observations are sorted and a

probability is assigned by its rank. The distribution generated by this ranking is known as

the empirical distribution. The estimated distribution function is also compared on the

same plot. The estimated distribution function is calculated with the mean and standard

deviation of the original data. If the distance between the empirical and the estimated

distributions is higher than a critical value d_α or D_{max}, the hypothesis of lognormality is

rejected. Notice in Figure 30 that the horizontal axis has a logarithmic scale.
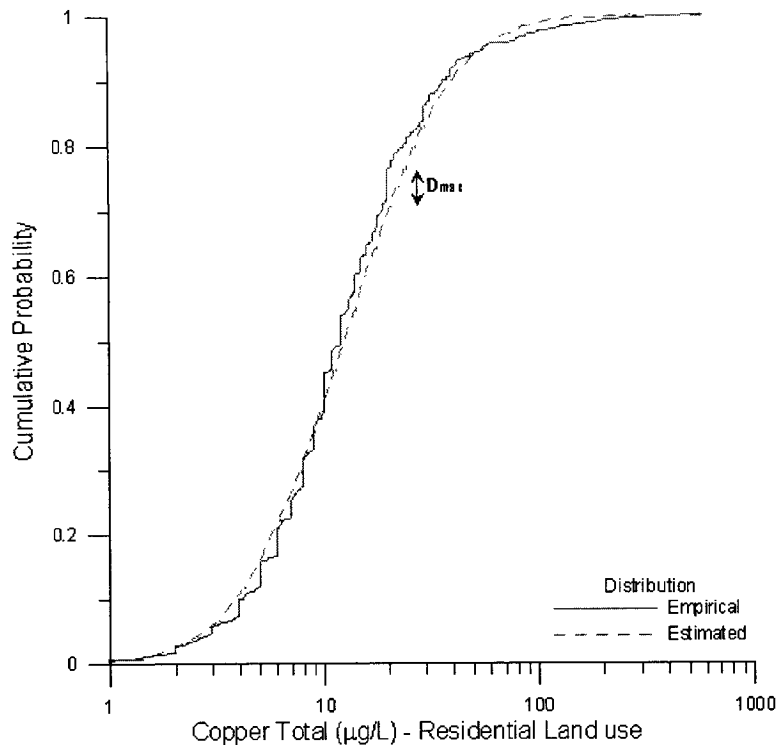
Fig. 30. Cumulative and empirical probability distributions of total copper for residential land use data (Goodness of fit test, Kolmogorov-Smirnov)

There are many options to assign the probability based on the ranks. Most methods assign the probability as a percentage of the total range. The probability of the observation is calculated as its rank divided by the number of observations. Kottegoda (1998) suggested that for extreme events analysis the plotting position can be calculated as:

(4.1) $$p = \frac{i - 0.5}{n}$$

Where p is the cumulative probability of the observation, i is the rank of the observation and n is the total number of observations. We used this plotting position for these analyses because it does not restrict the probability of the largest observation to be one.

In the Kolmogorov-Smirnov test, the null hypothesis is that the observed data follow a lognormal distribution. If the sample size is small, and the distance between the empirical and the observed distributions is smaller than the critical value D_{max}, the test is interpreted as "there is not enough evidence to reject the hypothesis that the distribution is lognormal." In most cases, the NSQD contains enough samples to be able to accept or reject the null hypothesis with acceptable levels of confidence and power.

The NSQD contains many factors for each sampled event that likely affect the observed concentrations. These include such factors as seasons, geographical zones, rain intensities, etc. These factors may affect the shape of the probability distribution. As more data become available, the critical value D_{max} is reduced in the test. There will always be a specific number of samples that will lead to rejection of the null hypothesis because the maximum distance between the empirical and estimated probability distributions became larger than the critical value D_{max}. The only way to evaluate the required number of samples in each category is using the power of the test. Power is the probability that the test statistic will lead to a rejection of the null hypothesis when it is false (Gibbons and Chakraborti 2003).

Masey (1950) states that the power of the Kolmogorov-Smirnov test can be written as:

$$(4.2) \qquad power = 1 - \Pr\left(\frac{-d_\alpha \pm \Delta\sqrt{n}}{\sqrt{F_1(x_0)(1-F_1(x_0))}} < \frac{\{S_n(x_0)-F_1(x_0)\}\sqrt{n}}{\sqrt{F_1(x_0)(1-F_1(x_0))}} < \frac{d_\alpha \pm \Delta\sqrt{n}}{\sqrt{F_1(x_0)(1-F_1(x_0))}} \right)$$

where:

d_α = Dmax: critical distance at the level of significance

α (confidence of the test),

S_n = Cumulative empirical probability distribution,

F_1 = Cumulative alternative probability distribution,

Δ = Maximum absolute difference between the

cumulative estimated probability distribution and

the alternative cumulative probability distribution.

Massey (1951) also found that for large sample sizes, the power can be never be smaller than

(4.3)
$$power > 1 - \int_{2(\Delta\sqrt{n}-d_\alpha\sqrt{n})}^{2(\Delta\sqrt{n}+d_\alpha\sqrt{n})} \frac{1}{\sqrt{2\pi}} e^{-\frac{t^2}{2}} dt$$

This reduced expression can be used to calculate the number of samples required to reject the null hypothesis with a desired power. Figure 31 shows the power of the D test for 1%, 5%, and 10 % levels of confidence of the test for samples size larger than 35 (Massey 1951). For example, assume that the maximum distance between the alternative cumulative and the estimated cumulative probability distributions is 0.2, and we want an 80% power against the alternative at a 5% level of confidence. To calculate the number of required samples, we read that $D(N)^{0.5}$ is 1.8 for a power of 0.8 and 5% level of confidence. Solving for N = $(1.8/0.2)^2$ = 81 samples. If we want to calculate the number of samples when the difference between the alternative cumulative and the estimated

cumulative probability function is 0.05, with the same power and level of confidence, then 1,296 samples would be required. When the lines are very close together, it is obviously very difficult to statistically show that they are different.

Lower Bounds for the Power of the D test
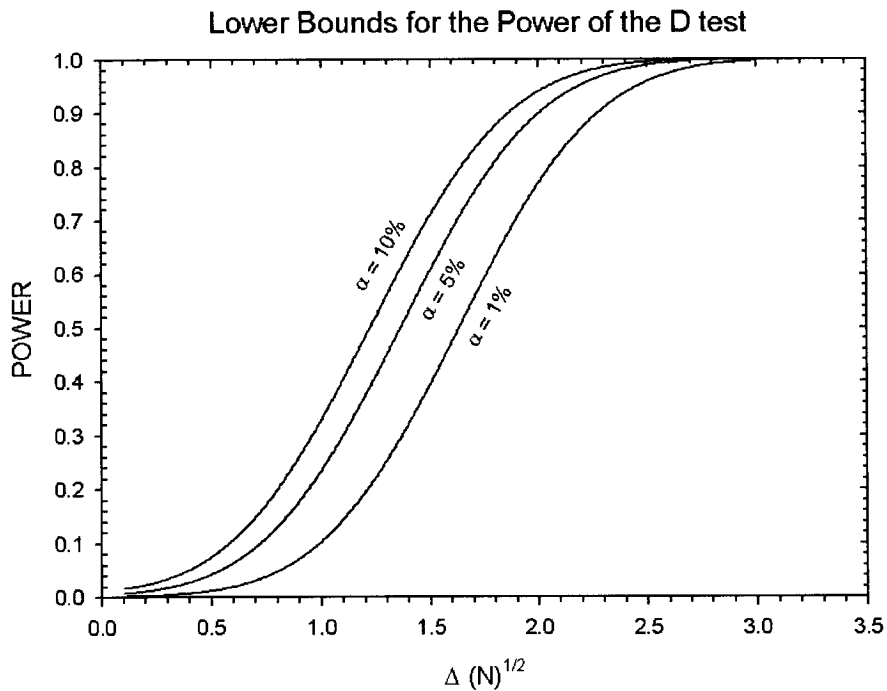
Fig. 31. Lower bounds for the power of the D test for $\alpha = 1\%$, 5% and 10% (N>35)

In the NSQD, most of the data were from residential land uses. The Kolmogorov-Smirnov test was used to indicate if the cumulative empirical probability distribution of the residential stormwater constituents can be adequately represented with a lognormal distribution. Table 19 shows the resulting power of the test for D=0.05 and D=0.1, when applied to selected constituents that had high levels of detection in residential land uses.

Table 19. Power of the Test When Applied to Selected Constituents in Residential Land Uses

CONSTITUENT	N	Percentage Detected	$\Delta N^{0.5}$ (Δ=0.05)	Power (Δ=0.05,α = 5%)	$\Delta N^{0.5}$ (Δ=0.1)	Power (Δ=0.1,α = 5%)
TDS (mg/L)	861	99.2	1.46	0.60	2.92	1
TSS (mg/L)	991	98.6	1.56	0.65	3.12	1
BOD (mg/L)	941	97.6	1.52	0.65	3.04	1
COD (mg/L)	796	98.9	1.40	0.55	2.80	1
NO2+NO3 (mg/L)	927	97.4	1.50	0.60	3.00	1
TKN (mg/L)	957	96.8	1.52	0.65	3.04	1
TP (mg/L)	963	96.9	1.53	0.65	3.06	1
Total Copper (μg/L)	799	83.6	1.29	0.50	2.58	1
Total Lead (μg/L)	788	71.3	1.19	0.40	2.38	1
Total Zinc (μg/L)	810	96.4	1.40	0.55	2.80	1

Table 19 shows that the number of collected samples is sufficient to detect if the empirical distribution is located inside an interval of width 0.1 above and below the estimated cumulative probability distribution. If the interval is reduced to 0.05, the power varies between 40 and 65%. To estimate the interval width, 10 cumulative distributions of 1,000 random data points, having a lognormal (1, 1) distribution, were compared with the estimated cumulative distribution for normal, gamma and exponential distributions. The maximum distance between the cumulative lognormal and the cumulative normal distributions was 0.25. The maximum distance with cumulative gamma (the same for exponential in this case) was 0.28. An interval width of 0.1 was considered appropriate for the analysis.

Another factor that must be considered is the importance of relatively small errors in the selected distribution and the problems of a false negative determination. It may not be practical to collect as many data observations as needed when the distributions are close (such as when the width interval is 0.05). Therefore, it is important to understand what types of further statistical and analysis problems may be caused by having fewer samples than optimal. For example, Figure 32 (total phosphorus in residential area)

shows that most of the data fall along the straight line (indicating a lognormal fit), with

fewer than 10 observations (out of 933) in the tails being outside of the obvious path of

the line.
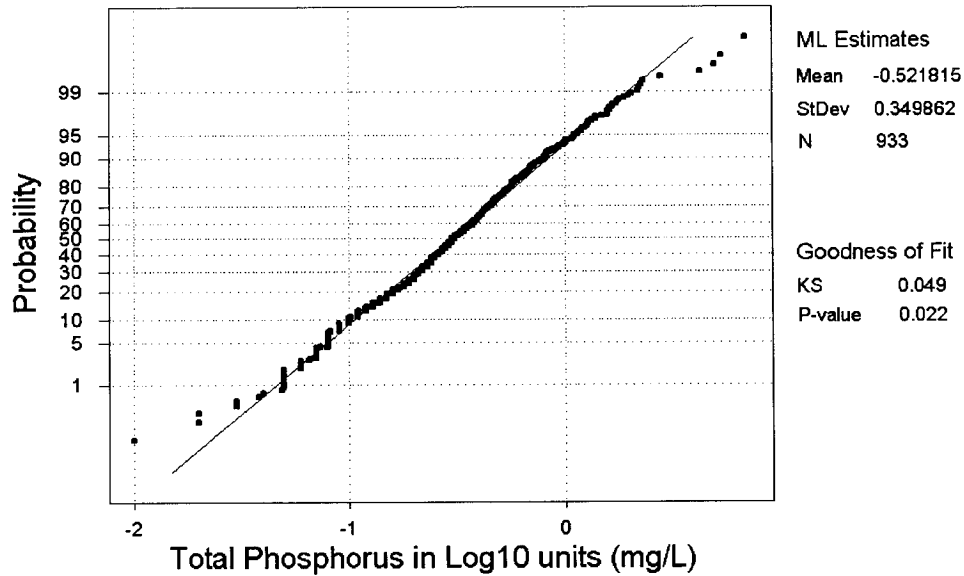
Fig. 32. Normality test for total phosphorus in residential land uses using the NSQD

The calculated p-value for the Kolmogorov-Smirnov test is 0.022, indicating that

the null hypothesis could be rejected and that there is not enough evidence that the

empirical distribution is adequately represented by a lognormal distribution. Notice that

errors in the tails are smaller than 0.049. However, the tails are not responsible for the

rejection of the null hypothesis (see Figure 33).
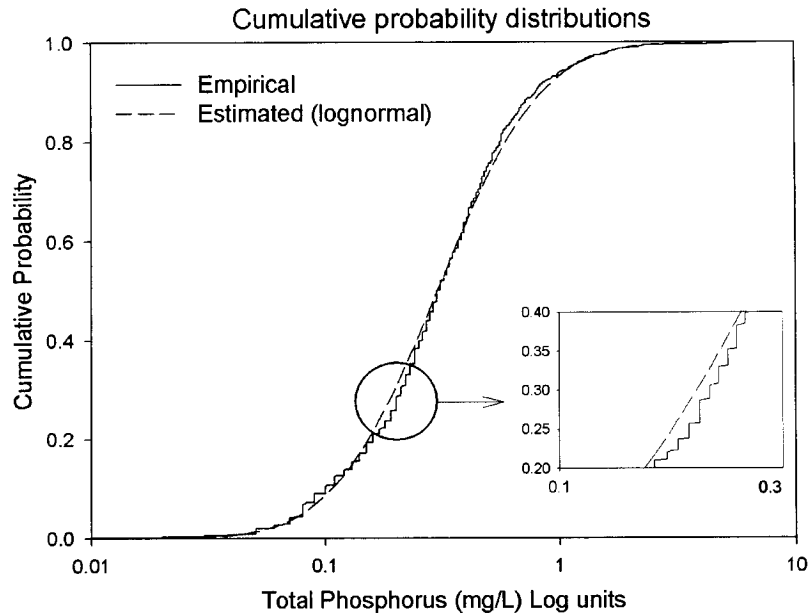
Fig. 33. D_{max} was located in the middle of the distribution

In this case, D_{max} is located close to a total phosphorus concentration of 0.2 mg/L (-0.7 in log scale). As in this case, the hypothesized distributions are usually rejected because of the departures in the middle of the distribution, not in the tails. However, as previously pointed out, a small number of observations in the upper tail can change the shape of the estimated cumulative probability distribution by affecting the mean and standard deviation of the data. The methods used previously by Van Buren and Beherra evaluated the probability distributions only using two parameters, the median and the standard deviation. They suggested the gamma and exponential distributions as alternatives to the lognormal for some stormwater constituents. Table 20 shows the comparison for the goodness of fit using the 2-parameter gamma, exponential and lognormal distributions using the method of moments.

Table 20. Comparison of Goodness of Fit for Gamma, Exponential and Lognormal Distributions Using the NSQD v.1.1

CONSTITUENT	PDF	RESIDENTIAL			COMMERCIAL			INDUSTRIAL			OPEN SPACE			FREEWAYS		
		N_Det	Dmax	P-value	N_Det	Dmax	P-value	N_Det	Dmax	P-value	N_Det	Dmax	P-value	N_Det	Dmax	P-value
Conductivity (mS/cm)	Gamma	106 / 100%	0.381	0	66 / 100	0.230	0.002	108 / 100	0.348	0	2 / 100	-	-	86 / 100	0.238	0
	Exponential		0.195	0.001		0.237	0.001		0.228	0		-	-		0.232	0
	Lognormal		0.081	0.493		0.100	0.530		0.074	0.619		-	-		0.129	0.113
Hardness (mg/L)	Gamma	250 / 100%	0.217	0	139 / 100	0.141	0.008	138 / 96.4	0.323	0	8 / 100	0.304	0.458	127 / 100	0.451	0
	Exponential		0.203	0		0.115	0.067		0.133	0.018		0.369	0.228		0.161	0.003
	Lognormal		0.071	0.166		0.090	0.206		0.080	0.369		0.354	0.268		0.077	0.447
Oil and Grease (mg/L)	Gamma	533 / 57.8%	0.876	0	308 / 70.8	0.629	0	327 / 65.1	0.939	0	19 / 36.84	0.210	1.080	60 / 71.7	0.103	0.810
	Exponential		0.514	0		0.304	0		0.697	0		0.265	0.750		0.286	0.002
	Lognormal		0.112	0.001		0.103	0.019		0.098	0.032		0.202	1.127		0.101	0.827
Total Dissolved Solids (mg/L)	Gamma	861 / 99.3%	0.234	0	399 / 99.5	0.457	0	413 / 99.5	0.645	0	45 / 97.8	0.109	0.698	97 / 99	0.082	0.553
	Exponential		0.207	0		0.150	0		0.172	0		0.195	0.070		0.171	0.007
	Lognormal		0.050	0.029		0.049	0.303		0.066	0.053		0.120	0.561		0.054	1.136
Total Suspended Solids (mg/L)	Gamma	991 / 98.6%	0.288	0	458 / 98.3	0.363	0	428 / 99.1	0.206	0	44 / 95.5	0.132	0.464	134 / 99.3	0.534	0
	Exponential		0.141	0		0.214	0		0.108	0		0.289	0.002		0.168	0.011
	Lognormal		0.032	0.280		0.064	0.053		0.029	0.995		0.113	0.683		0.066	0.627
BOD5 (mg/L)	Gamma	941 / 97.6%	0.321	0	432 / 97.5	0.191	0	406 / 95.3	0.921	0	44 / 86.4	0.112	0.770	26 / 84.6	0.272	0.076
	Exponential		0.140	0		0.142	0		0.355	0		0.261	0.011		0.168	0.580
	Lognormal		0.058	0.004		0.054	0.166		0.105	0		0.114	0.746		0.103	1.252
COD (mg/L)	Gamma	796 / 98.9%	0.129	0	373 / 98.4	0.137	0	362 / 98.9	0.216	0	43 / 76.7	0.373	0	67 / 98.5	0.163	0.061
	Exponential		0.161	0		0.136	0		0.119	0		0.168	0.312		0.139	0.157
	Lognormal		0.036	0.250		0.038	0.695		0.074	0.040		0.128	0.684		0.107	0.445
Fecal Coliform (Colonies/100 mL)	Gamma	446 / 88.3%	0.655	0	233 / 88	0.333	0	297 / 87.9	-	-	23 / 91.3	0.179	0.520	49 / 100	0.239	0.007
	Exponential		0.374	0		0.396	0		0.504	0		0.208	0.324		0.355	0
	Lognormal		0.080	0.013		0.076	0.192		0.051	0.510		0.181	0.503		0.105	0.677
Fecal Streptococcus (Colonies/100 mL)	Gamma	305 / 89.5%	0.158	0	181 / 91.7	0.354	0	195 / 93.8	-	-	22 / 90.9	0.144	0.869	25 / 100	0.096	1.262
	Exponential		0.202	0		0.278	0		0.399	0		0.142	0.892		0.164	0.518
	Lognormal		0.077	0.081		0.097	0.091		0.083	0.161		0.181	0.538		0.119	0.990
Ammonia (mg/L)	Gamma	595 / 81.5%	0.132	0	299 / 83.3	0.131	0	254 / 85.8	0.154	0	32 / 18.7	-	-	79 / 87.3	0.216	0.003
	Exponential		0.101	0		0.066	0.228		0.071	0.221		-	-		0.105	0.440
	Lognormal		0.044	0.305		0.050	0.589		0.047	0.758		-	-		0.133	0.173
NO2+NO3 (mg/L)	Gamma	927 / 97.4%	0.197	0	425 / 98.1	0.147	0	418 / 96.2	0.080	0.011	44 / 84.1	0.123	0.654	25 / 96	0.274	0.055
	Exponential		0.141	0		0.120	0		0.132	0		0.120	0.686		0.177	0.443
	Lognormal		0.070	0		0.040	0.531		0.080	0.011		0.141	0.463		0.139	0.789
TKN (mg/L)	Gamma	957 / 96.8%	0.203	0	449 / 97.3	0.127	0	440 / 95.9	0.195	0	45 / 71.1	0.169	0.323	125 / 96.8	0.280	0
	Exponential		0.182	0		0.156	0		0.134	0		0.141	0.556		0.138	0.020
	Lognormal		0.035	0.218		0.042	0.423		0.048	0.292		0.147	0.500		0.074	0.539

* P-values greater than one are used only for comparison. NDet: Number of collected samples and percentage detected

Table 20. Comparison of Goodness of Fit for Gamma, Exponential and Lognormal Distributions Using the NSQD – Continued

CONSTITUENT	PDF	RESIDENTIAL NDet	Dmax	P-value	COMMERCIAL NDet	Dmax	P-value	INDUSTRIAL NDet	Dmax	P-value	OPEN SPACE NDet	Dmax	P-value	FREEWAYS NDet	Dmax	P-value
Dissolved Phosphorus (mg/L)	Gamma	738 / 84.1	0.117	0	323 / 81.1	0.177	0	325 / 87.1	0.200	0	44 / 79.5	0.154	0.127	22 / 95.5	0.449	0
	Exponential		0.144	0		0.129	0		0.135	0		0.384	0.657		0.350	0.012
	Lognormal		0.043	0.199		0.075	0.104		0.124	0.682		0.124	0.682		0.170	0.593
Total Phosphorus (mg/L)	Gamma	963 / 96.9	0.184	0	446 / 95.7	0.179	0	434 / 96.3	0.227	0	46 / 84.8	0.666	0	128 / 99.2	0.456	0
	Exponential		0.129	0		0.114	0		0.107	0		0.320	0.001		0.187	0
	Lognormal		0.049	0.022		0.038	0.582		0.049	0.273		0.116	0.696		0.085	0.325
Total Antimony (µg/L)	Gamma	288 / 2.8	0.268	0.636	142 / 2.1	-	-	164 / 14.6	0.282	0.045	17 / 0	-	-	14 / 50	0.423	0.164
	Exponential		0.417	0.213		-	-		0.173	0.473		-	-		0.465	0.096
	Lognormal		0.233	0.841		-	-		0.096	1.279		-	-		0.419	0.171
Total Arsenic (µg/L)	Gamma	426 / 42	0.531	0	213 / 32.9	0.643	0	267 / 54.3	0.291	0	19 / 31.6	0.271	0.828	61 / 55.7	0.125	0.694
	Exponential		0.224	0		0.249	0		0.141	0.006		0.462	0.154		0.266	0.016
	Lognormal		0.154	0		0.164	0.046		0.129	0.016		0.273	0.819		0.149	0.441
Total Beryllium (µg/L)	Gamma	301 / 7.3	0.464	0	163 / 4.29	0.305	0.542	209 / 10.5	0.390	0.002	19 / 0	-	-	12 / 16.7	-	-
	Exponential		0.471	0		0.530	0.039		0.539	0		-	-		-	-
	Lognormal		0.200	0.342		0.205	1.108		0.163	0.620		-	-		-	-
Total Cadmium (µg/L)	Gamma	723 / 30.3	0.643	0	358 / 43	0.511	0	395 / 49.4	0.445	0	38 / 55.3	0.295	0.051	95 / 71.6	0.110	0.388
	Exponential		0.358	0		0.311	0		0.237	0		0.560	0		0.153	0.083
	Lognormal		0.120	0.004		0.113	0.039		0.083	0.136		0.206	0.338		0.052	1.380
Total Chromium (µg/L)	Gamma	435 / 55.4	0.292	0	235 / 58.7	0.151	0.004	256 / 72.7	0.122	0.008	36 / 36.1	0.252	0.386	76 / 98.7	0.058	1.208
	Exponential		0.132	0		0.201	0		0.067	0.381		0.272	0.290		0.176	0.019
	Lognormal		0.069	0.206		0.086	0.262		0.062	0.480		0.180	0.861		0.084	0.685
Total Copper (µg/L)	Gamma	799 / 83.6	0.394	0	387 / 92.8	0.296	0	416 / 89.9	0.408	0	39 / 74.4	0.107	0.226	97 / 99	0.451	0
	Exponential		0.149	0		0.137	0		0.177	0		0.127	0.092		0.231	0.090
	Lognormal		0.067	0.005		0.070	0.060		0.080	0.017		0.131	0.742		0.038	1.507
Total Lead (µg/L)	Gamma	788 / 71.3	0.300	0	377 / 85.4	0.297	0	412 / 76.5	0.276	0	45 / 42.2	0.177	0.608	107 / 100	0.203	0
	Exponential		0.173	0		0.136	0		0.225	0		0.389	0.006		0.125	0.072
	Lognormal		0.044	0.218		0.057	0.250		0.059	0.223		0.132	1.034		0.039	1.451
Total Nickel (µg/L)	Gamma	419 / 45.3	0.292	0	232 / 59.5	0.260	0	250 / 62.4	0.090	0.159	38 / 18.4	0.164	1.373	99 / 89.9	0.188	0.004
	Exponential		0.203	0		0.176	0		0.111	0.044		0.261	0.772		0.227	0
	Lognormal		0.081	0.160		0.056	0.831		0.065	0.525		0.166	1.360		0.091	0.460
Total Selenium (µg/L)	Gamma	318 / 6.9	0.263	0.095	169 / 7.7	0.169	0.952	203 / 5.9	0.434	0.022	19 / 21.1	-	-	16 / 6.3	-	-
	Exponential		0.254	0.117		0.174	0.907		0.256	0.416		-	-		-	-
	Lognormal		0.253	0.119		0.196	0.735		0.190	0.841		-	-		-	-
Total Silver (µg/L)	Gamma	406 / 12.6	0.421	0	222 / 11.3	0.143	0.718	287 / 17.4	0.263	0.002	19 / 5.3	-	-	21 / 19	-	-
	Exponential		0.333	0		0.159	0.563		0.340	0		-	-		-	-
	Lognormal		0.271	0.001		0.184	0.370		0.146	0.236		-	-		-	-
Total Zinc (µg/L)	Gamma	810 / 96.4	0.244	0	392 / 99	0.234	0	433 / 98.6	0.273	0	45 / 71.1	0.180	0.253	93 / 96.8	0.158	0.023
	Exponential		0.122	0		0.141	0		0.083	0.005		0.167	0.336		0.155	0.027
	Lognormal		0.054	0.020		0.040	0.585		0.044	0.389		0.105	0.981		0.063	0.985

* P-values greater than one are used only for comparison. NDet: Number of collected samples and percentage detected

Table 20 shows that for residential, commercial and industrial land uses, the lognormal distribution better fits the empirical data, except for selenium and silver in commercial land uses. In open space land uses, about 50% of the constituents were adequately fitted by the lognormal distribution, 30% by the gamma distribution and the remaining by the exponential distribution. In freeway areas, lognormal distributions better fit most of the constituents, except that fecal streptococcus, total arsenic and total chromium were better fitted by the gamma distribution and ammonia was better fitted by the exponential distribution. Also note in Table 20 that residential, commercial and industrial land uses had larger sample sizes than the other two land uses. It seems that for small sample sizes, gamma and exponential distributions better represent actual stormwater constituent distributions, but once the number of samples increases, the lognormal distribution is best. The few cases were the gamma distribution was a better fit was for NO_2+NO_3 in industrial land uses, and chromium in freeway areas. The exponential distribution better represents total ammonia in freeway areas (with around 70 detected samples) than the other two distribution types.

Other transformations were also tested, such as the square root, and other power functions, but the results were not improved. It was therefore decided to investigate if a three-parameter lognormal distribution function can be used to improve the overall goodness of fit for stormwater constituent probability distributions. As shown in the following section, this third parameter, in some cases, allows a much better fit of the cumulative empirical and estimated probability distributions.

4.4 Three Parameter Lognormal Calculations

Goodness of fit was evaluated using 3-parameter lognormal probability

distribution. The probability distributions were created for residential, commercial,

industrial, open space and freeways land uses. The distribution parameters were

calculated using the maximum likelihood and the L-moments methods. The maximum

likelihood method requires that it be solved iteratively using three equations (see

Appendix C). The results were compared with the 2-parameter standard model and the

actual data. The model with the smaller maximum distance between the empirical and the

estimated function was selected as the best model. All the calculations were made using

only the detected values. In general, the L-moments method provided a better fit for the

upper tail of the distribution whereas the maximum likelihood method provided a better

fit for the lower tail. Figure 34 shows the three estimated models for TSS in industrial.
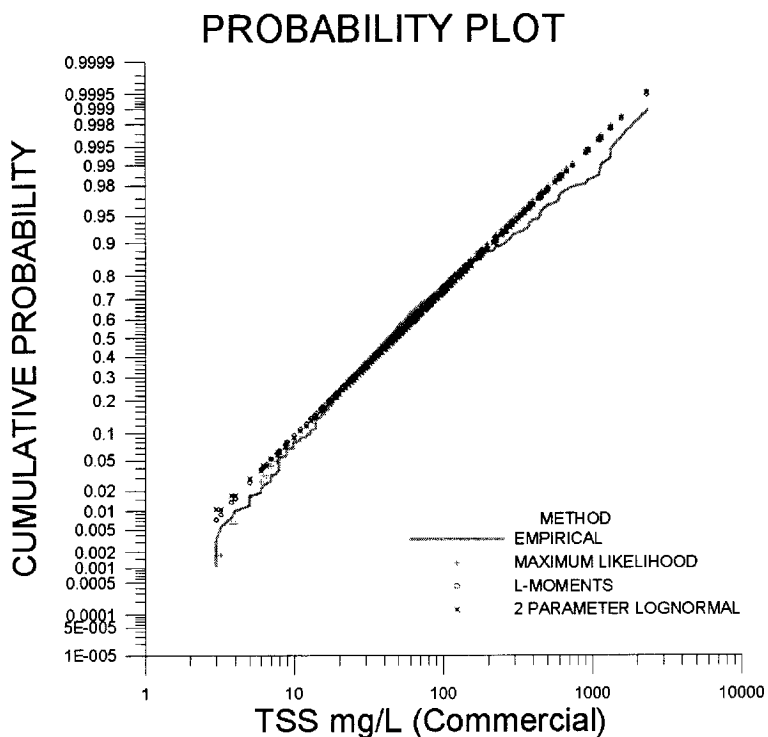
Fig. 34. Estimated models for TSS in commercial land uses

In this graph it is observed that the empirical distribution has higher values in the upper tail compared with any of the three models. In the lower tail, the maximum likelihood method using the 3-parameters better fit the observed values. In this case the maximum likelihood method was better than the other two models, although none of the methods adequately represented the extreme high values. The L-moments method generally betters fits the upper tail distribution, but typically trims or overestimate the lower tail. Figure 35 shows the results for TDS in industrial land uses. The L-moments better fits the empirical distribution in the upper tail, but it trims any observation smaller than 35 mg/L (almost 20% of the total dataset) in the lower tail. The 2-parameter lognormal and the maximum likelihood method provide better results although both were worse than the L-moments in the upper tail region.
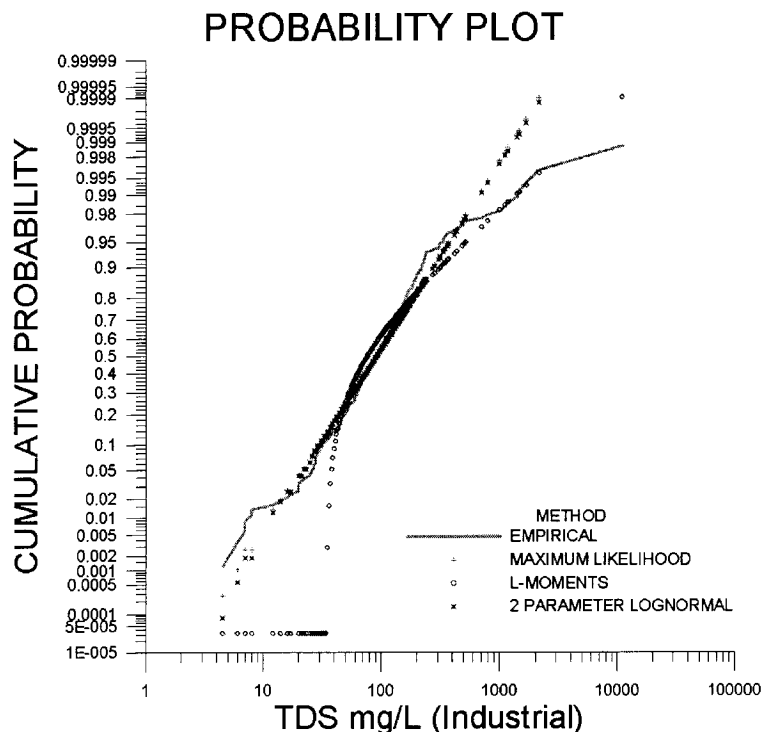
Fig. 35. Estimated models for TDS in industrial land use

Table 21 presents the results for 15 constituents in five land uses. For each of the three methods, the p-value was calculated. The higher the p-value, the better is the fit between the empirical and the estimated function. Some of the p-values in the table are larger than one. When the number of samples is large, the p-value is calculated as a chi square distribution with two degrees of freedom. This probability is calculated only with one tail of the chi square distribution. The p-value is two times this probability. The maximum p-value is one, but for effects of comparison this presents two times the probability calculated from a one tail chi square distribution.

The maximum likelihood method with 3-parameters, or the lognormal 2-parameter distribution produced the best descriptions for most of the constituents. For almost all constituents the function estimated by the L-moments method failed the lognormal assumption. Low p-values were obtained because the function was truncated and does not estimate the lower tail of the distribution.

It seems that when the numbers of samples increase, the L-moments method tends to truncate the function. The maximum likelihood method seems to improve the fit of the distribution, but when the number of samples is large, the cumulative estimated probability distribution is far from the cumulative empirical probability distribution, or no convergence is possible during the iteration process.

In commercial, industrial and freeways land uses, the numbers of samples available were between 100 and 500 samples. According to the prior discussion, this number of samples will result in an analysis having a power close or above 0.5. In these cases, most of the better fits were obtained using the L-moments method. In commercial

and industrial land uses, more than half of the constituents also had the highest p-values when the L-moments method was used.

In open space areas, there were not many samples available. The small number of samples results in a low power. In this case, the higher p-values results were observed when the 2-parameter lognormal distribution was used. The use of the third parameter in constituents having small numbers of sample observations did not improve the fit of the estimated cumulative probability distribution.

4.5 Summary

Most of the stormwater constituents can be assumed to follow a lognormal distribution with little error. The use of the third parameter does not show a significant improvement in estimating the empirical distribution compared with the 2-parameter lognormal. When the number of samples is large per category (approximately more than 400 samples) the maximum likelihood and the 2-parameter lognormal distribution better fit the empirical distribution. For large sample sizes, the L-moments method usually unacceptably truncates the distribution in the lower tail. When the sample size is small (<100 samples), the use of the third parameter does not improve the fit with the empirical distribution and the 2-parameter lognormal distribution produces a better fit than the other two methods.

The lognormal distribution is a skewed distribution when plotted in real space coordinates. When the sample size is small, the calculated skewness is smaller than the skewness of the real distribution. Insufficient sample sizes are not likely to accurately represent the actual distribution of the data.

Table 21. Goodness of Fit for Different Land Uses (p-values Larger than One are Used Only for Comparison)

RESIDENTIAL

CONSTITUENT	N	% Detected	Critical D_0.05	Observed μ	Observed σ	2 Parameter Log D2	2 Parameter Log p-value	3 param max likelihood μ	σ	c	D3m	p-value	3 param L-moments κ	α	ξ	D3l	p-value
Conductivity (µS/cm)	106	100	0.132	4.638	0.710	0.081	0.493	4.327	0.919	20.767	0.052	1.133	-1.240	61.461	88.366	0.087	0.398
Hardness (mg/L)	250	100	0.086	3.497	0.706	0.071	0.166	3.539	0.675	-1.114	0.066	0.231	-	-	-	0.635	0
Oil & Grease (mg/L)	533	57.8	0.077	1.428	1.204	0.112	0.001	1.356	1.267	0.164	0.102	0.003	-2.559	2.065	2.482	0.208	0
Fecal Coliform (Col/100 ml)	446	88.3	0.069	8.7205	2.40448	0.080	0.013	8.734	2.367	-2.992	0.078	0.017	-1.929	17283.541	8236.423	0.096	0.001
Fecal Streptococcus (C/100 ml)	305	89.5	0.082	9.8344	1.88029	0.077	0.081	9.907	1.725	-210.924	0.066	0.190	-1.309	38433.290	24834.760	0.077	0.078
Ammonia (mg/L)	595	81.5	0.062	-1.1672	0.9166	0.044	0.305	0.015	0.220	-0.684	0.139	0.000	-	-	-	-	-
DP (mg/L)	738	84.1	0.055	-1.8303	0.85689	0.043	0.199	0.003	0.056	-0.877	0.327	0.000	-	-	-	-	-
Antimony (µg/L)	288	2.8	0.454	1.3554	1.71904	0.233	0.841	6.569	0.023	-700.000	0.368	0.228	-1.117	9.631	5.281	0.270	0.626
Arsenic (µg/L)	426	42	0.102	1.2098	0.85043	0.154	0.000	1.047	0.971	0.356	0.166	0.000	-1.588	1.981	2.632	0.208	0
Beryllium (µg/L)	301	7.3	0.281	0.0283	1.5958	0.200	0.342	-0.423	2.040	0.136	0.186	0.436	-2.063	1.154	0.675	0.227	0.207
Total Cadmium (µg/L)	723	30.3	0.092	-0.3532	1.21891	0.120	0.004	-0.444	1.301	0.033	0.110	0.010	-1.880	0.626	0.511	0.093	0.045
Total Chromium (µg/L)	435	55.4	0.088	1.5794	0.89199	0.069	0.206	1.473	0.980	0.328	0.067	0.236	-1.157	4.003	4.382	0.066	0.242
Total Nickel (µg/L)	419	45.3	0.099	1.7909	0.75669	0.081	0.160	1.601	0.890	0.768	0.083	0.147	-1.078	4.163	5.384	0.084	0.137
Total Selenium (µg/L)	318	6.9	0.281	1.0969	0.83323	0.253	0.119	0.479	1.348	0.876	0.259	0.061	-1.178	2.103	2.577	0.257	0.11
Total Silver (µg/L)	406	12.6	0.19	1.0686	1.3707	0.271	0.001	0.984	1.469	0.089	0.278	0.001	-1.522	3.390	2.767	0.294	0

COMMERCIAL

CONSTITUENT	N	% Detected	Critical D_0.05	Observed μ	Observed σ	2 Parameter Log D2	2 Parameter Log p-value	3 param max likelihood μ	σ	c	D3m	p-value	3 param L-moments κ	α	ξ	D3l	p-value
Conductivity (µS/cm)	66	100	0.167	4.779	0.721	0.100	0.530	4.736	0.746	3.865	0.097	0.581	-1.011	76.386	108.055	0.093	0.633
Hardness (mg/L)	139	100	0.115	3.689	0.988	0.090	0.206	3.828	0.844	-3.808	0.063	0.653	-0.935	36.911	40.394	0.072	0.474
Oil & Grease (mg/L)	308	70.8	0.092	1.609	1.070	0.103	0.019	1.300	1.358	0.737	0.092	0.062	-1.853	3.648	3.638	0.126	0.002
TDS (mg/L)	399	99.5	0.068	4.332	0.791	0.049	0.303	4.393	0.741	-3.495	0.049	0.289	-1.066	55.106	70.423	0.071	0.035
TSS (mg/L)	458	98.3	0.064	3.883	1.180	0.064	0.053	3.735	1.218	1.988	0.042	0.416	-1.207	55.082	46.245	0.048	0.250
BOD (mg/L)	432	97.5	0.066	2.493	0.868	0.054	0.166	2.302	1.026	1.396	0.040	0.527	-1.002	10.321	11.447	0.044	0.380
COD (mg/L)	373	98.4	0.071	4.167	0.865	0.038	0.695	4.163	0.868	0.194	0.037	0.719	-0.911	54.903	63.127	0.034	0.860
Fecal Coliform (Col/100 ml)	233	88	0.095	8.202	2.380	0.076	0.192	8.191	2.398	1.870	0.077	0.175	-1.768	11370.330	5408.770	0.150	0.000
Fecal Streptococcus (C/100 ml)	181	91.7	0.106	8.940	2.061	0.097	0.091	8.936	2.061	2.494	0.096	0.093	-1.640	16795.600	9532.540	0.056	0.702
Ammonia (mg/L)	299	83.3	0.086	-0.706	1.083	0.050	0.589	-0.697	1.072	-0.002	0.048	0.632	-0.947	0.549	0.522	0.040	0.888
NO2+NO3 (mg/L)	425	98.1	0.067	-0.523	0.882	0.040	0.531	-0.432	0.800	-0.039	0.034	0.837	-0.849	0.510	0.600	0.030	0.954
TKN (mg/L)	449	97.3	0.065	0.471	0.828	0.042	0.423	0.575	0.734	-0.126	0.050	0.228	-0.866	1.246	1.571	0.032	0.816
DP (mg/L)	323	81.1	0.084	-2.077	1.016	0.075	0.104	-2.157	1.092	0.006	0.062	0.273	-1.077	0.127	0.121	0.059	0.315
TP (mg/L)	446	95.7	0.066	-1.473	0.881	0.038	0.582	-1.537	0.935	0.010	0.041	0.466	-0.991	0.196	0.220	0.049	0.264
Arsenic (µg/L)	213	32.9	0.163	0.9336	0.92361	0.164	0.046	0.729	1.098	0.301	0.195	0.010	-1.736	1.565	1.935	0.280	0
Total Cadmium (µg/L)	358	43	0.11	0.047	1.309	0.113	0.039	0.023	1.335	0.011	0.109	0.052	-1.640	1.180	0.879	0.063	0.591
Total Chromium (µg/L)	235	58.7	0.116	1.8134	0.71608	0.086	0.262	1.711	0.787	0.453	0.089	0.225	-0.734	4.462	6.100	0.083	0.295
Total Copper (µg/L)	387	92.8	0.072	2.8829	0.90117	0.070	0.060	2.807	0.965	0.874	0.063	0.117	-1.251	14.103	15.641	0.069	0.067
Total Lead (µg/L)	377	85.4	0.076	3.0328	1.03226	0.057	0.250	3.015	1.049	0.220	0.058	0.225	-1.251	19.633	18.805	0.053	0.329
Total Nickel (µg/L)	232	59.5	0.116	1.9782	0.8075	0.056	0.831	1.668	1.070	1.282	0.089	0.220	-0.969	5.691	6.842	0.076	0.406
Total Selenium (µg/L)	169	7.7	0.361	1.366	0.953	0.196	0.735	0.829	1.492	0.863	0.167	0.670	-0.803	3.940	3.621	0.210	0.638
Total Silver (µg/L)	222	11.3	0.272	0.9637	1.35108	0.184	0.370	1.080	1.174	-0.141	0.182	0.379	-0.911	3.587	3.133	0.165	0.513
Total Zinc (µg/L)	392	99	0.069	5.0388	0.84183	0.040	0.585	5.082	0.803	-4.834	0.039	0.619	-1.021	120.091	144.868	0.052	0.243

Table 21. Goodness of Fit for Different Land Uses (p-values Larger than One are Used Only for Comparison) – Continued

INDUSTRIAL CONSTITUENT	N	% Detected	Observed		D_0.05	2 Parameter Log		3 parameter maximum likelihood					3 parameter L-moments				
			μ	σ		D2	p-value	μ	σ	c	D3m	p-value	κ	α	ξ	D3l	p-value
Conductivity (µS/cm)	108	100	5.011	0.673	0.131	0.074	0.619	4.743	0.848	27.365	0.067	0.767	-1.197	83.673	129.715	0.106	0.174
Hardness (mg/L)	138	96.4	3.794	0.842	0.118	0.080	0.369	3.701	0.914	2.758	0.081	0.346	-1.272	31.221	38.285	0.119	0.047
Oil & Grease (mg/L)	327	65.1	1.623	1.153	0.093	0.098	0.032	1.456	1.298	0.447	0.093	0.048	-3.227	1.068	2.750	0.313	0.000
TDS (mg/L)	413	99.5	4.516	0.870	0.067	0.066	0.053	4.539	0.849	-1.484	0.065	0.064	-1.496	62.528	76.123	0.111	0.000
TSS (mg/L)	428	99.1	4.287	1.200	0.066	0.029	0.995	4.292	1.193	-0.169	0.028	1.023	-1.133	88.174	74.697	0.026	1.119
BOD (mg/L)	406	95.3	2.4121	0.992	0.069	0.105	0.000	2.303	1.085	0.729	0.095	0.002	-2.246	5.718	7.369	0.184	0.000
COD (mg/L)	362	98.9	4.2217	0.91441	0.072	0.074	0.040	4.237	0.899	-0.714	0.076	0.032	-1.096	57.774	63.159	0.046	0.437
Fecal Coliform (Col/100 ml)	297	87.9	7.6064	2.65965	0.084	0.051	0.510	7.574	2.732	1.560	0.055	0.417	-2.356	5638.993	2369.488	0.045	0.688
Fecal Streptococcus (C/100 ml)	195	93.8	9.1491	1.82906	0.101	0.083	0.161	9.190	1.741	-64.211	0.073	0.280	-2.253	11045.202	7378.455	0.199	0.000
Ammonia (mg/L)	254	85.8	-0.7071	1.00903	0.092	0.047	0.758	-0.685	0.985	-0.007	0.049	0.715	-0.864	0.518	0.524	0.046	0.789
NO2+NO3 (mg/L)	418	96.2	-0.3857	0.93116	0.068	0.080	0.011	-0.142	0.703	-0.132	0.043	0.454	-0.689	0.608	0.739	0.045	0.406
TKN (mg/L)	440	95.9	0.4238	0.88424	0.066	0.048	0.292	0.471	0.840	-0.050	0.050	0.239	-1.023	1.256	1.444	0.040	0.502
DP (mg/L)	325	87.1	-2.1766	0.87102	0.081	0.124	0.682	-2.141	0.837	-0.003	0.051	0.450	-1.002	0.093	0.108	0.063	0.211
TP (mg/L)	434	96.3	-1.2683	0.98202	0.067	0.049	0.273	-1.299	1.010	0.005	0.044	0.387	-1.068	0.271	0.271	0.035	0.724
Antimony (µg/L)	164	14.6	1.4793	1.01264	0.269	0.096	1.279	1.275	1.183	0.479	0.113	1.088	-1.334	3.747	3.661	0.150	0.684
Arsenic (µg/L)	267	54.3	1.5218	0.95205	0.113	0.129	0.016	1.191	21.2628	0.752	0.128	0.018	-1.069	4.359	4.359	0.116	0.039
Beryllium (µg/L)	209	10.5	-0.3588	1.94765	0.281	0.163	0.620	-0.892	2.658	0.060	0.197	0.362	-2.074	1.346	0.568	0.231	0.191
Total Cadmium (µg/L)	395	49.4	0.7417	1.12552	0.097	0.083	0.136	0.588	1.276	0.161	0.095	0.060	-1.611	1.898	1.686	0.115	0.012
Total Chromium (µg/L)	256	72.7	2.5512	1.06906	0.1	0.062	0.480	2.599	1.015	-0.359	0.059	0.543	-0.911	13.859	13.657	0.050	0.803
Total Copper (µg/L)	416	89.9	3.2275	1.02866	0.07	0.080	0.017	3.179	1.076	0.716	0.073	0.030	-1.343	22.969	22.081	0.057	0.172
Total Lead (µg/L)	412	76.5	3.3651	1.32824	0.077	0.059	0.223	3.333	1.367	0.379	0.057	0.263	-1.374	38.025	27.991	0.054	0.316
Total Nickel (µg/L)	250	62.4	2.8058	0.97034	0.109	0.065	0.525	2.802	0.971	0.042	0.066	0.512	-0.772	17.094	17.834	0.088	0.182
Total Selenium (µg/L)	203	5.9	1.1472	1.27671	0.375	0.190	0.841	6.916	0.014	-1000.000	0.364	0.083	-1.851	2.381	2.025	0.202	0.753
Total Silver (µg/L)	287	17.4	0.113	1.80819	0.192	0.146	0.236	0.158	1.708	-0.010	0.153	0.194	-1.358	2.181	1.394	0.157	0.170
Total Zinc (µg/L)	433	98.6	5.305	0.970	0.066	0.044	0.389	5.359	0.915	-7.026	0.034	0.743	-0.899	198.198	208.408	0.030	0.951

Table 21. Goodness of Fit for Different Land Uses (p-values Larger than One are Used Only for Comparison) – Continued

OPEN SPACE

CONSTITUENT	N	% Detected	Observed		2 Parameter Log			3 parameter maximum likelihood					3 parameter L-moments				
			μ	σ	D0.05	D2	p-value	μ	σ	c	D3m	p-value	κ	α	ξ	D3l	p-value
TDS (mg/L)	45	97.8	4.762	0.744	0.205	0.120	0.561	4.480	0.962	20.400	0.115	0.621	-			0.759	0
TSS (mg/L)	44	95.5	3.945	1.717	0.21	0.113	0.683	3.672	2.096	2.777	0.095	0.942	-1.184	121.932	72.238	0.173	0.162
BOD (mg/L)	44	86.4	1.6211	0.66954	0.215	0.114	0.746	1.624	0.659	-0.013	0.115	0.733	-0.614	3.421	5.098	0.110	0.801
COD (mg/L)	43	76.7	3.548	0.785	0.231	0.128	0.684	3.323	0.946	5.000	0.151	0.441	-1.221	23.447	-16.159	0.184	0.215
Fecal Coliform (Col/100 ml)	23	91.3	8.9527	1.43209	0.287	0.181	0.503	8.725	1.692	534.506	0.217	0.278	-0.791	13914.660	10684.240	0.187	0.458
Fecal Streptococcus (C/100 ml)	22	90.9	9.6472	1.62819	0.294	0.181	0.538	9.847	1.248	-1070.380	0.165	0.676	-0.599	27514.897	24175.705	0.139	0.921
NO2+NO3 (mg/L)	44	84.1	-0.478	1.07398	0.234	0.141	0.463	-0.378	1.173	0.135	0.141	0.464	-0.579	0.746	0.759	0.122	0.664
TKN (mg/L)	45	71.1	-0.097	0.92495	0.234	0.147	0.500	-0.378	1.173	0.135	0.140	0.569	-0.716	0.918	0.976	0.160	0.389
DP (mg/L)	44	79.5	-2.1844	1.08407	0.224	0.124	0.682	-2.137	1.017	-0.003	0.131	0.606	-0.595	0.135	0.137	0.142	0.484
TP (mg/L)	46	84.8	-1.4264	1.12592	0.215	0.116	0.696	-1.499	1.188	0.009	0.126	0.584	-1.999	0.161	0.171	0.211	0.062
Total Cadmium (μg/L)	38	55.3	0.0291	2.62317	0.287	0.206	0.338	6.921	0.244	-1000.000	0.327	0.022	-1.561	7.207	2.856	0.319	0.028
Total Chromium (μg/L)	36	36.1	-1.79	1.66154	0.325	0.799	0.000	6.926	0.030	-1000.000	0.275	0.279	-1.446	10.225	5.944	0.193	0.756
Total Copper (μg/L)	39	74.4	2.1184	1.20044	0.246	0.131	0.742	6.926	0.035	-1000.000	0.334	0.003	-1.539	8.380	6.792	0.187	0.262
Total Lead (μg/L)	45	42.2	1.8882	1.95431	0.301	0.132	1.034	1.634	2.355	0.188	0.203	0.417	-1.406	16.381	8.735	0.174	0.632
Total Zinc (μg/L)	45	71.1	3.6048	1.19833	0.234	0.105	0.981	3.315	1.527	3.980	0.132	0.658	-1.142	44.515	36.591	0.113	0.885

FREEWAYS

CONSTITUENT	N	% Detected	Observed		2 Parameter Log			3 parameter maximum likelihood					3 parameter L-moments				
			μ	σ	D0.05	D2	p-value	μ	σ	c	D3m	p-value	κ	α	ξ	D3l	p-value
Conductivity (μS/cm)	86	100	4.586	0.681	0.147	0.129	0.113	4.404	0.795	12.765	0.113	0.226	-1.060	58.592	87.231	0.096	0.411
Hardness (mg/L)	127	100	3.604	0.791	0.121	0.077	0.447	3.485	0.875	3.045	0.077	0.440	-1.395	22.761	30.345	0.114	0.074
Oil & Grease (mg/L)	60	71.7	1.974	0.575	0.207	0.101	0.827	1.395	0.976	2.433	0.161	0.214	-0.543	4.299	7.228	0.098	0.881
TDS (mg/L)	97	99	4.279	0.771	0.139	0.054	1.136	4.342	0.720	-3.500	0.051	1.210	-0.660	56.217	74.574	0.044	1.368
TSS (mg/L)	134	99.3	4.464	1.068	0.118	0.066	0.627	4.513	1.010	-2.484	0.059	0.797	-1.447	76.753	75.337	0.120	0.043
BOD (mg/L)	26	84.6	2.262	0.894	0.281	0.103	1.252	2.013	1.084	1.385	0.099	1.304	-1.212	7.478	8.170	0.117	1.097
COD (mg/L)	67	98.5	4.552	0.980	0.167	0.107	0.445	4.798	0.715	-17.925	0.073	0.999	-0.881	81.577	97.057	0.082	0.817
Fecal Coliform (Col/100 ml)	49	100	7.585	1.716	0.194	0.105	0.677	7.514	1.811	30.900	0.128	0.398	-1.569	3543.87	2053.58	0.108	0.635
Fecal Streptococcus (C/100 ml)	25	100	9.263	1.662	0.264	0.119	0.990	8.854	2.363	540.500	0.211	0.215	-0.896	21116.59	15451.91	0.132	0.833
Ammonia (mg/L)	79	87.3	0.009	1.094	0.164	0.133	0.173	0.141	0.942	-0.083	0.106	0.423	-1.065	1.00	1.01	0.123	0.252
NO2+NO3 (mg/L)	25	96	-1.097	0.868	0.269	0.139	0.789	-1.628	1.301	0.087	0.114	1.073	-1.275	0.241	0.275	0.117	1.038
TKN (mg/L)	125	96.8	0.750	0.887	0.124	0.074	0.539	0.723	0.907	0.038	0.075	0.512	-1.159	1.669	1.907	0.071	0.598
DP (mg/L)	22	95.5	-1.226	1.188	0.287	0.170	0.593	-1.670	1.595	0.053	0.207	0.333	-1.992	0.181	0.190	0.255	0.131
TP (mg/L)	128	99.2	-1.266	0.772	0.121	0.085	0.325	-1.545	0.961	0.051	0.065	0.686	-1.394	0.171	0.231	0.089	0.27
Antimony (μg/L)	14	50	0.967	0.284	0.483	0.419	0.171	4.415	0.008	-80.000	0.397	0.220	0.660	0.563	2.922	0.270	0.719
Arsenic (μg/L)	61	55.7	0.952	0.666	0.227	0.149	0.441	0.290	1.180	0.886	0.106	0.937	-0.679	1.787	2.547	0.096	1.07
Total Cadmium (μg/L)	95	71.6	0.016	0.838	0.165	0.052	1.380	-0.108	0.937	0.081	0.067	1.087	-0.781	0.028	1.574	0.051	1.409
Total Chromium (μg/L)	76	98.7	2.096	0.734	0.157	0.084	0.685	2.165	0.680	-0.450	0.075	0.854	-0.555	6.127	8.575	0.055	1.272
Total Copper (μg/L)	97	99.0	3.525	0.842	0.139	0.038	1.507	3.433	0.915	2.047	0.048	1.295	-0.857	28.576	33.493	0.041	1.443
Total Lead (μg/L)	107	100	3.226	1.164	0.131	0.039	1.451	3.178	1.212	0.587	0.045	1.285	-1.155	28.993	24.965	0.040	1.424
Total Nickel (μg/L)	99	89.9	2.325	0.673	0.144	0.091	0.460	1.989	0.896	2.228	0.062	1.013	-0.960	6.390	9.308	0.073	0.769
Total Zinc (μg/L)	93	96.8	5.273	0.877	0.143	0.063	0.985	5.392	0.757	-17.226	0.054	1.000	-0.920	156.885	189.601	0.064	0.959

Experimental design procedures enable the required sample size to be estimated, according to desired confidence and power of the experimental results. It may be possible, without being able to identify the real skewness, that the best distribution fit could be the gamma or exponential distribution.

The utility of the third parameter has been questioned, especially because one of the objectives in modeling is to be parsimonious. Only in cases where it is important to include the effect of unusual elevated values in the model, is the third parameter recommended. In all the other cases, the use of the 2-parameter distribution is adequate to explain the distribution of most of the contaminants.

When the mean and the standard deviation values are not known, Lilieford's test is recommended to evaluate the goodness of fit to a specific distribution. During this research, the Kolmogorov-Smirnov test was used based on the assumption that the large sample sizes minimized errors associated with small sample sizes and uncertainty in the mean and standard deviation values.

Some constituents (such as TKN, TP, COD and Cu) show an increase in the p-value when the number of samples is acceptable and the 3-parameter lognormal probability distribution is used. The use of the lognormal distribution also has an advantage over the other distributions because it can be easily transformed to a normal distribution.

The few cases where gamma distribution seems to be a better model was for cases with low counts (constituents in open space or arsenic, chromium and fecal streptococcus in freeways areas; for example). The exponential distribution better fit total ammonia in

freeway areas. The remaining constituents were well represented by the lognormal distribution.

The 2-parameter lognormal distribution is considered the most appropriate distribution to represent stormwater constituents. Its use facilitates statistical analyses of the data, because procedures such as ANOVA or regression require the errors to be normally distributed. If the number of observations is small, the use of nonparametric methods will be required, as the distributions cannot be accurately determined. Some nonparametric methods require symmetry in the data distribution. The log transformed constituent concentrations usually satisfy these assumptions.

Chapter 5
Identification of Significant Factors Affecting Stormwater Quality Using the NSQD

5.1 Introduction

The normal approach to classify urban sites for estimating stormwater

characteristics is based on land use. This approach is generally accepted because it is

related to the activity in the watershed, plus many site features are generally consistent

within each land use. Two drainage areas with the same size, percentage of

imperviousness, ground slope, sampling methods, and stormwater controls will produce

different stormwater concentrations if the main activity in one watershed is an automobile

manufacturing facility (industrial land use) while the other is a shopping center

(commercial land use) for example. There will likely be higher concentrations of metals

at the industrial site due to the manufacturing processes, while the commercial site may

have higher concentrations of PAHs (polycyclic aromatic hydrocarbons) due to the

frequency and numbers of customer automobiles entering and leaving the parking lots.

The results from the previous chapter indicated that there are significant

differences in stormwater constituents for different land use categories. This is supported

for other databases like NURP (EPA 1983), CDM (Smullen and Cave 2002) and USGS

(Driver et al. 1985). The main question to be addressed in this chapter is if there is a

different classification method that better describes stormwater quality, possibly by also

considering such factors as geographical area (EPA Rain Zone), season, percentage of

128

imperviousness, type of conveyance, controls in the watershed, sampling method, and

type of sample compositing, and possible interactions between these factors.

This chapter presents several approaches to explain the variability of stormwater

quality by considering these additional factors. As shown in Chapter 3, ignoring the non-

detected observations can adversely affect the mean, median and standard deviations of

the dataset, and the resulting statistical test results. Therefore, the calculations presented

in this chapter used the censored observations using the Cohen's maximum likelihood

method.

5.2 Main Factors Affecting Stormwater Quality

The EPA Rain Zone, percentage of imperviousness, watershed size, land use, type

of conveyance, controls in the watershed, sample analysis method, and type of sampling

procedures were selected as potential influencing factors affecting stormwater quality for

the preliminary analyses. Data from sites having single land use will be used in the basic

analyses, while data from the mixed land use sites could be used for verification. The first

step was to inventory the total number of events in each of the possible combinations of

these factors. The EPA Rain Zone, land use, type of conveyance, type of controls present

in the watershed, sampling methods and type of compositing procedures are discrete

variables, while percentage of imperviousness and watershed area are continuous

variables. The total counts and percentage for each discrete variable option is shown in

Table 22.

Table 22. Numbers and percentage of samples by discrete site variable category

Land use	Events	%
Residential	1042	27.68
Mixed Residential	611	16.23
Commercial	527	14.00
Mixed Commercial	324	8.61
Industrial	566	15.03
Mixed Industrial	249	6.61
Institutional	18	0.48
Open Space	49	1.30
Mixed Open Space	168	4.46
Freeways	185	4.91
Mixed Freeways	26	0.69

EPA Rain Zone	Events	%
1	69	1.83
2	2000	53.12
3	266	7.07
4	212	5.63
5	485	12.88
6	356	9.46
7	229	6.08
8	24	0.64
9	124	3.29

Controls	Events	%
Channel Weirs (CW)	30	0.80
Dry Pond (DP)	50	1.33
Detention Storage (enlarged pipe) (DS)	17	0.45
Wet Pond at Outfall (WP)	113	3.00
WP in Watershed (WP_W)	182	4.83
WP in Series at Outfall (WP_S)	42	1.12
None	3331	88.47

Sample Analysis	Events	%
Composite, type not specified	718	19.07
Flow Composite	2752	73.09
Time Composite	295	7.84

Type of Conveyance	Events	%
Curb and gutter	2454	65.18
Grass swale	344	9.14
Not specified	967	25.68

Sampler	Events	%
Automatic	3055	81.14
Manual	393	10.44
Not specified	317	8.42

About 80% of the samples were collected using automatic samplers. It was observed that manual sampling can result in lower TSS concentrations compared to automatic sampling procedures. This may occur, for example, if the manual sampling team arrives after the start of runoff and therefore misses the first flush (if it exists for the site), resulting in reduced event mean concentrations. For those sites using automatic

samplers, about 73% of the events were collected using flow-composite samplers, 8% were collected using time-composite samplers, and about 19% did not have any designation available. Flow-composite samples are considered more accurate than time-composite samples when obtaining data for event mean concentrations, unless very large numbers of subsamples are obtained (Roa-Espinosa and Bannerman 1995).

Almost 66% of the events were collected at sites drained with conventional curbs and gutters, 9% were collected at sites having roadside grass swales, and it was not possible to determine the drainage system for about 25% of the samples. Grass swales can reduce the concentrations of suspended solids and metals, especially during low flows. They can also infiltrate large quantities of the stormwater, reducing pollutant mass discharges, runoff volume, and peak flows.

5.3 Effects of Stormwater Controls on Stormwater Quality

It is hoped that stormwater controls located in a watershed, or at an outfall, would result in significant reductions in stormwater pollutant concentrations. Figure 36 shows the effects on effluent TSS concentrations when using various controls in residential area watersheds in EPA Rain Zone 2 (Maryland, Virginia, North Carolina, Tennessee and Kentucky), the area having large enough numbers of samples for an effective statistical analysis. The controls noted for these locations included:

1. Channel weir: a flow measurement weir in an open channel that forms a small pool (a very small wet pond).
2. Dry pond (DP): a dry detention pond that drains completely between each storm event.
3. Wet pond (WP): a wet detention pond that retains water between events, forming a small lake or pond. If the pond is in the watershed but not at the outfall, this will

be considered a wet pond inside of the watershed (WPW), which would only treat a fraction of the total stormwater from the site

4. Detention storage (DS): Oversize pipes with small outlet orifices, usually under parking lots.
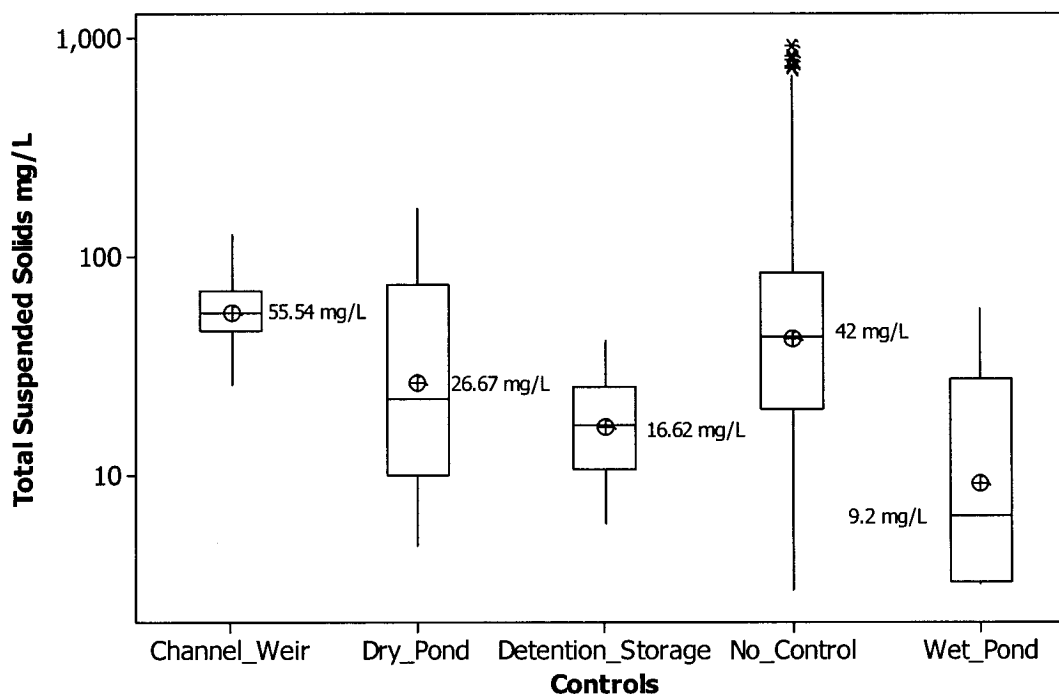
Fig. 36. TSS distribution by controls in residential areas and EPA Rain Zone 2 (the cross circles indicate the average concentrations, while the median concentrations are written next to the median bar in the box diagrams)

The stormwater monitoring was conducted at the outfalls of the drainage areas, after the stormwater controls. Wet ponds are seen to reduce the TSS concentration in the stormwater more than the other controls (about 78%) compared to the "no control" median value. Detention storage units and dry ponds also reduced the TSS concentrations, but to a smaller extent (about 60% and 37% respectively). Only one site (located in Virginia Beach) had a channel weir control, but that site did not reduce the observed TSS concentrations compared to the "no control" category.

The effectiveness of the stormwater controls was evaluated for each constituent separately. The effects of sample analysis method, sampler instrument, and type of conveyance were also examined.

The first step was to identify the suitable subsets that could be examined, based on suitable numbers of samples in each category. The following four land uses and EPA Rain Zones had suitable numbers of sites having controls that could be examined: residential, commercial and industrial in EPA Rain Zone 2 and industrial in EPA Rain Zone 3. For each group, one-way ANOVA analyses were used to identify if there were any differences in the concentrations of 13 constituents (after log-transformations and substitutions for non-detectable values) for those sites that included different controls. Dunnet's method was also used to compare sites with each specific stormwater control type to sites without stormwater controls, using a family error rate of 5%. Table 23 shows the results for these analyses for each of these groups.

Tables 23 through 26 show that there are no significant differences between sites with or without wet ponds for all constituents having observations in industrial land uses in EPA Rain Zone 3. Nitrite-nitrate, total phosphorus, total copper and total zinc were significantly lower in concentrations at sites located in EPA Rain Zone 2, having wet ponds before the outfall, compared to sites without stormwater controls. Wet ponds did not reduce the TKN concentrations in any of the four groups. Significant reductions in TSS concentrations were also observed for sites having wet ponds in residential and commercial land uses, but not in industrial land uses.

Table 23. One-Way ANOVA Results by Control Type in Residential Land Use, Rain Zone 2

Hardness mg/L — p-value 0.024*

	n	median	Dunnet
Weir			
DP			
DS	7	44.38	=
No Control	61	30.77	
WP	10	66.45	>

Oil and Grease mg/L — p-value 0.999

	n	median	Dunnet
Weir	3	2.50	=
DP	3	2.68	=
DS	9	2.19	=
No Control	202	2.38	
WP	13	2.50	=

TDS mg/L — p-value 0

	n	median	Dunnet
Weir	29	112.80	>
DP	3	58.88	=
DS	8	98.45	=
No Control	424	62.42	
WP	12	120.39	>

TSS mg/L — p-value 0

	n	median	Dunnet
Weir	29	55.54	=
DP	21	26.67	=
DS	9	14.46	<
No Control	559	40.10	
WP	12	9.25	<

BOD mg/L — p-value 0

	n	median	Dunnet
Weir	29	6.16	<
DP	21	3.44	<
DS	9	3.66	<
No Control	533	11.07	
WP	12	3.10	<

COD mg/L — p-value 0

	n	median	Dunnet
Weir	29	49.02	=
DP	3	33.45	=
DS	9	22.17	<
No Control	418	56.91	
WP	12	24.58	<

Ammonia mg/L — p-value 0

	n	median	Dunnet
Weir	29	0.05	<
DP	3	0.41	=
DS	9	0.40	=
No Control	409	0.24	
WP	12	0.07	<

NO$_2$ + NO$_3$ mg/L — p-value 0

	n	median	Dunnet
Weir	29	0.05	<
DP	21	0.59	=
DS	9	0.98	=
No Control	546	0.54	
WP	12	0.28	<

TKN mg/L — p-value 0.012

	n	median	Dunnet
Weir	29	1.49	=
DP	21	0.79	<
DS	9	1.38	=
No Control	549	1.34	
WP	12	1.04	<

Dissolved Phosphorus mg/L — p-value 0

	n	median	Dunnet
Weir	29	0.04	<
DP	3	0.15	=
DS	8	0.11	=
No Control	404	0.14	
WP	12	0.03	<

Total Phosphorus mg/L — p-value 0

	n	median	Dunnet
Weir	29	0.23	=
DP	21	0.12	<
DS	9	0.15	<
No Control	550	0.30	
WP	12	0.07	<

Total Copper µg/L — p-value 0

	n	median	Dunnet
Weir	3	2.69	<
DP	21	6.16	<
DS	9	20.75	>
No Control	403	11.01	
WP	4	3.13	<

Total Lead µg/L — p-value 0

	n	median	Dunnet
Weir	3	6.41	<
DP	21	1.50	<
DS	9	1.16	<
No Control	364	7.73	
WP	4	1.00	<

Total Zinc µg/L — p-value 0

	n	median	Dunnet
Weir	3	4.11	<
DP	21	29.63	<
DS	9	103.25	>
No Control	405	67.56	
WP	4	10.44	<

Table 24. One-Way ANOVA Results by Control Type in Commercial Land Use, Rain Zone 2

Hardness mg/L — p-value 0.717

	n	median	Dunnet
DS	8	58.17	=
No Control	35	58.97	
WP	11	71.80	=
WPW	9	47.11	=

Oil and Grease mg/L — p-value 0.082

	n	median	Dunnet
DS	8	1.84	=
No Control	100	4.20	
WP	17	2.84	=
WPW	13	3.36	=

TDS mg/L — p-value 0.477

	n	median	Dunnet
DS	8	100.69	=
No Control	174	74.89	
WP	26	89.99	=
WPW	13	71.12	=

TSS mg/L — p-value 0

	n	median	Dunnet
DS	8	19.54	v
No Control	244	48.13	
WP	26	19.47	v
WPW	13	16.85	v

BOD mg/L — p-value 0

	n	median	Dunnet
DS	8	4.44	v
No Control	241	14.66	
WP	26	7.06	v
WPW	12	5.41	v

COD mg/L — p-value 0

	n	median	Dunnet
DS	8	27.18	v
No Control	174	73.62	
WP	26	35.99	v
WPW	13	23.88	v

Ammonia mg/L — p-value 0

	n	median	Dunnet
DS	8	0.30	=
No Control	174	0.39	
WP	26	0.13	v
WPW	13	0.16	v

NO$_2$ + NO$_3$ mg/L — p-value 0

	n	median	Dunnet
DS	8	1.18	=
No Control	242	0.60	
WP	26	0.48	=
WPW	13	0.22	v

TKN mg/L — p-value 0.057

	n	median	Dunnet
DS	8	1.04	=
No Control	241	1.59	
WP	26	1.19	=
WPW	13	1.03	=

Dissolved Phosphorus mg/L — p-value 0

	n	median	Dunnet
DS	7	0.09	=
No Control	161	0.11	
WP	25	0.05	=
WPW	13	0.03	=

Total Phosphorus mg/L — p-value 0

	n	median	Dunnet
DS	8	0.16	=
No Control	238	0.25	
WP	26	0.13	v
WPW	13	0.08	v

Total Lead µg/L — p-value 0

	n	median	Dunnet
DS	8	1.61	v
No Control	194	16.41	
WP	7	4.90	v
WPW	4	2.49	v

Total Copper µg/L — p-value 0

	n	median	Dunnet
DS	8	14.14	=
No Control	194	17.53	
WP	6	5.57	v
WPW	4	6.00	v

Total Zinc µg/L — p-value 0

	n	median	Dunnet
DS	8	82.57	v
No Control	197	188.02	
WP	7	44.26	v
WPW	4	39.68	v

Table 25. One-Way ANOVA Results by Control Type in Industrial Land Use, Rain Zone 2

Hardness mg/L			
p-value	none		
	n	median	Dunnet
No Control			
WP			

Oil and Grease mg/L			
p-value	0		
	n	median	Dunnet
No Control	81	3.85	
WP	37	1.43	<

TDS mg/L			
p-value	none		
	n	median	Dunnet
No Control			
WP			

TSS mg/L			
p-value	0.693		
	n	median	Dunnet
No Control	205	51.96	
WP	29	48.05	=

BOD mg/L			
p-value	0.466		
	n	median	Dunnet
No Control	200	10.63	
WP	29	9.30	=

COD mg/L			
p-value	none		
	n	median	Dunnet
No Control			
WP			

Ammonia mg/L			
p-value	none		
	n	median	Dunnet
No Control			
WP			

$NO_2 + NO_3$ mg/L			
p-value	0		
	n	median	Dunnet
No Control	197	0.61	
WP	29	0.30	<

TKN mg/L			
p-value	0.166		
	n	median	Dunnet
No Control	198	1.22	
WP	29	0.98	=

Dissolved Phosphorus mg/L			
p-value	none		
	n	median	Dunnet
No Control			
WP			

Total Phosphorus mg/L			
p-value	0		
	n	median	Dunnet
No Control	200	0.23	
WP	29	0.09	<

Total Copper µg/L			
p-value	0		
	n	median	Dunnet
No Control	150	16.00	
WP	29	7.38	<

Total Lead µg/L			
p-value	0.353		
	n	median	Dunnet
No Control	142	11.16	
WP	29	8.66	=

Total Zinc µg/L			
p-value	0		
	n	median	Dunnet
No Control	157	180.01	
WP	29	60.44	<

Table 26. One-Way ANOVA Results by Control Type in Industrial Land Use, Rain Zone 3

Hardness mg/L

	Hardness mg/L		
p-value	None		
	n	median	Dunnet
No Control			
WP			

TDS mg/L

	TDS mg/L		
p-value	0.112		
	N	median	Dunnet
No Control	44	69.53	
WP	25	49.84	=

TSS mg/L

	TSS mg/L		
p-value	0.281		
	n	median	Dunnet
No Control	44	48.35	
WP	25	70.40	=

BOD mg/L

	BOD mg/L		
p-value	0.221		
	n	median	Dunnet
No Control	44	6.41	
WP	23	5.14	=

Oil and Grease mg/L

	Oil and Grease mg/L		
p-value	None		
	n	median	Dunnet
No Control			
WP			

COD mg/L

	COD mg/L		
p-value	0.395		
	n	median	Dunnet
No Control	44	37.00	
WP	25	43.06	=

$NO_2 + NO_3$ mg/L

	$NO_2 + NO_3$ mg/L		
p-value	0.193		
	n	median	Dunnet
No Control	30	0.57	
WP	25	0.40	=

TKN mg/L

	TKN mg/L		
p-value	0.807		
	n	median	Dunnet
No Control	43	1.18	
WP	25	1.12	=

Ammonia mg/L

	Ammonia mg/L		
p-value	0.165		
	N	median	Dunnet
No Control	3	0.12	
WP	25	0.03	=

Total Copper µg/L

	Total Copper µg/L		
p-value	0.106		
	n	median	Dunnet
No Control	38	16.66	
WP	25	12.58	=

Total Lead µg/L

	Total Lead µg/L		
p-value	0.454		
	n	median	Dunnet
No Control	31	8.49	
WP	25	6.73	=

Dissolved Phosphorus mg/L

	Dissolved Phosphorus mg/L		
p-value	0.191		
	n	median	Dunnet
No Control	39	0.07	
WP	25	0.06	=

Total Phosphorus mg/L

	Total Phosphorus mg/L		
p-value	0.438		
	N	median	Dunnet
No Control	43	0.16	
WP	25	0.19	=

Total Zinc µg/L

	Total Zinc µg/L		
p-value	0.608		
	n	median	Dunnet
No Control	38	143.28	
WP	25	156.93	=

Note. The bold, italicized probability values indicate "statistically significant" findings at the 0.05 level, or better. Dunnet test compared if sites with control produces larger concentrations ">", smaller concentrations "<" or not statistically difference "=" than sites without control at a family error of 5%.
"None" indicates no samples were collected for this constituent in the group.

Dry ponds were only available for evaluation in the residential land use category in EPA Rain Zone 2. No significant differences were found for TSS or nitrite-nitrate for sites having dry ponds. However, significant reductions of BOD_5, TKN, total phosphorus, total copper, total lead and total zinc were noted.

Some communities have installed detention-storage facilities (enlarged pipes) under parking lots to reduce runoff flow rates. More than 400 of these underground pipes are located in Arlington, Virginia, for example. A significant reduction in the TSS, BOD_5, COD, total lead, and total zinc concentrations were observed at sites with these underground devices. On the other hand, these controls did not indicate a significant difference in the concentrations of nutrients (ammonia, nitrite-nitrate, TKN, dissolved phosphorus and total phosphorus), compared to comparable sites not having stormwater controls. A conflicting situation was observed in EPA Rain Zone 2 for total zinc for sites having underground enlarged pipes. Zinc concentrations at residential land uses were significantly higher, while zinc concentrations at commercial areas were significantly lower, compared to sites with no stormwater controls. It is possible that the sites having elevated zinc concentrations used galvanized metal enlarged pipe systems.

5.4 Sampling Method Effects on Stormwater Concentrations

The use of manual or automatic sampling is a factor that is sometimes mentioned as having a possible effect on the quality of the collected samples. Manual sampling is usually preferred when the number of samples is small and when there are not available resources for the purchase, installation, operation, and maintenance of automatic samplers. Manual sampling may also be required when the constituents being sampled

require specific handling (such as for bacteria, oil and grease, and volatile organic compounds) (ASCE/EPA 2002). Automatic samplers are recommended for larger sampling programs, when better representations of the flows are needed, and especially when site access is difficult or unsafe. In most cases, where a substantial number of samples are to be collected and when composite sampling is desired, automatic sampling can be much less expensive. Automatic samples also improve repeatability by reducing additional variability induced by the personnel from sample to sample (Bailey 1993). Most importantly, automatic samplers can be much more reliable compared to manual sampling, especially when the goal of a monitoring project is to obtain data for as many of the events that occur as possible, and sampling must start near the beginning of the rainfall (Burton and Pitt 2002).

Residential, commercial and industrial sites located in EPA Rain Zone 2 were used to evaluate any significant differences between the two sampling methods. One-way ANOVA analyses were used to identify any statistical differences between the two groups. Dunnet's test was used to compare manual sampling against automatic sampling. Tables 27 through 29 show the results from the ANOVA analyses.

Table 27. One-Way ANOVA Results by Type of Sampler in Residential Land Use, Rain Zone 2

Hardness mg/L

	n	median	Dunnet
p-value		0	
Automatic	23	51.9	
Manual	28	22.4	<

Oil and Grease mg/L

	n	median	Dunnet
All manual			

TSS mg/L

	n	median	Dunnet
p-value		0	
Automatic	420	45.5	
Manual	78	19.2	<

BOD mg/L

	n	median	Dunnet
p-value		0.162	
Automatic	396	11.3	
Manual	78	9.8	=

COD mg/L

	n	median	Dunnet
p-value		0	
Automatic	312	62.2	
Manual	66	36.4	<

TDS mg/L

	n	median	Dunnet
p-value		0.004	
Automatic	318	65.4	
Manual	66	50.0	<

Ammonia mg/L

	n	median	Dunnet
p-value		0.909	
Automatic	310	0.229	
Manual	66	0.233	=

NO$_2$ + NO$_3$ mg/L

	n	median	Dunnet
p-value		0.005	
Automatic	410	0.51	
Manual	75	0.66	>

TKN mg/L

	n	median	Dunnet
p-value		0.048	
Automatic	410	1.40	
Manual	78	1.16	<

Dissolved Phosphorus mg/L

	n	median	Dunnet
p-value		0.308	
Automatic	302	0.136	
Manual	63	0.120	=

Total Phosphorus mg/L

	n	median	Dunnet
p-value		0	
Automatic	416	0.325	
Manual	73	0.230	<

Total Lead µg/L

	n	median	Dunnet
p-value		0	
Automatic	247	9.74	
Manual	71	4.14	<

Total Copper µg/L

	n	median	Dunnet
p-value		0.025	
Automatic	256	11.57	
Manual	77	8.80	<

Total Zinc µg/L

	n	median	Dunnet
p-value		0.02	
Automatic	256	73.71	
Manual	76	53.22	<

Table 28. One-Way ANOVA Results by Type of Sampler in Commercial Land Use, Rain Zone 2

Hardness mg/L

p-value		0	
	n	median	Dunnet
Automatic	23	97.86	
Manual	12	22.34	<

TSS mg/L

p-value		0	
	n	median	Dunnet
Automatic	179	52.29	
Manual	24	20.55	<

Oil and Grease mg/L

p-value		0.009	
	n	median	Dunnet
Automatic	70	4.75	
Manual	19	2.30	<

TDS mg/L

p-value		0.25	
	n	median	Dunnet
Automatic	123	76.36	
Manual	18	60.80	=

BOD mg/L

p-value		0.189	
	n	median	Dunnet
Automatic	178	14.86	
Manual	23	11.70	=

COD mg/L

p-value		0.003	
	n	median	Dunnet
Automatic	123	79.74	
Manual	18	44.02	<

NO$_2$ + NO$_3$ mg/L

p-value		0.137	
	n	median	Dunnet
Automatic	178	0.55	
Manual	23	0.75	=

TKN mg/L

p-value		0.117	
	n	median	Dunnet
Automatic	177	1.63	
Manual	24	1.21	=

Dissolved Phosphorus mg/L

p-value		0.554	
	n	median	Dunnet
Automatic	113	0.097	
Manual	17	0.115	=

Total Phosphorus mg/L

p-value		0.003	
	n	median	Dunnet
Automatic	176	0.261	
Manual	23	0.157	<

Total Lead µg/L

p-value		0.422	
	n	median	Dunnet
Automatic	130	17.62	
Manual	20	13.66	=

Total Zinc µg/L

p-value		0.404	
	n	median	Dunnet
Automatic	130	208	
Manual	23	168	=

Total Copper µg/L

p-value		0.001	
	n	median	Dunnet
Automatic	127	20.27	
Manual	23	11.80	<

Table 29. One-Way ANOVA Results by Type of Sampler in Industrial Land Use, Rain Zone 2

Hardness mg/L

	n	median	Dunnet
p-value		none	
Automatic			
Manual			

Oil and Grease mg/L

	n	median	Dunnet
p-value		0.723	
Automatic	62	3.68	=
Manual	14	4.10	

TDS mg/L

	n	median	Dunnet
p-value		0.362	
Automatic	128	73.2	=
Manual	10	100.0	

TSS mg/L

	n	median	Dunnet
p-value		0.402	
Automatic	171	51.45	=
Manual	19	62.82	

BOD mg/L

	n	median	Dunnet
p-value		0.112	
Automatic	166	9.65	=
Manual	19	13.47	

COD mg/L

	n	median	Dunnet
p-value		0.371	
Automatic	127	55.02	=
Manual	10	67.68	

Ammonia mg/L

	n	median	Dunnet
p-value		*0*	
Automatic	122	0.243	>
Manual	10	1.54	

NO$_2$ + NO$_3$ mg/L

	n	median	Dunnet
p-value		*0.021*	
Automatic	163	0.558	>
Manual	19	0.904	

TKN mg/L

	n	median	Dunnet
p-value		**0.008**	
Automatic	164	1.135	>
Manual	19	1.944	

Dissolved Phosphorus mg/L

	n	median	Dunnet
p-value		0.870	
Automatic	109	0.091	=
Manual	10	0.086	

Total Phosphorus mg/L

	N	median	Dunnet
p-value		*0.056*	
Automatic	166	0.214	=
Manual	19	0.315	

Total Copper µg/L

	N	median	Dunnet
p-value		0.797	
Automatic	108	15.66	=
Manual	22	14.97	

Total Lead µg/L

	n	median	Dunnet
p-value		0.908	
Automatic	109	11.27	=
Manual	16	10.83	

Total Zinc µg/L

	n	median	Dunnet
p-value		*0.028*	
Automatic	115	156	>
Manual	22	233	

Note. The bold, italicized probability values indicate "statistically significant" findings at about the 0.05 level, or better. Dunnet test compared if sites with control produces larger concentrations ">", smaller concentrations "<" or not statistically difference "=" than sites without control at a family error of 5%.
"None" indicates no samples were collected for this constituent in the group.

Tables 27 through 29 indicated that BOD_5 and dissolved phosphorus measurements are not affected by differences in sampling methods used in residential, commercial or industrial areas in EPA Rain Zone2. In residential and commercial land uses, TSS and COD concentrations obtained using automatic samplers were almost twice the concentrations obtained when using manual sampling methods. Median total phosphorus concentrations were about 50% higher using automatic samplers, while no effects were noted for other nutrients. Figure 37 contains box and whisker plots comparing automatic versus manual sampling methods in residential land uses in EPA Rain Zone 2. TSS, total copper and total zinc have lower concentrations using manual sampling compared with automatic sampling (p-values of 0, 0.025 and 0.02 respectively). The opposite pattern was observed for nitrate-nitrate, manual sampling shows higher concentrations than samples collected with automatic samples (p-value of 0.005).

In industrial land uses, the pattern was found to be opposite. Ammonia, nitrate-nitrite, TKN and total zinc indicated higher concentrations when using manual sampling methods compared to using automatic samplers. Concentrations for these constituents were almost twice as high when using manual sampling, except for ammonia that was almost six times higher when manual sampling was used compared to automatic sampling methods. These elevated concentrations were observed in industrial sites located in Fairfax County Virginia, Howard County Maryland and the city of Charlotte in North Carolina. Sites with controls were not included in the previous analyses.
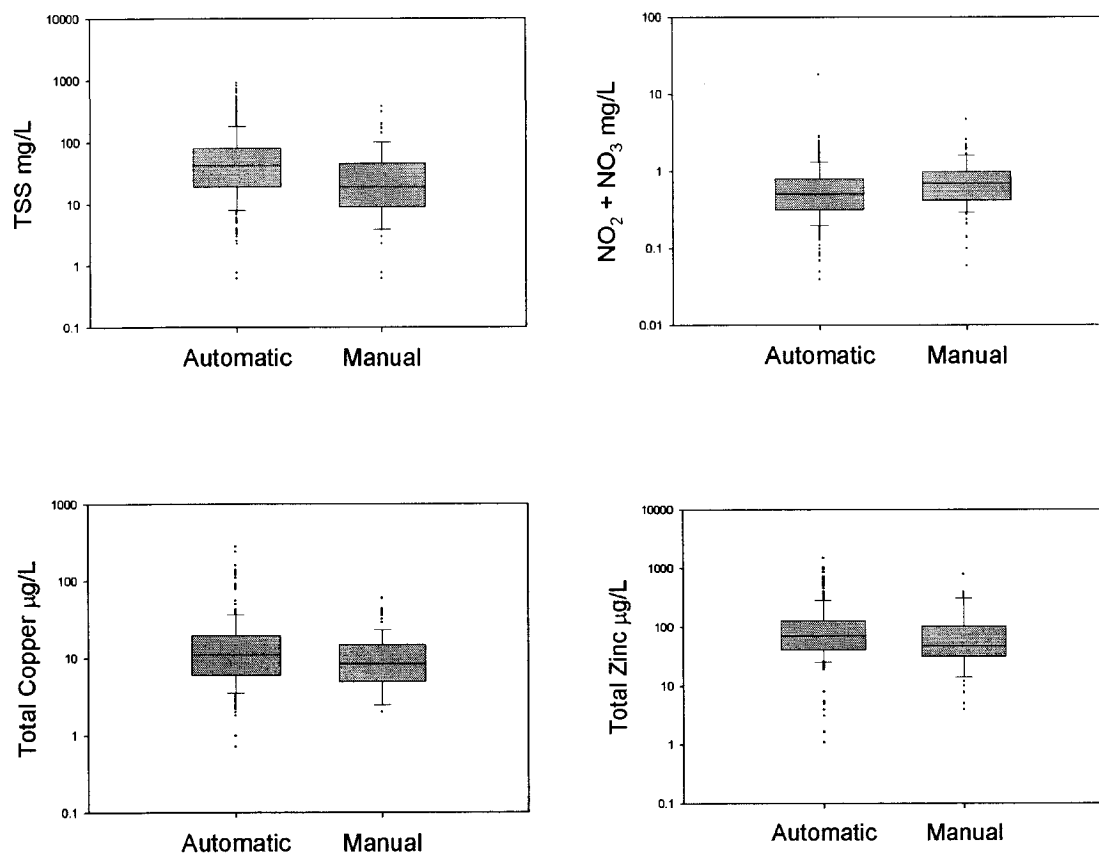
Fig. 37. Comparison of reported concentrations in residential land use and EPA Rain Zone 2 for automatic vs. manual sampling methods

5.5 Sample Compositing Procedures

Time and flow-weighted composite options were also evaluated in residential, commercial, and industrial land uses in EPA Rain Zone 2 and in industrial land uses in EPA Rain Zone 3. With time-compositing, individual subsamples are combined for even time increments. As an example, automatic samplers can be programmed to collect a subsample every 15 minutes for deposit into a large composite bottle. An automatic sampler can also collect discrete subsamples at even time increments, keeping each sample in a separate smaller sample bottle. After the sampled event, these samples can be

manually combined as a composite. With flow-weighted sampling, an automatic sampler can be programmed to deposit a subsample into a large composite bottle for each set increment of flow.

The Wisconsin Department of Natural Resources conducted a through evaluation of alternative sampling modes for stormwater sampling to determine the average pollutant concentrations for individual events (Roa-Espinosa and Bannerman 1995). Four sampling modes were compared at outfalls at five industrial sites, including: flow-weighted composite sampling, time-discrete sampling, time-composite sampling, and first flush sampling during the first 30 minutes of runoff. Based on many attributes, they concluded that time-composite sampling at outfalls is the best method due to simplicity, low cost, and good comparisons to flow-weighted composite sampling (assumed to be the most accurate). The time-composite sampling cost was about 25% of the cost of the time-discrete and flow-weighted sampling schemes, but was about three times the cost of the first flush sampling only. The accuracy and reproducibility of the composite samples were all good, while these attributes for the first flush samples were poor. Burton and Pitt (2001) stress that it is important to ensure that acceptable time-weighted composite sampling include many subsamples. Any sampling scheme is very inaccurate if too few samples are collected. Samples need to be collected to represent the extreme conditions during the event, and the total storm duration. Experimental design methods can be used to determine the minimum number of subsamples needed considering likely variations. It is more common to now include the use of "continuous" water quality probes at sampling locations, with in-situ observations obtained every few minutes. Unfortunately, these details were not available for the NSQD sampling sites; some sites may have had too few

subsamples to represent the storm conditions, while others may have had sufficient numbers of subsamples. Also, most of the NSQD samples only represented the first 3 hours of runoff events. If events were longer, the later storm periods were likely not represented. These issues are discussed more in the next subsection.

One-way ANOVA tests were used to evaluate the presence of significant differences between these two composite sampling schemes. Dunnet's comparison test was used to evaluate if concentrations associated with time-compositing were larger or lower than concentrations associated with flow- compositing. Tables 30 through 33 show the results of these tests.

Table 30. One-Way ANOVA Results by Sample Compositing Scheme in Residential Land Use, Rain Zone 2

Hardness mg/L				Oil and Grease mg/L				TDS mg/L				TSS mg/L		
p-value	none			p-value				p-value	0.229			p-value	0	
	n	median	Dunnet		n	median	Dunnet		n	Median	Dunnet		n	median	Dunnet
Flow composite				Flow composite	No composite			Flow composite	351	64.02		Flow composite	398	36.08	
Time composite				Time composite				Time composite	14	76.90	=	Time composite	80	90.30	>

BOD mg/L				COD mg/L				Ammonia mg/L				NO$_2$ + NO$_3$ mg/L		
p-value	0.785			p-value	0.416			p-value	0			p-value	0.097	
	n	median	Dunnet		n	median	Dunnet		n	median	Dunnet		n	median	Dunnet
Flow composite	379	11.04		Flow composite	348	56.28		Flow composite	345	0.24		Flow composite	388	0.52	
Time composite	78	10.75	=	Time composite	14	47.93	=	Time composite	14	0.62	>	Time composite	80	0.60	=

TKN mg/L				Dissolved Phosphorus mg/L				Total Phosphorus mg/L				Total Copper µg/L		
p-value	0.215			p-value	0.832			p-value	0			p-value	0	
	n	median	Dunnet		n	median	Dunnet		n	median	Dunnet		n	median	Dunnet
Flow composite	391	1.30		Flow composite	334	0.139		Flow composite	392	0.292		Flow composite	228	9.99	
Time composite	80	1.46	=	Time composite	14	0.132	=	Time composite	80	0.426	>	Time composite	85	16.89	>

Total Lead µg/L				Total Zinc µg/L		
p-value	0			p-value	0	
	n	median	Dunnet		n	median	Dunnet
Flow composite	222	5.94		Flow composite	227	50.77	
Time composite	85	19.62	>	Time composite	85	142	>

Table 31. One-Way ANOVA Results by Sample Compositing Scheme in Commercial Land Use, Rain Zone 2

Hardness mg/L — p-value: none

	n	median	Dunnet
Flow composite			
Time composite			

Oil and Grease mg/L — p-value: Few samples

	n	median	Dunnet
Flow composite			
Time composite			

TSS mg/L — p-value: 0

	n	median	Dunnet
Flow composite	163	38.18	
Time composite	30	135.6	>

BOD mg/L — p-value: 0.563

	n	median	Dunnet
Flow composite	162	13.43	
Time composite	30	14.56	=

COD mg/L — p-value: Few samples

	n	median	Dunnet
Flow composite			
Time composite			

Ammonia mg/L — p-value: Few samples

	n	median	Dunnet
Flow composite			
Time composite			

$NO_2 + NO_3$ mg/L — p-value: 0.875

	n	median	Dunnet
Flow composite	163	0.583	
Time composite	30	0.567	=

TKN mg/L — p-value: 0.637

	n	median	Dunnet
Flow composite	163	1.47	
Time composite	30	1.36	=

Dissolved Phosphorus mg/L — p-value: Few samples

	n	median	Dunnet
Flow composite			
Time composite			

Total Phosphorus mg/L — p-value: 0.118

	n	median	Dunnet
Flow composite	161	0.242	
Time composite	30	0.194	=

Total Lead µg/L — p-value: 0

	n	median	Dunnet
Flow composite	115	11.96	
Time composite	30	52.23	>

Total Copper µg/L — p-value: 0

	n	median	Dunnet
Flow composite	115	14.91	
Time composite	30	36.34	>

Total Zinc µg/L — p-value: 0

	n	median	Dunnet
Flow composite	115	156	
Time composite	30	408	>

Table 32. One-Way ANOVA Results by Sample Compositing Scheme in Industrial Land Use, Rain Zone 2

Hardness mg/L

p-value		none	
	n	median	Dunnet
Flow composite			
Time composite			

Oil and Grease mg/L

p-value		Few samples	
	n	median	Dunnet
Flow composite			
Time composite			

TSS mg/L

p-value		0	
	n	median	Dunnet
Flow composite	116	44.2	
Time composite	40	84.6	>

TDS mg/L

p-value		0.076	
	n	median	Dunnet
Flow composite	101	68.5	
Time composite	9	132.9	=

BOD mg/L

p-value		0.861	
	n	median	Dunnet
Flow composite	112	9.67	
Time composite	39	9.94	=

COD mg/L

p-value		0.519	
	n	median	Dunnet
Flow composite	100	53.93	
Time composite	9	63.04	=

$NO_2 + NO_3$ mg/L

p-value		0.488	
	n	median	Dunnet
Flow composite	109	0.547	
Time composite	39	0.614	=

Ammonia mg/L

p-value		0	
	n	median	Dunnet
Flow composite	96	0.25	
Time composite	9	1.11	>

TKN mg/L

p-value		0.672	
	n	median	Dunnet
Flow composite	109	1.06	
Time composite	40	1.13	=

Dissolved Phosphorus mg/L

p-value		0.601	
	n	median	Dunnet
Flow composite	82	0.087	
Time composite	9	0.074	=

Total Phosphorus mg/L

p-value		0.338	
	n	median	Dunnet
Flow composite	111	0.208	
Time composite	40	0.242	=

Total Copper µg/L

p-value		0.070	
	n	median	Dunnet
Flow composite	72	15.75	
Time composite	40	21.27	=

Total Lead µg/L

p-value		0.001	
	n	median	Dunnet
Flow composite	66	9.34	
Time composite	40	22.23	>

Total Zinc µg/L

p-value		0.930	
	n	median	Dunnet
Flow composite	72	189.7	
Time composite	40	186.8	=

Table 33. One-Way ANOVA Results by Sample Compositing Scheme in Industrial Land Use, Rain Zone 3

Hardness mg/L — p-value: none

	n	median	Dunnet
Flow composite			
Time composite			

TDS mg/L — p-value: *0.012*

	n	median	Dunnet
Flow composite	16	93.93	
Time composite	20	43.11	<

TSS mg/L — p-value: 0.103

	n	median	Dunnet
Flow composite	16	22.44	
Time composite	20	52.38	=

Oil and Grease mg/L — p-value: none

	n	median	Dunnet
Flow composite			
Time composite			

BOD mg/L — p-value: 0.265

	n	median	Dunnet
Flow composite	16	5.32	
Time composite	20	7.25	=

COD mg/L — p-value: 0.088

	n	median	Dunnet
Flow composite	16	33.38	
Time composite	20	50.36	=

Ammonia mg/L — p-value: none

	n	median	Dunnet
Flow composite			
Time composite			

NO$_2$ + NO$_3$ mg/L — p-value: *0.006*

	n	median	Dunnet
Flow composite	9	0.248	
Time composite	19	0.844	>

TKN mg/L — p-value: *0.002*

	n	median	Dunnet
Flow composite	15	0.64	
Time composite	20	1.63	>

Dissolved Phosphorus mg/L — p-value: *0*

	n	median	Dunnet
Flow composite	11	0.035	
Time composite	20	0.116	>

Total Phosphorus mg/L — p-value: 0.556

	N	median	Dunnet
Flow composite	15	0.161	
Time composite	20	0.189	=

Total Copper µg/L — p-value: 0.143

	N	median	Dunnet
Flow composite	15	12.56	
Time composite	20	18.19	=

Total Lead µg/L — p-value: 0.279

	n	median	Dunnet
Flow composite	10	9.75	
Time composite	19	6.08	=

Total Zinc µg/L — p-value: 0.163

	n	median	Dunnet
Flow composite	15	108.9	
Time composite	20	161.4	=

Note. Dunnet test compared if sites with control produces larger concentrations ">", smaller concentrations "<" or not statistically difference "=" than sites without control at a family error of 5%.
"None" indicates no samples were collected for this constituent in the group.

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Tables 30 through 33 show that no significant differences were observed for

BOD_5 concentrations using either of the compositing schemes for any of the four

categories.

A similar result was observed for COD except for commercial land uses in EPA Rain

Zone 2, where not enough samples were collected to detect a significant difference. TSS

and total lead median concentrations in EPA Rain Zone 2 were two to five times higher

in concentration when time-compositing was used instead of flow-compositing.

Nutrients in EPA Rain Zone 2 collected in residential, commercial and industrial

areas showed no significant differences using either compositing method. The only

exceptions were for ammonia in residential and commercial land use areas and total

phosphorus in residential areas where time-composite samples had higher concentrations.

Metals were higher when time-compositing was used in residential and commercial land

use areas. No differences were observed in industrial land use areas, except for lead.

Figure 38 shows box and whiskers plots for TSS using both methods.
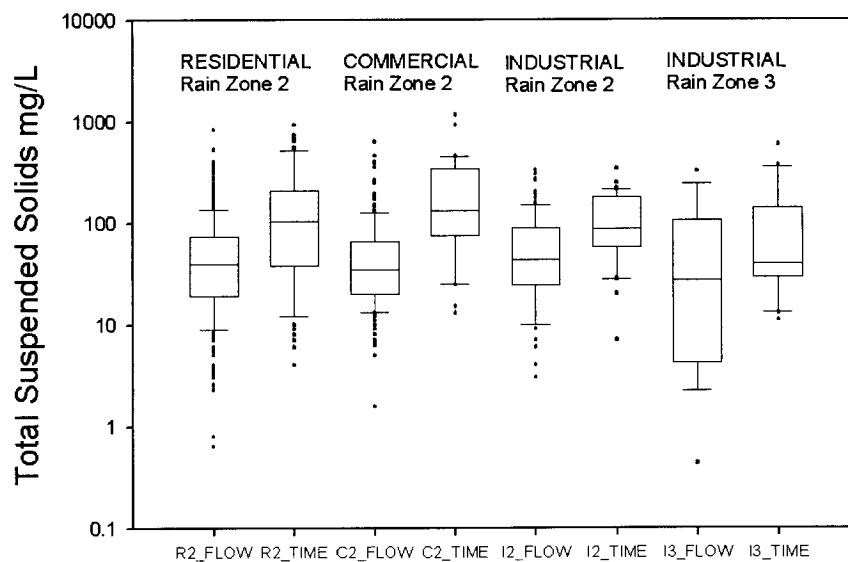
Fig. 38. Comparisons between time- and flow-composite options for TSS

5.6 Sampling Period During Runoff Event and Selection of Events to Sample

Another potential factor that may affect stormwater quality is the sampling period during the runoff event. Automatic samplers can initiate sampling very close to the beginning of flow, while manual sampling usually requires travel time and other delays before sampling can be started. It is also possible for automatic samplers to represent the complete storm, if of very long duration, as long as proper sampler setup programming is performed (Burton and Pitt 2001). However, automatic samplers are not capable of sampling bed load material, and are less effective in sampling larger particles (>500 μm) than typically suspended solids. Manual sampling, if able to collect a sample from a cascading flow, can collect from the complete particle size distribution.

The NPDES stormwater sampling protocols only required collecting composite samples over the first three hours of the event instead of during the whole event. Truncating the sampling before the runoff event ended may have adversely affected the measured stormwater quality.

Selecting a small subset of the annual events can also bias the monitoring results. In most stormwater research projects, the goal is to sample and analyze as many events as possible during the monitoring period. As a minimum, about 30 samples are usually desired in order to adequately determine the stormwater characteristics with an error level of about 25% (assuming 95% confidence and 80% power) (Burton and Pitt 2001). With only three events per year required per land use for the NPDES stormwater permits, the accuracy of the calculated EMC is questionable until many years have passed. Also, the three storms need to be randomly selected from the complete set of rains in order to be most statistically representative.

Flagstaff Street, in Prince George MD, had the most events collected for any site in the NSQD. They collected 28 events during two years of sampling (1998 and 1999). A statistical test was made choosing 6 events (three for each year) from this set, creating 5,600 different possibilities. Figure 39 shows the histogram of these possibilities. The median TSS of the 28 events was 170 mg/L, with a 95% confidence interval between 119 and 232 mg/L. Only 60% of the 5,600 possibilities were inside this confidence interval. Almost half (40%) of the possibilities for the observed EMC would therefore be outside the 95% confidence interval for the true median concentration if only three events were available for two years. As the number of samples increase, there will be a reduction in the bias of the EMC estimates. In Southern California, Leecaster (2002) determined that ten years of collecting three samples per year was required in order to reduce the error to 10% (Leecaster 2002).
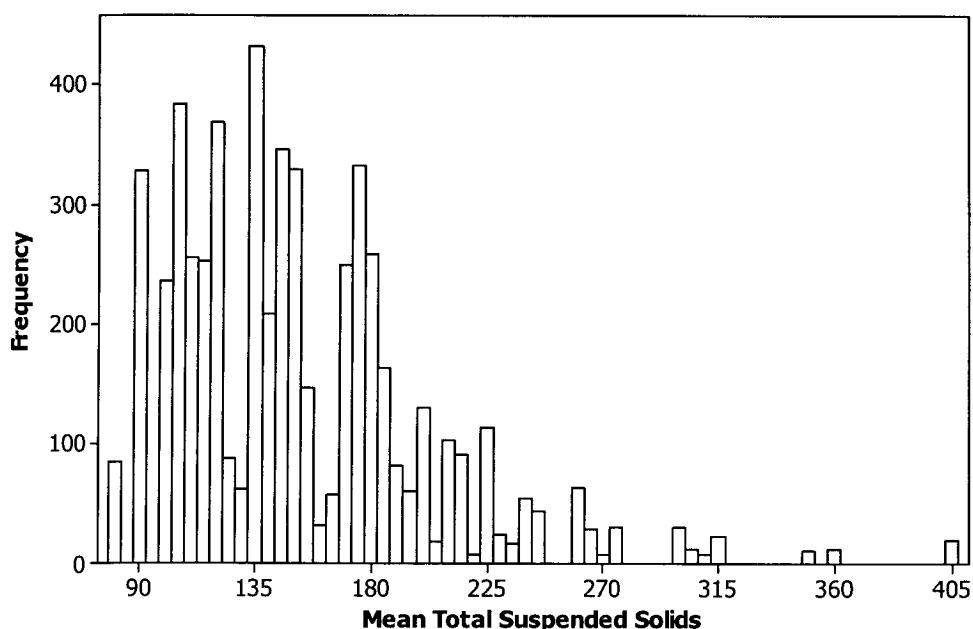
Fig. 39. Histogram of possible TSS concentrations in Flagstaff Street based on collecting three samples per year for two years (the measured median TSS concentration was 170 mg/L)

5.7 Type of Conveyance

Almost all of the samples in the NSQD were collected using automatic samplers and flow compositing. Statistical tests investigating the effects of the type of conveyance only used information from flow-weighted composite samples to reduce potential errors associated with other sampling schemes, as discussed above. Grass swales are considered to be effective stormwater controls compared to conventional curb and gutter stormwater collection systems. Grass swales are commonly found in residential areas with low levels of imperviousness, especially in low density residential areas. NSQD data from residential and mixed residential sites in Virginia, Georgia, and Texas were used to compare stormwater concentrations in areas drained by grass swales and by concrete curbs and gutters.

Historical swale performance tests usually focused on pollutant mass discharges and not concentrations. Swales normally infiltrate significant amounts of the flowing water, resulting in large mass discharge decreases. Most swales operate with relatively deep water, and any "filtering" benefits of the grass (and hence concentration reductions) are usually minimal. Very shallow flows in swales do have particulate pollutant concentration reductions, but these are rarely observed during moderate to large flows (Nara and Pitt 2005).

One-way ANOVA analyses were used to identify any significant differences in stormwater pollutant concentrations between watersheds drained with grass swales or with curbs and gutters. Dunnett's test was used to determine if grass swales produced different concentrations than curbs and gutters. The results are shown in Tables 34 through 37.

Table 34. One-Way ANOVA Results by Type of Conveyance in Residential Land Use, Rain Zone 2

Hardness mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

Oil and Grease mg/L

p-value	0.824		
	n	median	Dunnet
Curb and Gutter	59	3.11	
Grass Swale	7	2.95	=

TDS mg/L

p-value	none		
	n	Median	Dunnet
Curb and Gutter			
Grass Swale			

TSS mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

BOD mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

COD mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

Ammonia mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

$NO_2 + NO_3$ mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

TKN mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

Dissolved Phosphorus mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

Total Phosphorus mg/L

p-value	none		
	n	median	Dunnet
Curb and Gutter			
Grass Swale			

Total Copper µg/L

p-value	*0*		
	n	median	Dunnet
Curb and Gutter	82	10.67	
Grass Swale	7	3.11	<

Total Lead µg/L

p-value	0.112		
	n	median	Dunnet
Curb and Gutter	77	11.7	
Grass Swale	7	5.67	=

Total Zinc µg/L

p-value	*0.002*		
	n	median	Dunnet
Curb and Gutter	82	59.46	
Grass Swale	7	17.85	<

Table 35. One-Way ANOVA Results by Type of Conveyance in Industrial Land Use, Rain Zone 2

Hardness mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

Oil and Grease mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

TSS mg/L

	n	median	Dunnet
p-value		0.023	
Curb and Gutter	69	37.52	
Grass Swale	7	97.70	>

TDS mg/L

	n	median	Dunnet
p-value		0	
Curb and Gutter	67	45.5	
Grass Swale	77	184	>

BOD mg/L

	n	median	Dunnet
p-value		0	
Curb and Gutter	67	6.84	
Grass Swale	5	39.98	>

COD mg/L

	n	median	Dunnet
p-value		0.035	
Curb and Gutter	66	50.16	
Grass Swale	7	85.64	>

Ammonia mg/L

	n	median	Dunnet
p-value		0.492	
Curb and Gutter	61	0.223	
Grass Swale	7	0.285	=

NO$_2$ + NO$_3$ mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

TKN mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

Dissolved Phosphorus mg/L

	n	median	Dunnet
p-value		0.012	
Curb and Gutter	50	0.07	
Grass Swale	4	0.23	>

Total Phosphorus mg/L

	n	median	Dunnet
p-value		0.468	
Curb and Gutter	64	0.174	
Grass Swale	7	0.232	=

Total Lead µg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

Total Copper µg/L

	n	median	Dunnet
p-value		0.905	
Curb and Gutter	20	13.0	
Grass Swale	7	12.36	=

Total Zinc µg/L

	n	median	Dunnet
p-value		0.447	
Curb and Gutter	20	225.7	
Grass Swale	7	188.4	=

156

Table 36. One-Way ANOVA Results by Type of Conveyance in Residential Land Uses, Rain Zone 3

Hardness mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

Oil and Grease mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

TDS mg/L

	n	median	Dunnet
p-value		*0.049*	
Curb and Gutter	11	94.06	
Grass Swale	6	47.84	<

TSS mg/L

	n	median	Dunnet
p-value		0.425	
Curb and Gutter	12	19.2	
Grass Swale	6	29.6	=

BOD mg/L

	n	median	Dunnet
p-value		0.749	
Curb and Gutter	11	7.56	
Grass Swale	5	6.63	=

COD mg/L

	n	median	Dunnet
p-value		*0.027*	
Curb and Gutter	11	29.36	
Grass Swale	5	67.27	>

Ammonia mg/L

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

NO₂ + NO₃ mg/L ($NO_2 + NO_3$ mg/L)

	n	median	Dunnet
p-value		none	
Curb and Gutter			
Grass Swale			

TKN mg/L

	n	median	Dunnet
p-value		0.17	
Curb and Gutter	11	1.22	
Grass Swale	6	0.94	<

Dissolved Phosphorus mg/L

	n	median	Dunnet
p-value		0.324	
Curb and Gutter	8	0.07	
Grass Swale	6	0.04	=

Total Phosphorus mg/L

	n	median	Dunnet
p-value		0.319	
Curb and Gutter	12	0.22	
Grass Swale	6	0.14	=

Total Copper µg/L

	n	median	Dunnet
p-value		0.007	
Curb and Gutter	11	19	
Grass Swale	6	5	<

Total Lead µg/L

	n	median	Dunnet
p-value		0.154	
Curb and Gutter	9	12.9	
Grass Swale	6	4.20	=

Total Zinc µg/L

	n	median	Dunnet
p-value		0.781	
Curb and Gutter	11	49.5	
Grass Swale	6	43.0	=

Table 37. One-Way ANOVA Results by Type of Conveyance in Industrial Land Use, Rain Zone 3

Hardness mg/L

	n	median	Dunnet
p-value	none		
Curb and Gutter			
Grass Swale			

Oil and Grease mg/L

	n	median	Dunnet
p-value	none		
Curb and Gutter			
Grass Swale			

TSS mg/L

	n	median	Dunnet
p-value	***0.014***		
Curb and Gutter	10	9.68	
Grass Swale	6	91.2	>

BOD mg/L

	n	median	Dunnet
p-value	0.461		
Curb and Gutter	10	4.68	
Grass Swale	6	6.61	=

COD mg/L

	n	median	Dunnet
p-value	0.446		
Curb and Gutter	10	29.40	
Grass Swale	6	41.26	=

TDS mg/L

	n	median	Dunnet
p-value	0.134		
Curb and Gutter	10	76.74	
Grass Swale	6	131.6	=

NO$_2$ + NO$_3$ mg/L

	n	median	Dunnet
p-value	none		
Curb and Gutter			
Grass Swale			

TKN mg/L

	n	median	Dunnet
p-value	0.299		
Curb and Gutter	9	0.515	
Grass Swale	6	0.885	=

Dissolved Phosphorus mg/L

	n	median	Dunnet
p-value	0.077		
Curb and Gutter	5	0.046	
Grass Swale	6	0.027	=

Total Phosphorus mg/L

	N	median	Dunnet
p-value	0.460		
Curb and Gutter	9	0.138	
Grass Swale	6	0.202	=

Ammonia mg/L

	n	median	Dunnet
p-value	none		
Curb and Gutter			
Grass Swale			

Total Lead µg/L

	n	median	Dunnet
p-value	0.157		
Curb and Gutter	4	4.86	
Grass Swale	6	15.5	=

Total Zinc µg/L

	n	median	Dunnet
p-value	***0.007***		
Curb and Gutter	9	72.86	
Grass Swale	6	198.9	>

Total Copper µg/L

	N	median	Dunnet
p-value	0.098		
Curb and Gutter	9	8.57	
Grass Swale	6	22.32	=

Note. Dunnet test compared if sites with control produces larger concentrations ">", smaller concentrations "<" or not statistically difference "=" than sites without control at a family error of 5%. "None" indicates no samples were collected for this constituent in the group.

Total lead and total phosphorus did not have any significant differences in concentrations when comparing the two conveyance systems in both land use areas. Total copper concentrations from residential land uses in EPA Rain Zones 2 and 3 were lower when grass swale was used instead of curbs and gutters. No copper concentrations differences were observed at industrial land uses having different conveyance systems.

Figure 40 shows box and whiskers plots for TSS in industrial land uses, EPA Rain Zones 2 and 3 and residential areas in EPA Rain Zone 2. The median concentrations in industrial land uses were smaller in locations where curbs and gutters were used compared to sites having grass swales. The statistical tests did not identify a significant difference between the median concentrations in residential areas in EPA Rain Zone 3 (the residential boxes have much more overlap than for the industrial sites).

5.8 Concentration Effects Associated with Varying Amounts of Impervious Cover

The reported values for imperviousness do not reflect the amount of pavement and roofs that are not directly connected to the drainage system. Directly connected impervious areas (DCIA) are also referred to as effective impervious areas (EIA). For example, imagine a park with a single paved basketball court surrounded by turf; the area of the court will be counted as part of the total impervious area, but would not be considered as part of the effective impervious area. The runoff from the paved court would likely be totally infiltrated by the grass and will be not discharged to the drainage system. In this case, even if we have a value for "total imperviousness," the "effective percentage of imperviousness" is zero.
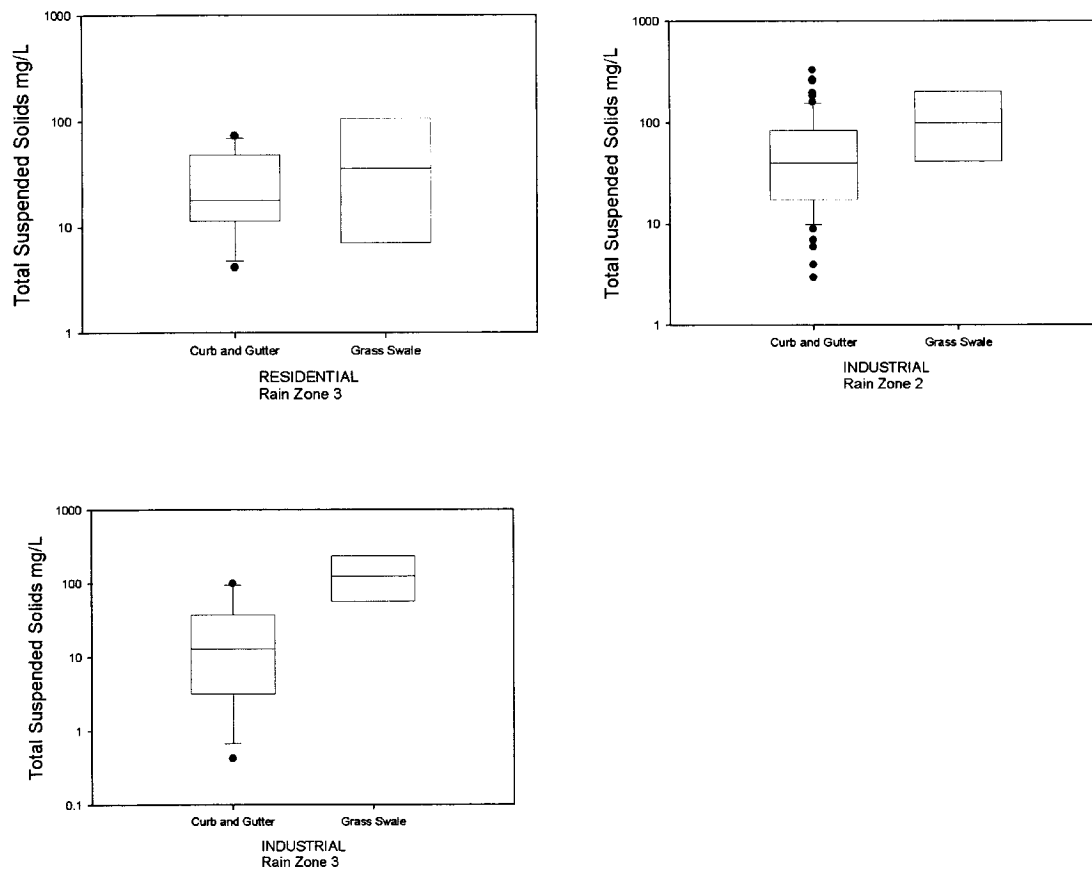
Fig. 40. TSS concentration by type of conveyance (Significant differences were observed in industrial land uses)

It is therefore difficult to compare database concentrations with the

imperviousness values due to these potential uncertainties in the actual effective

imperviousness. Figure 41 is an example plot of the percent imperviousness values of

different land uses for COD. Each vertical set of observations represent a single

monitoring location (all of the events at a single location have the same percent

imperviousness). The variation of COD at any one monitoring location is seen to vary

greatly, typically by about an order of magnitude. These large variations will make trends

difficult to identify. All of the lowest percentage imperviousness sites are open space land

uses, while all of the highest percentage imperiousness sites are freeway and commercial

land uses. This plot shows no apparent trend in concentration that can be explained by

imperviousness. However, it is very likely that a significant and important trend does

exist between percent effective imperviousness and pollutant mass that is discharged.

While the relationship between imperviousness and concentration is not clear, the

relationship between effective imperviousness and total runoff volume is much clearer

and more obvious as the non-paved areas can infiltrate much water.
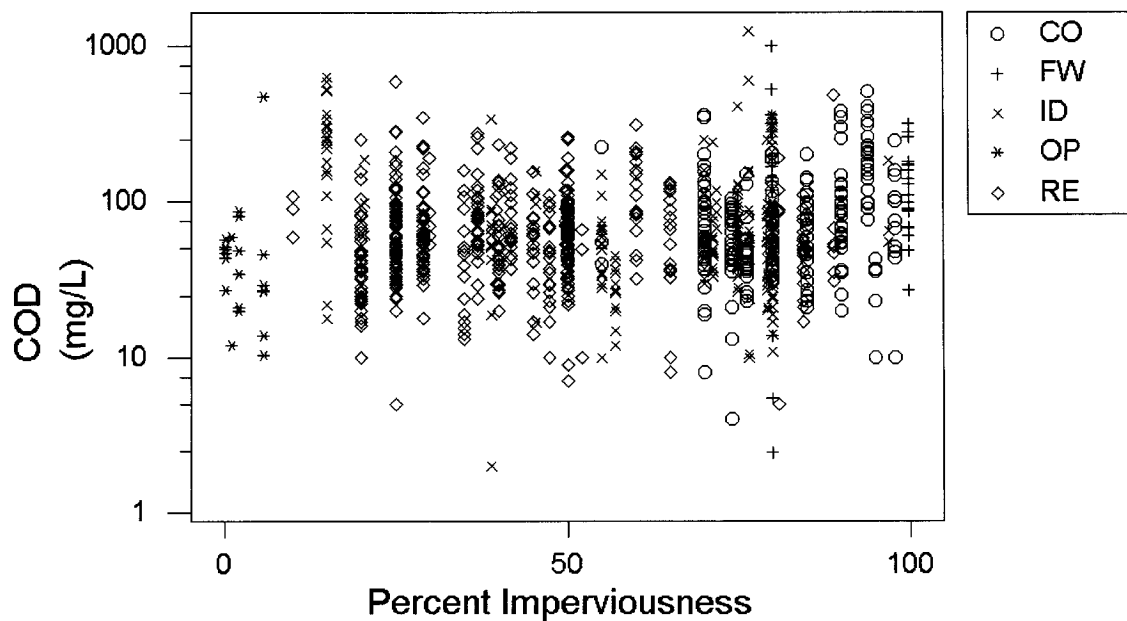
Fig. 41. Plot of COD concentrations against watershed area percent imperviousness values for different land uses (CO: commercial; FW: freeway; ID: industrial; OP: open space; and RE: residential)

One important feature in the percentage of imperviousness is that most of the

residential sites have low levels of imperviousness, while commercial and industrial sites

usually have high percentages of imperviousness. Figure 42 shows the mean TSS

concentration for residential, commercial and industrial land uses in the database. Only

four of the residential watershed has percentage of imperviousness values larger than

60%. Two commercial sites have less than 60% imperviousness, with the remaining

commercial sites above this value. Analyses concerning the effects of impervious cover

on stormwater concentrations for each land use separately are difficult as there are

limited ranges of impervious cover within each land use category.
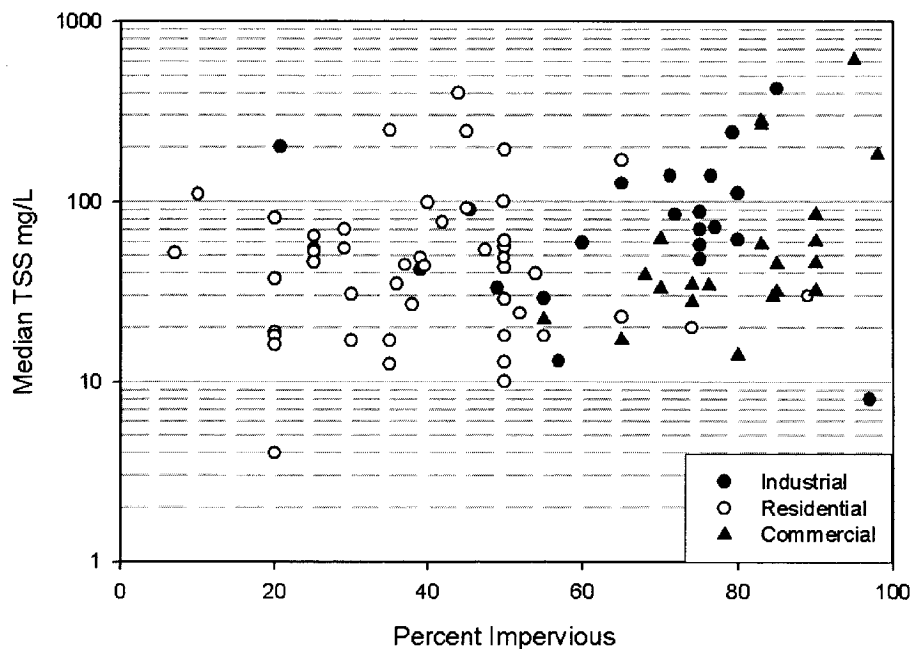
Fig. 42. TSS concentrations by impervious cover and single land use

Regression analyses were used to identify possible relationships between

constituent concentrations and the percentage of imperviousness for residential land use

data. Table 38 shows the results from these regression analyses. Residential land uses in

EPA Rain Zone 2 were examined during these analyses. Median concentrations from

sites using automatic, flow-weighted samplers, and not having any controls and with curb

and gutter conveyance systems were selected for analyses. Data from the site

KYLOTSR3 was not used during these analyses because sewage disposal facilities were

located in the test watershed. Solids and heavy metal median concentrations were higher

at this location than for the remaining residential sites in the same Rain Zone.

Only nitrate-nitrite indicated a significant regression relationship between

percentage of imperviousness and constituent concentration for these sites, as shown in

Figure 43. In this case, the slope was negative, indicating a reduction in the concentration

as the level of imperviousness increased. One possible explanation is that the nutrients

are associated with landscaped areas and the use of fertilizers which all decrease with

increasing impervious areas. This does not indicate that the total mass of nitrate-nitrite

will be reduced. The load of this constituent depends on the total runoff volume that is

discharged during the event. As the percentage of imperviousness increases, the runoff

volume also increases due to lack of infiltration. Even if the concentration is shown to

decrease, the total mass discharge may still increase with increasing amounts of

pavement or roofs. There was not enough evidence to indicate a relationship between

concentration and percentage of imperviousness for the other 11 constituents examined.

Table 38. Regression of Median Concentrations by Percentage of Impervious in Residential land Use, EPA Rain Zone 2

Constituent	n	Constant		Impervious		R^2 adjusted	Result
		Coefficient	p-value	Coefficient	p-value		
TDS mg/L	10	71.94	0.002	-0.386	0.446	0	Not significant
TSS mg/L	10	74.44	0.002	-0.715	0.172	0.121	Not significant
BOD_5 mg/L	10	8.74	0.117	0.076	0.619	0	Not significant
COD mg/L	10	53.94	0.027	0.332	0.578	0	Not significant
Ammonia mg/L	10	0.319	0.052	-0.002	0.639	0	Not significant
NO_3-NO_2 mg/L	9	0.756	0	-0.009	*0.013*	0.556	Significant
TKN mg/L	9	1.817	0.003	-0.016	0.247	0.069	Not significant
DP mg/L	10	0.237	0.033	-0.003	0.349	0	Not significant
TP mg/L	10	0.561	0.002	-0.006	0.13	0.171	Not significant
Cu µg/L	11	16.51	0.005	-0.140	0.225	0.065	Not significant
Pb µg/L	11	46.64	0.336	-0.337	0.767	0	Not significant
Zn µg/L	11	98.13	0.027	-0.572	0.542	0	Not significant
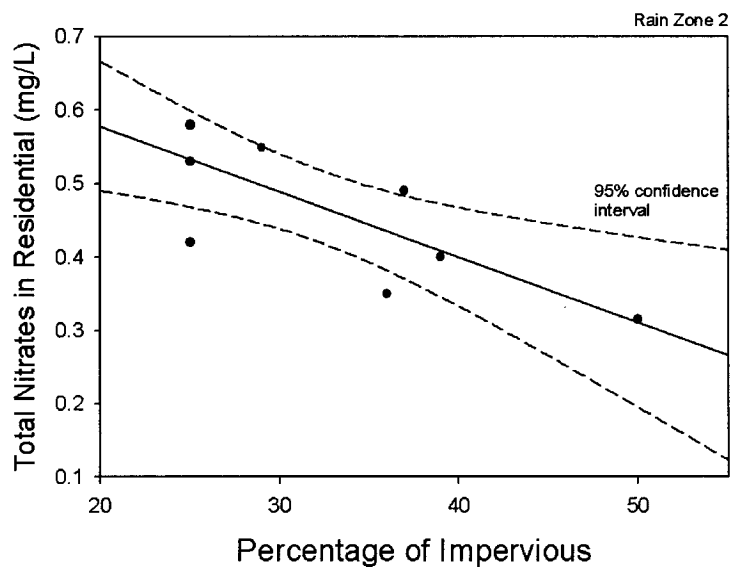
Fig. 43. Total nitrates regression at different percentages of impervious

The same regression analysis was performed for commercial and industrial land uses in EPA Rain Zone 2. The results of the regression analyses are shown in Table 39.

Table 39. Regression of Median Concentrations by Percentage of Impervious in
Commercial and Industrial land use, EPA Rain Zone 2

Constituent	n	Constant		Impervious		R^2 adjusted	Result
		Coefficient	p-value	Coefficient	p-value		
TDS mg/L	5	-4.80	0.854	0.821	0.103	0.523	Not significant
TSS mg/L	5	-22.01	0.406	0.805	0.097	0.541	Not significant
BOD_5 mg/L	5	-1.80	0.879	0.153	0.410	0	Not significant
COD mg/L	5	1.41	0.968	0.748	0.215	0.268	Not significant
Ammonia mg/L	5	-0.05	0.906	0.005	0.439	0	Not significant
NO_3-NO_2 mg/L	5	0.01	0.985	0.007	0.438	0	Not significant
TKN mg/L	5	-0.84	0.467	0.030	0.140	0.426	Not significant
DP mg/L	5	-0.02	0.858	0.001	0.516	0	Not significant
TP mg/L	5	-0.10	0.649	0.004	0.271	0.168	Not significant
Cu μg/L	5	4.26	0.759	0.089	0.679	0	Not significant
Pb μg/L	6	15.69	0585	-0.021	0.961	0	Not significant
Zn μg/L	6	247.9	0.269	-0.949	0.765	0	Not significant

None of the median stormwater constituents in commercial and industrial areas
seem to be affected by changes in impervious cover. There is not enough evidence to
indicate a significant relationship between constituent concentration and percentage of
imperviousness. More samples will be required to identify those regressions.

5.9 Seasonal Effects on Stormwater Quality

Another factor that may affect stormwater quality is the season when the sample
was obtained. If the few samples collected for a single site were all collected in the same
season, the results may not be representative of the whole year. The NPDES sampling
protocols were designed to minimize this effect by requiring the three samples per year to
be separated by at least 1 month. The few samples still could be collected within a single
season, but at least not within the same week. Seasonal variations for residential
stormwater data are shown in Figure 44. These variations are not as obvious as the land
use or geographical variations, except for bacteria which appear to be lowest during the

winter season and highest during the summer and fall (a similar conclusion was obtained

during the NURP, EPA 1983, data evaluations). The database does not contain any

snowmelt data, so all of the data corresponds to rain-related runoff only.
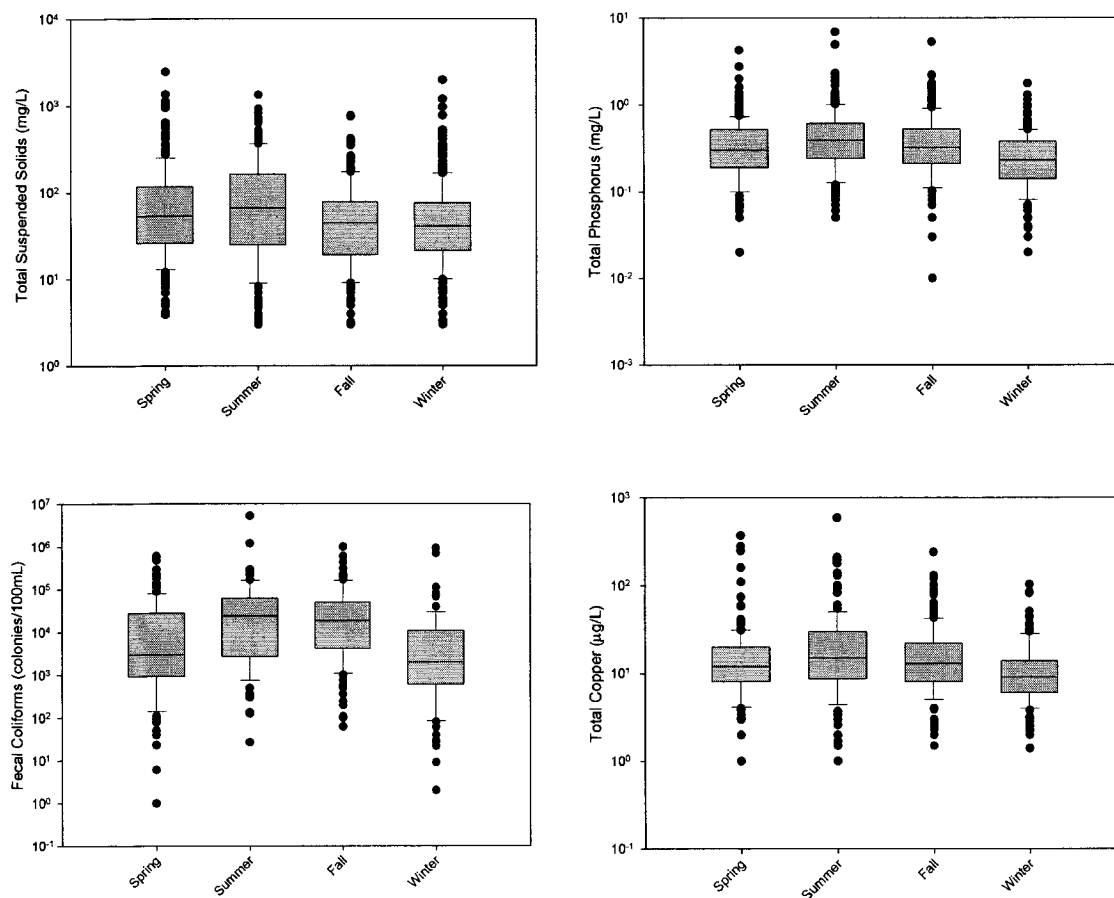
Fig. 44. Example residential area stormwater pollutant concentrations sorted by season

5.10   Precipitation Effects on Stormwater Quality

A common assumption is that higher runoff concentrations are associated with

smaller rain events. While this has been shown to be true during controlled washoff

studies (Pitt 1987), or for sheetflows taken from relatively small paved areas during rains

(see Chapter 6 discussion about first flush observations), this has not been frequently

detected for samples collected at outfalls for areas having a mixture of surfaces and for

typical random periods of high rain intensities. Figure 45 contains several scatter plots

showing concentrations plotted against rain depth. There are no obvious trends of

concentrations associated with rain depth for the NSQD data.
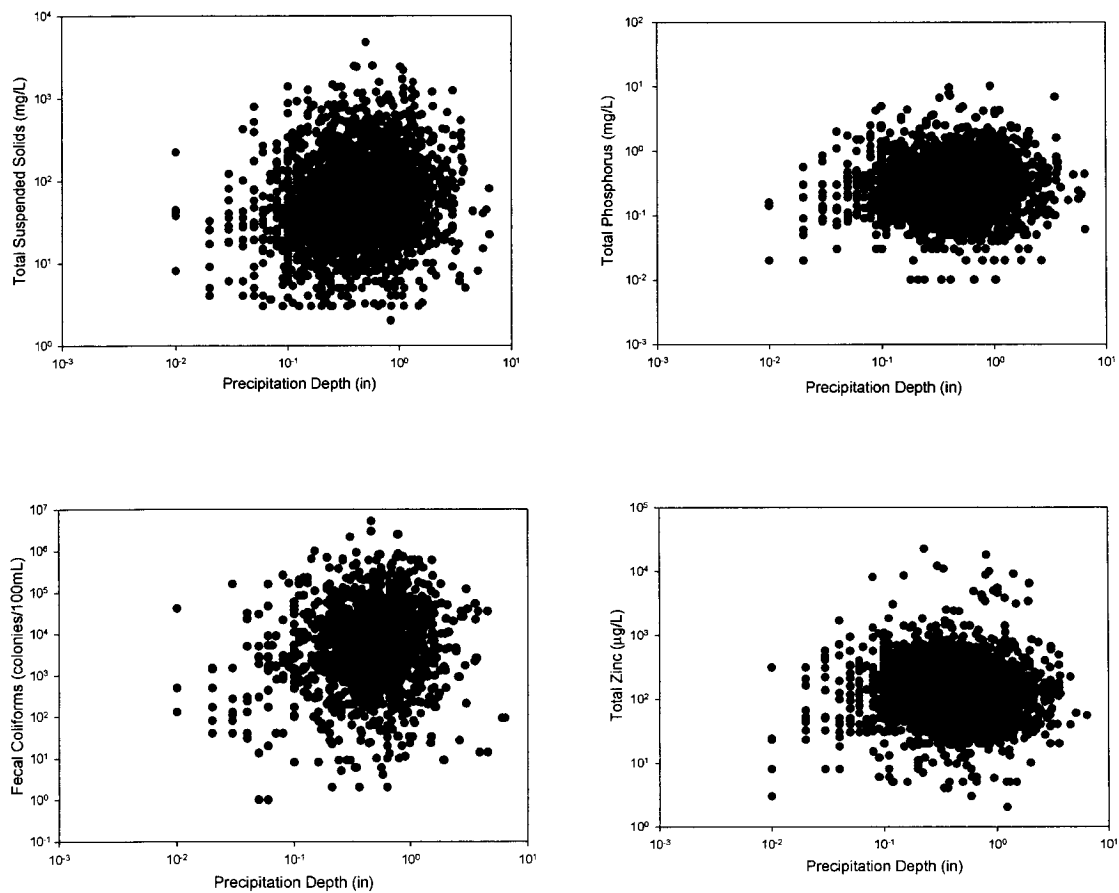
Fig. 45. Examples of scatter plots by precipitation depth

Figure 46 shows scatter plots of rainfall and runoff depth for each land use. These

should follow a 45 degree line for areas having very large amounts of directly connected

impervious areas.

Scatterplot of Runoff vs Precipitation Depth

Panel variable: Landuse
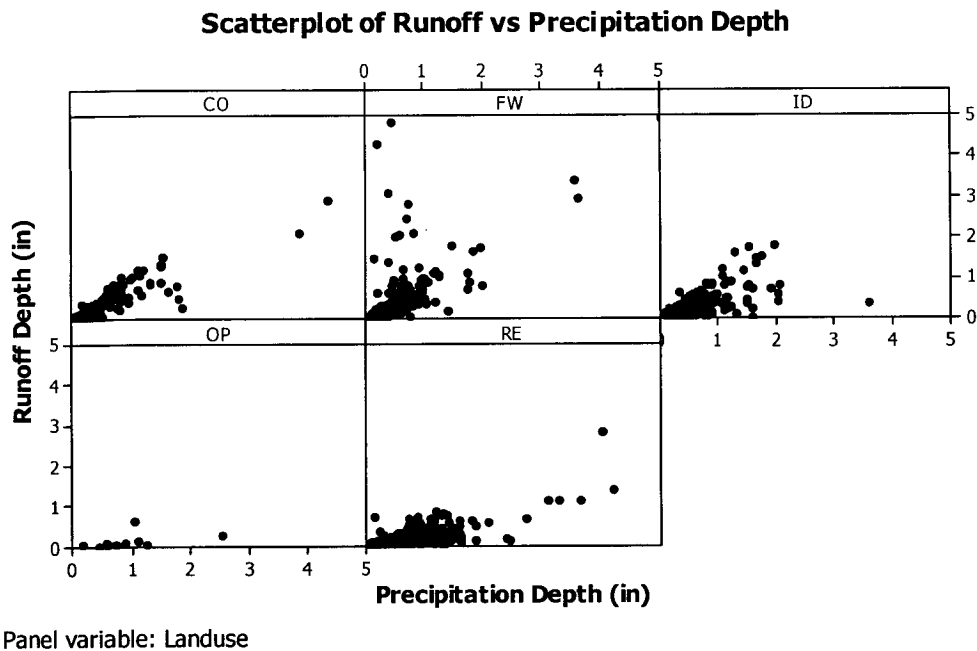
Fig. 46. Precipitation depth and runoff depth plotted by land use

These plots show much greater scatter than expected. Some of the plots even indicated larger amounts of runoff than precipitation. This may have occurred due to several reasons: (1) the rainfall was not representative of the drainage area being monitored (especially possible for those sites that relied on off-site rain data); (2) the runoff monitoring was inaccurate (possible when the runoff monitoring relied on stage recording devices and the Manning's equation was applied without local calibration); (3) the drainage area was inaccurately delineated; or (4) when base flows contributed significant amounts of runoff during the event. When reviewing the runoff plots provided in some of the annual reports, significant base flows were observed. It was also apparent that these base flows were not subtracted from the total flows recorded during the rain event. The magnitude of the error would be greater for smaller rain events when the base flows could be much larger than the direct runoff quantity. Base flows commonly occur

when a local spring or high groundwater levels enter the storm drainage system. In addition, runoff may still be occurring from a prior large event that ended soon before the current event started (the 3 day antecedent dry period requirement for monitored events was intended to minimize this last cause of base flows).

5.11   Days without Rain

The EPA Rain Zones with the longest reported dry interevent periods having data in the NSQD are zones 6 (southern California) and 7 (Oregon). In these EPA Rain Zones, some antecedent dry periods were reported to be longer than 100 days. Monitored events with the shortest interevent periods of no rains were monitored along the east and south east coasts of the country (EPA Rain Zones 2 and 3). The mean interevent dry period in the western states was about 18 days, while eastern states had mean interevent dry periods of about 5 days. Figure 47 shows box and whisker plots of the number of days having no rain before the monitored event by each EPA Rain Zone.

Samples collected using automatic flow-weighted samplers from watersheds having curbs and gutters and without stormwater controls were used during the following analyses. Only EPA Rain Zone 2 has enough observations to evaluate possible effects of the antecedent dry period on the concentration of stormwater pollutants. Table 40 shows the results from the regression analyses. In residential land uses, 7 out of 12 constituents indicated that antecedent dry period has a significant effect on the median concentrations. All the regression slope coefficients were positive, indicating that as the number of days having no rain increased the concentrations also increased.
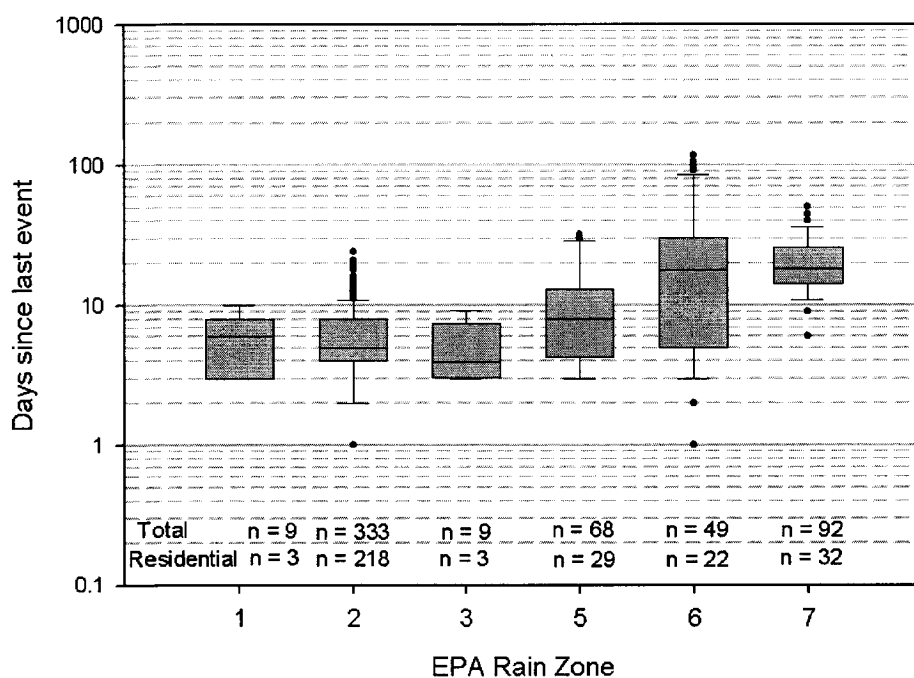
Fig. 47. Box and whisker plot of days since preceding event by Rain Zone

Table 40. Regression of Logarithm of Constituent Concentrations by Logarithm of
Antecedent Dry Period for Residential Land Use, EPA Rain Zone 2

Constituent	n	Constant		Days since last event		R^2 adjusted	Significant
		Coefficient	p-value	Coefficient	p-value		
Oil - Grease mg/L	35	0.737	0	-0.364	0.062	0.074	No
TDS mg/L	208	1.761	0	0.094	0.120	0.007	No
TSS mg/L	214	1.524	0	0.116	0.254	0.001	No
BOD_5 mg/L	211	0.887	0	0.211	*0.004*	0.035	Yes
COD mg/L	206	1.682	0	0.151	*0.032*	0.018	Yes
Ammonia mg/L	204	-0.826	0	0.300	*0.003*	0.039	Yes
NO_3-NO_2 mg/L	208	-0.428	0	0.160	*0.014*	0.024	Yes
TKN mg/L	208	-0.066	0.193	0.232	*0.001*	0.049	Yes
DP mg/L	203	-1.061	0	0.282	*0.002*	0.043	Yes
TP mg/L	214	-0.629	0	0.183	*0.005*	0.031	Yes
Cu µg/L	58	1.082	0	0.025	0.830	0	No
Pb µg/L	53	1.305	0	-0.311	0.277	0.004	No
Zn µg/L	58	1.872	0	-0.058	0.764	0	No

All nutrients in residential land uses showed a positive correlation between days since last event and constituent concentration. In all cases, the coefficients of determination (R^2) were smaller than 0.05, indicating that relatively little of the total variation was explained by percent imperviousness. Solids and metals were not affected by the antecedent dry period. Figure 48 shows the regression lines and 95% confidence intervals for four nutrients in residential land uses.
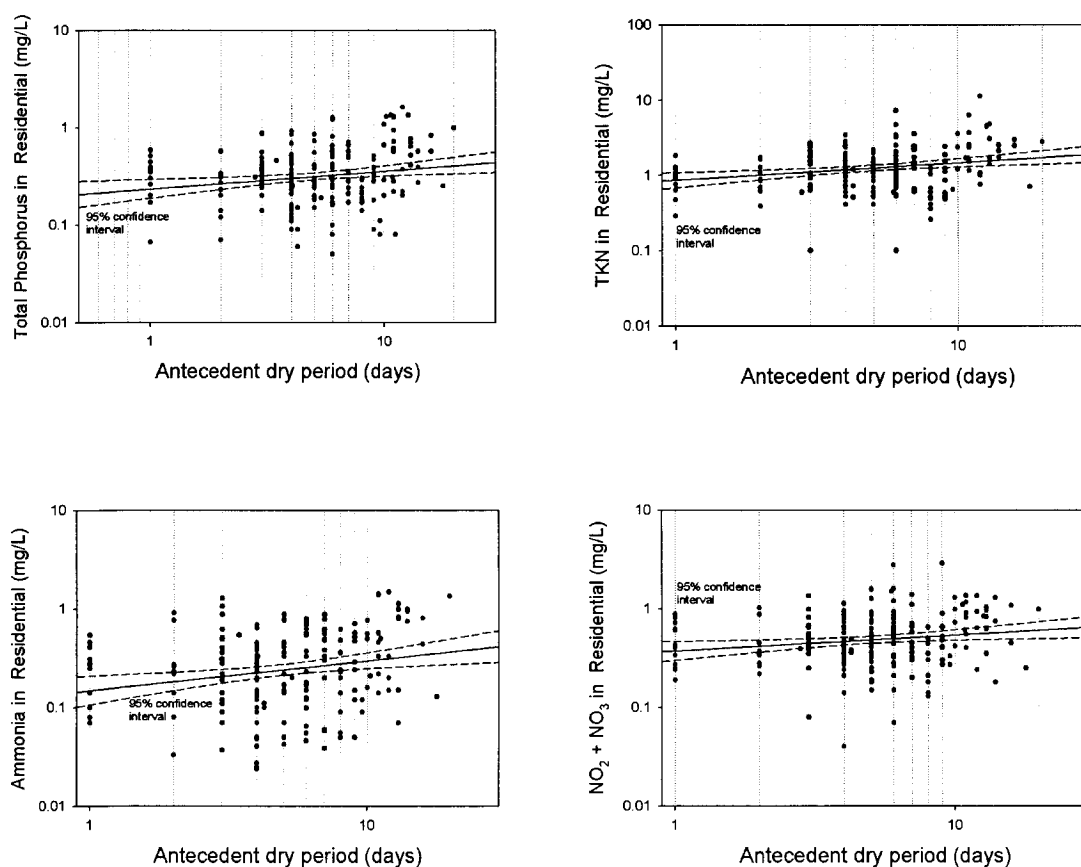
Fig. 48. Nutrient concentrations affected by dry periods since last rain in residential land use

Table 41 shows the results from the regression analyses in commercial land uses. Except for nitrates, all the nutrients have positive regressions inside the 95% confidence

interval. In commercial land uses, the effects of antecedent dry periods on the median

concentrations were less important. Only total phosphorus and total lead had significant

regression results. As in the residential case, phosphorus has a positive coefficient with a

small coefficient of determination. However, lead decreases with the number of dry days

before the storm.

Table 41. Regression of Logarithm of Constituent Concentrations by Logarithm of
Antecedent Dry Period for Commercial Land Use, EPA Rain Zone 2

| Constituent | n | Constant | | Impervious | | R^2 adjusted | Significant |
		Coefficient	p-value	Coefficient	p-value		
Oil - Grease mg/L	25	0.783	0.001	-0.202	0.402	0	No
TDS mg/L	64	1.715	0	0.215	0.169	0.015	No
TSS mg/L	82	1.506	0	0.018	0.872	0	No
BOD$_5$ mg/L	83	0.971	0	0.149	0.176	0.01	No
COD mg/L	64	1.670	0	0.221	0.093	0.029	No
Ammonia mg/L	64	-0.591	0	0.258	0.175	0.014	No
NO2 mg/L	83	-0.235	0	-0.208	0.176	0.01	No
TKN mg/L	83	-0.006	0.949	0.196	0.109	0.019	No
DP mg/L	61	-1.329	0	0.241	0.160	0.017	No
TP mg/L	83	-0.784	0	0.198	*0.028*	0.047	Yes
Cu μg/L	33	1.081	0	0.959	0.501	0	No
Pb μg/L	33	1.498	0	-1.02	*0.001*	0.261	Yes
Zn μg/L	32	2.21	0	-0.082	0.527	0	No

Figure 49 shows the regression equations for total phosphorus and total lead for

data from commercial land uses. The 95% confidence interval of the regression line for

total phosphorus can include zero slope lines. This indicates that there is not a strong

correlation between antecedent dry period and total phosphorus concentrations. For total

lead, the reduction in concentrations with increasing dry periods is more obvious, but not

very explicable.
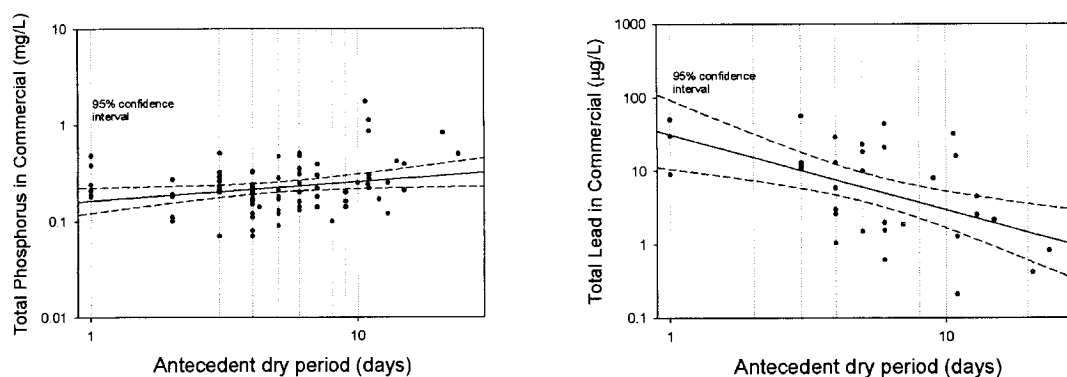
Fig. 49. Total phosphorus and total lead as function of antecedent dry period in commercial land use

The effect of the antecedent dry period on stormwater concentrations at industrial

land uses was not significant, except for TSS, as shown on Table 42. Figure 50 is a plot

of the TSS concentrations increasing with increasing dry periods.

Table 42. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period in Industrial Land Use, EPA Rain Zone 2

Constituent	n	Constant		Impervious		R^2 adjusted	Significant
		Coefficient	p-value	Coefficient	p-value		
Oil - Grease mg/L	3	0.271	0.773	-0.451	0.700	0	No
TDS mg/L	30	1.651	0	-0.009	0.958	0	No
TSS mg/L	31	1.190	0	0.656	*0.025*	0.134	Yes
BOD$_5$ mg/L	32	0.780	0	0.201	0.202	0.022	No
COD mg/L	29	1.685	0	0.071	0.622	0	No
Ammonia mg/L	27	-0.487	0.014	-0.084	0.753	0	No
NO2 mg/L	32	-0.154	0.233	-0.124	0.493	0	No
TKN mg/L	32	-0.151	0.215	0.218	0.207	0.021	No
DP mg/L	28	-1.176	0	0.190	0.406	0	No
TP mg/L	32	-0.966	0	0.373	0.11	0.053	No
Cu μg/L	3	1.109	0.124	0.216	0.565	0	No
Pb μg/L	3	0.882	0.197	0.119	0.787	0	No
Zn μg/L	3	2.072	0.056	0.186	0.555	0	No
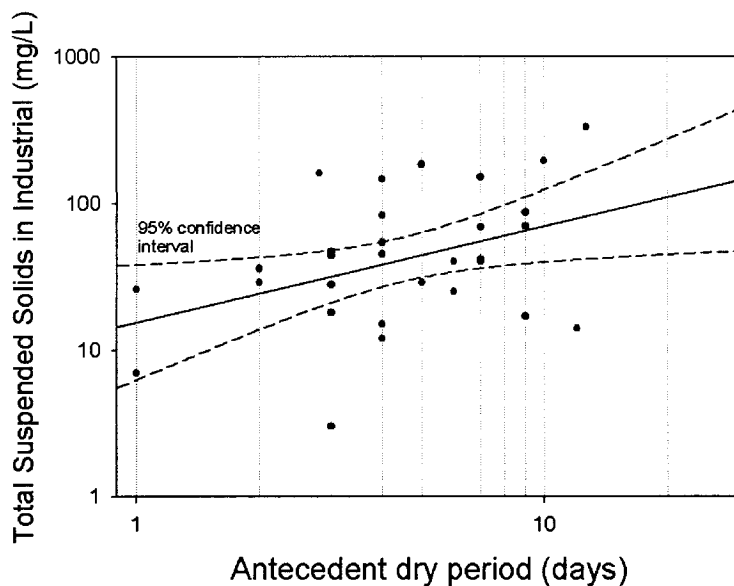
Fig. 50. TSS concentrations for days since preceding event in industrial land use

5.12 Trends in Stormwater Quality with Time

In an effort to recognize why differences were observed between the NURP and

NSQD databases (see Chapter 2), further examinations of two communities that

monitored stormwater during both NURP and the Phase I NPDES program were made.

As part of their MS4 Phase I application, Denver and Milwaukee both returned to some

of their earlier sampled monitoring stations used during the local NURP projects (EPA

1983). In the time between the early 1980s (NURP) and the early 1990s (MS4 permit

applications), they did not detect any significant differences, except for large decreases in

lead concentrations. Figure 51 compares suspended solids, copper, lead, and zinc

concentrations at the Wood Center NURP monitoring site in Milwaukee. The average site

concentrations remained the same, except for lead, which decreased from about 450 to

about 110μg/L, as expected due to the decrease in leaded gasoline during this period.

EMC Trends at Wood Center
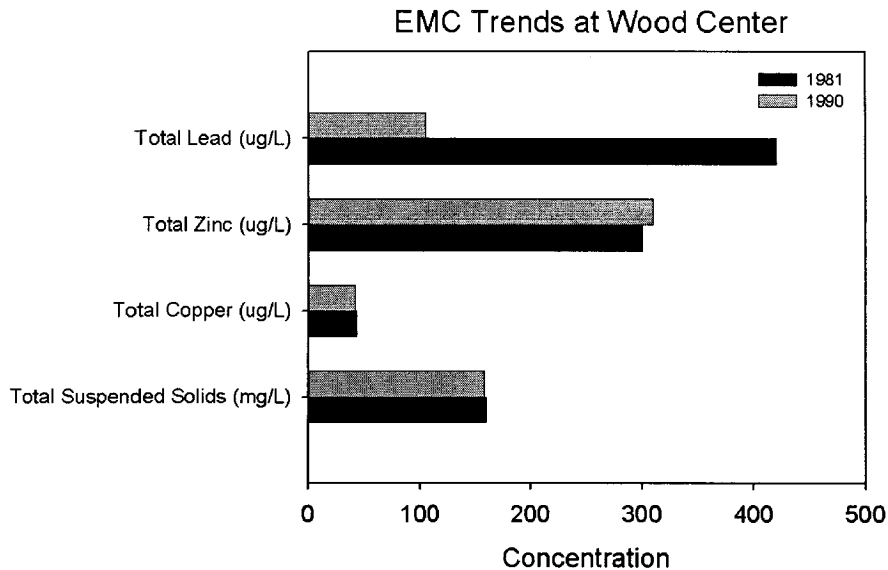
Figure 51. Comparison of pollutant concentrations collected during NURP (1981) to MS4 application data (1990) at the same location (personal communication, Roger Bannerman, WI DNR)

Urban Drainage and Flood Control District performed similar comparisons in the

Denver Metropolitan area. Table 43 compares stormwater quality for commercial and

residential areas for 1980/81 (NURP) and 1992/93 (MS4 application). Although there

was an apparent difference in the averages of the event concentrations between the

sampling dates, they concluded that the differences were all within the normal range of

stormwater quality variations, except for lead, which decreased by about a factor of four.

Trends of stormwater concentrations with time can also be examined using the

NSQD data. A classical example would be for lead, which is expected to decrease over

time with the increased use of unleaded gasoline. Older stormwater samples from the

1970s typically have had lead concentrations of about 100 to 500µg/L, or higher (as

indicated above for Milwaukee and Denver), while most current data indicate

concentrations as low as 1 to 10µg/L.

Table 43. Comparison of Commercial and Residential Stormwater Runoff Quality from
1980/81 to 1992/93 (Doerfer, 1993)

Constituent	Commercial		Residential	
	1980 - 1981	1992 - 1993	1980 - 1981	1992 - 1993
Total suspended solids (mg/L)	251	165	226	325
Total nitrogen (mg/L)	3.0	3.9	3.2	4.7
Nitrate plus nitrite (mg/L)	0.80	1.4	0.61	0.92
Total phosphorus (mg/L)	0.46	0.34	0.61	0.87
Dissolved phosphorus (mg/L)	0.15	0.15	0.22	0.24
Copper, total recoverable (µg/L)	27	81	28	31
Lead, total recoverable (µg/L)	200	59	190	53
Zinc, total recoverable (µg/L)	220	290	180	180

Figure 52 shows a plot of lead concentrations for residential areas only (in EPA

Rain Zone 2), for the time period from 1991 to 2002. This plot shows likely decreasing

lead concentrations with time. Statistically however, the trend line is not significant due

to the large variation in observed concentrations ($p = 0.41$; there is insufficient data to

show that the slope term is significantly different from zero). Likewise the COD

concentrations have an apparent downward trend with time, but again, the slope term is

not significant ($p = 0.12$).
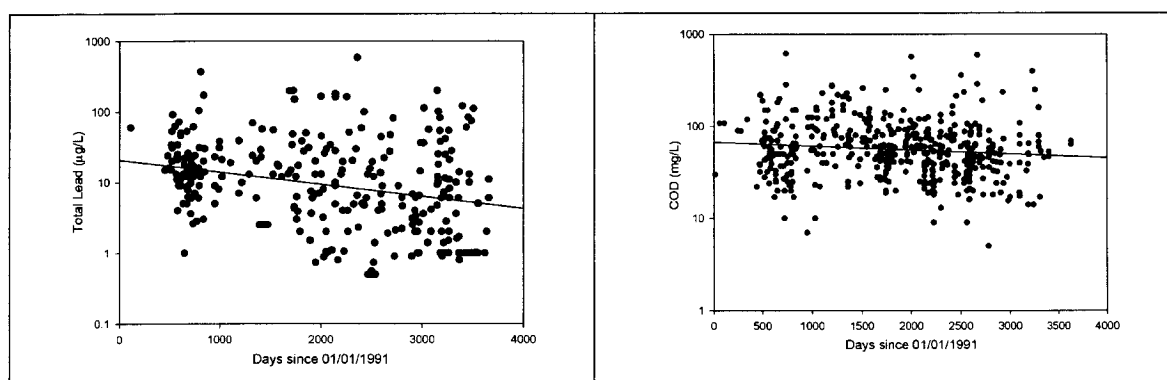
Fig. 52. Residential lead and COD concentrations with time (EPA Rain Zone 2 data only)

Except for lead, it is not likely that time between the data collection efforts is the reason why the NURP and NSQD databases have different values.

5.13   Summary

Several factors were evaluated using data from the NSQD. Only residential, commercial and industrial land uses in EPA Rain Zone 2 and industrial areas in EPA Rain Zone 3 have enough numbers of samples to evaluate factors affecting stormwater concentrations. The effect of each factor cannot be extrapolated to the rest of the country. However they can be used as guidance for communities in other EPA Rain Zones. Additional data from communities that were not included in this first phase of the NSQD database would enable more complete and sensitive analyses. Also, this chapter examined most of these factors in isolation, more as sensitivity analyses and to help identify significant factors. These analyses did not consider factors together and possible interactions.

There is a significant reduction in TSS, nitrite-nitrate, total phosphorus, total copper, and total zinc concentration at sites having wet ponds, the control practice having the largest concentration reductions. No reductions in TKN concentrations were found using wet ponds, but TKN seems to be reduced by dry ponds. Locations with detention storage facilities had smaller reductions of TSS, BOD_5, COD, total lead and total zinc concentrations. Unfortunately, there were few sites in the database having grass swales that could be compared with data from sites having curbs and gutters.

The use of automatic or manual sampling methods is a concern. There were statistical differences found between both methods in residential areas for several

constituents. Most communities calculate their EMC values using flow-composited

sample analyses. If first flush effects are present, manual sampling may likely miss these

more concentrated flows due to delays in arriving at the site to initiate sampling. If the

first flush is for a very short duration, time-composited samples may overly emphasize

these higher flows. Flow compositing produces more accurate EMC values than time

composite analyses. An automatic sampler with flow-weighted samples, in conjunction

with a bed load sampler, is likely the most accurate sampling alternative.

There is a certain amount of redundancy (self-correlation) between land use and

the percentage of impervious areas, as each land use category generally has a defined

narrow range of paved and roof areas. Therefore, it is no possible to test the hypothesis

that different levels of impervious (surface coverage) are more important than differences

in land use (activities within the area). Residential land uses cover only the lower range

of imperviousness, while commercial sites have imperviousness amounts larger than

50%. In order to perform a valid comparison test, the range of imperviousness needs to

be similar for both test cases.

Antecedent dry periods were found to have a significant effect for residential land

uses, at the 6% level of significance, for BOD_5, COD, ammonia, nitrates, TKN,

dissolved, and total phosphorus. As the number of days increased there was an increase in

the concentrations of the stormwater constituents. This relationship was not observed for

freeway sites. This may be associated with the very small drainage areas associated with

the freeway sites (drainage areas close to 1 acre), while the drainage areas for residential,

commercial and industrial areas ranged between 50 and 100 acres (Figure 2).

No seasonal effects on concentrations were observed, except for bacteria levels that appear to be lower in winter and high in summer. No effects on concentration were observed according to precipitation depth. Rainfall energy determines erosion and washoff of particulates, but sufficient runoff volume is needed to carry the particulate pollutants to the outfalls. Different travel times from different locations in the drainage areas results in these materials arriving at different times, plus periods of high rainfall intensity occur randomly throughout the storm. The resulting outfall stormwater concentration patterns for a large area having various surfaces is therefore complex and rain depth is just one of the factors involved.

Chapter 6

Nonparametric Statistical Tests Comparing First Flush and Composite Samples from the National Stormwater Quality Database

6.1 Introduction

The NPDES MS4 stormwater permit program was developed by EPA in response to the United States Congress to protect United States' receiving waters from contaminated stormwater discharges. In 1972, the Congress passed the Federal Water Pollution Control Act (CWA) expanding the federal role of water pollution control. Some of the effects of the CWA were to increase the federal funding for construction of publicly owned wastewater treatment works (POTW), and to develop community participation and a permit for each point discharge, among other activities. The National Pollutant Discharge Elimination System (NPDES) established effluent guidelines for point discharges that contaminate the nation's water.

The first stormwater regulation was issued in 1973 (38 FR 13530, May 22), but EPA believed that the traditional end-of-pipe controls used for process discharges and treatment works could not be used to control stormwater pollution. In addition, it would require a tremendous effort to issue NPDES permits for each of the stormwater sources in the United States. The initial stormwater regulations were developed for large municipalities (>100,000 population) and for certain industrial categories. Current regulations associated with Phase II of the stormwater permit program now require stormwater management for all urban areas in the US.

180

The CWA of 1972 provided an important tool for communities. Section 208 provided the capability to implement stormwater management plans at the regional level. The task was welcomed by planning offices, which in some cases received advice from the U.S. Army Corps of Engineers. In 1976, EPA enlarged the planning initiative through the "Section 208: Areawide Assessment Procedures Manual". However, in the late 1970s, some problems arose with the 208 planning projects due to inadequate data and lack of technological development (Whipple, as quoted in Pitt 1999).

Between 1978 and 1983, the EPA conducted the National Urban Runoff Program (NURP) that sought to determine water quality from separate storm sewers for different land uses. This program studied 81 outfalls at 28 sites, monitoring approximately 2300 storm events (EPA 1983).

NURP is still an important reference for water quality characteristics of urban stormwater; however, there are other important studies that characterize stormwater pollutant generation. The USGS created a database with more than 1100 storms at 98 sites in 20 metropolitan areas as part of their monitoring for Phase I communities (Smullen and Cave 2003)

In 1987, the amendments to the CWA established a two-phase program to regulate 13 classes of stormwater discharges. Two of these classifications were discharges from large and medium Municipal Separate Storm Sewer Systems (MS4). A large MS4 serves an urban population of 250,000 or more, while a medium MS4 serves communities between 100,000 and 250,000. EPA set up a permit strategy for communities complying with NPDES requirements. Monitoring data from this program have been included in some databases. The CDM National Stormwater Quality database

included 816 NPDES storm events in a database that includes approximately 3100 total

events. The Rouge River National Wet Weather Demonstration Program office in Detroit

included their NPDES data in their database (Smullen and Cave 2003). Recently EPA

granted the University of Alabama and the Center of Watershed Protection funding to

collect and evaluate NPDES MS4 municipal stormwater permits. By the end of 2002 this

project had collected 3765 storm events from 66 agencies and municipalities in 17 states.

This database (National Stormwater Quality Database) includes geographic and seasonal

information that can be useful for various analyses (Pitt et al. 2003).

Sample collection conducted for some of the NPDES MS4 Phase I permits

required both a grab and a composite sample for each event. A grab sample was taken

during the first 30 minutes of discharge, and a flow-weighted composite sample for the

entire time of discharge (up to three hours). The initial grab sample was used for the

analysis of the "first flush effect," which assumes that more of the pollutants are

discharged during the first period of runoff than during later periods. The composite

sample was obtained with aliquots collected about every 15 to 20 minutes for at least 3

hours, or until the event ended.

6.2 First Flush

First flush refers to an assumed elevated load of pollutants discharged during the

beginning of a runoff event. The first flush effect has been observed more often in small

catchments than in large catchments (Thompson et al, 1995, cited by WEF and ASCE

1998). In another study was found that large catchments (>162 Ha, 400 acres) had the

highest concentrations observed at the times of flow peak (Soeur, et al. 1994; Brown et

al. 1995). The presence of a first flush also has been reported to be associated with runoff duration by the City of Austin, TX (Swietlik et al.. 1995). Peak pollutant concentrations can occur after the peak discharge, thus some pollutant discharges can be significant for events longer than the time of concentration (Ellis 1986). Adams and Papa (2000), and Deletic (1998) both concluded that the presence of a first flush depends on numerous site and rainfall characteristics.

In this chapter, pollutant characteristics are evaluated using the NSQD database for events that included separate samples collected during both the first 30 minutes and for the entire event (the composite sample), using nonparametric statistical methods. A better analysis of first flush conditions could be performed by using mass discharge curves that relate the total mass discharge as a function of the total runoff volume; however, this procedure requires high resolution flow and concentration information. The NSQD database only contains concentration data from composite samples (and selected first flush samples) and few flow data.

6.3 Methodology

A total of 417 storm events having paired first flush and composite samples were available from the NPDES MS4 database. The majority of the events were located in North Carolina (76.2%), but some events were also from Alabama (3.1%), Kentucky (13.9%) and Kansas (6.7%). Table 44 shows the events that were used for this analysis, separated by land use and community. All the events correspond to end-of-pipe samples in separate storm drainage systems.

Table 44. Preliminary Number of Storm Events Selected

State	Community	CO	FW	ID	IS	OP	RE	Total Events	%
AL	Jefferson County	5	2	0	0	0	6	13	3.1
NC	City of Charlotte	8	0	8	0	3	16	35	8.4
NC	City of Fayetteville	18	0	18	18	6	46	106	25.4
NC	City of Greensboro	33	0	33	0	15	33	114	27.3
KY	City of Lexington	12	3	2	0	2	18	37	8.9
KY	City of Louisville	0	0	7	0	0	14	21	5.0
NC	City of Raleigh	18	0	18	0	9	18	63	15.1
KA	City of Wichita	7	0	7	0	0	14	28	6.7
Total Events		101	5	93	18	35	165	417	
%		24.2	1.2	22.3	4.3	8.4	39.6		100

Note: CO (commercial), FW (freeway), ID (industrial), IS (institutional), OP (Open Space) and RE (residential) land uses

The initial task was to select the constituents and land uses that meet the requirements of the statistical comparison tests. Probability plots, box and whiskers plots, concentration vs. precipitation and standard descriptive statistic calculations were performed for 22 constituents for each land use and all land uses combined. Nonparametric statistical analyses were performed after these initial analyses. Mann-Whitney and Fligner-Policello tests were most commonly used. Minitab and Systat statistical programs, along with Word and Excel macros, were used during the analysis.

6.3.1 Initial Analyses

One of the conclusions of the NURP program was that most of the constituents in stormwater generally follow a log-normal distribution, especially between the 5th and 95th percentiles (EPA, 1983). This characteristic was validated using probability plots during the initial analyses. Results from first flush and composite samples were log-transformed, for different pollutant types, in each land use category.

Figure 53 shows initial statistical results for both phosphorus and COD. Elevated

first flush concentrations were evident for COD compared to phosphorus. Probability

plots provide useful information about the characteristics of the sample population.
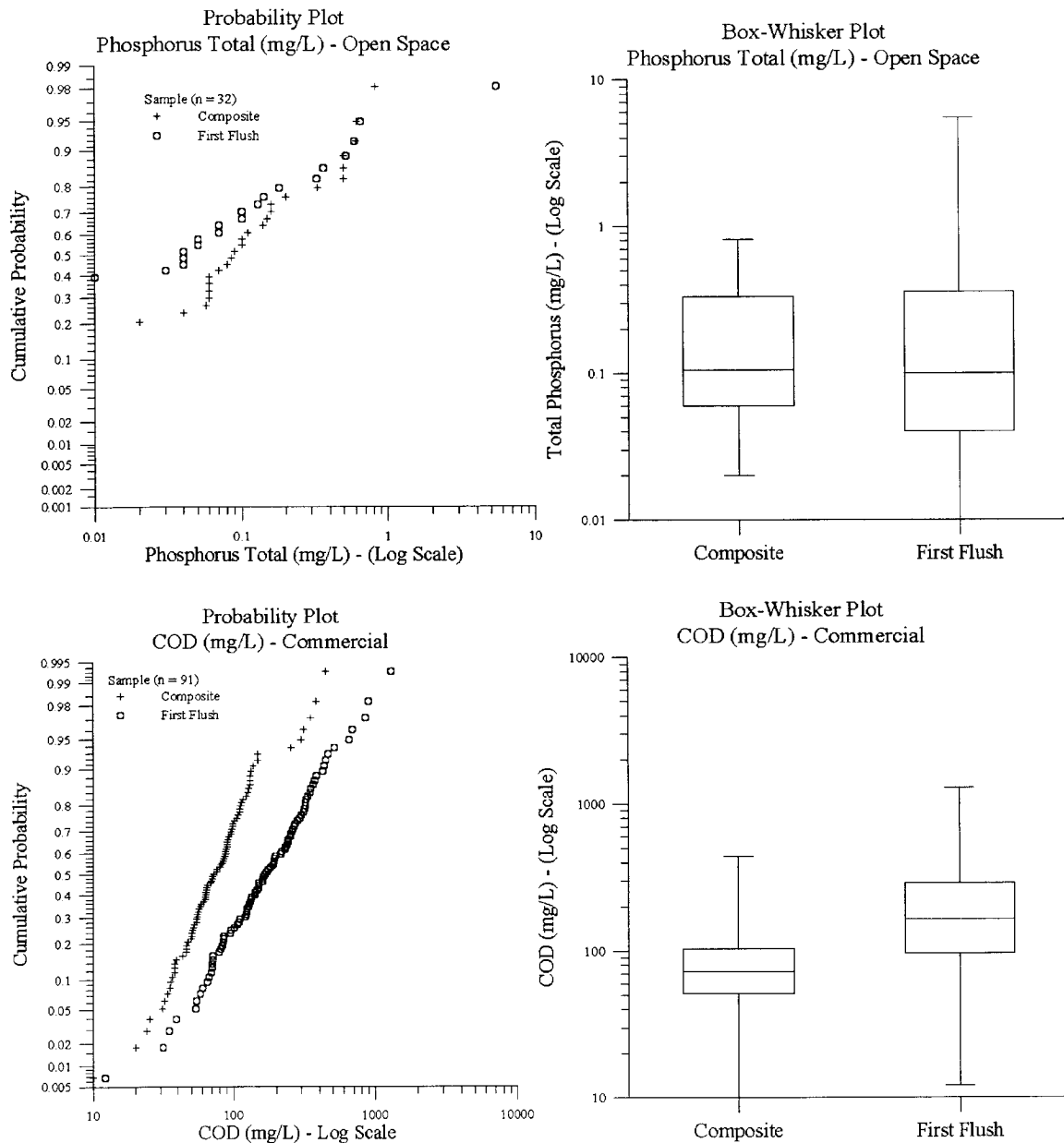
Fig. 53. Cumulative probability and box and whiskers plots

Figure 53 is an example for total phosphorus observations from the open space

land use. Both sample sets follow a lognormal distribution because most of the points lie

on a straight line. The slopes of the lines are different, indicating unequal variances. Different communities used different methods having different detection limits. In this case, about 40% of the first flush samples did not have detected concentrations for phosphorus, while about 20% of the composite samples had non-detected phosphorous concentrations. This plot also indicated that the median concentration of the composite samples is almost twice the median value for the first flush samples.

The next initial analysis used box plots. These plots also represent the distribution of the data, but only show the detectable concentrations. The middle line inside the box represents the median of the data. The top of the box represents the third quartile, and the bottom the first quartile. The whiskers are extended from the 5th to the 95th percentile limits. Values outside these limits are represented with asterisks. The exclusion of the non-detected values changes the median of the data compared to the probability plots. In this example, both of the medians are similar, in contrast with the results of the probability plot. In this example, the variability of the first flush observations is also seen to be larger than the composite data set.

Descriptive statistics for each constituent and land use were calculated to determine if the distributions were symmetrical and if they had the same variance (see Appendix E). This evaluation is needed to select the most appropriate statistical tests. In some conditions, the number of sample pairs was not large enough to allow further analyses. Table 45 shows the results of the initial analysis. Samples having lognormal probability distributions and sufficient data sets were selected for further analyses.

Figure 54 shows the steps that were followed during the nonparametric analysis. The most useful test was the Fligner-Policello test. This test requires independent random

samples symmetric about the medians for each data set. The advantage of this test is that

does not require normality or the same variance in each data set (Fligner and Policello,

1981). The U statistic and the p-value are shown in the attached appendix for some

constituents. Chakraborti (2003) presents a definition and explanation of the Mann-

Whitney U test. P-values smaller than 5% (<0.05) indicate that the first flush and

composite sample sets have different median concentrations at the 95%, or greater,

confidence level.

Table 45. Initial Analyses

Constituent	CO	ID	IS	OP	RE	ALL
Turbidity, NTU	Selected	No data	No data	Ned	Selected	Selected
pH, S.U.	Selected	Selected	No data	Ned	Selected	Selected
BOD5, mg/L	Selected	Selected	Box plot FF > Com	Selected	Selected	Selected
COD, mg/L	Selected	Selected	Selected	Selected	Selected	Selected
TSS, mg/L	Selected	Selected	Selected	Selected	Selected	Selected
TDS, mg/L	Selected	Selected	Selected	Selected	Selected	Selected
O&G, mg/L	Selected	Ned	Ned	Ned	Selected	Selected
Fecal Coliform, col/100mL	Selected	Ned	Ned	Ned	Selected	Selected
Fecal Streptococcus, col/100 mL	Selected	Ned	Ned	Ned	Selected	Selected
Ammonia, mg/L	Selected	Selected	Box plot FF > com.	Ned	Selected	Selected
$NO_2 + NO_3$, mg/L	Selected	Selected	Selected	Selected	Selected	Selected
N Total, mg/L	Selected	Selected	Ned	Selected	Selected	Selected
TKN, mg/L	Selected	Selected	Box plot FF > com.	Selected	Selected	Selected
P Total, mg/L	Selected	Selected	Selected	Selected	Selected	Selected
P Dissolved, mg/L	Selected	Selected	Selected	Selected	Selected	Selected
Ortho-P, mg/L	Ned	Selected	Ned	Ned	Selected	Selected
Cadmium Total, µg/L	Selected	Selected	Ned	Selected	Selected	Selected
Chromium Total, µg/L	Selected	Selected	Ned	Selected	Selected	Selected
Copper Total, µg/L	Selected	Selected	Selected	Selected	Selected	Selected
Lead Total, µg/L	Selected	Selected	Selected	Selected	Selected	Selected
Mercury, µg/L	Ned	Ned	Ned	Ned	Ned	Ned
Nickel, µg/L	Selected	Selected	Ned	Ned	Selected	Selected
Zinc, µg/L	Selected	Selected	Selected	Selected	Selected	Selected

* Ned: Not enough data. CO (commercial), FW (freeway), ID (industrial), IS (institutional), OP (Open Space) and RE (residential)

6.3.2 Nonparametric Analyses

If the number of samples is large, and the distributions are normal and have the

same variance, a paired Student's t-test is usually a better test to evaluate the hypothesis

and support the results of the Fligner-Policello test. To verify that the data distributions

are normal, the Anderson-Darling normality test was used (Kottegoda and Rosso, 1997).

This method uses an empirical cumulative distribution function to check normality. In the attached appendix, the p-values of the paired differences are shown. P-values larger than 5% (> 0.05) indicate that the normality requirement was met at the 95% or greater confidence level.

Finally, if the first flush and composite sample distributions are symmetrical (but not necessarily normal), and if they have the same variance, the Mann-Whitney test can be used. If the p-value is larger than 5% (>0.05), the medians of the sample distribution are assumed to be the same, at the 95% or greater confidence level. The preferred test would be the Student's t-test, if the sample characteristics warrant, followed by the Mann-Whitney test and finally the Fligner-Policello test. The selected cases are only for pairs with concentration values above the detection limits. The ratios between the first flush and composite sample medians are also shown. Commercial and residential areas have the highest ratios for most constituents. The smallest ratios were found for open space sites.

Null Hypothesis:
median first flush and composite
concentrations are the same

Alternative Hypothesis:
median first flush and composite
concentrations are different
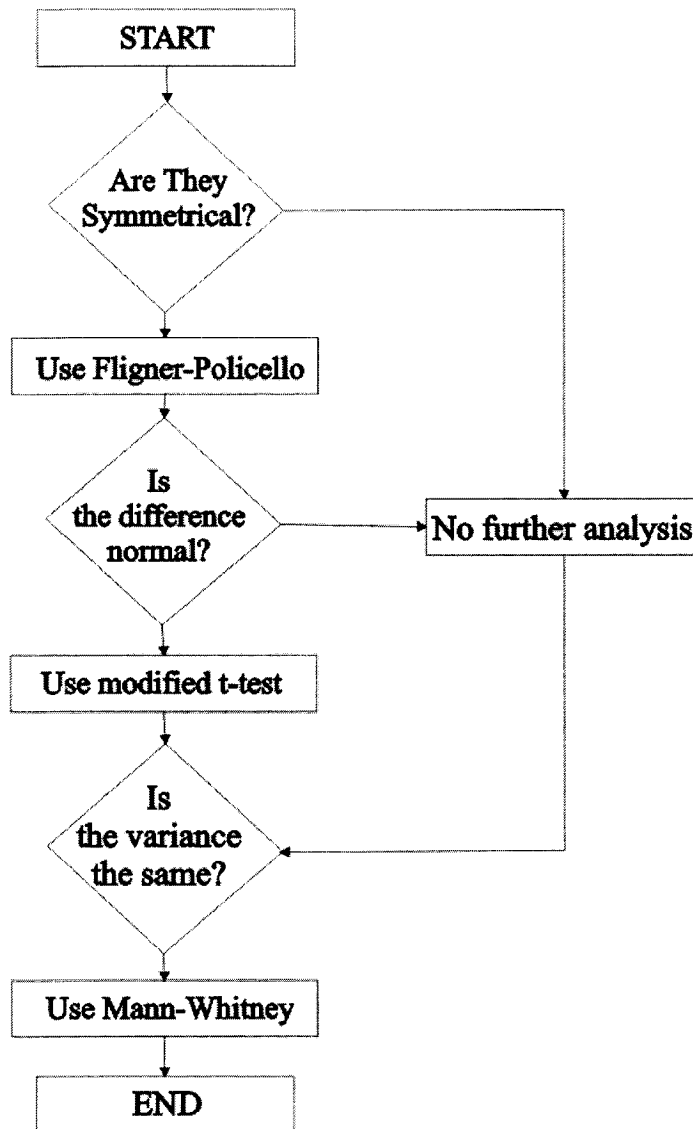
Fig. 54. Analysis flow chart

6.4 Results

About 83% of the possible paired cases were successfully evaluated. The

remaining cases could not be evaluated because the data set did not have enough paired

data or they were not symmetrical. Table 46 shows the results of the analysis.

Table 46. Significant First Flushes Ratios (first flush to composite median concentration)

Parameter	Commercial				Industrial				Institutional			
	n	sc	R	ratio	n	sc	R	ratio	n	sc	R	ratio
Turbidity, NTU	11	11	=	1.32			X				X	
pH, S.U.	17	17	=	1.03	16	16	=	1.00			X	
COD, mg/L	91	91	≠	2.29	84	84	≠	1.43	18	18	≠	2.73
TSS, mg/L	90	90	≠	1.85	83	83	=	0.97	18	18	≠	2.12
BOD5, mg/L	83	83	≠	1.77	80	80	≠	1.58	18	18	≠	1.67
TDS, mg/L	82	82	≠	1.83	82	81	≠	1.32	18	18	≠	2.66
O&G, mg/L	10	10	≠	1.54			X				X	
Fecal Coliform, col/100mL	12	12	=	0.87			X				X	
Fecal Streptococcus, col/100 mL	12	11	=	1.05			X				X	
Ammonia, mg/L	70	52	≠	2.11	40	33	=	1.08	18	16	≠	1.66
NO2 + NO3, mg/L	84	82	≠	1.73	72	71	≠	1.31	18	18	≠	1.70
N Total, mg/L	19	19	=	1.35	19	16	=	1.79			X	
TKN, mg/L	93	86	≠	1.71	77	76	≠	1.35			X	
P Total, mg/L	89	77	≠	1.44	84	71	=	1.42	17	17	=	1.24
P Dissolved, mg/L	91	69	=	1.23	77	50	=	1.04	18	14	=	1.05
Ortho-P, mg/L			X		6	6	=	1.55			X	
Cadmium Total, µg/L	74	48	≠	2.15	80	41	=	1.00			X	
Chromium Total, µg/L	47	22	≠	1.67	54	25	=	1.36			X	
Copper Total, µg/L	92	82	≠	1.62	84	76	≠	1.24	18	7	=	0.94
Lead Total, µg/L	89	83	≠	1.65	84	71	≠	1.41	18	13	≠	2.28
Nickel, µg/L	47	23	≠	2.40	51	22	=	1.00			X	
Zinc, µg/L	90	90	≠	1.93	83	83	≠	1.54	18	18	≠	2.48
Turbidity, NTU			X		12	12	=	1.24	26	26	=	1.26
pH, S.U.			X		26	26	=	1.01	63	63	=	1.01
COD, mg/L	28	28	=	0.67	140	140	≠	1.63	363	363	≠	1.71
TSS, mg/L	32	32	=	0.95	144	144	≠	1.84	372	372	≠	1.60
BOD5, mg/L	28	28	=	1.07	133	133	≠	1.67	344	344	≠	1.67
TDS, mg/L	31	30	=	1.07	137	133	≠	1.52	354	342	≠	1.55
O&G, mg/L			X				X		18	14	≠	1.60
Fecal Coliform, col/100mL			X		10	9	=	0.98	22	21	=	1.21
Fecal Streptococcus, col/100 mL			X		11	8	=	1.30	26	22	=	1.11
Ammonia, mg/L			X		119	86	≠	1.36	269	190	≠	1.54
NO2 + NO3, mg/L	30	21	=	0.96	121	118	≠	1.66	324	310	≠	1.50
N Total, mg/L	6	6	=	1.53	31	30	=	0.88	77	73	=	1.22
TKN, mg/L	32	14	=	1.28	131	123	≠	1.65	335	301	≠	1.60
P Total, mg/L	32	20	=	1.05	140	128	≠	1.46	363	313	≠	1.45
P Dissolved, mg/L	32	14	=	0.69	130	105	≠	1.24	350	254	=	1.07
Ortho-P, mg/L			X		14	14	=	0.95	22	22	=	1.30
Cadmium Total, µg/L	30	15	=	1.30	123	33	≠	2.00	325	139	≠	1.62
Chromium Total, µg/L	16	4	=	1.70	86	31	=	1.24	218	82	≠	1.47
Copper Total, µg/L	30	22	=	0.78	144	108	≠	1.33	368	295	≠	1.33
Lead Total, µg/L	31	16	=	0.90	140	93	≠	1.48	364	278	≠	1.50
Nickel, µg/L			X		83	18	=	1.20	213	64	≠	1.50
Zinc, µg/L	21	21	=	1.25	136	136	≠	1.58	350	350	≠	1.59

Note: n = number of total possible events. sc = number of selected events with detected values. R = result. Not enough data (X); not enough evidence to conclude that median values are different (=); median values are different (≠).

The "$\neq$" sign indicates that the medians of the first flush and the composite data set are different. The "=" sign indicates that there is not enough information to reject the null hypothesis at the desired level of confidence (at least at the 95% level). Events without enough data are represented with an "X".

Also, shown on this table are the ratios of the medians of the first flush to the composite data for each constituent and land use combination. Generally, a statistically significant first flush is associated with a median concentration ratio of about 1.4, or greater (the exceptions are where the number of samples in a specific category is much smaller). The largest ratios are about 2.5, indicating that for these conditions, the first flush sample concentrations are about 2.5 times greater than the composite sample concentrations. More of the larger ratios are found for the commercial and institutional land use categories, areas where larger paved areas are likely to be found. The smallest ratios are associated with the residential, industrial, and open spaces land uses, locations where there may be larger areas of unpaved surfaces.

Results indicate that for 55% of the evaluated cases, the median of the first flush data set were different than the composite sample set. In the remaining 45% of the cases, both medians were likely the same, or the concentrations were possibly greater later in the events.

Approximately 70% of the constituents in the commercial land use category had elevated first flush concentrations, about 60% of the constituents in the residential, institutional and the mixed (mostly commercial and residential) land use categories had elevated first flushes, and only 45% of the constituents in the industrial land use category

had elevated first flushes. In contrast, no constituents were found to have elevated first flushes in the open space category.

COD, BOD$_5$, TDS, TKN and Zn all had first flushes in all areas (except for the open space category). In contrast, turbidity, pH, fecal coliform, fecal streptococcus, total N, dissolved and ortho-P never showed a statistically significant first flush in any category. The different findings for TKN and total nitrogen imply that there may be other factors involved in the identification of first flushes besides land use. If additional paired data becomes available during later project periods, it may be possible to extend this analysis to consider rain effects, drainage area, and geographical location.

6.5 Summary

It is expected that peak concentrations generally occur during periods of peak flows (and highest rain energy). On relatively small paved areas, however, it is likely that there will always be a short initial period of relatively high concentrations associated with washing off of the most available material (Pitt, 1987). This peak period of high concentrations may be overwhelmed by periods of high rain intensity that may occur later in the event. In addition, in more complex drainage areas, the routing of these short periods of peak concentrations may blend with larger flows and may not be noticeable. A first flush in a separate storm drainage system is therefore most likely to be seen if a rain occurs at relatively constant intensity over a paved area having a simple drainage system.

If the peak flow (and highest rain energy) occurs later in the event, then there likely will not be a noticeable first flush. However, if the rain intensity peak occurs at the beginning of the event, then the effect is exaggerated. Figure 55 shows an example storm

in Lexington, KY. Note that in this event there are two periods of elevated peaks, the first occurs one hour after the rain started, the second two hours later. If the concentration remains the same during the entire event, the maximum load will occur during the later periods having the maximum flows (the two peaks), and not during the initial period of the storm. Another factor that needs to be considered is the source of the contaminants and how fast they travel through the watershed. Streets and other impervious areas will contribute flows to the outfall monitoring location before the pervious areas in the drainage area.
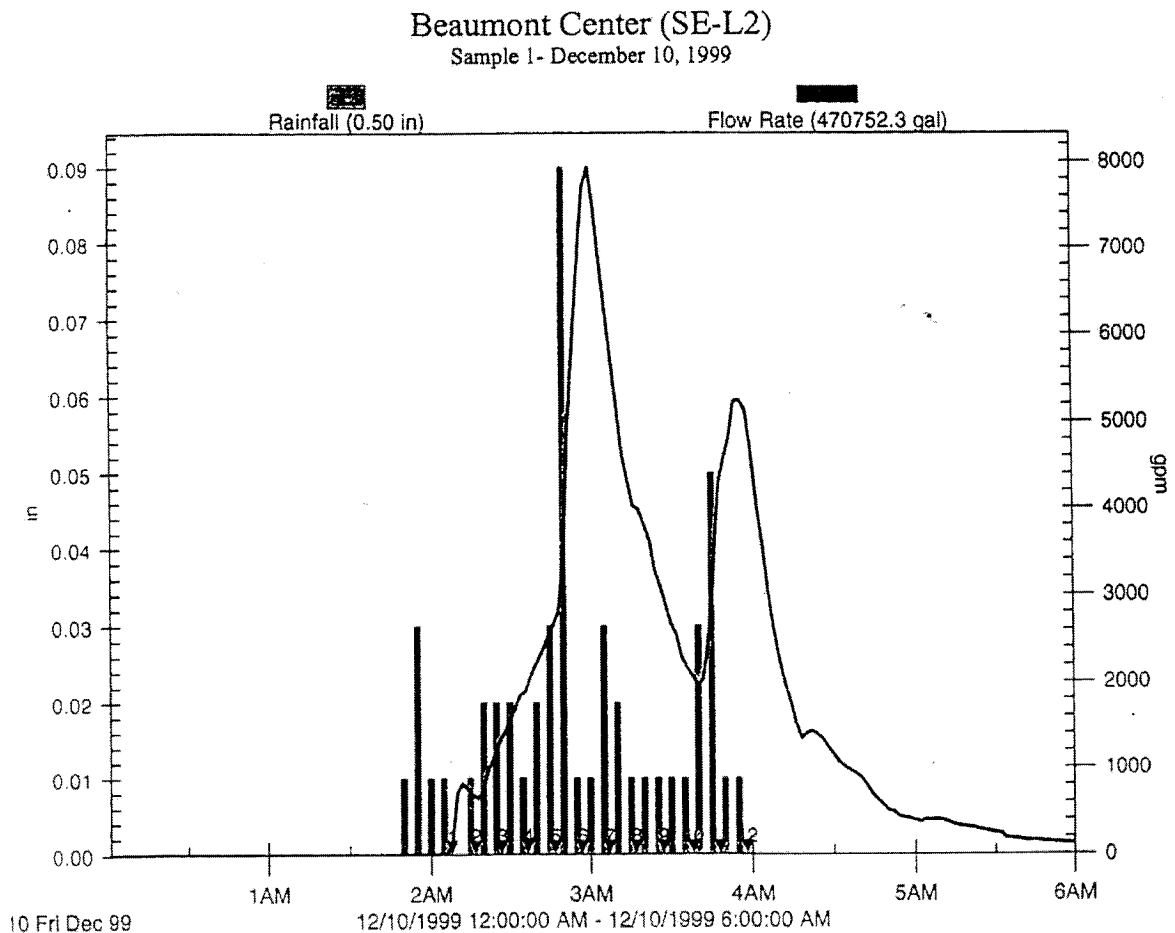
Fig. 55. Hydrograph for a storm event (Source: NPDES permit Lexington–KY 2000)
(1 in = 25.4 mm, 1 m3 = 264.17 gal)
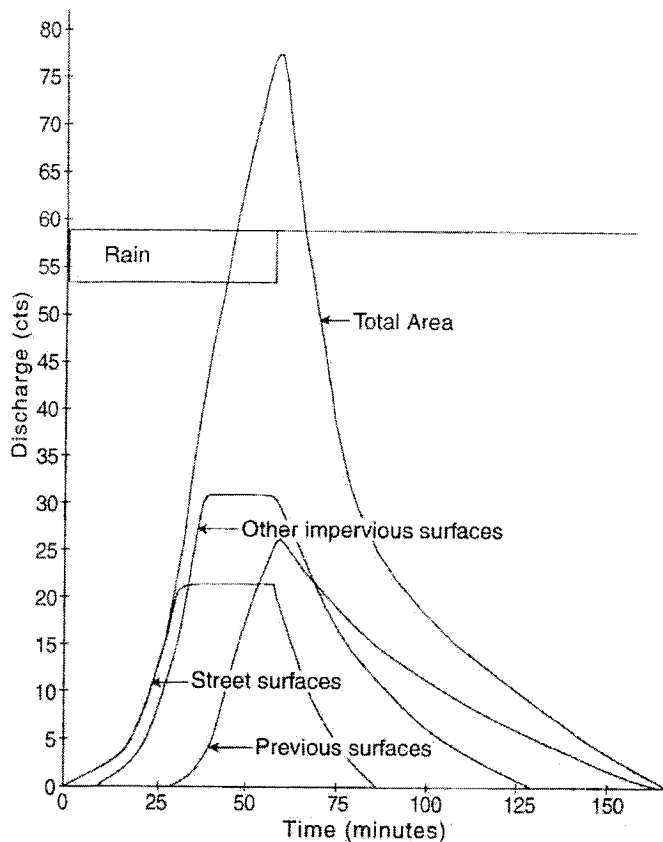
Fig. 56. Contributing areas in urban watersheds (Pitt, 1999) (1 m3/s = 35.32 cfs)

Figure 56 (Pitt, 1999) shows that for an example constant rainfall, the source area flow contribution changes for different rain conditions in an area. If the percentage of impervious surfaces is high, many of the constituents will be discharged faster. This observation agrees with the results observed from the statistical analysis. Commercial areas have a larger frequency of high concentrations at the beginning of the event in contrast to open space areas.

Figure 57 shows that for events (< 12mm, or 0.5 in) in this example medium density residential area, most of the runoff is generated by impervious areas. The average percentage of imperviousness for the monitoring sites was examined. Commercial areas had an average of 83% imperviousness, followed by industrial areas at 70%

imperviousness. Institutional and residential land uses were very similar, with 45% and

42% imperviousness respectively. The open space land use category had the smallest

imperviousness area, at about 4%. As indicated in Figure 57, larger events can generate

more runoff from previous areas than impervious areas. However, it is likely that most of

the runoff during the MS4 monitoring activities was associated with the more common

small events, and hence, impervious areas were more important.
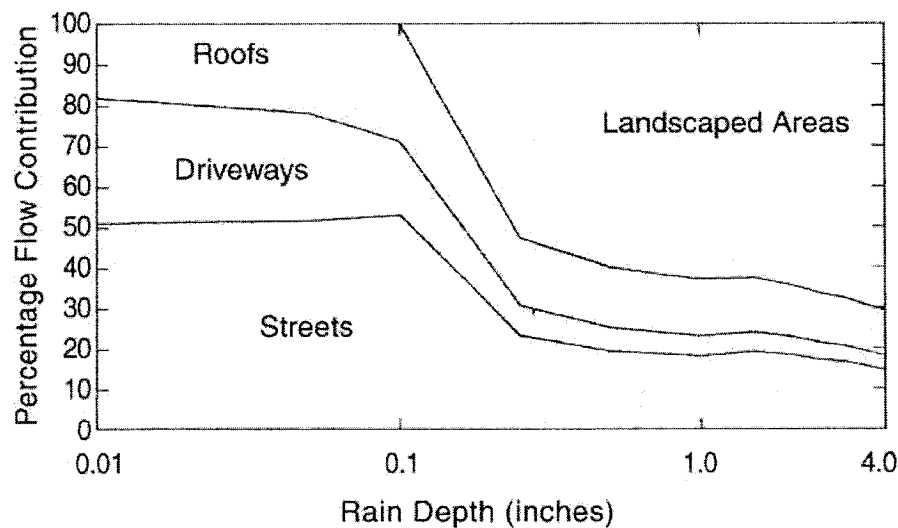
Fig. 57. Contributing areas in urban watersheds (Pitt and Voorhees, 1995)
(1 in = 25.4 mm)

Probability plots of the precipitation associated with each monitored event for

each land use category were prepared to see if there were any significant differences in

the ranges of rains observed within each land use category that could have influenced the

results. Figure 58 shows that precipitation has the same distribution for almost all the

different land uses. The institutional land use category shows a slightly smaller median

rain, but this is likely because of the smaller number of events observed in that land use

category (18 events). The median precipitation observed during the monitoring at all land

uses was about 8 mm (0.3 in), indicating the importance of the runoff from the

impervious areas.
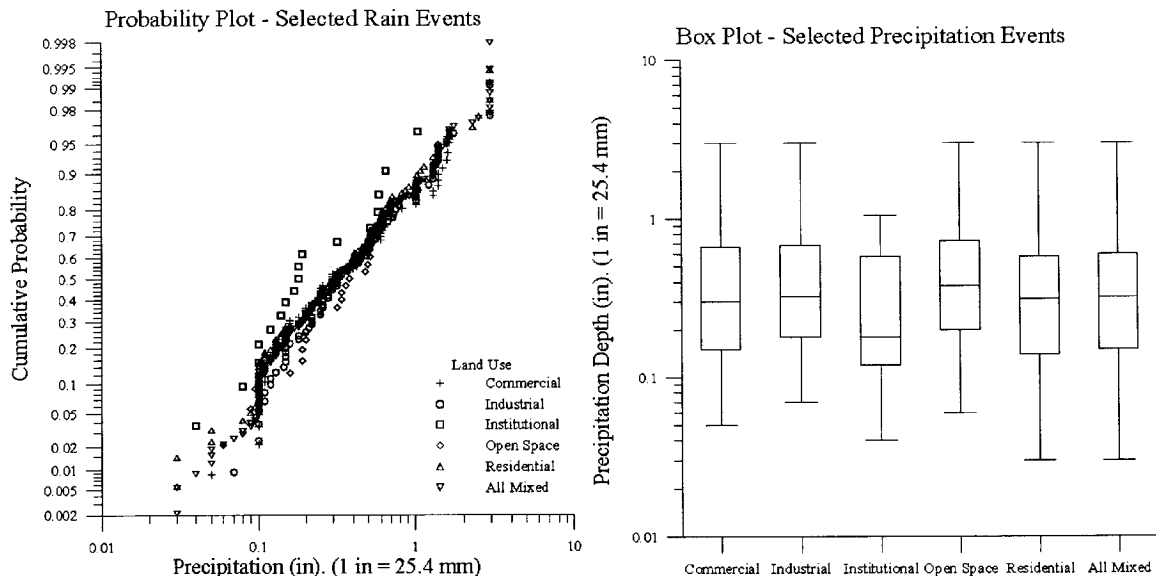
Fig. 58. Probability and box and whiskers plot of selected rain events

Finally, another factor that must be considered is the effect of the sampling

duration. The guidance provided for monitoring during the Phase I activities was to

collect a sample during the first 30 minutes of the event, and a composite sample only

during the first three hours of the event (or the complete event, if shorter than three

hours). Figure 59 shows an example case when these conditions can lead to inappropriate

conclusions for longer duration events.
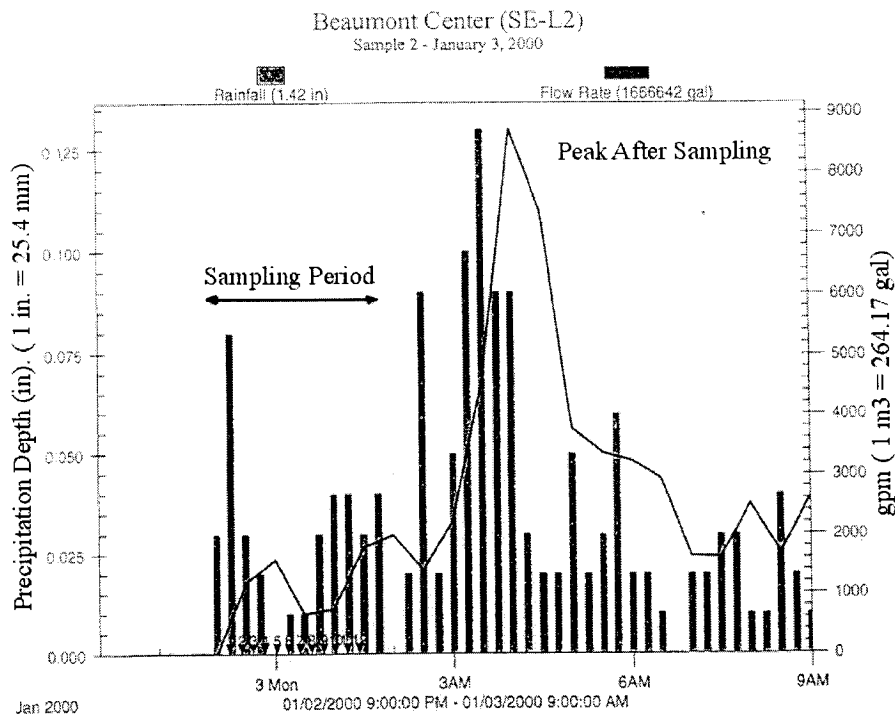
Fig. 59. Example of an event with peaks after the sampling period
(Source: NPDES permit Lexington-KY, 2000)

The 12 aliquots sampled during the first three hours are shown on the left side of Figure 59. The peak discharge occurred four hours after the event started, as shown on the right side of the figure, and was not represented in the sampling effort. Missing these later storm periods can lead to inappropriate conclusions. It is suggested that for stormwater monitoring, samples should be collected during the complete event and composited before laboratory analyses.

Another sampling example was presented by Roa-Espinosa and Bannerman (1995) who collected samples from five industrial sites using five different monitoring methods. Table 47 shows the ranking of the best methods of sampling based in six criteria. In this table a value between one and five points is assigned to each criterion. Five points indicates that the method is excellent in the specific criterion. Rao-Espinosa

and Bannerman concluded that many time-composite subsamples combined for a single

composite analysis can provide improved accuracy compared to fewer samples associated

with flow-weighted sampling. They also found that time composite subsamples provide

better results than samples collected during the first 30 minutes of the event.

Table 47. Ranking by Methods of Sampling (Roa-Espinosa, Bannerman, 1995)

Criteria	Flow Composite	Time Discrete	Time Composite	Old Source Sample	New Source Sample	First 30 Minutes
Site Selection	1	1	1	5	5	3
Cost	1	1	3	5	5	5
Technical difficulty	1	1	3	5	5	5
Accuracy	5	5	4	1	5	1
Reproducibility	5	5	5	1	5	1
Representativeness	1	1	3	5	5	1
TOTAL POINTS	14	14	19	22	30	16

6.6 Conclusion

A major goal of the present study is to provide guidance to stormwater managers

and regulators. Especially important will be the use of this data as an updated benchmark

for comparison with locally collected data. In addition, this data may be useful for

preliminary calculations when using the "simple method" for predicting mass discharges

for unmonitored areas. These data can also be used as guidance when designing local

stormwater monitoring programs (Burton and Pitt, 2002), especially when determining

the needed sampling effort based on expected variations. Additional analyses reported in

other dissertation chapters expand on these preliminary examples and also investigate

other stormwater data and sampling issues.

This investigation of first flush conditions indicated that a first flush effect was

not present for all the land use categories, and certainly not for all constituents.

Commercial and residential areas were more likely to show this phenomenon, especially if the peak rainfall occurred near the beginning of the event. It is expected that this effect will be more likely to occur in a watershed with a high level of imperviousness, but even so, the data indicated first flushes for less than 50% of the samples for the most impervious areas. This reduced frequency of observed flush flushes in these areas most likely to have first flushes is likely associated with the varying rain conditions during the different events, including composite samples that did not represent the complete runoff duration.

Groups of constituents showed different behaviors for different land uses. All the heavy metals evaluated showed higher concentrations at the beginning of the event in the commercial land use category. Similarly, all the nutrients showed higher concentrations in residential land use areas, except for total nitrogen and ortho-phosphorus. This phenomenon was not found in the bacteria analyses. None of the land uses showed a higher number of colonies during the beginning of the event. Conventional constituents showed elevated concentrations in commercial, residential and institutional land uses.

Chapter 7

Example Application of the National Stormwater Quality Database: TSS and Nutrient Export Calculations for Chesapeake Bay Watersheds

7.1 Overview

This chapter is a demonstration of how the data contained in the NSQD can be used, especially in conjunction with additional urban area flow data, and rural runoff data to estimate the relative contributions of pollutants in a region. This chapter first summarizes the data used, the statistical tests performed, and the results obtained, as part of our effort to identify the most appropriate nonpoint source runoff characteristics for the Chesapeake Bay watershed, the area having most of the collected data in the NSQD.

7.2 Data Availability

Two sources of data were used to estimate nonpoint sources of pollution. The first data source corresponding to discharges from urban areas was obtained from the NSQD for the area. The second data source corresponding to discharges from agricultural land uses and forested land cover was obtained from regional data summaries provided by the EPA's Chesapeake Bay Program, "Smart Growth" project group (Office of Policy, Economics, and Innovation).

7.2.1 Urban Data

Data from within the Chesapeake Bay watershed, as contained in the National

Stormwater Quality Database (NSQD version 1.1), were used to determine the most

appropriate concentrations for urban stormwater nutrients and suspended solids. The

NSQD contains information of stormwater discharge concentrations for 19 counties in

Virginia and Maryland (Table 48). More than 1,300 events were monitored in these areas

representing residential, commercial and industrial land uses. There were no data

reported for open space or freeway land uses. The watersheds monitored in Maryland and

Virginia ranged from 3.5 and 882 acres and were between 7 and 90% impervious.

Reported events used in these analyses were monitored from October 1990, through

December 2000.

Table 48. Urban Monitoring Locations in the Chesapeake Bay Watershed
Represented in the National Stormwater Quality Database (NSQD, version 1.1)

Virginia	Maryland
Arlington County	Hartford County
Norfolk County	Baltimore County
Virginia Beach County	Baltimore City
Chesapeake County	Carroll County
Portsmouth County	Howard County
Hampton County	Anne Arundel County
Newport News County	Price George's County
Henrico County	Charles County
Chesterfield County	Montgomery County
Fairfax County	

Data for total nitrogen (the sum of total Kjeldahl nitrogen, TKN, and nitrite plus

nitrate, NO_2+NO_3), total phosphorus, and total suspended solids (TSS) were evaluated for

use in the Chesapeake Bay watershed.

7.2.2 Rural Data

Chesapeake Bay rural water quality information was reported by the USGS in: Synthesis of Nutrient and Sediment Data for Watersheds within the Chesapeake Bay Drainage Basin (Langland, et al. 1995), prepared in corporation with the EPA. This report describes the comprehensive database of nutrient and sediment data collected from 1972 through 1992 from 1,058 non-tidal monitoring stations in the Chesapeake Bay watershed. Annual discharge loads were calculated at 48 locations for total nitrogen, at 99 locations for total phosphorus, and at 33 locations for suspended sediment. Many of the stations did not have sufficient samples, or flow data to enable load calculations. The fewer locations available for suspended sediment reflect those stations that evaluated suspended sediment, and not suspended solids. Gray, et al. (2000) concluded that suspended solids data are not reliable indicators of suspended sediment due to the laboratory processing associated with TSS analyses. The typical pipetting, or pouring, of a subsample for gravimetric analyses typically under predicts the mass associated with sand-sized particles (> 63 micrometers). If cone or churn splitters were used, then the TSS analyses were found to be reasonable. They also found that the results using the two methods are comparable if the mass of these larger particles comprise less than about 25% of the total sample mass. Since no particle size data was available for the TSS samples, they only used information for locations that had total sediment concentrations. Outfall urban runoff samples typically have less than 20% sand, although some early season samples in northern areas where sand is used for traction control may periodically have close to 50% sand, and some source area samples can also have large sand fractions. The TSS values used in the urban component of the analyses, described previously, are

expected to be acceptable, as Chesapeake Bay region samples should not be influenced

by appreciable winter sand applications, and these are all outfall samples.

Langland, et al. (1995) calculated annual nutrient and sediment loads for the

selected locations using an unbiased log-linear regression model. This model enabled

them to extrapolate the results to annual conditions, and to recognize both base flow

conditions (groundwater recharge to the rivers is a major nitrate source, for example) and

higher flows associated with surface runoff during storm periods. This analysis also

enabled them to consider the potential septic tank and atmospheric deposition

contributions to the annual soluble nitrogen loads. Numerous correlation analyses of

annual yields of sediment and nutrients with respect to land use, physiographic province,

and rock type. They found that river basins having larger percentages of agricultural land

had larger nutrient and sediment yields, and that basins that were urbanized had

substantially less yields. Table 49 shows the amount of each major land use category in

the watershed, and in the portions of the major states within the watershed. In all cases,

the land is dominated by forest and agricultural lands, with all urban lands making up

about 12% for Maryland and 9% for Virginia portions of the watershed.

Table 49. Land Uses in the Chesapeake Bay Watershed (Landland, et al. 1995)

	Percent of Bay Basin	Pennsylvania	Maryland	Virginia
Woody (forest)	53.9	62.5	32.6	52.4
Herbaceous (agriculture)	30.6	31.1	31.3	28.3
High intensity urban	0.6	0.3	1.2	0.6
Low intensity urban	4.0	2.9	6.4	4.5
Woody urban	1.1	0.6	1.9	1.4
Herbaceous urban	1.6	0.8	2.7	2.3
Water	7.2	11.1	20.8	9.6
Exposed	0.3	0.7	0.2	0.1
Herbaceous wetland	0.8	0	2.9	0.8
Total	100	100	100	100

Langland, et al. (1995) used Kendall's tau test to examine simple linear

correlations between annual nutrient and sediment yields and land use, physiographic

province, and rock type in the river basins above each station where the annual loads

were calculate. They found that land use was the most important variable for predicting

nutrient and sediment yield from a river basin. The strongest, most significant and most

consistent correlations were between nutrient and sediment yields and agricultural land

use. Table 50 shows selected annual yield and land use data for ten of the "load" stations

evaluated by the USGS in their Chesapeake Bay report (Langland, et al. 1995).

Unfortunately, they did not determine the unit area yields corresponding to separate land

uses. They presented these stations as representing the range of land uses for separate

locations.

Table 50. Reported Mean Annual Yields and Land Use (Landland, et al. 1995)

Basin	Area (mi²)	Percentage Land Use			Mean Annual Yields (lb/acre/year)	
		Urban	Agriculture	Forest	Total Nitrogen (TKN plus nitrates)	Total Phosphorus
All 127 "load" basins					6.8	0.70
Predominantly Urban Basins						
01571000	11.2	46.0	26.8	28	8.2	0.80
01589300	32.5	54.4	16.8	27	8.1	na
01593500	38.0	54.5	16.8	21	na	0.67
01646000	57.9	50.9	11.4	28	5.9	0.61
01657655	4.0	48.6	22.5	27	na	0.28
Agricultural and Urban Basins						
01586000	56.6	42.4	51.0	3.4	na	0.41
01616000	16.5	43.8	41.9	13	29.7	4.0
Predominantly Agricultural Basins						
01573810	0.38	1.4	91.0	6.7	42.1	6.3
0157608335	1.42	1.1	63.4	26	26.4	4.5
01639500	102	1.1	69.9	29	14.6	na

7.3 Summary of Data and Load Calculations

The "simple" model (Schueler 1987) was used to calculate the nonpoint discharges of TSS, total phosphorus, and total nitrogen for Anne Arundel County, Maryland, for the EPA's Chesapeake Bay Program, "Smart Growth" project (Office of Policy, Economics, and Innovation). The simple model was developed by Schueler to enable rapid calculations of pollutant discharges by multiplying the event mean concentration values for a specific land use, the volumetric runoff coefficient for that land use, and the annual rainfall. With appropriate unit conversions, the result can be expressed as the unit area annual discharge for a specific pollutant. When multiplied by the area corresponding to each of the land uses in the area of concern, the total area pollutant discharges can be calculated, and the relative sources of the discharges can be identified. When working with large watersheds, these calculated values are usually much greater than the monitored in-stream values observed at the watershed outlet, because pollutant transport in the stream or river is not considered. However, it is a suitable method to identify the relative pollutant contributions of different land uses in a county, as in this example.

The volumetric runoff coefficients for each land use category were based on analyses of typical land use surface configurations (mostly the impervious area characteristics) and the rain depth was determined from 50 years of rain records from the Baltimore (BWI) airport. The urban area concentration values were obtained by statistical evaluations of the Maryland and Virginia urban area data contained in the National Stormwater Quality Database, as described in the following subsections of this chapter. The urban runoff and concentration factors are assumed to have excellent reliability.

However, some of the urban categories were not represented with regional Chesapeake Bay region data, so these factors were obtained from the national averaged values in the database and are labeled with a moderate reliability. The non-urban values are labeled as having poor to very good reliability, depending on the availability of local data. The agriculture values are from regional information summarized by Staver (1995) and Hartigan (1983) and are assumed to be of very good reliability. The forestlands data are from regional Chesapeake Bay regional data collected by the EPA, Office of Policy, Economics, and Innovation (Richards, personal communication) and are assumed to be of moderate reliability. The other land categories and the extraction lands data are unknown and are of poor reliability. Fortunately, as shown on the following summary tables, the best information is associated with the agricultural, forest, and urban categories which are responsible for almost the entire calculated discharges for the county.

The runoff factors are also indicated with varying reliabilities. The urban lands data all have excellent reliability due to the use of calibrated urban data for varying conditions. The agricultural runoff data is of the poorest reliability due to the uncertainties associated with the many agricultural operations that can have dramatic effects on these values. The natural land runoff values are expected to have moderate reliabilities. The USGS (Langland, et al. 1995) reported values are not comparable to these discharge values due to a number of reasons, most specifically because they are in-stream values and are affected by sediment and pollutant transport. The USGS report also did not report unit area loadings for specific land uses and the preliminary calculations resulted in unrealistic results that were highly variable. Tables 51, 52 and 53 list the

nutrient and suspended solids data applicable for Chesapeake Bay watershed analyses,

based on the analyses performed and outlined later in this chapter.

Table 51. Commercial TSS (mg/L), Mean and COV

	<0.1 inches	0.1 to 0.35 inches	0.35 to 1 inch	>1 inch
Spring	18 (0.72)	31 (0.67)	75 (1.5)	no data
Summer	75 (1.5)	18 (0.72)	75 (1.5)	75 (1.5)
Fall	18 (0.72)	75 (1.5)	18 (0.72)	18 (0.72)
Winter	18 (0.72)	75 (1.5)	75 (1.5)	75 (1.5)

Table 51 shows that storm events with precipitation depths larger than 0.35 inches

are more likely to discharge higher TSS concentrations in spring, summer and winter than

fall. Table 52 shows the expected total nitrogen concentrations in commercial land uses.

There is a clear variation among the seasons and precipitation depth. Storm events

smaller than 0.1 inch are expected to have higher total nitrogen discharges during the fall

and winter than during the summer and spring seasons. For rain events between 0.35 and

1 inch, the highest concentrations were observed during the summer and fall. Table 53

shows the average concentrations and coefficients of variation for TSS, total phosphorus

and total nitrogen for residential, commercial and industrial urban land use areas. This

table also includes the expected concentrations in agricultural and forested areas.

Table 52. Commercial Total Nitrogen (mg/L), Mean and COV

	<0.1 inch	0.1 to 0.35 inches	0.35 to 1 inch	>1 inch
Spring	2.0 (0.49)	3.2 (0.50)	2.0 (0.49)	no data
Summer	2.0 (0.49)	2.0 (0.49)	3.2 (0.50)	3.2 (0.50)
Fall	3.2 (0.50)	3.2 (0.50)	3.2 (0.50)	2.0 (0.49)
Winter	3.2 (0.50)	2.0 (0.49)	2.0 (0.49)	3.2 (0.50)

Table 53. Average Concentrations by Land Use

Land Use	Constituent	Conditions	Average (COV)
Urban - Residential	TSS	Summer rains (between 0.1 and 0.35 inches in depth)	143 (0.71)
		All other rains	58 (0.70)
	TP	Sites having <27% impervious cover:	
		Winter rains	0.28 (0.59)
		All other rains	0.41 (0.65)
		Sites having >27% impervious cover:	
		Winter rains (less than 0.1 inches in depth)	0.16 (0.86)
		All other rains	0.30 (0.63)
	TN	Fall rains (less than 0.1 and greater than 1 inch in depth)	1.4 (0.57)
		Winter rains (0.35 and 1 inch in depth)	1.5 (0.30)
		Fall rains (0.35 and 1 inch) and Winter rains (between 0.1 and 0.35 inches in depth)	1.9 (0.51)
		All other rains	2.4 (0.62)
		Spring and summer rains (between 0.1 and 0.35 inches in depth)	2.6 (0.38)
Urban - Commercial	TSS and TN	See tables 7.4 and 7.5	
	TP	Summer rains >1 inch and fall rains between 0.1 and 0.35 inch	0.46 (0.36)
		All other rains	0.23 (0.71)
Urban - Industrial	TSS	Fall, spring, and summer	77 (1.48)
		Winter	81 (0.93)
	TP	Rains less than 0.35 inches	0.29 (0.81)
		Rains greater than 0.35 inches	0.22 (1.05)
	TN	All conditions	2.1 (0.79)
Rural - Agricultural	Sediment		1115 lb/ac/yr (unreliable estimate)
	TN		40 lb/ac/yr
	TP		5.4 lb/ac/yr
Rural - Forest	Sediment		4500 lb/ac/yr (unreliable estimate)
	TN		0 lb/ac/yr
	TP		0 lb/ac/yr

The total runoff discharges for the county can be determined based on the
calculated total mass discharges for each land use, and the areas for each land use area.
Table 54 shows the percentage of total annual runoff volume produced for each land use
by season and rain depth range. About 61% of the total annual runoff volume was
produced by events having more than 1 inch of rain, followed by rain events in the range
0.36 to 1 inches (31% of the annual runoff volume), rain events in the range of 0.1 to 0.35
inches (7% of the annual runoff volume), and rain events less than 0.1 inch in depth (with
1% of the annual runoff volume).

Table 54. Fraction of Annual Flow Associated with Season and Rain Depth Categories (based on 50 years of rain records at Baltimore, BWI)

	rain range	Fraction of rain depth in category	Ultra low density residential	Low density residential	Medium density residential	High density residential	Commercial	Industrial	Institutional	Open urban land	Freeways
Winter	<0.1 inch	0.0078	0.0008	0.0011	0.0013	0.0016	0.0017	0.0014	0.0039	0.0004	0.0022
	0.1 to 0.35	0.0275	0.0109	0.0137	0.0171	0.0193	0.0184	0.0173	0.0190	0.0086	0.0180
	0.36 to 1	0.0917	0.0727	0.0786	0.0855	0.0860	0.0870	0.0874	0.0857	0.0727	0.0846
	>1	0.1025	0.1394	0.1319	0.1230	0.1202	0.1202	0.1212	0.1188	0.1420	0.1220
Spring	<0.1 inch	0.0091	0.0010	0.0013	0.0016	0.0018	0.0020	0.0017	0.0046	0.0005	0.0026
	0.1 to 0.35	0.0387	0.0153	0.0193	0.0240	0.0272	0.0259	0.0243	0.0267	0.0121	0.0253
	0.36 to 1	0.0999	0.0793	0.0857	0.0932	0.0937	0.0948	0.0953	0.0934	0.0792	0.0922
	>1	0.1075	0.1463	0.1383	0.1290	0.1261	0.1261	0.1271	0.1246	0.1489	0.1280
Summer	<0.1 inch	0.0089	0.0010	0.0010	0.0015	0.0018	0.0020	0.0017	0.0045	0.0005	0.0026
	0.1 to 0.35	0.0366	0.0145	0.0183	0.0228	0.0258	0.0245	0.0230	0.0253	0.0115	0.0240
	0.36 to 1	0.0857	0.0680	0.0735	0.0799	0.0804	0.0813	0.0817	0.0801	0.0679	0.0791
	>1	0.1438	0.1955	0.1849	0.1724	0.1686	0.1685	0.1699	0.1665	0.1991	0.1711
Fall	<0.1 inch	0.0064	0.0007	0.0009	0.0011	0.0013	0.0014	0.0012	0.0032	0.0004	0.0018
	0.1 to 0.35	0.0248	0.0098	0.0124	0.0154	0.0174	0.0166	0.0156	0.0171	0.0078	0.0163
	0.36 to 1	0.0691	0.0548	0.0593	0.0645	0.0648	0.0656	0.0659	0.0646	0.0548	0.0638
	>1	0.1391	0.1892	0.1789	0.1668	0.1631	0.1631	0.1644	0.1611	0.1926	0.1655

The flow weighting factors in Table 54 were used with the statistical analyses of

the concentration data to obtain calculated long term averaged concentrations for mass

loading calculations. Table 55 shows the urban area concentrations developed for Anne

Arundel County using the Chesapeake Bay regional data contained in the National

Stormwater Quality Database, along with concentrations and runoff quantities for other

county land uses.

Table 55. Total Suspended Solids Concentrations for Land Use Categories in
Anne Arundel County, Maryland

Land Use Description	# of acres in 2000	TSS (mg/L)	Concentration reliability?	R_v	R_v reliability?
Large lot subdivision (1 unit/ 5- 10 ac)	0	60	excellent	0.09	excellent
Low-density residential (1 unit/ 5 acres to 2 units/acre)	33,337	60	excellent	0.14	excellent
Medium-density residential (2 to 8 units/acre)	33,791	60	excellent	0.23	excellent
High-density residential (8+ units/acre)	6,274	60	excellent	0.34	excellent
Commercial	11,670	58	excellent	0.72	excellent
Industrial	3,249	80	excellent	0.52	excellent
Institutional (schools, churches, military institutions, etc.)	9,813	58	moderate	0.49	excellent
Open urban land	4,139	50	moderate	0.08	excellent
Transportation	1,557	99	moderate	0.41	excellent
Extractive	1,686	350	poor	0.3	moderate
Deciduous forest	43,901	90	moderate	0.08	moderate
Evergreen forest	4,891	90	moderate	0.08	moderate
Mixed forest	56,621	90	moderate	0.08	moderate
Brush	2,565	90	poor	0.08	moderate
Wetlands	1,643	0	poor	0.65	moderate
Beaches	29	0	poor	0.1	moderate
Bare ground	224	1000	poor	0.3	moderate
Row and garden crops	300	357	very good	0.2	poor
Cropland	42,368	357	very good	0.2	poor
Orchards / vineyards / horticulture	63	357	very good	0.15	poor
Pasture	4,690	145	very good	0.08	moderate
Feeding operations	49	145	very good	0.2	poor
Agricultural building, breeding and training facilities	163	145	very good	0.5	poor

Urban land uses produced slightly lower TSS concentrations compared with those observed in forest areas. However the volumetric runoff coefficients for forests are smaller than any other use, except for open urban land, resulting in the likely lowest annual yields. Bare ground, cropland, vineyards, horticulture, row and garden crops and extractive activities have the highest estimated concentrations amongst the land uses examined. Total phosphorus concentrations are presented in Table 56.

Table 56. Total Phosphorus Concentrations for Land Use Categories in
Anne Arundel County, Maryland

Land Use Description	# of acres in 2000	TP (mg/L)	Concentration reliability?
Large lot subdivision (1 unit/5- 10 acres)	0	0.38	excellent
Low-density residential (1 unit/ 5 acres to 2 units/acre)	33,337	0.38	excellent
Medium-density residential (2 to 8 units/acre)	33,791	0.3	excellent
High-density residential (8+ units/acre)	6,274	0.3	excellent
Commercial	11,670	0.25	excellent
Industrial	3,249	0.23	excellent
Institutional (schools, churches, military institutions, etc.)	9,813	0.27	moderate
Open urban land	4,139	0.25	moderate
Transportation	1,557	0.25	moderate
Extractive	1,686	0.5	poor
Deciduous forest	43,901	0.1	moderate
Evergreen forest	4,891	0.1	moderate
Mixed forest	56,621	0.1	moderate
Brush	2,565	0.38	poor
Wetlands	1,643	0.38	poor
Beaches	29	0.1	poor
Bare ground	224	0.38	poor
Row and garden crops	300	1.00	very good
Cropland	42,368	1.00	very good
Orchards / vineyards / horticulture	63	1.00	very good
Pasture	4,690	0.38	very good
Feeding operations	49	0.38	very good
Agricultural building, breeding and training facilities	163	0.38	very good

In this case, the highest concentrations were assumed for croplands, row and garden crops, orchards, vineyards and horticulture. The lowest concentrations were assumed for forested areas. One order of magnitude separates the minimum and maximum concentrations. This difference can be associated with the use of fertilizers and associated nutrient discharges. For urban areas, industrial and commercial land use areas had lower phosphorus concentrations than residential land use areas. Table 57 shows the average urban area concentrations for long term analyses, based on statistical analyses examining site factors for this regional data. Only phosphorus had different concentrations associated with different site categories that were tested.

Table 57. Urban Areas Concentrations

Land Use	Constituent	Conditions	Average value for long-term analyses (mg/L)
Urban – Residential	TSS		60
	TP	Sites having <27% impervious cover (ultra low and low density areas)	0.38
		Sites having >27% impervious cover (medium and high density areas)	0.30
	TN		2.1
Urban – Commercial	TSS		58
	TP		0.25
	TN		2.6
Urban – Industrial	TSS		80
	TP		0.23
	TN		2.1

Table 58 shows the summary for total nitrogen. Similar to the total phosphorus case, the largest nitrogen concentrations were predicted for croplands, vineyards, row and garden crops orchards and horticulture activities. The lowest concentrations were

observed in open urban land and forested areas. The ratio between largest and smallest concentrations was approximately 2 to 1.

Table 58. Total Nitrogen Calculated Concentrations for Land Use Categories in Anne Arundel County, Maryland

Land Use Description	# of acres in 2000	TN (mg/L)	Concentration reliability?
Large lot subdivision (1 unit/5- 10 acres)	0	2.1	excellent
Low-density residential (1 unit/ 5 acres to 2 units/acre)	33,337	2.1	excellent
Medium-density residential (2 to 8 units/acre)	33,791	2.1	excellent
High-density residential (8+ units/acre)	6,274	2.1	excellent
Commercial	11,670	2.6	excellent
Industrial	3,249	2.1	excellent
Institutional (schools, churches, military institutions, etc.)	9,813	2	moderate
Open urban land	4,139	1.3	moderate
Transportation	1,557	2.3	moderate
Extractive	1,686	1.5	poor
Deciduous forest	43,901	1.5	moderate
Evergreen forest	4,891	1.5	moderate
Mixed forest	56,621	1.5	moderate
Brush	2,565	1.5	poor
Wetlands	1,643	1.5	poor
Beaches	29	1.5	poor
Bare ground	224	1.5	poor
Row and garden crops	300	2.92	very good
Cropland	42,368	2.92	very good
Orchards / vineyards / horticulture	63	2.92	very good
Pasture	4,690	2.2	very good
Feeding operations	49	2.2	very good
Agricultural building, breeding and training facilities	163	2.2	very good

Using the simple model, it is possible to calculate the total annual discharges from these different non point sources. Table 59 shows the total estimated runoff discharged by year, and the total discharges of suspended solids, total nitrogen and total phosphorus for each of the major land use categories. Urban sites produced most of the runoff and total nitrogen, followed by agricultural and forested areas. Half of the total suspended solids

were produced by agricultural activities, followed by urban areas (30%), forested areas

(12%), and other lands (10%). Urban and agricultural sites combined (in about equal

fractions) produced almost 90% of the phosphorus loads. Forested areas only produced

about 4% of the total phosphorus annual loads. The remaining phosphorus discharges

were produced by other land uses.

Table 59. Discharges by Major Land Use Categories in
Anne Arundel County, Maryland

	Runoff Yield		Total Suspended Solids	
	Total Discharge (ft^3/year)	Percent of Total	Total Discharge (kg/year)	Percent of Total
Urban	4.2×10^9	60.6	7.4×10^6	28.7
Agricultural	1.3×10^9	18.7	1.3×10^7	50.0
Forest	1.2×10^9	17.1	3.0×10^6	11.9
Other lands	2.6×10^8	3.7	2.4×10^6	9.4
Total County	7.0×10^9		2.6×10^7	

	Total Phosphorus		Total Nitrogen	
	Total Discharge (kg/year)	Percent of Total	Total Discharge (kg/year)	Percent of Total
Urban	3.6×10^4	46.4	2.6×10^5	60.5
Agricultural	3.6×10^4	45.8	1.1×10^5	25.0
Forest	3.4×10^3	4.3	5.1×10^4	11.9
Other lands	2.8×10^3	3.6	1.1×10^4	2.6
Total County	7.8×10^4		4.3×10^5	

The final values used during for the calculations are summarized in Tables 60 and

Table 61. For each of the main land uses, the percentage of impervious areas (indicating

the percentage connected and disconnected), the volumetric runoff coefficient and the

TSS, total phosphorus, and total nitrogen concentrations are shown. The volumetric

runoff coefficients, and curve numbers, were calculated using 50 years of precipitation

data from the BWI airport in Baltimore.

Table 60. Urban Land Use Categories Used in Anne Arundel County, Maryland

Description	Note	Average percentage of Impervious areas	TSS (mg/L)	Total P (mg/L)	Total N (mg/L)	Comments
High density residential	Rv = 0.34, 47% pervious, 39.9% dir con imp, and 13.1% dis con impervious	53	60	0.3	2.1	Rv and CN calculated using 50 yrs of BWI rains and concentration factors from MD and VA MS4 data
Medium density residential	Rv = 0.23, 62.3% pervious, 24.2% directly con imp, and 13.5% disconnected impervious	37.8	60	0.3	2.1	Rv and CN using 50 yr BWI rain and concentration factors from MD and VA MS4 data
Low density residential	Rv = 0.14, 79.6% pervious, 14.9% dir con impervious, and 5.5% disconnected imp.	20.4	60	0.38	2.1	1 unit/5 ac to 2 units/ac. Calc Rv and CN using 50 years BWI rains and concentration factors from MD and VA MS4
Ultra low den residential	Rv=0.09, 90.4% pervious, 5.6%directly con imp and 4% discon impervious	9.6	60	0.38	2.1	1 unit/5 to 10 ac, calc 50 yr, concentration factors from MD and VA MS4 data
Freeways and other main roads with paved drainage	Rv = 0.41, 49.5% pervious, 50.5% dir con impervious.	50.5	99	0.25	2.3	Calc using 50 yrs of BWI rains and concentration factors from national MS4 data
Commercial (shopping centers)	Rv=0.72, 8.3% pervious, and 91.7% dir con imp.	91.7	58	0.25	2.6	50 yr of BWI rains and concentration factors from MD and VA MS4 data
Institutional (schools, churches, military, etc.)	Rv=0.49, 36.4% pervious, 61.3%dir con imp, and 2.3%discon imp.	63.6	57.9	0.35	1.57	Calculated from 50 yr BWI rains and concentration factors from national average institutional MS4 data.
Industrial (medium)	Rv=0.52, 16.7% pervious, 62.8% dir con imp, and 20.5% discon con imp.	83.3	80	0.23	2.1	CN calc using 50 yr BWI rain and concentration factors from MD and VA MS4 data.
Open urban area	Rv=0.08, 95.1% pervious and 4.9% dir con impervious.	4.9	70	0.12	1.5	CN calc from 50 yr BWI rains and concentration factors from national average urban open area MS4 data

The land uses having the largest amounts of directly connected impervious surfaces were the commercial, institutional, and industrial land use areas. Urban TSS concentrations ranged between 57 and 99 mg/L, total phosphorus concentrations ranged

between 0.12 and 0.40 mg/L, and total nitrogen ranged between 1.5 and 2.6 mg/L. Table

61 shows the summaries for the other land uses.

Table 61. Other Land Use Categories Used in Anne Arundel County Calculations

Description	Note	TSS (mg/L)	Total Phosphorus (mg/L)	Total Nitrogen (mg/L)
Fallow	Straight Row. Concentration factors from prior regional data	107	1.3	4.4
Row Crops	Straight Row, small grain. Concentration factors from prior regional data	357	1	2.92
Row and garden Crops	Straight Row. Concentrations factors from prior regional data	357	1	2.92
Orchards, vineyards, horticultural	Concentration factors from prior regional data	357	1	2.92
Pasture or Range	Concentration factors from prior regional data	145	0.38	2.2
Feeding operations	Continuous forage, poor. Concentration factors from prior regional data	145	0.38	2.2
Woods or Forest Land	Deciduous forest (woods, good). Concentration factors from prior regional data	90	0.1	1.5
Woods or Forest Land	Evergreen forest (woods, good condition). Concentration factors from prior regional data.	90	0.1	1.5
Woods or Forest Land	Mixed forest (woods, good). Concentration factors from prior regional data.	90	0.1	1.5
Farmsteads	Agricultural buildings, breeding and training facilities	163	0.38	2.2
Brush	Herbaceous, fair.	90	0.38	1.5
Extractive		1000	0.38	1.5
Wetlands		0	0.38	1.5
Beaches		0	0.1	1.5
Bare ground		1000	0.38	1.5

The largest TSS concentrations were observed in extractive activities and for bare ground, or exposed soil sites. Land uses where the intensive use of fertilizers is most frequent had the largest total phosphorus and total nitrogen concentrations. The lowest nutrient concentrations were observed in forested areas.

Figure 60 shows the area distributions and the relative contributions for major sources of runoff, total suspended solids, total nitrogen and total phosphorus for sites located in Anne Arundel County in Maryland. Forested and urban land use areas represent almost 80% of the total land uses in the county. About 15% of the area is agricultural and the remaining of 5% is associated with other activities.

Urban land use areas produce almost 65% of the total runoff volume for the county, followed by agricultural and forested areas (about 15% each). As expected, impervious surfaces in urban land use areas were responsible for most of the total discharged runoff volume. Agricultural land uses produce almost half of the total TSS discharges, although they make up only about 15% of the county area. Urban land uses are the second major source of TSS in the county, contributing about 28% of the total annual TSS discharges in the county. Forested areas and other land uses contribute the smallest fractions of the total load, with almost 11% each. Urban and agricultural areas combined produced almost 90% of the total phosphorus load, in about equal percentages. Forested areas and other land uses contribute about 10% of the total countywide phosphorus load. Finally, urban land uses contributed almost 60% of the total nitrogen load for the county, followed by agricultural activities (25%), forested areas (13%) and other land uses (2%).
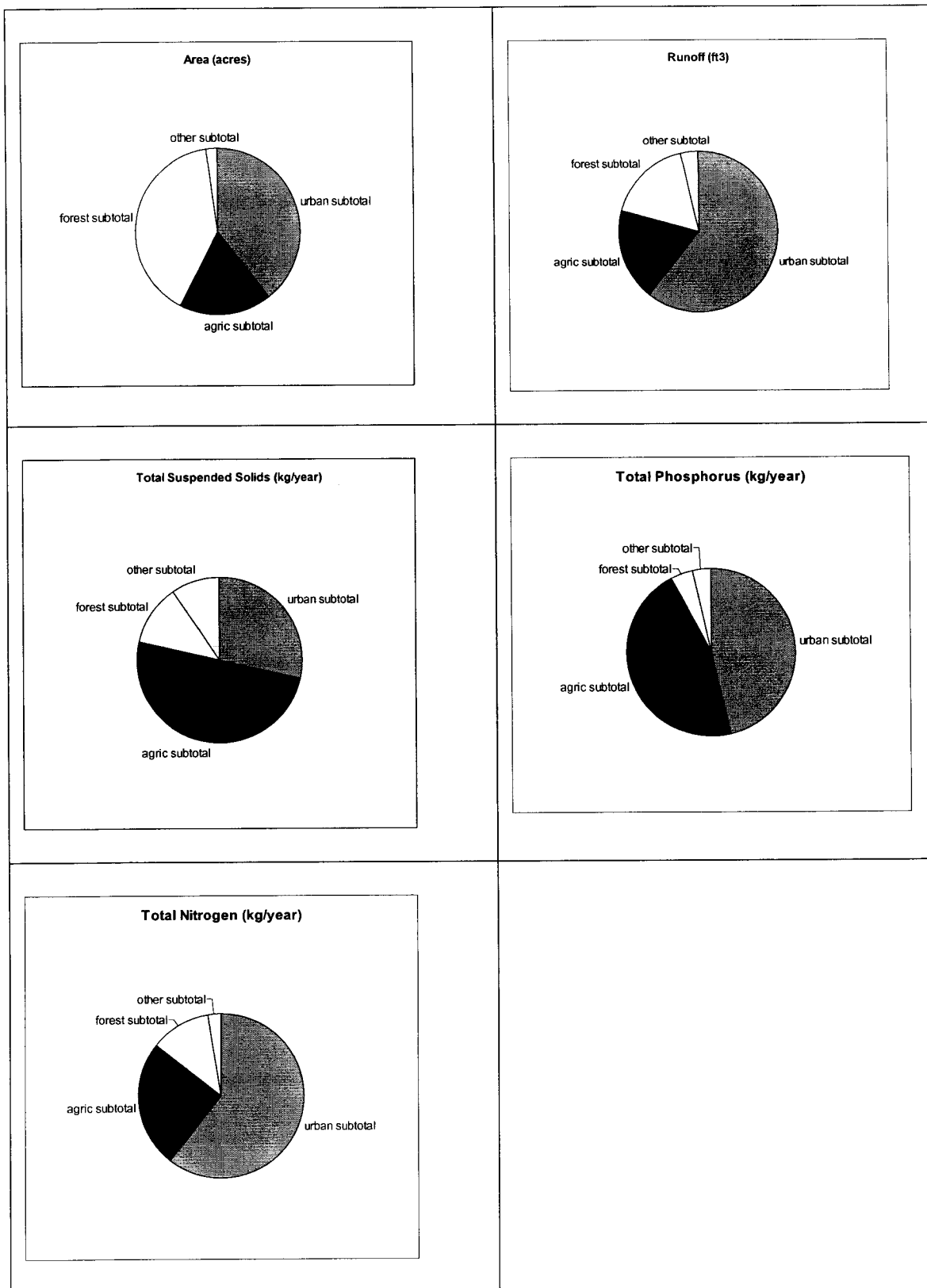
Fig. 60. Sources of runoff, TSS and nutrients for different sources in Anne Arundel

7.4 Statistical Analyses Performed

The following discussion describes the statistical analyses performed to identify the different groups in TSS, total nitrogen and total phosphorus by season and precipitations depth for the Chesapeake Bay region data. The objective of the statistical tests was to identify significantly distinct categories of the Chesapeake Bay regional data. Specifically, land use, season, precipitation, percent imperviousness, and watershed drainage area, were considered potentially important factors that would affect the concentration values. In addition, variations of reported concentrations with time were also examined. After appropriate normalization of the data, three-way and two-way ANOVA tests were used to identify the significant factors and interactions between these potentially important factors, while one-way ANOVA tests, along with parametric and nonparametric comparison tests, were used to identify groupings within the range of any one factor. As an example, one-way ANOVA analyses were performed to identify any ranges of percentage of imperviousness that produce different distributions of stormwater constituent concentrations from other ranges, while two-way ANOVA analyses were used to identify any seasonal-total precipitation interactions in the distribution of the stormwater constituent concentrations

Before ANOVA analyses can be conducted, the first step is to examine the data to ensure that it fits a normal probability distribution. If not, the data needs to be transformed. Prior tests (reported in this dissertation) found that most all of the stormwater constituents in the NSQD fit lognormal distributions. In this case, the base 10 logarithm of the original observations adequately followed a normal distribution. Therefore, data from the same population group will fall along the same straight line.

Groups in either tail that do not fall on the line can be considered different. This procedure was used in the ANOVA analyses to identify if the concentration values were statistically different for different levels of the factor, or factors, being examined. For example, if the expected values are different for different levels of imperviousness, or different seasons, then those data groupings will not follow the main probability distribution, and the ANOVA test results will indicate a likely significantly different data population. The significant ANOVA coefficients were then used to create a model to predict the concentration values for the different groups. All of the observed conditions within each group will have the same expected concentration value. Once the groups were identified, the mean and standard deviations were calculated from the original observations in the database for each observation in each group, and the data for each group are plotted on probability and box and whisker plots. The following discussion is a detailed description of the tests conducted using the Chesapeake data for total suspended solids.

7.4.1 Residential Area Total Suspended Solids Analyses

The ANOVA tests did not identify any significant groupings for either drainage area, or percentage imperviousness variations. Trends with time since the last rain, and for time since the initiation of the watershed monitoring were also examined, but these analyses did not identify any apparent, or significant, trends for any of the test sites. Initial data evaluations indicated a possible significant variation due to the level of imperviousness in the test watersheds, but when evaluated in conjunction with season and rain depth, these other factors were found to be the only significant factors to describe the

variations in TSS concentrations in residential areas. Obviously, the percentage

imperviousness values will have a large effect on the amount of runoff volume expected,

so the imperviousness will be very important in affecting the mass of pollutants

discharged. This is similar to data evaluations for other regions. The Maryland and

Virginia data provided a great opportunity to test this hypothesized effect, because there

were 13 residential area test watersheds having imperviousness ranging from 7 to 65%

(although most of the data were represented in six watersheds ranging in imperviousness

from 20 to 50%). The statistical tests identified two distinct groups of residential TSS

data, as represented in the following plots and tables: small summer rains (in the range of

0.1 to 0.35 inches) which had an average TSS concentration of about 143 mg/L, and all

other residential conditions which had an average TSS concentration of about 58 mg/L.

The following plots and data summaries describe these two data groupings.
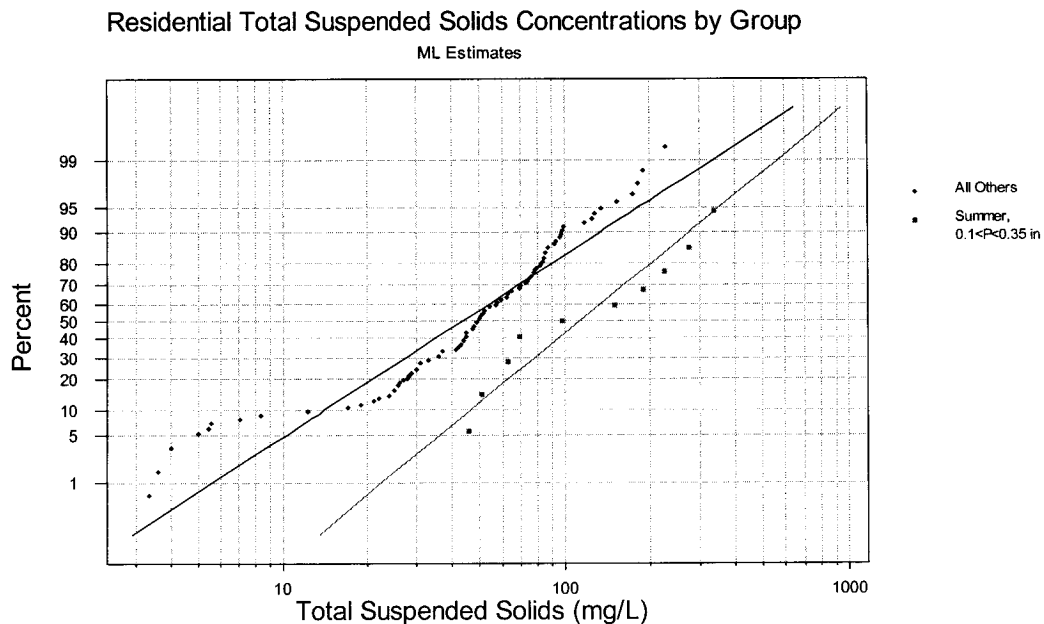
Fig. 61. Residential TSS distributions by groups

Residential Total Suspended Solids Concentrations by Groups
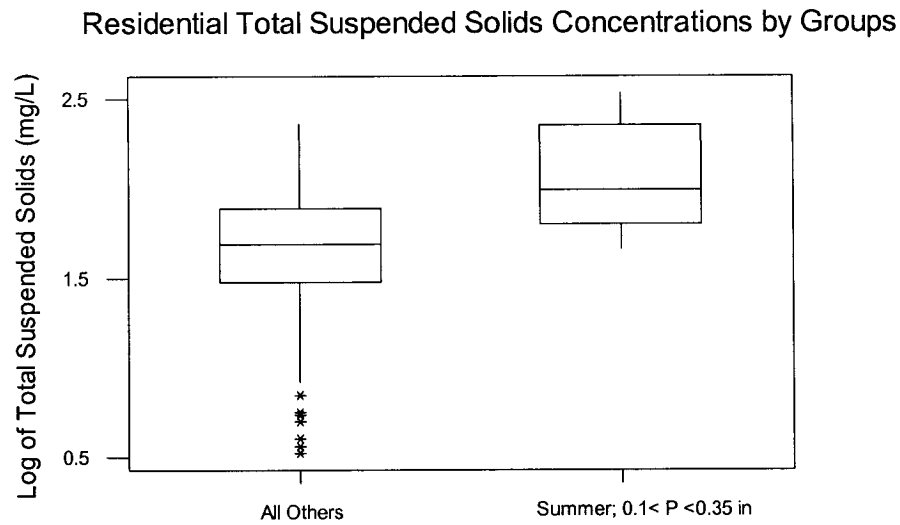
Fig. 62. Residential TSS box and whiskers plot distribution by groups

Table 62. Results for Residential TSS (mg/L)

Descriptive statistics of residential TSS (mg/L) by groups:					
Groups	N	Mean	Median	StDev	COV
All other rains	111	57.8	49.0	40.5	0.70
Small summer rains	11	143.0	98.0	101.6	0.71
Groups	SE Mean	Minimum	Maximum	Q1	Q3
All other rains	3.85	3.33	229.00	30.0	78.0
Small summer rains	30.6	46.0	337.0	63.0	227.0

Groups
Small summer rains (0.1 to 0.35 inches in depth):   Overall 95% confidence interval of all observed data: 29 to 439 mg/L (from fitted probability distribution)   95% confidence interval of reported median: 75 to 170 mg/L (from fitted probability distribution)   All other rains: Overall 95% confidence interval of all observed data: 8 to 243 mg/L (from fitted probability distribution)   95% confidence interval of reported median: 37 to 51 mg/L (from fitted probability distribution)

7.4.2 Residential Area Total Phosphorus Data Analyses

The statistical tests of the residential total phosphorus data indicated a significant

effect associated with the amount of imperviousness cover in the monitored watersheds.

Sites having small amounts of impervious cover (7 to 25%) had significantly higher total

phosphorus concentrations than sites having larger amounts of impervious cover (29 to

65%). Winter rains had lower total phosphorus concentrations in each group (all winter

rains in the first group, and small winter rains of less than 0.1 inch in the second group).

The following plots and data summaries describe these data groupings, separated by the

two impervious cover categories.
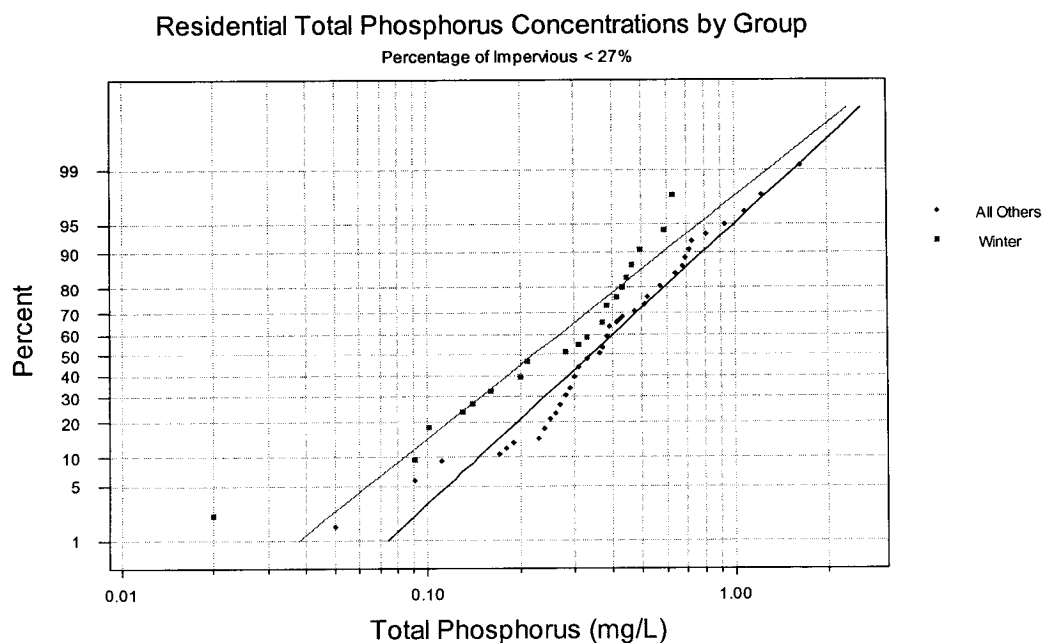
Fig. 63. Residential total phosphorus concentrations for sites having < 27% impervious surfaces

Residential Total Phosphorus Concentrations by Groups for Impervious Cover less than 27%
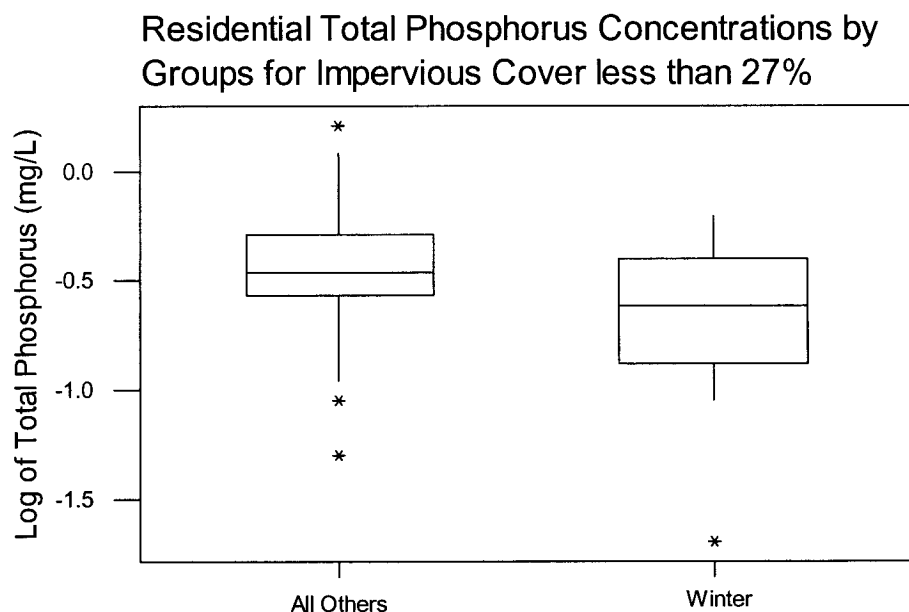
Fig. 64. Box and whiskers plot for residential total phosphorus concentrations for sites having < 27% impervious surfaces

Table 63. Results for Residential Total Phosphorus (Impervious < 27%)

Descriptive statistics of residential TSS (mg/L) by groups:					
Season	N	Mean	Median	StDev	COV
All other seasons	72	0.41	0.35	0.27	0.65
Winter	28	0.28	0.25	0.16	0.59
Season	SE Mean	Minimum	Maximum	Q1	Q3
All other seasons	0.031	0.05	1.62	0.27	0.52
Winter	0.031	0.02	0.63	0.13	0.40

Groups
All winter rains: Overall 95% confidence interval of all observed data: 0.050 to 0.98 mg/L (from fitted probability distribution) 95% confidence interval of reported median: 0.17 to 0.29 mg/L (from fitted probability distribution) All other seasons: Overall 95% confidence interval of all observed data: 0.093 to 1.23 mg/L (from fitted probability distribution) 95% confidence interval of reported median: 0.29 to 0.39 mg/L (from fitted probability distribution)
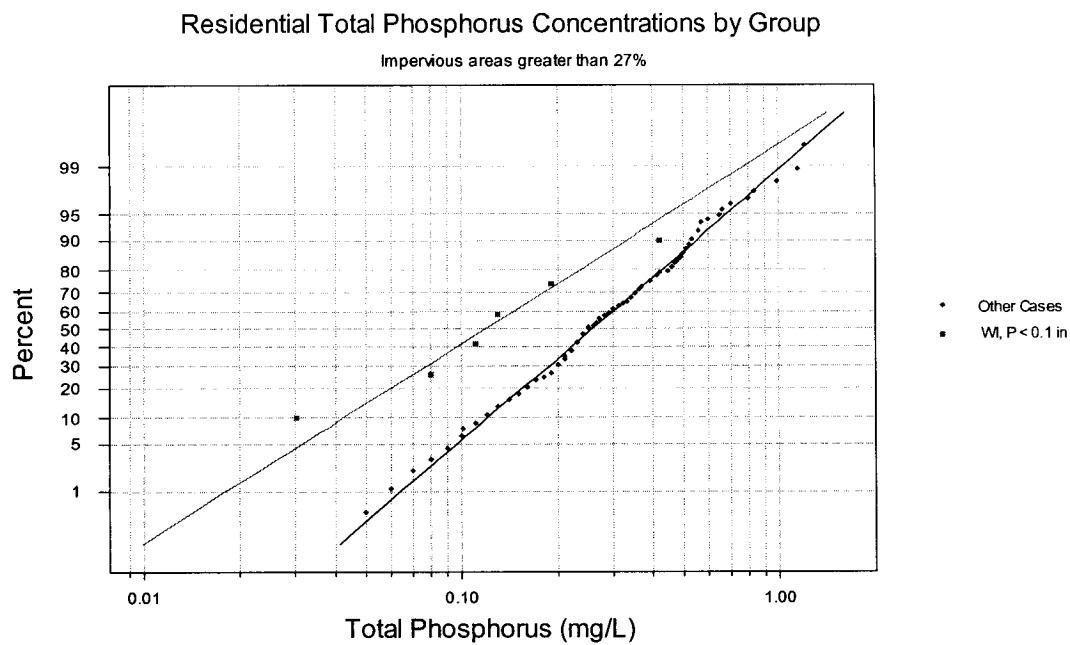
Fig. 65. Residential total phosphorus concentrations for sites having > 27% impervious surfaces
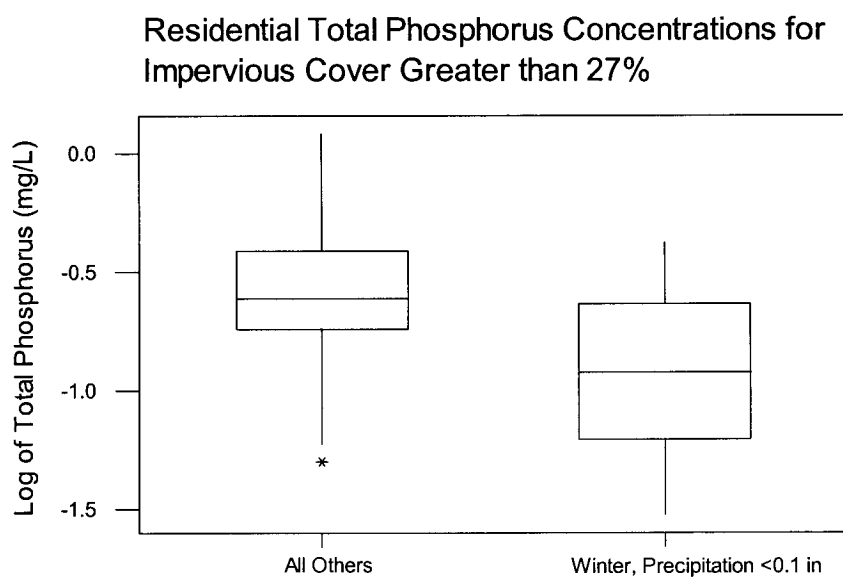
Fig. 66. Box and whiskers plot for residential total phosphorus concentrations for sites having > 27% impervious surfaces

Table 64. Results for Residential Total Phosphorus (Impervious > 27%)

Descriptive statistics of residential TSS (mg/L) by groups:					
Group	N	Mean	Median	StDev	COV
All others	152	0.30	0.24	0.19	0.63
Small winter rains	6	0.16	0.12	0.14	0.86
Group	SE Mean	Minimum	Maximum	Q1	Q3
All others	0.017	0.05	1.2	0.18	0.39
Small winter rains	0.057	0.03	0.42	0.068	0.25

Groups
Small winter rains <0.1 inch: Overall 95% confidence interval of all observed data: 0.024 to 0.58 mg/L (from fitted probability distribution)   95% confidence interval of reported median: 0.062 to 0.23 mg/L (from fitted probability distribution)   All other conditions: Overall 95% confidence interval of all observed data: 0.080 to 0.83 mg/L (from fitted probability distribution)   95% confidence interval of reported median: 0.23 to 0.28 mg/L (from fitted probability distribution)

7.4.3 Residential Area Total Nitrogen Data Analyses

The statistical analysis of the residential total nitrogen data identified several important interactions between season and rain depth. There were no significant factors associated with drainage area, percent imperviousness, or trend with time. Five significant groups were identified for residential total nitrogen concentrations:

1) Fall rains <0.1 and > 1 inch

2) Winter rains between 0.35 and 1 inch

3) Fall for rains between 0.35 and 1 inch, and winter rains between 0.1 and 0.35 inches

4) All other conditions

5) Spring and summer rains between 0.1 and 0.35 inches

The following plots and data summaries describe these five data groupings for residential area total nitrogen concentrations.
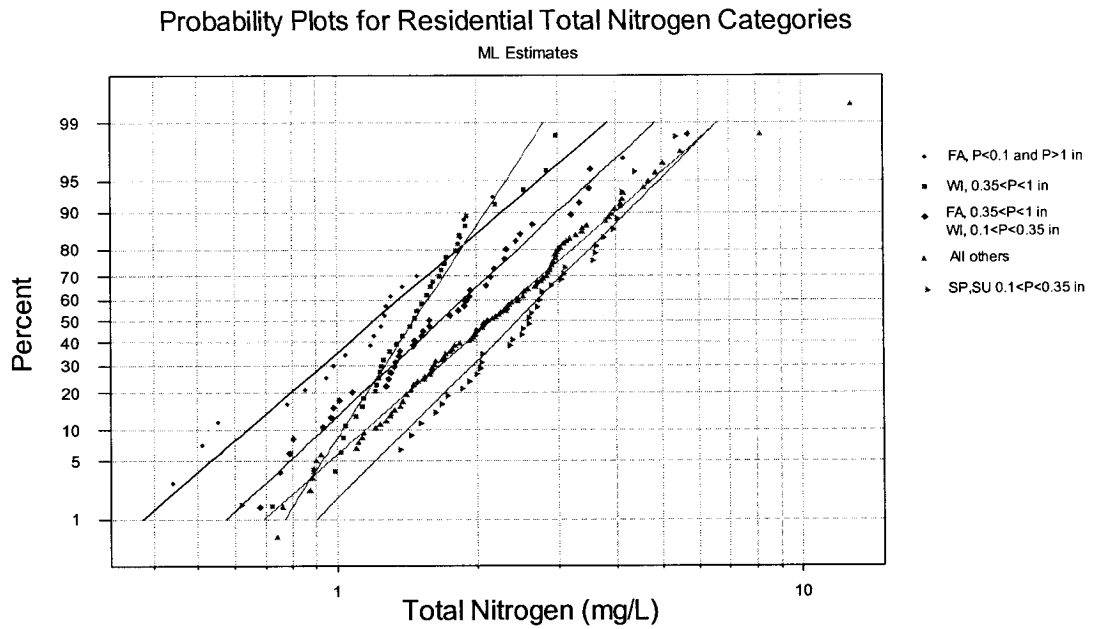
Fig. 67. Residential total nitrogen concentration groups
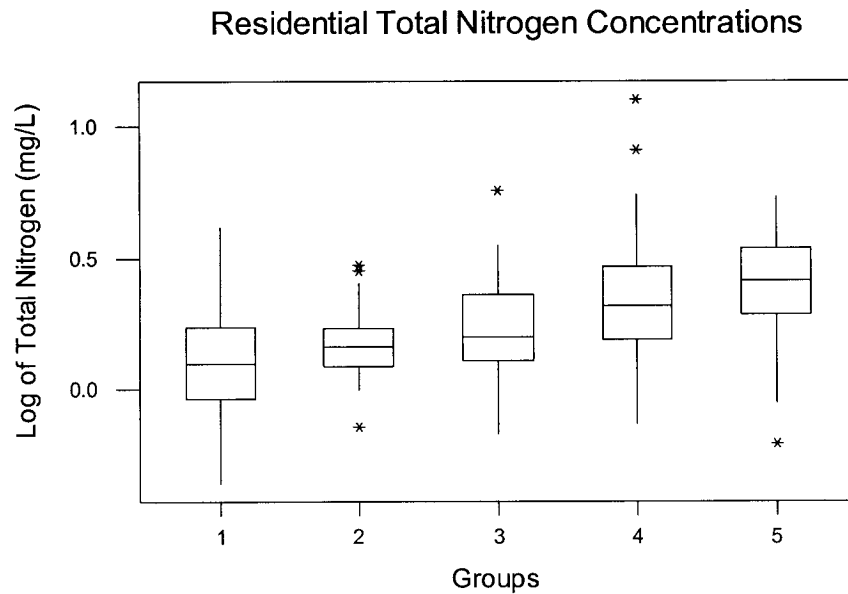
Fig. 68. Box and whiskers plot for residential total nitrogen

Table 65. Results for Residential Total Nitrogen

Descriptive statistics of residential TSS (mg/L) by groups:					
Groups	N	Mean	Median	StDev	COV
Fall, P<0.1 and P>1	22	1.4	1.3	0.77	0.57
Winter, 0.35<P<1	42	1.5	1.5	0.46	0.30
Fall, 0.35<P<1;					
WI, 0.1<P<0.35	43	1.9	1.6	0.95	0.51
All other conditions	112	2.4	2.1	1.5	0.62
Sp and Su, 0.1<P<0.35	40	2.6	2.6	1.0	0.38

Groups	SE Mean	Minimum	Maximum	Q1	Q3
Fall, P<0.1 and P>1	0.17	0.44	4.1	0.93	1.7
Winter, 0.35<P<1	0.072	0.72	3.0	1.23	1.7
Fall, 0.35<P<1;					
WI, 0.1<P<0.35	0.14	0.68	5.7	1.3	2.3
All other conditions	0.14	0.74	13	1.5	2.9
Sp and Su, 0.1<P<0.35	0.16	0.62	5.4	1.9	3.4

Groups

Fall, rains <0.1 inches and rains >1 inch:
Overall 95% confidence interval of all observed data: 0.45 to 3.2 mg/L (from fitted probability distribution)
95% confidence interval of reported median: 0.98 to 1.5 mg/L (from fitted probability distribution)

Winter, rains between 0.35 and 1 inch:
Overall 95% confidence interval of all observed data: 0.85 to 2.5 mg/L (from fitted probability distribution)
95% confidence interval of reported median: 1.3 to 1.6 mg/L (from fitted probability distribution)

Fall, rains between 0.35 and 1 inch; and Winter rains between 0.1 and 0.35 inches:
Overall 95% confidence interval of all observed data: 0.68 to 4.1 mg/L (from fitted probability distribution)
95% confidence interval of reported median: 1.5 to 1.9 mg/L (from fitted probability distribution)

Spring and Summer rains between 0.1 and 0.35 inches:
Overall 95% confidence interval of all observed data: 1.1 to 5.7 mg/L (from fitted probability distribution)
95% confidence interval of reported median: 2.1 to 2.8 mg/L (from fitted probability distribution)

All other conditions:
Overall 95% confidence interval of all observed data: 0.82 to 5.9 mg/L (from fitted probability distribution)
95% confidence interval of reported median: 2.0 to 2.3 mg/L (from fitted probability distribution)

7.4.4 Commercial Area Total Suspended Solids Analyses

The commercial area total solids data appears to be affected by season and rain depth interactions, plus season and rain depth main factors. No affects associated with drainage area, or trends with time or interevent period, were identified.

7.4.5 Commercial Area Total Phosphorus Analyses

The commercial area total phosphorus data appears to be affected by season and rain depth interactions, plus season main factors. No affects associated with drainage area, or trends with time or interevent period, were identified.

7.4.6 Commercial Area Total Nitrogen Analyses

The commercial area total nitrogen data appears to be affected by season and rain depth interactions alone. No affects associated with drainage area, or trends with time or interevent period, were identified.

7.4.7 Industrial Area Total Suspended Solids Analyses

The industrial area total suspended solids data appears to be affected by season main factors alone. No affects associated with rain depth, drainage area, or trends with time or interevent period, were identified.

7.4.8 Industrial Area Total Phosphorus Analyses

The industrial area total phosphorus solids data appears to be affected by rain depth main factors alone. No affects associated with season, drainage area, or trends with time or interevent period, were identified.

7.4.9 Industrial Area Total Nitrogen Analyses

The commercial area total nitrogen solids data does not appear to be affected by any of the factors, or interactions examined. No affects associated with rain depth, season, drainage area, or trends with time or interevent period, were identified.

7.5 Summary

In this chapter, the NSQD was used to estimate the expected total suspended solids and nutrient mass discharges from urban, agricultural and forested sources in Anne Arundel County, Maryland, in a typical year. The parameters used in Schueler's simple method are the mean concentrations from each of these sources, the areas associated with each source, the volumetric runoff coefficient for these sources, and the rain depth associated with the period of calculation. The NSQD includes several catchments and more than 1,000 storm events in the Chesapeake Bay area which were used to determine the most appropriate urban area mean stormwater concentrations for residential, commercial and industrial land uses.

The effects associated with different seasons and rain depths on the urban area concentrations of solids and nutrients were also addressed for these regional data in this chapter. ANOVA analyses indicated that there are significant differences in the

concentrations according to the seasonal period when the samples were collected, the total precipitation depth, and the interaction between these two factors for some of these pollutants and urban land uses. A stronger influence of the interactions between these factors was observed in residential areas compared with commercial or industrial land use areas.

The data summaries indicated that solids concentrations from forested and urban areas are similar, however the total runoff volume produced in forest areas is very small compared with the urban areas. For this reason, annual mass discharges from forested areas are less than half of the annual mass discharges produced from urban areas, even though the areas for these two main land use categories are similar.

Annual agricultural mass discharges of suspended solids are almost twice those calculated for urban areas. In urban areas, lower TSS concentrations occur, but a much larger fraction of the precipitation is transformed to runoff. Total urban area nitrogen mass discharges are expected to be almost twice the loads discharged from agricultural areas.

Chapter 8

Conclusions

8.1 Introduction

The purpose of this dissertation was to examine several commonly accepted assumptions concerning stormwater characteristics (and associated management decisions) by stormwater managers and researchers. These included assumptions relating to the existence of "first flushes"; the effect of the abundance of impervious areas and the length of antecedent dry period on stormwater constituent concentrations; the influences of non-detected observations on stormwater characteristics; among others. These assumptions were evaluated using information contained in the National Stormwater Quality Database (NSQD). More than 3765 events were monitored at 360 sites throughout the U.S. and the monitored water quality and associated information was included in the first version of the database. Most of the data was collected from residential, commercial and industrial land use areas in the eastern part of the U.S. (according to the original study design), although most geographical areas are represented. The evaluation results from investigating each of the dissertation hypotheses presented in Chapter 1 are presented in the following discussions.

232

8.2 Dissertation Research Hypotheses

The main hypothesis for this dissertation was that commonly accepted

assumptions concerning stormwater characteristics are correct and applicable for a wide

range of conditions, including different land uses, precipitation depths, seasons,

watershed area and geographic locations throughout the U.S. This assumption was

evaluated by testing the following hypotheses:

8.2.1 Research Hypothesis 1. Lognormal distributions are robust descriptions of

stormwater quality data and a few unusual values have little effect on dataset summary

statistical descriptions

- A total of 25 constituents in 5 land uses were evaluated using the NSQD

database. In 71% of the cases, lognormal distributions better described

the stormwater constituent concentrations compared with gamma and

exponential distributions. These last two distributions better represented

10% and 4% of the cases, respectively. In 15% of the cases, lognormal,

gamma and exponential distributions did not adequately represent

stormwater constituent concentrations. Constituents that mostly were not

well described by any of these three distributions included: BOD_5, total

arsenic, total cadmium and total copper in residential, commercial and

industrial land uses.

- Gamma and exponential distributions better described bacteria and

nutrient concentrations in open space land use areas.

- The use of the 3-parameter lognormal distribution, did not improve the description information compared with the simpler 2-parameter distribution. The 2-parameter lognormal distribution is therefore recommended for those constituents were the use of lognormal distributions produced a better fit of the data.

- Unusually elevated values have a significant effect in the mean, median, standard deviation and coefficient of variation of the sample distribution. When 0.5% of the data is affected by a factor of a thousand (such as may occur when heavy metals are incorrectly expressed with mg/L units when they should be µg/L units), the coefficient of variation will be increased almost 15 times compared to the value when the extreme observations are not present. The effect on the coefficient of variation is larger as the percentage of extreme samples is reduced.

- Unusual low values do not have a significant effect on the mean, median, standard deviation, or coefficient of variation, unless the percentage of samples having the low values is higher than 25%.

8.2.2 Research Hypothesis 2. Censored data can be adequately adjusted by substituting half of the detection limit, with little resulting effects on the mean and variance of stormwater datasets.

- Replacing non-detected observations by half of the detection limit is appropriate when the percentage of left censored observations (those having concentrations lower than the detection limit) is lower than 15%

of the total data set. Replacing the non-detected values with zero will have more extreme effects on these distribution summary values.

- Ignoring the non-detected values will result in higher means, medians and standard deviations, and lower coefficients of variation than the true values for the distributions.

- The use of the Cohen's maximum likelihood method is recommended to replace the censored observations for those constituents that have lognormal distributions. This is an appropriate method when non-paired statistical tests are to be performed, as the assignment of replacement values for specific tests in not important. However, no replacements are suitable when paired comparison tests are to be conducted, and these tests should only be conducted on the data sets having complete pairs (not using pairs where one or both parts of the pair are below detection). When calculating percentage reductions, or other comparison tests, non-detected effluent concentrations can be used, without substitutions, in calculations to determine the lower limit of removal.

- When the number of non-detected observations exceed about 40% of the total number of observations, no substitution method (the maximum likelihood method or half of the detection limit) is suitable.

8.2.3 Research Hypothesis 3. Different levels of imperviousness are more important than differences in land use categories when predicting stormwater constituent concentrations.

- The use of the impervious area information alone did not reduce the uncertainty about the variability of stormwater constituents. One of the main factors associated with land use concerns the activities that occur in the land use. It is expected that the use of both factors (land use and information about the surface covers in the area, such as the percentage of impervious areas) will reduce the variability of the stormwater concentrations observed, rather than when only one of these factors is considered. However, these tests were only conducted on stormwater concentrations, not on mass discharges. Increases in impervious cover are directly associated with increases in runoff volumes, and therefore in pollutant mass discharges.

- When only residential area data from EPA Rain Zone 2 were used, the percentage of impervious areas was found to have a significant effect on the concentration of nitrates. The concentrations of nitrates were reduced as the percentage of impervious cover increased. This is an expected finding; when the impervious areas increase, less landscaping is likely (a major source of nutrient discharges).

- No significant relationships were observed between the amount of impervious cover and any stormwater constituent concentration that was examined for industrial and commercial land use areas.

8.2.4 Research Hypothesis 4. Antecedent dry periods have a significant effect on stormwater constituent concentrations.

- Antecedent dry periods are not the same for all the EPA Rain Zones in the country. Longer antecedent dry periods occur at west coast sites compared to other locations.

- The antecedent dry periods had a positive significant ($\alpha = 5\%$) effect in the concentration of 7 of 13 constituents examined: nutrients (ammonia, nitrates, TKN, total and dissolved phosphorus), COD and BOD_5 at residential sites located in EPA Rain Zone 2. It was not significant in oil and grease, TDS, TSS, total copper, total lead and total zinc.

- Only total phosphorus and total lead concentrations were affected by the antecedent dry period at commercial sites located in EPA Rain Zone 2. Total phosphorus concentrations increased with increasing days before the sampled storm. An opposite relation was observed for total lead at commercial sites.

- Only TSS was affected by the antecedent dry period at industrial sites located in EPA Rain Zone 2. A positive relationship was observed, with TSS concentrations increasing as the number of antecedent dry days increased.

8.2.4 Research Hypothesis 5. Outfall samples collected during the "first flush" periods of storms have significantly greater concentrations than total storm composite samples.

- The first flush effect was not present for all constituents and all land uses. The phenomenon was most likely to occur in commercial and high density residential land uses, watersheds having high percentages of impervious areas. It was not observed in open space areas, watersheds having low percentages of impervious areas.

- TSS, COD, TDS, total copper, total lead, total zinc and TKN had observed flush concentrations that were significantly higher than the composite sample concentrations in those areas where the "first flush" was most likely to occur. pH was the only constituent that did not indicate a first flush effect. Observed elevated first flush concentrations were less than 3 times higher than the corresponding storm composite concentrations.

8.3 Recommendations for Future Stormwater Permit Monitoring Activities

The NSQD is an important tool for the analysis of stormwater discharges at outfalls. About a third of the total existing information from the NPDES Phase I program is included in the database. Most of the analyses in this dissertation were performed for residential, commercial and industrial land uses in EPA Rain Zone 2 (the area of emphasis according to the terms of the EPA funded research). Much more data are available from other stormwater permit holders that were not included in this database. Acquiring this additional data for inclusion in the NSQD is a recommended and cost-

effective activity and should be accomplished as additional data are also being obtained from on-going monitoring projects.

The use of automatic samplers, coupled with bedload samplers, is preferred over manual sampling procedures. In addition, flow monitoring needs to be included as part of all stormwater characterization monitoring. The additional information associated with flow data will greatly enhance the usefulness of the much more expensive water quality monitoring. Flow monitoring must also be correctly conducted, with adequate verification and correct base-flow subtraction methods applied. A related issue frequently mentioned by the monitoring agencies is the lack of on-site rainfall information for many of the sites. Using regional rainfall data from locations distant from the monitoring location is likely to be a major source of error when rainfall factors are being investigated.

Many of the stormwater permits also only required monitoring during the first three hours of the rain event. This may have influenced the event mean concentrations if the rain event continued much beyond this time. Monitoring should continue for the complete rain duration. Monitoring only three events per year from each monitoring location requires many years before statistically adequate numbers of observations are obtained. In addition, it is much more difficult to ensure that such a small fraction of the total number of annual events is representative. Also, there is minimal value in obtaining continued data from an area after sufficient information is obtained. It is recommended that a more concentrated monitoring program to be conducted for a two or three periods, with a total of about 30 events monitored for each site. Periodic checks can be made in

future years, such as repeating concentrated monitored every 10 years, or so (and for only 15 events during the follow-up surveys).

Finally, better watershed area descriptions, especially accurate drainage area delineations, are needed for all monitored sites. While the data contained in the NSQD is extremely useful, it is believed that future monitoring information obtained as part of the stormwater permit program would be greatly enhanced with these additional considerations.

During the analyses of the NSQD data, a number of areas where further research would be useful were identified. Some of these are listed below:

- Ammonia concentrations were higher at the freeway sites than in any of the other land use areas. Ammonia may therefore be related to the emissions of nitrogen compounds from automobile exhaust. However, no literature information was found relating nutrient sources to automobile use. The role of atmospheric sources on stormwater quality is a general area of research need.

- Another important research need is the role of bed load and sediment transport in urban areas. Current sampling methods, that do not consider bed load sampling, may not be accurate. However, most bed load material is not expected to travel long distances in conventional drainage systems. The significance of this material in urban drainage needs additional research, along with investigations of the most appropriate sampling methods for this coarse material.

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241

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Appendix A
Sites Included in the Database

A1. Description

The following table shows the number of samples, land use and community for each site.

Table A1. Site Name and Land Use

State	LOCATION ID	Land use	Jurisdiction	Site ID
AL	ALHUCHIP	ID	City of Huntsville	Chase Industrial Park
AL	ALHUDRAV	RE	City of Huntsville	Drake Avenue
AL	ALHUHURI	RE	City of Huntsville	Hunters Ridge
AL	ALHUMASM	CO	City of Huntsville	Madison Square Mall
AL	ALHUWERP	CO	City of Huntsville	Western Research Park
AL	ALJC004L	CO FW ID RE IS	Jefferson County	C004L
AL	ALJC004R	CO FW ID RE IS	Jefferson County	C004R
AL	ALJCC001	FW ID CO	Jefferson County	C001
AL	ALJCC002	RE OP ID CO	Jefferson County	C002
AL	ALJCC009	RE FW	Jefferson County	C009
AL	ALJCC010	RE FW	Jefferson County	C010
AL	ALJCC012	CO FW RE OP	Jefferson County	C012
AL	ALMOCREO	RE	City of Mobile	Creola
AL	ALMODAPH	CO	City of Mobile	Daphne
AL	ALMOSARA	RE	City of Mobile	Saraland
AL	ALMOSIIV	ID	City of Mobile	Mobile Site IV
AL	ALMOSITV	CO	City of Mobile	Mobile Site V
AL	ALMOSIVI	RE	City of Mobile	Mobile Site VI
AL	ALMOTHEO	ID	City of Mobile	Theodore
AZ	AZMCA001	ID	Maricopa Cnty	48th Street Drain
AZ	AZMCA002	OP	Maricopa Cnty	South Mountain Park
AZ	AZMCA003	ID	Maricopa Cnty	27th Ave at Salt River
AZ	AZMCA004	RE CO	Maricopa Cnty	Aqua Fria at Youngtown
AZ	AZMCA005	CO	Maricopa Cnty	43rd Ave at Peoria
AZ	AZMCA006	RE	Maricopa Cnty	67th Ave Olive Ave at Glendale
AZ	AZTUA001	RE	Tucson	Grant Road and Wilson Avenue
AZ	AZTUA002	RE	Tucson	Greenlee Road

246

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
AZ	AZTUA003	CO	Tucson	El Con Mall
AZ	AZTUA004	ID	Tucson	17th Street
CA	CAALAL03	ID RE	Alameda County	Woods Street
CA	CAALAL04	RE CO FW	Alameda County	Alice Street and 4th
CA	CAALAL07	CO RE	Alameda County	Cotter Way
CA	CAALAL09	ID	Alameda County	Pacific Street
CA	CAALAL10	ID CO RE	Alameda County	37TH ST 8TH AVE
CA	CACTA001	FW	Caltrans	3 07 Sacramento
CA	CACTA002	FW	Caltrans	4 35 Solano
CA	CACTA003	FW	Caltrans	6 205 Fresno
CA	CACTA004	FW	Caltrans	6 209 Fresno
CA	CACTA005	FW	Caltrans	7 01 Los Angeles
CA	CACTA006	FW	Caltrans	7 127 Los Angeles
CA	CACTA007	FW	Caltrans	7 128 Los Angeles
CA	CACTA008	FW	Caltrans	7 201 Los Angeles
CA	CACTA009	FW	Caltrans	7 202 Los Angeles
CA	CACTA010	FW	Caltrans	7 203 Los Angeles
CA	CACTA011	FW	Caltrans	8 01 Riverside
CA	CACTA012	FW	Caltrans	8 02 San Bernandino
CA	CACTA013	FW	Caltrans	8 03 Riverside
CA	CACTA014	FW	Caltrans	12 01 Orange
CA	CACTA015	FW	Caltrans	12 02 Orange
CO	COCSA001	CO OP RE	Colorado Springs	Sixteenth Hole Valley Hi Golf Course
CO	COCSA002	ID OP	Colorado Springs	Chestnut Street at Douglas Creek
CO	COCSA003	ID CO	Colorado Springs	Beacon Street at Buchanan Street
CO	COCSA004	RE OP	Colorado Springs	Wahsatch Street at Cross Lane
CO	COCSA005	OP CO ID	Colorado Springs	Walmart at Eighth Street
CO	CODEA001	CO	Denver Metro	Cherry Creek Storm Drain at Colfax Ave
CO	CODEA002	CO	Denver Metro	Cherry Creek Storm Drain at University Blvd
CO	CODEA003	RE	Denver Metro	North Sanderson Gulch Tributary at Lakewood
CO	CODEA004	ID	Denver Metro	Sand Creek Tributary at 34th and Havana
CO	CODEA005	RE	Denver Metro	Shop Creek at Parker Road
CO	CODEA006	ID	Denver Metro	South Platte River Storm Drain at 54th and Steele
CO	CODEA007	ID	Denver Metro	South Platte River Storm Drain at 7th Ave
CO	CODEA008	CO	Denver Metro	Villa Italia Storm Drain at Lakewood
GA	GAATAT01	ID	City of Atlanta	Ellsworth Industrial Drive
GA	GAATAT02	RE	City of Atlanta	Beverly Road Doncaster Drive
GA	GACLCOSI	ID	Clayton County	Southridge Industrial Park
GA	GACLCOTR	RE	Clayton County	Tara Road

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
GA	GACOC1A2	RE CO	Cobb County	Cobb Long Term 1 Pebble Creek Lot989
GA	GACOC1A3	RE	Cobb County	Cobb Long Term 1 Sewell Mill Creek Roswell Road
GA	GACOCOL2	RE CO	Cobb County	Cobb Long Term 2 Worley Rd Noonday Creek
GA	GADKCOTD	ID CO	Dekalb County	Truman Drive
GA	GAFUCOS1	RE OP	Fulton County	Johns Creek Buice Road
GA	GAFUCOS2	ID OP	Fulton County	Boat Road Blvd Grange Blvd
GA	GAFUCOS3	RE CO	Fulton County	Long Island Creek Northside Drive
ID	IDADA001	CO ID	Ada County Highway District	Koppels Site
ID	IDADA002	RE	Ada County Highway District	Lucky Drive Site
ID	IDADA003	RE FW	Ada County Highway District	Franklin Road Site
ID	IDADA004	CO ID	Ada County Highway District	Production Avenue Site
KA	KATOATWO	RE	City of Topeka	Atwood
KA	KATOBROO	RE	City of Topeka	Brookfield
KA	KATOJACK	CO	City of Topeka	Jackson
KA	KATOSTFE	ID	City of Topeka	SantaFe
KA	KAWIHUNT	RE	City of Wichita	Huntington
KA	KAWIMCLE	ID	City of Wichita	McLean
KA	KAWISBWY	RE CO	City of Wichita	Broadway
KA	KAWITOWN	CO	City of Wichita	Towne East
KY	KYLOTSR1	RE	City of Louisville	Buechel
KY	KYLOTSR2	ID	City of Louisville	Obannon
KY	KYLOTSR3	RE	City of Louisville	St Matthews
KY	KYLOTSR4	ID	City of Louisville	Okolona
KY	KYLOTSR5	RE CO	City of Louisville	Pleasure Ridge Park
KY	KYLOTSR6	RE CO	City of Louisville	Hurstbourne Acres
KY	KYLXEHL4	OP	City of Lexington	Lakeside golf
KY	KYLXEHL5	OP	City of Lexington	Walnut Hill Chilesburg
KY	KYLXEHL6	FW	City of Lexington	Alumni ManOwar
KY	KYLXEHL7	RE	City of Lexington	Squires Road
KY	KYLXNEL1	RE	City of Lexington	Greenbrier East
KY	KYLXNEL2	RE OP	City of Lexington	Greenbrier
KY	KYLXNEL3	CO	City of Lexington	Eastland
KY	KYLXTBL1	RE	City of Lexington	Mt Vernon
KY	KYLXTBL2	ID	City of Lexington	Leestown
KY	KYLXTBL3	OP	City of Lexington	Viley Road
KY	KYLXWHL1	CO	City of Lexington	Wilhite Drive
MA	MABOA001	RE OP	Boston	Charlestown 29J212
MA	MABOA002	RE	Boston	West Roxebury 13D077 078
MA	MABOA003	CO	Boston	Dorchester 8J102

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
MA	MABOA004	ID	Boston	Brighton 25E037
MA	MABOA005	FW ID	Boston	Hyde Park 2F120
MA	MABOA006	RE	Boston	Mount Vernon 26K099
MA	MABOA007	OP	Boston	Wesley G Ross 6G108
MD	MDAACOMW	ID	Anne Arundel	Midway industrial park MW
MD	MDAACOOD	RE	Anne Arundel County	Odenton OD
MD	MDAACOPP	CO	Anne Arundel	Parole Plaza PP
MD	MDAACORK	RE	Anne Arundel	Rolling Knolls RK
MD	MDAACOSC	CO ID	Anne Arundel	Science Drive SC
MD	MDBACOBC	ID	Baltimore County	Brien Run BC
MD	MDBACOLC	CO RE OP	Baltimore County	Long Quarter Branch LC
MD	MDBACOSC	RE	Baltimore County	Spring Branch SC
MD	MDBACOTC	ID	Baltimore County	Tobasoo creek TC
MD	MDBACOWC	RE	Baltimore County	White Marsh Run WC
MD	MDBCTYBO	ID	Baltimore City	BO
MD	MDBCTYFM	ID	Baltimore City	FM
MD	MDBCTYHA	RE CO	Baltimore City	Hamilton HA
MD	MDBCTYHO	RE	Baltimore City	Home land HO
MD	MDBCTYHR	RE	Baltimore City	Herring Run HR
MD	MDBCTYKO	CO	Baltimore City	Coppers Avenue KO
MD	MDCHCOIP	ID	Charles County	IP
MD	MDCHCOPA	RE	Charles County	PA
MD	MDCHCOPF	RE	Charles County	PF
MD	MDCHCOTG	CO ID	Charles County	TG
MD	MDCLCOBP	CO ID	Carroll County	Route 97 airportindustrial BP
MD	MDCLCOCE	RE	Carroll County	Candice estates CE
MD	MDCLCOJS	CO	Carroll County	John street JS
MD	MDCLCOKW	OP RE	Carroll County	Kate wagner KW
MD	MDCLCOSD	RE ID	Carroll County	Sunset Drive SD
MD	MDHACFBA	XX	Harford County	FBA
MD	MDHACOBP	RE	Harford County	Brentwood Park Woodland Hills
MD	MDHACOCF	CO	Harford County	Constant Friendship CF
MD	MDHACOCS	RE	Harford County	Cool Spring CS
MD	MDHACOGR	RE	Harford County	Green Ridge-II GR
MD	MDHACOIC	ID	Harford County	Greater harford industrial centre IC
MD	MDHOCODC	CO	Howard County	Dobbin center DC
MD	MDHOCOFM	ID	Howard County	Food market FM
MD	MDHOCOGM	RE	Howard County	Green Moon GM
MD	MDHOCOMH	RE	Howard County	Murray Hill MH
MD	MDHOCOOC	ID	Howard County	Oak land centre OC
MD	MDMOCOBC	CO	Montgomery County	Burtons ville crossing BC
MD	MDMOCOCV	ID	Montgomery County	Coles villeCV

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
MD	MDMOCONV	RE	Montgomery County	Venture V
MD	MDMOCOQA	RE	Montgomery County	Quaint Acres QA
MD	MDMOCOSL	ID	Montgomery County	Southlawn lane SL
MD	MDMOCOWP	CO	Montgomery County	Wheaten plaza WP
MD	MDPGCOS1	CO	Prince Georges County	Brightseat Rd S1
MD	MDPGCOS2	RE	Prince Georges County	Flagstaff Street S2
MD	MDPGCOS3	CO ID	Prince Georges County	Maryland 50 industrail park S3
MD	MDPGCOS4	RE	Prince Georges County	wayne place S4
MD	MDPGCOS5	ID	Prince Georges County	John Hanson S5
MD	MDPGCOS6	ID	Prince Georges County	Pennsy Dr N3
MD	MDSHDTDV	OP ID	State Highway	DV
MD	MDSHDTPS	OP ID	State Highway	PS
MN	MNMISD01	RE	City of Minneapolis	E Harriet Pkwy W44 St
MN	MNMISD02	RE	City of Minneapolis	Luella St Orange Ave
MN	MNMISD03	ID	City of Minneapolis	Vandalia st
MN	MNMISD04	RE CO	City of Minneapolis	Charles Ave
MN	MNMISD05	RE CO	City of Minneapolis	E 29 St 31 Ave S
NC	NCCHBREV	ID	City of Charlotte	Brevi1
NC	NCCHHIDD	RE	City of Charlotte	Hiddr2
NC	NCCHHOSK	ID	City of Charlotte	Hoski2
NC	NCCHNANC	RE	City of Charlotte	Nancr1
NC	NCCHROSE	RE ID OP CO	City of Charlotte	Rosem1
NC	NCCHSHEF	OP RE	City of Charlotte	Shefo1
NC	NCCHSIMS	RE	City of Charlotte	Simsr3
NC	NCCHSTAR	CO	City of Charlotte	Starc1
NC	NCCHYARD	CO RE	City of Charlotte	Yardc2
NC	NCFV71ST	IS	City of Fayetteville	71 ST High School 100ft NE Raeford SR1409
NC	NCFVCLEA	RE	City of Fayetteville	3606 Clearwater Drive
NC	NCFVELMS	CO	City of Fayetteville	ELM Street Eutaw Shopping Center
NC	NCFVROSE	RE OP CO	City of Fayetteville	Rose Apartments 225 Tiffany Court

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
NC	NCFVSTRK	OP	City of Fayetteville	Strickland Bridge Road
NC	NCFVTRYO	RE	City of Fayetteville	1740 Tryon Rd
NC	NCFVWINS	ID	City of Fayetteville	Winslow Pine Railroad tracks
NC	NCGRATHE	CO	City of Greensboro	Athena
NC	NCGRCOUN	OP	City of Greensboro	Country Park
NC	NCGRHUST	ID	City of Greensboro	Husbands Street
NC	NCGRMERR	CO	City of Greensboro	Merrit Drive
NC	NCGRRAND	RE	City of Greensboro	Randlem Road
NC	NCGRUNIO	ID CO IS RE	City of Greensboro	Union Street
NC	NCGRWILL	RE	City of Greensboro	Willoughby
NC	NCRASIT1	OP RE	City of Raleigh	I40 400ft east S State Street
NC	NCRASIT2	RE CO	City of Raleigh	Williamson Drive Pineview Street
NC	NCRASIT3	RE CO OP	City of Raleigh	I40 Dandridge Drive Bunche Drive
NC	NCRASIT4	CO RE	City of Raleigh	Williamson Drive Wade Avenue
NC	NCRASIT5	ID OP	City of Raleigh	Pylon Drive 100ft North Hutton Street
NC	NCRASIT6	ID RE	City of Raleigh	South Wilmington Street City Farm Road
NC	NCRASIT7	CO RE ID OP	City of Raleigh	50ft east N West Street Peace Street Dortch Street
OR	ORCCA001	RE	Clackamas County	Bell Station
OR	ORCCA002	RE	Clackamas County	Lake Oswego
OR	ORCCA003	RE	Clackamas County	Milwaukie
OR	ORCCA004	RE	Clackamas County	Oregon City
OR	ORCCA005	CO	Clackamas County	Wilson Road
OR	OREUA001	CO	City of Eugene	C1 Olive Ave
OR	OREUA002	XX	City of Eugene	M1 Bailey Ave
OR	OREUA003	RE	City of Eugene	R1 Coetivy Ave
OR	ORGRA001	ID RE	City of Gresham	E 3 Boeing
OR	ORGRA002	RE CO	City of Gresham	I 13 Riverview St
OR	ORGRA003	RE	City of Gresham	K 4 Fairview Park
OR	ORGRA004	CO RE	City of Gresham	M 16
OR	ORODA001	FW	ODOT	Ashland
OR	ORODA002	FW	ODOT	Astoria

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
OR	ORODA003	FW	ODOT	Eugene
OR	ORODA004	FW	ODOT	Neskowin
OR	ORODA005	FW	ODOT	Portland
OR	ORPOA001	CO	City of Portland	C 1 Jantzen Beach
OR	ORPOA002	CO RE	City of Portland	C 2 Salmon Street
OR	ORPOA003	ID	City of Portland	I 1 Yeon Ave 35th Ave
OR	ORPOA004	ID	City of Portland	I 2 Swan Island
OR	ORPOA005	RE CO	City of Portland	M 1 Columbia Slough
OR	ORPOA006	RE	City of Portland	R 2 Sandy Doulevard
OR	ORPOA007	FW	City of Portland	T 1
OR	ORSAA001	CO RE	City of Salem	Commercial
OR	ORSAA002	CO	City of Salem	Cottage
OR	ORSAA003	ID	City of Salem	Edgewater
OR	ORSAA004	RE	City of Salem	Redleaf
PA	PAPH0864	RE	Philadelphia	Cresheim Creek
PA	PAPH0891	RE	Philadelphia	Tacony Creek
PA	PAPH1014	RE	Philadelphia	Byberry Creek
PA	PAPH1051	RE CO	Philadelphia	Wooden Bridge Run
PA	PAPH1182	OP RE	Philadelphia	North Byberry Creek
TN	TNKXTYAP	ID OP RE	City of Knoxville	Acker Place
TN	TNKXTYFC	RE FW	City of Knoxville	First Creek
TN	TNKXTYGV	RE OP	City of Knoxville	Gallaher View
TN	TNKXTYTC	OP ID IS RE	City of Knoxville	Third Creek
TN	TNKXTYWE	CO RE IS OP	City of Knoxville	Wellington Drive
TN	TNMET207	OP	City of Memphis	207 Walnut Grove
TN	TNMET211	ID	City of Memphis	211 Warford
TN	TNMET231	RE	City of Memphis	231 Raleigh Lagrange
TN	TNMET260	CO RE	City of Memphis	260 Austin Peay
TN	TNMET410	RE	City of Memphis	410 Whitehaven
TX	TXARA001	CO	City of Arlington	The Parks mall AC603
TX	TXARA002	RE	City of Arlington	R Legacy PK AR602
TX	TXARA003	RE	City of Arlington	Trib to W FK Tri AR601
TX	TXARA004	RE	City of Arlington	Trib To Johnson Creek AI604
TX	TXDAA001	ID	City of Dallas	Joes Cr 138
TX	TXDAA002	ID	City of Dallas	Bastille St 325
TX	TXDAA003	RE ID CO	City of Dallas	Knights Branch 34
TX	TXDAA004	RE	City of Dallas	White Rock Creek 86
TX	TXDAA005	RE	City of Dallas	Ash Creek 55
TX	TXDAA006	RE OP	City of Dallas	Newton Creek 189
TX	TXDCA001	OP FW	TXDOT Dallas	Mountain Creek DH902
TX	TXDCA002	OP FW	TXDOT Dallas	Bachman Branch DH901
TX	TXFWA001	OP RE	City of Fort Worth	Clear FK Trin R TRI STG1
TX	TXFWA002	OP IS	City of Fort Worth	Pylon St PY1
TX	TXFWA003	CO RE	City of Fort Worth	West Fk Trinity R BEL1

Table A1.Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
TX	TXFWA004	ID	City of Fort Worth	Dry Branch CRA1
TX	TXFWA005	RE CO OP	City of Fort Worth	Estrn Hills HS EH1
TX	TXGAA001	RE CO ID OP	City of Garland	Mills Branch Tributary GM404
TX	TXGAA002	ID OP	City of Garland	Trib to Duck Creek GI401
TX	TXGAA003	RE CO	City of Garland	Sleepy Hollow St GR402
TX	TXGAA004	CO RE	City of Garland	I635 Outfall at CE GC603
TX	TXHCA001	RE	Harris County	Overbluff
TX	TXHCA002	RE OP	Harris County	Cypress Trace Station
TX	TXHCA003	CO	Harris County	Steeplechase
TX	TXHCA004	ID	Harris County	Bayport
TX	TXHCA005	CO	Harris County	WillowBrook Mall
TX	TXHCA006	RE	Harris County	Little Cypress Creek
TX	TXHCA007	OP	Harris County	Hadden Road
TX	TXHOA001	OP	City of Houston	Briar Forest
TX	TXHOA002	ID	City of Houston	Eleventh Street
TX	TXHOA003	RE	City of Houston	Lazybrook
TX	TXHOA004	CO	City of Houston	Memorial City Mall
TX	TXHOA005	RE	City of Houston	Tanglewilde
TX	TXIRA001	RE	City of Irving	Bear Cr IR501
TX	TXIRA002	ID RE OP	City of Irving	Cottonwood Branch Trib IM504
TX	TXIRA003	ID CO	City of Irving	Hereford Rd II503
TX	TXIRA004	ID CO	City of Irving	Trib to ELM FK II502
TX	TXMEA001	CO FW	City of Mesquite	South mesquite I635 MC801
TX	TXMEA002	RE	City of Mesquite	South Mesquite South Parkway MC802
TX	TXMEA003	RE	City of Mesquite	South Mesquite Bruton Road MC803
TX	TXPLA001	RE CO	City of Plano	Rowlett Cr PR701
TX	TXPLA002	OP FW	City of Plano	Beck Brach PU704
TX	TXPLA003	CO OP	City of Plano	Spring Creek PC702
TX	TXPLA004	CO ID RE	City of Plano	Spring Creek PI703
TX	TXTCA001	FW OP ID	TXDOT Tarrant County	Deer Creek TH904
VA	VAARLCV2	RE	Arlington	Colonial Village CV2
VA	VAARLLP1	RE	Arlington	Little Pimmet LP1
VA	VAARLRS3	CO	Arlington	Randolph Street RS3
VA	VAARLTC4	ID	Arlington	Trades Center TC4
VA	VACHCCC4	CO	Chesterfield County	CoverLeaf Mall CC4
VA	VACHCCC5	RE	Chesterfield County	Buck Rub Drive CC5
VA	VACHCN1A	RE	Chesterfield County	Gates bluff 1A
VA	VACHCN2A	RE	Chesterfield County	Helmsley road 2A
VA	VACHCOF1	ID	Chesterfield County	unnamed OF1

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
VA	VACHCOF2	OP RE	Chesterfield County	Oak river drive OF2
VA	VACHCOF3	RE	Chesterfield County	Kings mill road OF3
VA	VACHCOF4	RE	Chesterfield County	OF4
VA	VACHCOF5	RE	Chesterfield County	Laurel oak road OF5
VA	VACPTC1A	RE	Chesapeake	Briarfield Drive C1A
VA	VACPTSF2	RE	Chesapeake	Woodards Mill SF2
VA	VACPTYC1	RE	Chesapeake	Etheridge rd Mt Pleasant Rd C1
VA	VACPTYC2	RE OP	Chesapeake	Hunningdon Lakes C2
VA	VACPTYC3	RE	Chesapeake	Horse Run Ditch C3
VA	VACPTYC4	CO	Chesapeake	Woodford Square Along Battlefield Blvd C4
VA	VACPTYC5	ID	Chesapeake	Cavalier Industrial Park C5
VA	VACPTYO1	ID	Chesapeake	Paramont Avenue O1
VA	VAFFCOF1	RE	Fairfax County	Apple Ridge Road
VA	VAFFCOF2	RE CO ID	Fairfax County	Sunset Hills Road
VA	VAFFCOF3	RE	Fairfax County	Onley Road
VA	VAFFCOF4	CO	Fairfax County	Green Look Place
VA	VAFFCOF5	RE	Fairfax County	Oakton Terrace Road
VA	VAFFCOF6	CO	Fairfax County	Fairview Park Drive
VA	VAFFCOF7	RE	Fairfax County	Lakeview Drive
VA	VAFFCOF8	RE	Fairfax County	Pumphrey Drive
VA	VAFFCOF9	RE	Fairfax County	Rock Ridge Road
VA	VAFFOF10	ID	Fairfax County	Boston Boulevard
VA	VAFFOF11	ID	Fairfax County	Prosperity Avenue
VA	VAHAHMS2	ID	Hampton	Copeland Industrial Park HMS2
VA	VAHAHMS5	RE OP	Hampton	Grays Landing HMS5
VA	VAHATYH1	CO	Hampton	Commerce Drive H1
VA	VAHATYH2	ID	Hampton	Mingee Drive H2
VA	VAHATYH3	RE	Hampton	Hampton Club H3
VA	VAHATYH4	RE	Hampton	Bay Avenue H4
VA	VAHATYH5	RE	Hampton	Willow Oaks Boulevard H5
VA	VAHCCOC1	CO	Henrico County	Dickens Place C1
VA	VAHCCOC2	CO	Henrico County	Carousel Lane C2
VA	VAHCCON1	ID	Henrico County	Tomlyn Street N1
VA	VAHCCON2	ID	Henrico County	Impala Drive and Galaxy Road N2
VA	VAHCCOR1	RE	Henrico County	Prestwick Circle R1
VA	VAHCCOR2	RE	Henrico County	Westbury Drive R2
VA	VANFTMS5	CO	Norfolk	Viillage avenue MS5
VA	VANFTMS6	RE	Norfolk	Robin hood road MS6
VA	VANFTMS8	CO	Norfolk	North Hampton MS8
VA	VANFTMS9	CO	Norfolk	Bay side road MS9
VA	VANFTYN1	CO ID	Norfolk	Armisted Avenue N1

Table A1. Site Name and Land Use - *Continued*

State	LOCATION ID	Land use	Jurisdiction	Site ID
VA	VANFTYN2	RE	Norfolk	Modoc Avenue N2
VA	VANFTYN3	RE	Norfolk	Liitle creek road N3
VA	VANFTYN4	CO	Norfolk	Military circle N4
VA	VANFTYN5	RE	Norfolk	Sewel's point N5
VA	VANNTMF1	RE	Newport News	Marshall Avenue MF1
VA	VANNTMF4	RE	Newport News	Chasapeake Bay Apertments MF4
VA	VANNTNN1	RE	Newport News	Glendale Road NN1
VA	VANNTNN2	RE OP	Newport News	Shields Road NN2
VA	VANNTNN3	CO	Newport News	Patrick Henry Mall NN3
VA	VANNTNN4	CO ID	Newport News	Oyster Point Park Jefferson Ave NN4
VA	VANNTNN5	CO RE ID	Newport News	Oyster Point Park Thimble Shoals Blvd NN5
VA	VANNTSF4	RE	Newport News	Central Parkway SF4
VA	VANNTSF6	RE	Newport News	Jefferson Avenue SF6
VA	VANNTYI2	ID OP	Newport News	City Line Rd I2
VA	VAPMTYP1	CO	Portsmouth	Cradock Shopping center P1
VA	VAPMTYP2	RE	Portsmouth	West park homes P2
VA	VAPMTYP3	RE CO	Portsmouth	Church land shopping center P3
VA	VAPMTYP4	RE	Portsmouth	Edgefield apartmentsP4
VA	VAPMTYP5	RE	Portsmouth	South hampton P5
VA	VAVBTYA1	OP ID	Virginia Beach	Morris Neck Road A1
VA	VAVBTYI1	ID	Virginia Beach	Airport Industrial Park I1
VA	VAVBTYM2	RE CO OP ID	Virginia Beach	Ketlam Road M2
VA	VAVBTYR1	RE	Virginia Beach	Homestead Drive R1
VA	VAVBTYV1	RE	Virginia Beach	Bow creek V1
VA	VAVBTYV2	RE	Virginia Beach	Salem Road V2
VA	VAVBTYV3	CO OP	Virginia Beach	Haygood V3
VA	VAVBTYV4	ID	Virginia Beach	Viking Drive V4
VA	VAVBTYV5	RE OP	Virginia Beach	Holland road V5

Table A2. Site Characteristics

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
ALHUCHIP	19.5	3			GS		08/27/92	09/12/00	8
ALHUDRAV	20	3			GS		09/19/01	09/19/01	1
ALHUHURI	78.3	3			CG		08/27/92	09/19/01	9
ALHUMASM	87	3			CG		08/27/92	09/19/01	9
ALHUWERP	130	3			CG		08/27/92	09/19/01	9
ALJC004L	2564	3			CG		09/19/01	01/19/02	2
ALJC004R	1047	3			CG		09/20/01	01/19/02	2
ALJCC001	336	3			CG		11/27/01	01/19/02	2
ALJCC002	750	3			CG		11/27/01	03/20/02	2
ALJCC009	112	3			CG		08/31/01	03/09/02	2
ALJCC010	167	3			CG		08/31/01	03/09/02	2
ALJCC012	244	3			CG		12/17/01	12/17/01	1
ALMOCREO	74	3			GS		02/10/93	04/15/93	3
ALMODAPH	14	3					02/16/93	04/20/93	3
ALMOSARA	64	3			GS		01/24/93	04/04/93	3
ALMOSIIV	450	3					02/11/93	04/15/93	3
ALMOSITV	304	3			CG		01/24/93	04/04/93	3
ALMOSIVI	194	3			CG		01/24/93	04/04/93	3
ALMOTHEO	27	3			GS		01/24/93	03/30/93	3
AZMCA001	39	6	80				11/10/91	07/22/98	27
AZMCA002	1120	6	1		GS		01/12/92	02/07/92	2
AZMCA003	45	6	15				12/10/91	07/22/98	27
AZMCA004	81	6	33		CG		10/27/91	08/22/92	6
AZMCA005	3.4	6	94		CG		12/04/92	08/07/98	26
AZMCA006	17.8	6	60		CG		03/07/94	09/11/98	20
AZTUA001	103	6			CG		07/25/96	12/04/01	13
AZTUA002	48.3	6			CG		08/26/96	12/04/01	12
AZTUA003	29	6			CG		08/14/96	12/04/01	11
AZTUA004	83	6			CG		09/24/96	12/11/01	11
CAALAL03	168	6			CG		02/15/90	03/25/93	20
CAALAL04	20	6			CG		03/02/90	02/27/91	5
CAALAL07	78	6			CG		01/13/90	03/17/91	5
CAALAL09	260	6			CG		01/13/90	03/17/91	9
CAALAL10	144	6			CG		03/02/90	03/19/91	8
CACTA001	0.69	6	95		CG		01/23/01	03/10/02	14
CACTA002	1.61	6	100		CG		10/28/00	03/06/02	16
CACTA003	1.85	6	70		CG		01/23/01	03/23/02	10
CACTA004	0.44	6	70		CG		01/23/01	03/23/02	11
CACTA005	0.99	6	100		CG		11/26/97	03/25/99	8
CACTA006	0.99	6	100		CG		01/25/99	03/20/99	3
CACTA007	0.99	6	100		CG		01/25/99	03/25/99	4

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
CACTA008	3.16	6	80		CG		01/17/00	01/27/02	19
CACTA009	4.18	6	80		CG		11/20/99	03/17/02	24
CACTA010	0.96	6	80		CG		11/08/99	03/17/02	26
CACTA011	0.4	6	100		CG		11/10/97	02/09/99	4
CACTA012	0.99	6	100		CG		11/13/97	02/09/99	4
CACTA013	1.41	6	100		CG		11/10/97	02/09/99	3
CACTA014	0.99	6	100		CG		11/13/97	01/25/99	4
CACTA015	0.99	6	100		CG		11/13/97	01/25/99	4
COCSA001	80	9	58.1		CG		06/03/92	11/21/92	7
COCSA002	105.6	9	37.5		CG		05/31/92	07/29/02	7
COCSA003	110.72	9	55.9		CG		06/05/92	11/21/92	7
COCSA004	209.28	9	34.2		CG		05/26/92	11/21/92	7
COCSA005	31.36	9	40.1		CG		06/10/92	12/06/92	7
CODEA001	150	9	83		CG		06/05/92	07/12/92	3
CODEA002	55	9	83		CG		04/14/92	07/12/92	3
CODEA003	269	9	20		CG		03/22/92	08/23/92	4
CODEA004	498	9	85		CG		05/21/92	07/10/92	3
CODEA005	495	9	44		CG		06/06/92	08/23/92	3
CODEA006	636	9	85		CG		06/08/92	07/10/92	3
CODEA007	56	9	85		CG		03/28/92	07/02/92	3
CODEA008	146	9	83		CG		03/28/92	05/31/92	3
GAATAT01	28	3			CG		03/08/95	02/16/97	10
GAATAT02	95	3			CG		10/04/95	02/16/97	9
GACLCOSI	18	3			CG		11/29/95	03/19/00	20
GACLCOTR	125	3			CG		05/01/95	03/16/00	24
GACOC1A2	63.6	3					01/27/96	02/28/00	17
GACOC1A3	7590.4	3					08/24/00	03/19/01	6
GACOCOL2	2947	3					01/19/95	03/12/01	22
GADKCOTD	115	3			CG	WP	12/13/93	06/06/00	25
GAFUCOS1	10339	3			GS		11/10/94	04/25/01	22
GAFUCOS2	3915	3			CG		10/30/94	04/25/01	19
GAFUCOS3	6257	3			GS		01/06/95	04/25/01	22
IDADA001	10.9	8			CG		08/11/99	04/19/01	7
IDADA002	105	8			CG		04/29/99	04/11/01	7
IDADA003	17	8			CG		04/29/99	07/30/01	9
IDADA004	18	8					04/11/01	04/11/01	1
KATOATWO	38	4	55		CG		04/27/98	09/13/02	15
KATOBROO	18.5	4	25		CG		04/27/98	09/13/02	16
KATOJACK	218	4	65		CG		04/27/98	09/13/02	16
KATOSTFE	39.5	4	75		CG		04/27/98	08/16/02	17
KAWIHUNT	36	4	50				02/09/98	10/05/01	16

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
KAWIMCLE	30	4	65				02/09/98	10/05/01	16
KAWISBWY	250	4	60				02/09/98	10/05/01	16
KAWITOWN	40	4	90				02/09/98	10/05/01	16
KYLOTSR1	96.64	2	39.6		CG		01/15/91	10/05/91	3
KYLOTSR2	108.16	2	20.6		GS		02/05/91	10/05/91	3
KYLOTSR3	134.4	2	35		CG		03/01/91	12/12/91	3
KYLOTSR4	43.52	2	45.5		GS		03/12/91	04/15/92	4
KYLOTSR5	84.48	2	68.9		CG		03/27/91	05/12/92	5
KYLOTSR6	180.48	2	63.5		CG		04/04/91	12/12/91	3
KYLXEHL4	13	2	10		GS		10/07/98	10/07/98	1
KYLXEHL5	550	2			GS		07/30/98	07/30/98	1
KYLXEHL6	1.3	2			CG		10/24/97	01/05/98	3
KYLXEHL7	4.8	2			CG		10/24/97	01/05/98	3
KYLXNEL1	32	2			GS		10/07/98	10/07/98	1
KYLXNEL2	580	2			CG	WP	07/30/98	10/07/98	2
KYLXNEL3	73	2			CG		06/03/92	09/27/96	12
KYLXTBL1	71	2			CG		06/03/92	09/27/96	12
KYLXTBL2	94	2			GS		06/30/92	09/27/96	12
KYLXTBL3	205	2			GS		06/19/92	09/21/96	5
KYLXWHL1	38	2			CG		06/03/92	09/27/96	13
MABOA001	40.4	1	74		CG		04/11/92	08/14/92	5
MABOA002	86.7	1	52		CG		04/17/92	06/24/92	3
MABOA003	5	1	55		CG		04/11/92	06/24/92	3
MABOA004	32	1	97		CG		04/11/92	06/24/92	3
MABOA005	102.7	1	38		CG		04/17/92	06/24/92	3
MABOA006	3.3	1	74		CG		06/02/01	07/17/01	3
MABOA007	12.2	1			GS		09/25/01	09/25/01	1
MDAACOMW	5	2	94	E	CG		07/31/92	09/25/92	3
MDAACOOD	28	2	41	E	CG		08/11/92	10/09/92	3
MDAACOPP	25	2	85	E	CG		08/11/92	11/14/00	26
MDAACORK	12	2	41	E	CG		08/28/92	10/30/92	3
MDAACOSC	26	2	41	E	CG		08/11/92	10/09/92	3
MDBACOBC	25.3	2	60				12/15/93	03/08/94	3
MDBACOLC	225	2	70				10/20/93	01/15/98	19
MDBACOSC	83.5	2	30				12/15/93	06/19/98	26
MDBACOTC	144.06	2					01/12/94	04/07/94	3
MDBACOWC	73	2	7				01/12/94	03/21/94	3
MDBCTYBO	48.43	2					04/16/93	03/21/94	3
MDBCTYFM	45.96	2					05/30/92	11/04/93	3
MDBCTYHA	104.4	2	32	E			05/17/95	12/14/00	66
MDBCTYHO	354.09	2					06/05/92	09/25/92	3

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
MDBCTYHR	38.8	2	54				05/30/92	10/20/93	3
MDBCTYKO	54.36	2					04/26/93	03/21/94	3
MDCHCOIP	11	2			CG		10/04/95	01/02/96	3
MDCHCOPA	10	2			CG		09/22/95	01/24/96	3
MDCHCOPF	50	2			GS		01/02/96	04/30/96	3
MDCHCOTG	10	2			CG		10/14/95	10/14/95	1
MDCLCOBP	15	2					07/21/94	12/10/94	3
MDCLCOCE	22.35	2	26	E			12/10/93	11/21/94	3
MDCLCOJS	20	2	91	E			12/21/93	11/01/94	3
MDCLCOKW	66.2	2	11	E			03/21/94	09/22/94	3
MDCLCOSD	36	2	49	E			02/21/94	11/10/94	3
MDHACFBA		2					11/05/93	11/10/94	3
MDHACOBP	69.7	2	16	E		WP	02/17/99	12/16/00	18
MDHACOCF	14.4	2					01/05/93	04/10/93	2
MDHACOCS	51	2					04/21/93	09/22/94	3
MDHACOGR	80	2					04/26/93	04/10/94	2
MDHACOIC	10	2					08/06/93	09/22/94	2
MDHOCODC	7.5	2	90		CG	WP	12/15/93	04/13/94	3
MDHOCOFM	3.5	2	77		CG	GS	12/15/93	11/01/94	3
MDHOCOGM	29.5	2	38		CG	WP	12/10/93	11/01/94	3
MDHOCOMH	19	2	65		CG	WP	12/10/93	04/13/94	3
MDHOCOOC	11.7	2	49		CG	WP	11/17/93	03/21/94	3
MDMOCOBC	14.2	2	83	E	CG		05/25/94	09/22/95	3
MDMOCOCV	11.5	2	55	E		WP	08/13/96	09/25/00	37
MDMOCONV	75.4	2	57	E	CG		05/04/94	03/08/95	3
MDMOCOQA	34.2	2	45	E	CG		05/04/94	10/27/95	3
MDMOCOSL	81	2	92	E	CG	OT	09/22/94	09/22/95	3
MDMOCOWP	70	2	96	E	CG	OT	05/25/94	10/27/95	3
MDPGCOS1	19.7	2	47	E	CG		08/11/92	01/22/97	26
MDPGCOS2	57.3	2	45	E	CG		08/11/92	09/25/00	63
MDPGCOS3	34.4	2	96	E	CG		08/11/92	03/04/93	3
MDPGCOS4	102.5	2	33	E	CG		08/11/92	03/04/93	3
MDPGCOS5	41.3	2	83	E	CG		08/11/92	03/04/93	3
MDPGCOS6	42.4	2			GS		10/23/94	08/20/97	28
MDSHDTDV	4	2			CG		06/14/99	06/06/00	8
MDSHDTPS	20	2			GS		02/11/98	06/21/00	13
MNMISD01	143	1			CG		05/06/01	10/13/01	10
MNMISD02	95	1			CG		05/06/01	10/13/01	9
MNMISD03	80	1			CG		05/20/01	11/12/01	10
MNMISD04	63	1			CG		05/06/01	11/12/01	9
MNMISD05	100	1			CG		05/20/01	11/12/01	10

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
NCCHBREV	15.1	2	75		CG		05/13/92	03/03/93	4
NCCHHIDD	20	2	30		CG		05/13/92	12/10/93	5
NCCHHOSK	17.4	2	71.83				05/13/92	03/03/93	4
NCCHNANC	10.9	2	20		GS		08/27/92	02/21/94	4
NCCHROSE	78.87	2	42.55		CG		05/13/92	08/12/92	3
NCCHSHEF	42.5	2	20.68				06/04/92	02/21/94	3
NCCHSIMS	6.8	2	50		CG		05/13/92	03/03/93	4
NCCHSTAR	14.1	2	70		GS		05/13/92	03/03/93	4
NCCHYARD	88.6	2	68.21		CG		05/13/92	03/03/93	4
NCFV71ST	36	2	45		CG		01/21/93	06/15/99	18
NCFVCLEA	12	2	20				01/04/93	04/01/99	14
NCFVELMS	40	2	90		CG		01/04/93	04/01/99	18
NCFVROSE	39.27	2	50		CG		12/17/92	06/15/99	14
NCFVSTRK	85	2	1				02/07/93	06/16/96	6
NCFVTRYO	25	2	50				01/21/93	04/01/99	18
NCFVWINS	12	2	75				01/04/93	06/15/99	18
NCGRATHE	23	2	90		CG		07/06/95	04/15/99	17
NCGRCOUN	18.5	2	2		GS		06/19/95	04/01/99	15
NCGRHUST	13	2	75		CG		06/01/95	05/14/99	16
NCGRMERR	21	2	74		CG		05/10/95	04/15/99	16
NCGRRAND	26	2	50		CG		06/01/95	06/15/99	17
NCGRUNIO	33	2	75		CG		06/01/95	01/23/99	17
NCGRWILL	13	2	20				05/19/95	05/14/99	16
NCRASIT1	21	2					05/19/93	03/16/00	9
NCRASIT2	42	2					05/19/93	03/16/00	9
NCRASIT3	110	2					05/19/93	03/16/00	9
NCRASIT4	30	2					05/19/93	03/16/00	9
NCRASIT5	32	2					05/19/93	03/16/00	9
NCRASIT6	58	2					05/19/93	03/16/00	9
NCRASIT7	467	2					05/19/93	03/16/00	9
ORCCA001	15	7			CG		11/18/92	11/18/92	1
ORCCA002	120	7			CG		11/18/92	10/14/94	6
ORCCA003	165	7			CG		03/01/93	10/14/94	5
ORCCA004	50	7			CG		11/18/92	10/14/94	6
ORCCA005	41	7			CG		11/18/92	10/14/94	6
OREUA001	380	7			CG		09/23/92	05/21/96	16
OREUA002	886	7			CG		09/23/92	05/21/96	15
OREUA003	377	7			CG		09/23/92	05/21/96	15
ORGRA001	292	7			CG		03/02/93	04/11/96	6
ORGRA002	789	7			CG		03/02/93	04/11/96	6
ORGRA003	73	7			CG		03/02/93	04/11/96	6

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
ORGRA004	64	7			CG		03/02/93	04/11/96	6
ORODA001	22.4	7			CG		12/04/95	12/04/95	1
ORODA002	1.2	7			CG		10/02/95	10/02/95	1
ORODA003	18.2	7			CG		01/07/95	01/18/96	5
ORODA004	3.6	7			CG		10/12/95	10/12/95	1
ORODA005	23.1	7			CG		01/07/95	01/18/96	6
ORPOA001	35	7			CG		05/07/91	10/25/95	13
ORPOA002	75	7			CG		05/07/91	03/03/96	16
ORPOA003	46	7			CG		05/07/91	03/03/96	14
ORPOA004	49	7			CG		08/09/91	10/25/95	13
ORPOA005	91	7			CG		05/07/91	03/03/96	14
ORPOA006	85	7			CG		05/07/91	03/03/96	13
ORPOA007	10	7			CG		05/07/91	03/03/96	14
ORSAA001	31	7			CG		01/07/95	01/14/96	6
ORSAA002	40	7			CG		01/07/95	01/14/96	6
ORSAA003	35	7			CG		01/07/95	01/14/96	6
ORSAA004	72	7			CG		01/07/95	01/14/96	6
PAPH0864	22	2	84	E	CG		09/10/92	09/25/92	2
PAPH0891	35	2	83	E	CG		09/22/92	10/09/92	2
PAPH1014	22	2	82	E	CG		09/10/92	10/09/92	3
PAPH1051	223	2	87	E	CG		09/22/92	10/09/92	2
PAPH1182	31	2	57	E	CG		09/10/92	10/09/92	3
TNKXTYAP	582.4	2	44		GS	WP	03/27/91	06/30/01	63
TNKXTYFC	2880	2	40			WP	03/06/92	06/07/01	47
TNKXTYGV	224	2	37		GS		08/14/91	08/25/99	39
TNKXTYTC	352	2	34				02/13/92	04/11/00	54
TNKXTYWE	364.8	2	60				04/08/91	05/03/00	51
TNMET207	157	2			GS	WP	06/21/00	04/23/01	5
TNMET211	45	2			CG		01/11/00	04/23/01	4
TNMET231	26	2			CG		01/11/00	04/23/01	4
TNMET260	294	2			CG		07/20/00	05/17/01	4
TNMET410	154	2			CG		06/21/00	04/23/01	4
TXARA001	38.8	5	76.2		CG		10/28/92	03/08/01	22
TXARA002	160.6	5	47.4		CG		10/28/92	03/08/01	21
TXARA003	77	5	89		CG		10/29/92	04/14/93	7
TXARA004	85.5	5	80.9			WP	12/09/92	03/28/93	7
TXDAA001	9	5	80		CG		03/03/92	09/21/92	7
TXDAA002	49.5	5	80		CG		03/03/92	03/08/01	19
TXDAA003	486.7	5			CG		12/02/97	05/04/01	21
TXDAA004	59.1	5	84.5		CG		02/22/92	04/11/01	20
TXDAA005	71.3	5	50		CG		02/12/92	09/21/92	7

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
TXDAA006	38.9	5	44.9		CG		02/22/92	03/24/01	20
TXDCA001	115.36	5	10				09/03/97	03/08/01	17
TXDCA002	12.05	5	33				01/11/98	03/08/01	9
TXFWA001	61.7	5	21.9				02/22/92	08/12/92	7
TXFWA002	151.6	5	27.7	* I	CG		02/03/92	03/08/01	21
TXFWA003	136	5	66.5		CG		03/09/92	10/28/92	7
TXFWA004	73.7	5	79.3				03/24/92	03/08/01	21
TXFWA005	150.8	5	61.4				04/17/92	03/24/01	23
TXGAA001	268	5			CG		12/02/97	03/27/01	23
TXGAA002	33.9	5	67.3				06/20/92	03/27/01	22
TXGAA003	67.3	5	55.4				09/01/92	01/23/93	7
TXGAA004	36.2	5	84.6		CG		09/01/92	01/23/93	7
TXHCA001	560	4			GS		01/29/99	03/27/01	8
TXHCA002	95	4	65				07/20/92	04/16/01	14
TXHCA003	32	4			CG		02/11/99	06/22/01	8
TXHCA004	99	4	71.25				04/07/93	03/08/01	14
TXHCA005	81	4	95		CG		06/30/92	04/07/93	6
TXHCA006	401	4	45				06/30/92	04/29/93	6
TXHCA007	872	4				WP	07/15/92	11/09/93	6
TXHOA001	44	4	5.7		GS		06/30/92	11/19/92	7
TXHOA002	232	4	76.5				06/22/92	05/31/01	16
TXHOA003	65	4	45		GS		06/22/92	03/27/01	14
TXHOA004	24	4	98		CG		07/19/92	11/22/99	12
TXHOA005	38	4	65		CG		06/22/92	07/19/01	16
TXIRA001	65.3	5	41.9		CG		09/03/92	03/24/01	22
TXIRA002	127.7	5			CG		03/18/99	05/28/01	22
TXIRA003	43.4	5	77.3			WP	08/24/92	01/09/93	7
TXIRA004	43.9	5	77.8				09/21/92	01/28/93	7
TXMEA001	45.9	5	89.4		CG		02/24/93	03/24/01	22
TXMEA002	45.4	5	49.8		CG		03/11/93	06/25/93	7
TXMEA003	46.2	5	49.9		CG	WP	02/10/93	05/23/93	7
TXPLA001	51.4	5	54.3		CG		12/09/92	04/14/93	7
TXPLA002	73.5	5					11/09/98	04/11/01	22
TXPLA003	22.7	5	73.5		CG		12/09/92	05/04/01	25
TXPLA004	49	5	81.6				01/09/93	06/09/93	7
TXTCA001	63.13	5	27				02/06/97	03/24/01	15
VAARLCV2	24.7	2	35			DS	02/11/98	01/19/01	9
VAARLLP1	38.7	2	35				10/20/99	03/04/01	8
VAARLRS3	14	2	74			DS	09/21/99	01/19/01	8
VAARLTC4	36	2	39				02/03/98	06/01/01	13
VACHCCC4	60	2	80		CG		08/12/96	12/10/01	13

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Conveyance	Control	First Sample	Last Sample	Number of Samples
VACHCCC5	10	2	50		CG		08/12/96	12/10/01	13
VACHCN1A	10	2	10		CG		08/19/99	01/08/01	4
VACHCN2A	60	2	20		CG		08/19/99	01/08/01	4
VACHCOF1	22.5	2					04/16/93	10/26/93	3
VACHCOF2	19.05	2	10		CG		04/16/93	10/08/98	8
VACHCOF3	13.5	2	20		CG		04/16/93	02/01/99	11
VACHCOF4	38.5	2					04/16/93	12/15/93	3
VACHCOF5	55.6	2	50		CG		04/16/93	12/10/01	16
VACPTC1A	130	2	25	E	CG GS		11/01/97	06/15/99	8
VACPTSF2	91	2	10	E	GS		04/16/93	10/26/93	3
VACPTYC1	57	2	25	E	CG		02/26/93	12/05/96	7
VACPTYC2	188	2	25	E	CG	WP	02/26/93	01/24/99	15
VACPTYC3	32	2	50	E	CG		02/26/93	02/18/99	15
VACPTYC4	28	2	85	E	CG		04/16/93	02/02/99	14
VACPTYC5	16	2	57	E	CG		03/27/93	01/15/99	15
VACPTYO1	14	2	50	E	CG		02/26/93	04/16/93	3
VAFFCOF1	32.3	2			CG		03/18/92	08/11/92	3
VAFFCOF2	20.1	2	50	E	CG	DP	07/03/92	08/01/00	14
VAFFCOF3	63.9	2			CG		06/18/92	09/02/92	3
VAFFCOF4	108.8	2	70	E	CG	WP	04/21/92	09/03/00	13
VAFFCOF5	39.7	2			CG	DP	04/16/92	09/22/92	3
VAFFCOF6	213.4	2	21	E	CG	WP	07/12/92	11/10/00	14
VAFFCOF7	49.9	2	25	E	CG		06/24/92	11/29/00	15
VAFFCOF8	57.5	2			CG		04/21/92	09/02/92	3
VAFFCOF9	63.8	2	50	E	CG	WP	07/21/92	09/02/00	13
VAFFOF10	82	2			CG		04/21/92	08/11/92	3
VAFFOF11	37.9	2	66	E	CG		06/26/97	11/29/00	11
VAHAHMS2	793	2	67	E	CG		11/26/92	01/21/93	3
VAHAHMS5	53	2	28	E			11/12/92	02/12/93	3
VAHATYH1	115	2	80	E	CG		11/12/92	05/14/99	18
VAHATYH2	47	2	70	E	CG		11/26/92	04/24/99	19
VAHATYH3	18	2	40	E	CG		11/12/92	06/20/99	17
VAHATYH4	134	2	25	E	CG		11/12/92	04/24/99	17
VAHATYH5	35	2	25	E	CG		11/12/92	04/24/99	17
VAHCCOC1	65	2	89	E	CG		11/13/92	12/20/92	2
VAHCCOC2	70	2	87	E	CG		10/30/92	01/05/93	3
VAHCCON1	75	2	89	E	CG		12/18/92	01/22/93	2
VAHCCON2	23	2	89	E	CG		11/22/92	01/22/93	3
VAHCCOR1	40	2	61	E	CG		11/03/92	01/05/93	3
VAHCCOR2	70	2	57	E	CG		11/03/92	01/05/93	3
VANFTMS5	56	2			CG		04/22/92	07/27/92	3

Table A2. Site Characteristics - *Continued*

LOCATION ID	Area (acres)	EPA Rain Zone	% Impervious	Q	Convey ance	Control	First Sample	Last Sample	Number of Samples
VANFTMS6	68	2					05/05/92	07/27/92	3
VANFTMS8	65	2					04/22/92	07/27/92	3
VANFTMS9	40	2					05/05/92	08/27/92	3
VANFTYN1	43	2	47		CG		04/22/92	02/12/00	28
VANFTYN2	97	2	25		CG		05/30/92	12/14/99	30
VANFTYN3	27	2	37		CG		04/22/92	12/14/99	28
VANFTYN4	43	2	70		CG		04/22/92	12/14/99	28
VANFTYN5	39	2	25		CG		06/09/92	02/18/00	28
VANNTMF1	39	2	50		CG		10/04/92	01/21/93	3
VANNTMF4	12	2	73		CG		12/28/92	03/13/93	3
VANNTNN1	75	2	40		OT		10/31/92	04/02/99	12
VANNTNN2	397	2	24		CG	DP	12/10/92	04/16/99	15
VANNTNN3	24	2	85		CG	WP	10/04/92	04/16/99	15
VANNTNN4	294	2	58		CG	WP	10/04/92	04/16/99	16
VANNTNN5	83	2	62		OT		12/28/92	04/02/99	11
VANNTSF4	111	2	30		GS		12/10/92	02/26/93	3
VANNTSF6	207	2	37		GS		10/04/92	03/03/93	4
VANNTYI2	49	2	73		GS		10/04/92	01/21/93	3
VAPMTYP1	27.2	2	68		CG		01/16/93	05/14/99	18
VAPMTYP2	101.1	2	36		CG		02/26/93	06/20/99	17
VAPMTYP3	46	2			CG		01/16/93	05/14/99	17
VAPMTYP4	35.3	2	39		CG		12/20/92	06/20/99	17
VAPMTYP5	53.5	2	14	E	CG		12/20/92	05/23/99	17
VAVBTYA1	225	2	7				07/01/92	10/30/92	5
VAVBTYI1	8	2	90				06/09/92	10/04/92	3
VAVBTYM2	310	2	35				10/04/90	10/30/92	4
VAVBTYR1	49	2	25				05/07/92	09/19/92	5
VAVBTYV1	63	2	29		OT		03/26/92	02/28/99	27
VAVBTYV2	260	2	29		OT	WP	05/07/92	02/18/99	30
VAVBTYV3	25	2	25		CG		04/12/92	02/28/99	33
VAVBTYV4	29	2	55		CG		04/12/92	03/14/99	30
VAVBTYV5	882	2	47		OT	WP	05/07/92	03/14/99	28

Appendix B
Modified Values in the Database

B1. Description

The following table indicates the modified values in the database. The column "Order" corresponds to the row number in the table. The column "Problem" indicates the reason why the value was deleted or modified. In the case that the information available can solve the problem the action was described in the column "action". The last column indicates the community where the event was located.

Table B1. Modified Values in the NSQD

Order	Constituent	Original value	Problem	Action	Location_ID
1373	TSS	10100	High	Delete	Acker Place
890	TSS	53000	High	Delete	Philadelphia
1909	TP	80.1	High - Ortho very Low	Delete	Louisville
1707	TP	35	High	Delete	Lexington
1629	TP	15.4	High	Delete	
1907	Ortho P	60.1	High	Delete	Louisville
3135	Dis Zn / Tot Zn		High Ratio	Delete Total	Boston
3118	TDS	17900	High TDS	Deleted	Boston
2893	Dis Cu / Tot Cu		High Ratio	Delete Dissolved Copper	Portland
2883	Dis Cu / Tot Cu		High Ratio	Delete Dissolved Copper	Ada County
561	Dis P / Tot P		Wrong Dissolved Values	corrected	Portsmouth
562	Dis P / Tot P		Wrong Dissolved Values	corrected	Portsmouth
563	Dis P / Tot P		Wrong Dissolved Values	corrected	Portsmouth

265

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
221	Dis P / Tot P		Wrong Dissolved Values	Corrected	Hampton
223	Dis P / Tot P		Wrong Dissolved Values	Corrected	Hampton
1707	Dis P / Tot P		High Values	Deleted	Lexington
1999	Dis P / Tot P		High Dissolved Value	Deleted	Cobb
2301	Dis P / Tot P		Low Total	Delete both	Fayetteville
2268	Dis P / Tot P		Low Total	Delete both	Fayetteville
2293	Dis P / Tot P		Low Total	Delete both	Fayetteville
4315	Dis P / Tot P		Low Total	Delete both	Raleigh
4306	Dis P / Tot P		Low Total	Delete both	Raleigh
4351	Dis P / Tot P		Low Total	Delete both	Raleigh
4342	Dis P / Tot P		Low Total	Delete both	Raleigh
4055	Dis P / Tot P		Low Total	Delete both	Greensboro
4070	Dis P / Tot P		Inverted	Corrected	Greensboro
4197	Dis P / Tot P		Low Total	Delete both	Greensboro
4249	Dis P / Tot P		Low Total	Delete both	Greensboro
4085	Dis P / Tot P		Low Total	Delete both	Greensboro
4217	Dis P / Tot P		Low Total	Delete both	Greensboro
4068	Dis P / Tot P		Wrong values	Corrected	Greensboro
4038	Dis P / Tot P		Low Total	Delete both	Greensboro
4134	Dis P / Tot P		Low Total	Delete both	Greensboro
4233	Dis P / Tot P		Low Total	Delete both	Greensboro
4149	Dis P / Tot P		Low Total	Delete both	Greensboro
3698	Dis P / Tot P		Low Total	Delete both	Greensboro
4024	Dis P / Tot P		Low Total	Delete both	Greensboro
2150	Dis P / Tot P		Low Total	Delete both	Fulton
1449	Dis P / Tot P		Low Total	Delete both	Knoxville
1617	Dis P / Tot P		Low Total	Delete both	Knoxville
1596	Dis P / Tot P		Low Total	Delete both	Knoxville
1616	Dis P / Tot P		Low Total	Delete both	Knoxville
1460	Dis P / Tot P		Low Total	Delete both	Knoxville
1707	TKN	290	High	Deleted	Lexington
1000	TKN	250	High	Deleted	Baltimore City
4149	TKN	147	High	Deleted	Greensboro
2699	TKN	120	High	Deleted	Maricopa
3136	NO2 NO3	1690	High	Deleted	Atlanta
3281	NO2 NO3	50	High	Deleted	Houston 06/30/92
3331	NO2 NO3	48	High	Deleted	Houston 06/30/92
3289	NO2 NO3	32.1	High	Deleted	Houston 06/30/92
3305	NO2 NO3	28	High	Deleted	Houston 06/30/92
48	COD BOD	< 5	COD low	Deleted	Chesterfield 02/03/98
737	COD	5050		Deleted	Bow creek V1

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
97	BOD		Weird Value <30	Deleted	Chesterfield 09/27/99
1895	BOD	610		Deleted	Pleasure_Ridge_P ark
2676	COD BOD	4300		Deleted	Maricopa
3299	COD BOD	1260		Retyped	Houston
4087	BOD	545		Deleted	Husbands_Street
4343	COD BOD			Deleted	Williamson_Drive_ Wade_Avenue
4399	COD BOD			Deleted	50ft_east_N_West_ Street_Peace_Stre et_Dortch_Street
96	Cadmium		Weird Values Dissolved Higher	Deleted	Chesterfield 08/19/99
97	Cooper		Weird Values Dissolved Higher	Deleted	Chesterfield 09/27/99
85	Cooper		Weird Values Dissolved Higher	Deleted	Chesterfield 09/27/99
110	Lead		Weird Values Dissolved Higher	Deleted	Chesterfield 09/27/99
76	Zinc		Weird Values Dissolved Higher	Deleted	Chesterfield 09/27/99
110	Zinc		Weird Values Dissolved Higher	Deleted	Chesterfield 09/27/99
3288	Antimony		Elevated values for the same set of samples	Deleted	Houston
3304	Antimony		Elevated values for the same set of samples	Deleted	Houston
3330	Antimony		Elevated values for the same set of samples	Deleted	Houston
3281	Antimony		Elevated values for the same set of samples	Deleted	Houston
3289	Antimony		Elevated values for the same set of samples	Deleted	Houston
3305	Antimony		Elevated values for the same set of samples	Deleted	Houston
3331	Antimony		Elevated values for the same set of samples	Deleted	Houston
3276	Oil & Grease		Detection Limit is different	Replace as a detected value	Harris County
2836	Conductivity		Elevated Value. Two samples	Use mean value	Colorado Srpings
4077	Turbidity		NT in cell	Move to qualifier	Greensboro
446	TDS <46		Wrong Qualifier	Delete Qualifier..	Norfolk N2
2128	TDS	0.065	Factor of a thousand.	Change value from 0.065 to 65	Fulton County
2257	TDS		Wrong value	corrected (32 mg/L)	Fayetteville
3136	TDS		Value not clear in hardcopy	Delete value <31 mg/L	Atlanta

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
2649	TDS		TDS>TS	Delete pair	Maricopa
2658	TDS		TDS>TS	Delete pair	Maricopa
2660	TDS		TDS>TS	Delete pair	Maricopa
2664	TDS		TDS>TS	Delete pair	Maricopa
2665	TDS		TDS>TS	Delete pair	Maricopa
2672	TDS		TDS>TS	Delete pair	Maricopa
2686	TDS		TDS>TS	Delete pair	Maricopa
2699	TDS		TDS>TS	Delete pair	Maricopa
2704	TDS		TDS>TS	Delete pair	Maricopa
2706	TDS		TDS>TS	Delete pair	Maricopa
2709	TDS		TDS>TS	Delete pair	Maricopa
2713	TDS		TDS>TS	Delete pair	Maricopa
2714	TDS		TDS>TS	Delete pair	Maricopa
2715	TDS		TDS>TS	Delete pair	Maricopa
2716	TDS		TDS>TS	Delete pair	Maricopa
2717	TDS		TDS>TS	Delete pair	Maricopa
2719	TDS		TDS>TS	Delete pair	Maricopa
2720	TDS		TDS>TS	Delete pair	Maricopa
2721	TDS		TDS>TS	Delete pair	Maricopa
2725	TDS		TDS>TS	Delete pair	Maricopa
2726	TDS		TDS>TS	Delete pair	Maricopa
2729	TDS		TDS>TS	Delete pair	Maricopa
2730	TDS		TDS>TS	Delete pair	Maricopa
2734	TDS		TDS>TS	Delete pair	Maricopa
2740	TDS		TDS>TS	Delete pair	Maricopa
2742	TDS		TDS>TS	Delete pair	Maricopa
2750	TDS		TDS>TS	Delete pair	Maricopa
2751	TDS		TDS>TS	Delete pair	Maricopa
3321	TSS		Wrong Detection Limit	Change for value	Houston
3154	TSS		Wrong Detection Limit	Change for <1	Atlanta
1117	TSS		Wrong Detection Limit	Deleted	Montgomery County
2543	TSS		First Flush was compared with TDS	Change for 160	Los Angeles
32	COD		Weird Detection Limit	Delete Qualifier..	Arlington
5	COD		Weird Detection Limit	Delete Qualifier. Quantification limit = 5mg/L	Arlington
16	COD		Weird Detection Limit	Delete Qualifier. Quantification limit = 5mg/L	Arlington
859	Fecal Streptococcus		Atypical Growth	Delete value >6000	Fairfax
1401	Ammonia		Typo in detection limit	Change <2 to <0.2 mg/L	Knoxville

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
2299	Ammonia		Detection Limit is different	Change for <0.05 (other sites)	Fayetteville
2886	Ammonia		Different Detection Limit	Change <0.14 to <0.20	Portland
2899	Ammonia		Different Detection Limit	Change <0.14 to <0.20	Portland
2915	Ammonia		Different Detection Limit	Change <0.14 to <0.20	Portland
2942	Ammonia		Different Detection Limit	Change <0.14 to <0.20	Portland
2956	Ammonia		Different Detection Limit	Change <0.14 to <0.20	Portland
1999	TKN		Detection Limit is Weird	Delete <0.6	Cobb County
2097	TKN		Typo in Detection Limit	Delete <0.0 for <0.2	Fulton County
2257	TKN		Rows seem to be wrong	Change 46 for 1.2	Fayetteville
2257	TDS		Rows seem to be wrong	Change 1.2 for 46	Fayetteville
2300	Nitrogen Total Organic		Value = 0, Grab = 1.46	Delete value	Fayetteville
2336	Nitrogen Total Organic		Value = 0, Grab = 1.14	Delete value	Fayetteville
13	Phosphorus Dissolved		Low Value	Change 0.009 by 0.09	Arlington
1488	Phosphorus Dissolved		Values lower than DL	Change by <0.02	Knoxville
1527	Phosphorus Dissolved		Values lower than DL	Change by 0.02	Knoxville
1580	Phosphorus Dissolved		Values lower than DL	Change by 0.02	Knoxville
4079	Beryllium		Detection limit	Change <0.6 by <0.06	Greensboro
4245	Cadmium		Detection limit	Change <0.4 by <0.04	Greensboro
2150	Cadmium		Wrong Columns	Copy 16000 in TotCol 230 Fec	Fulton
2128	Cadmium		Wrong Columns	Cd, Tot col and Fec Col in correct columns	Fulton
1107	Cadmium		LD in cell	Replace by detection limit	Howard County
1110	Cadmium		LD in cell	Replace by detection limit	Howard County
2864	Cadmium		Detection limit	Replace <2.5 by <0.5	Ada County
1107	Chromium		LD in cell	Replace by detection limit	Howard County

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
1110	Chromium		LD in cell	Replace by detection limit	Howard County
2935-3419	Cyanide		Factor of a thousand.	Multiply by 1000	Texas
18	Cyanide		Wrong Detection Limit	Changed	Arlington
29	Cyanide		Wrong Detection Limit	Changed	Arlington
35	Cyanide		Wrong Detection Limit	Changed	Arlington
2460	Conductivity	1	The value is the detection limit	Deleted	CAALA001
2871	Conductivity	2.5	Conductivity was collected in grab sample	Deleted	IDADA002
2662	DO	11.6	Evaluated by temperature	Deleted	AZMCA001
3016	DO	10.2	Evaluated by temperature	Deleted	ORODA005
3076	DO	15	Evaluated by temperature	Deleted	OREUA001
3077	DO	12.2	Evaluated by temperature	Deleted	OREUA001
3078	DO	17	Evaluated by temperature	Deleted	OREUA001
3092	DO	12.1	Evaluated by temperature	Deleted	OREUA002
3093	DO	18.4	Evaluated by temperature	Deleted	OREUA002
3097	DO	14	Evaluated by temperature	Deleted	OREUA002
3107	DO	11.5	Evaluated by temperature	Deleted	OREUA003
3108	DO	19.2	Evaluated by temperature	Deleted	OREUA003
3115	DO	16.3	Evaluated by temperature	Deleted	MABOA001
3120	DO	21.8	Evaluated by temperature	Deleted	MABOA002
3122	DO	10.2	Evaluated by temperature	Deleted	MABOA002
3126	DO	15.4	Evaluated by temperature	Deleted	MABOA004
3129	DO	14.6	Evaluated by temperature	Deleted	MABOA005
3020	HARDNESS	<1	Weight of evidence compared with conductivity and TDS	Deleted	ORODA005
4065	TSS	66	Turbidity high but TSS low, checked with other parameters	TSS to 660	NCGRHUST

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
2619	Turbidity	5.2	Grab samples higher than composite	Calculate time composite	CACTA010
2620	Turbidity	4.3	Grab samples higher than composite	Calculate time composite	CACTA010
1350 - 1351	Various		Values don't seem correct. The community change lab after this samples	Delete event	GACLCOSI
65	TDS	1	Low value	Use <5	VACHCOF5
875	TDS	2	Low value	Use <5	VAFFOF10
1611	TDS	2	Low value	Use <5	TNKXTYWE
1645	TDS	5406	Elevated value without support	Deleted	TNMET410
2102	TDS	4000	Elevated value without support	Deleted	GAFUCOS1
2122	TDS	4100	Elevated value without support	Deleted	GAFUCOS2
2144	TDS	4200	Elevated value without support	Deleted	GAFUCOS3
2128 , 2150	Various		Values don't seem correct. The community change lab after this samples	Delete event	GAFUCOS2,GAFUCOS3
2155	TDS	<	Missing detection limit	Deleted	GAFUCOS3
2699	TDS	1290	Elevated value without support	Deleted	AZMCA003
2965	TDS	3	Low Value	Use <5	ORPOA006
2942	TDS	4	Low Value	Use <5	ORPOA005
3691	TDS	1	Low Value	Use <5	TXIRA002
3772	TDS	1	Low Value	Use <5	TXIRA002
3119	TDS	17900	Elevated value, but other samples support it.	Keep with ?	MABOA001
16	TSS	2	Low Value	Use <5	VAARLCV2
19	TSS	1	Low Value	Use <5	VAARLRS3
48	TSS	2	Low Value	Use <5	VACHCOF2
65	TSS	1	Low Value	Use <5	VACHCOF5
76	TSS	2	Low Value	Use <5	VACHCOF5
266	TSS	2	Low Value	Use <5	VAHATYH3
498	TSS	2	Low Value	Use <5	VANFTYN4
812	TSS	1	Low Value	Use <5	VAFFCOF3
842	TSS	2.5	Low Value	Use <5	VAFFCOF6
935	TSS	2	Low Value	Use <5	MDAACORK
965	TSS	1.8	Low Value	Use <5	MDBACOSC
1160	TSS	2.87	Low Value	Use <5	MDMOCOCV
1441	TSS	1	Low Value	Use <5	TNKXTYFC
1482	TSS	1	Low Value	Use <5	TNKXTYGV

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
2014	TSS	2	Low Value	Use <5	GACOC1A2
2028	TSS	2.2	Low Value	Use <5	GACOCOL2
2143	TSS	1	Low Value	Use <5	GAFUCOS3
2235	TSS	2	Low Value	Use <5	NCFVTRYO
3143	TSS	2.99	Low Value	Use <5	GAATAT02
3149	TSS	1.83	Low Value	Use <5	GAATAT01
3150	TSS	1.98	Low Value	Use <5	GAATAT01
3265	TSS	2	Low Value	Use <5	TXHCA005
3451	TSS	0.5	Low Value	Use <5	TXDAA004
3453	TSS	0.5	Low Value	Use <5	TXDAA004
3552	TSS	0.5	Low Value	Use <5	TXFWA003
3647	TSS	0.5	Low Value	Use <5	TXGAA003
3775	TSS	2	Low Value	Use <5	TXPLA002
3776	TSS	2	Low Value	Use <5	TXPLA002
3781	TSS	1	Low Value	Use <5	TXPLA002
1314	BOD	0.73	Low Value	Use <1	MDSHDTPS
1317	BOD	0.41	Low Value	Use <1	MDSHDTPS
1322	BOD	0.91	Low Value	Use <1	MDSHDTPS
3868	BOD	0.7	Low Value	Use <1	NCCHSHEF
32	COD	<150	Unusual Detection Limit	Deleted	VAARLTC4
2250	COD	1500	Unusual elevated value, no evidece	Deleted	NCFVTRYO
3479	COD	1300	Unusual elevated value, no evidece	Deleted	TXDAA006
1897	Ammonia	60.3	Unusual elevated value, no evidece	Deleted	KYLOTSR5
1907	Ammonia	60.5	Unusual elevated value, no evidece	Deleted	KYLOTSR6
1909	Ammonia	30.4	Unusual elevated value, no evidece	Deleted	KYLOTSR6
2699	Ammonia	64	Unusual elevated value, no evidece	Deleted	AZMCA003
8	NO2 NO3	13	Unusual elevated value, no evidece	Deleted	VAARLLP1
1314	NO2 NO3	7.05	Unusual elevated value, no evidece	Deleted	MDSHDTPS
2011	NO2 NO3	6.3	Unusual elevated value, no evidece	Deleted	GACOC1A2
2030	NO2 NO3	>0.2	Unusual Detection Limit	Deleted	GACOCOL2
2140	NO2 NO3	9.3	Unusual elevated value, no evidece	Deleted	GAFUCOS3
2966	NO2 NO3	6.5	Unusual elevated value, no evidece	Deleted	ORPOA006
1905	TN	0.39	TN < NH3	Deleted both	KYLOTSR6
1907	TN	0.9	TN < NH3	Deleted both	KYLOTSR6
1909	TN	3	TN < NH3	Deleted both	KYLOTSR6

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
1600	TN	1.5	TN < NH3	Deleted both	TNKXTYWE
4324	TN	3.42	TN < TKN	Deleted both	NCRASIT3
4387	TN	6.65	TN < TKN	Deleted both	NCRASIT6
3281	TN	50.2	Unusual elevated value, no evidece	Deleted	TXHOA001
3289	TN	33	Unusual elevated value, no evidece	Deleted	TXHOA002
3305	TN	28.9	Unusual elevated value, no evidece	Deleted	TXHOA003
3331	TN	49.7	Unusual elevated value, no evidece	Deleted	TXHOA005
1907	Nitrogen Nitrite	40	Unusual elevated value, no evidece	Deleted	KYLOTSR6
1907	Phosphate Ortho	60.1	Unusual elevated value, no evidece	Deleted	KYLOTSR6
2978	Phosphate Ortho	0.8	Unusual elevated value, no evidece	Deleted	ORPOA007
3419	Phosphorus Total	0.005	Low value	Deleted	TXDAA002
4073	Antimony	0.02	Grab sample	Deleted	NCGRHUST
2006	Cadmium Total	40.16	Unusual elevated value, no evidece	Deleted	GACOC1A2
2007	Cadmium Total	42.7	Unusual elevated value, no evidece	Deleted	GACOC1A2
2036	Cadmium Total	122	Unusual elevated value, no evidece	Deleted	GACOCOL2
2128	Cadmium Total	16000	Unusual elevated value, no evidece, seems to be wrong columns	Deleted	GAFUCOS2
1131	Chromium Total	120	Unusual elevated value, no evidece	Deleted	MDMOCOWP
797	Copper Total	396	Unusual elevated value, no evidece	Deleted	VAFFCOF2
889 - 900	Mercury Total	0		Deleted	PAPH
1790	Nickel Total	200	Unusual elevated value, Detection limit 20	Change by 20	KYLXWHL1
3299	Nickel Total	325	Unusual elevated value, no evidece	Deleted	TXHOA002
3321	Nickel Total	720	Unusual elevated value, no evidece	Deleted	TXHOA004
3504	Nickel Total	0.013	Low value	Deleted	TXDCA001
3515	Nickel Total	0.01	Low value	Deleted	TXDCA002
2456	Selenium Total	0.3	Low value	Deleted	CAALA001
2457	Selenium Total	0.4	Low value	Deleted	CAALA001
2458	Selenium Total	0.068	Low value	Deleted	CAALA001
2459	Selenium Total	0.2	Low value	Deleted	CAALA001
2460	Selenium Total	0.059	Low value	Deleted	CAALA001
2461	Selenium Total	0.13	Low value	Deleted	CAALA001

Table B1. Modified Values in the NSQD - *Continued*

Order	Constituent	Original value	Problem	Action	Location_ID
2462	Selenium Total	0.095	Low value	Deleted	CAALA001
1082	Silver Total	290	Unusual elevated value, no evidece	Deleted	MDCLCOJS
1262	Silver Total	90	Unusual elevated value, no evidece	Deleted	MDPGCOS4
2006	Zinc Total	0.11	Low value	Deleted	GACOC1A2
2007	Zinc Total	0.19	Low value	Deleted	GACOC1A2
3704	Zinc Total	1	Low value	Deleted	TXIRA002
3777	Zinc Total	1	Low value	Deleted	TXPLA002
3514	Runoff	2.296	High Value	Deleted	TXDCA002
3515	Runoff	0.909	High Value	Deleted	TXDCA002
3201	Runoff	3.25	High Value	Deleted	KATOBROO
1364	Runoff	1.318	High Value	Deleted	GACLCOSI
2401	Runoff	1.73	High Value	Deleted	NCFVWINS

Appendix C
Methods to Estimate Non-Detected Values in Stormwater Datasets

C1.Introduction

A few large stormwater quality databases have been prepared in the past 20 years (NURP 1983; Smullen 2002). The data collected generally shows that there are important variabilities in stormwater pollutant concentrations for different land uses. Other factors that some researchers have found to be important include: imperviousness, slope, and size of the watershed. However, these databases include numerous instances where the laboratory results are reported to be "below detection." Statistical analyses can be greatly affected by these uncertain values, depending on their number and percentage of occurrence. There are several schemes that have generally been used to overcome the problems associated with these non-detected values.

The NSQD database has collected data representing more than 3,700 storm events in the U.S., including information about the location of the monitoring station, watershed characteristics, hydrology, and chemical constituents. Each community has the flexibility to choose the equipment and analytical methods to detect the constituents in the stormwater. Chemical constituents in this database had been preliminary analyzed for different land uses (Pitt, Maestre, and Morquecho 2003).

275

It has been observed while preparing the NSQD database, that different methods and procedures had been used for the analyses of the samples. The use of different methods generates different detection limits in the database for the same constituent.

Datasets containing values below the detection limits (censored data) complicate the statistical analyses, even including the basic calculations of the means and variance. Most of the time, "left-censored" data are of concern (observations below the detection limit). However, there are situations where "right-censored" data may occur, especially for bacteria analyses, when the observations are greater than the upper limit of the dilution. Three main approaches to the analysis of censored data can be found in the literature: substitution, statistical estimation, and graphical methods. In this chapter, these methods will be presented using different data sets.

C2. Analysis of Multiple Censored Data

Estimation methods for single censored data have been widely discussed in the literature. However, in the case of multiple censored data the situation is not the same. Helsel and Cohn (1988) continued the previous work of Guilliom and Helsel but for the multiple censored data.

Eight methods were studied in the multiple censored cases:

1. ZE: Censored data are assumed to equal zero.

2. DL: Censored data are assumed to equal the detection limit.

3. HA: Censored data are assumed to equal half the detection limit.

4. LR: Entire data set is log transformed and is assumed to be normally distributed. Censored data is estimated using least squares regression.

5. MR: Plotting positions are calculated using equations given by Hirsch and Stedinger.

6. LM: Concentrations are assumed lognormally distributed with parameters using the Cohen method. The mean and standard deviation of the untransformed were estimated using the equations given by Aitchison and Brown (1969).

7. MM: This method uses the maximum likelihood method but for the case of multiple censored data. Cohen (1976)

8. AM: Adjusted maximum likelihood procedure of Cohn (1988). The AM method is the same as the MM but makes a first order correction in the bias.

When the LR and LM methods were used all the points below the highest of the censoring thresholds were treated as less than that censoring level. This will simplify the problem as a single censored. The authors also pointed out that in the last three methods assumed that the data distribution is lognormal.

The results indicate that the MM and MR methods improve the results than those obtained with the single threshold assumption. The MR, MM and AM were also compared. It was found a higher (root mean squared error) RMSE for the moments estimated by the MM method. The AM method present lower error than the plotting position method (MR) but is less robust for distributions different than lognormal. The substitution methods present a higher error than the MR or the AM methods.

One of the main problems using these methods was to assume that that the data is lognormal. There is not certainty that water quality follows this distribution, for that reason robust methods are considered very important in water quality analysis. When data depart from the lognormal distribution the RMSE in the MM and AM methods can be higher than 1000% in the mean and standard deviation. Helsel and Cohn indicate that in water quality data the lognormal distribution and the gamma with a coefficient of variation of two are very common. The MR model present better results when the distribution is not known.

They also evaluate the plotting position using the Weibull, Blom and Hazel equations. The differences between the methods indicate that there is not a really effect in the results when any of this equations were used.

The authors suggest that if the distribution is unknown the MR method should be chosen. If there is certainty that the distribution is lognormal the AM method is recommended. The previous methods were evaluated with copper observations in commercial areas during the fall. Table C1 shows the original observations, and table C2 show the log-transformed observations.

Table C1. Copper Observations in Commercial Areas

2	2	3	5	5	5	5	5.2	5.4	5.5	6	6
6	6.5	6.5	7	8	8	8.1	8.4	9	9	9	10
10	10	10	10	10	10	10	10	10	11	11	13
13	13.4	14	14	14	14	14	14	14.4	14.5	15	15
15	17	17	17	17	17.1	18	19	19	20	20	20
20	20	20	20	20	20	21	21	21	22	22	22.4
23	24	24	26	26.6	26.9	29	29	30	30	33	33.7
36	37	37	40	41.3	42	50	50	50.5	50.7	59.4	60
60	61	62	70	100	130	130	175	<10	<10	<10	<10
<10	<10	<20									

Table C2. Copper Observations in Commercial Areas – (Log Values)

0.3	0.3	0.5	0.7	0.7	0.7	0.7	0.7	0.73	0.74	0.78	0.78
0.8	0.81	0.8	0.9	0.9	0.9	0.9	0.9	0.95	0.95	0.95	1
1	1	1	1	1	1	1	1	1	1.04	1.04	1.11
1.1	1.13	1.2	1.2	1.15	1.15	1.2	1.2	1.16	1.16	1.18	1.18
1.2	1.23	1.2	1.2	1.23	1.23	1.3	1.3	1.28	1.3	1.3	1.3
1.3	1.3	1.3	1.3	1.3	1.3	1.3	1.3	1.32	1.34	1.34	1.35
1.4	1.38	1.4	1.4	1.43	1.43	1.5	1.5	1.48	1.48	1.52	1.53
1.6	1.57	1.6	1.6	1.62	1.62	1.7	1.7	1.7	1.7	1.77	1.78
1.8	1.79	1.8	1.9	2	2.11	2.1	2.2	<1	<1	<1	<1
<1	<1	<1.30									

Figure C1 shows the result in the probability plots when censored data was deleted, replaced by the detection limit or half of the detection limit and estimating the values below the highest detection limit using the LR method. The results indicated that there is a bias in the mean value when the censored data is deleted or replaced by the detection limit. When data is replaced by half of the detection limit or is estimated by the LR methods the results are very similar. Notice that in the LR method all the values below the highest detection limit are considered censored. This assumption changes the level of censoring from 6% to 58%, but even at this level of censoring the results are very close to those obtained with the substitution methods. Because the transformed data seems to follows a normal distribution, it is possible to estimate the moments using only the upper side of the line.

Probability Plot Copper - Commercial - Censored data deleted   Probability Plot Copper - Commercial - Censored data = DL

Probability Plot Copper - Commercial - Censored data = HA   Probability Plot Copper - Commercial - Censored data = LR
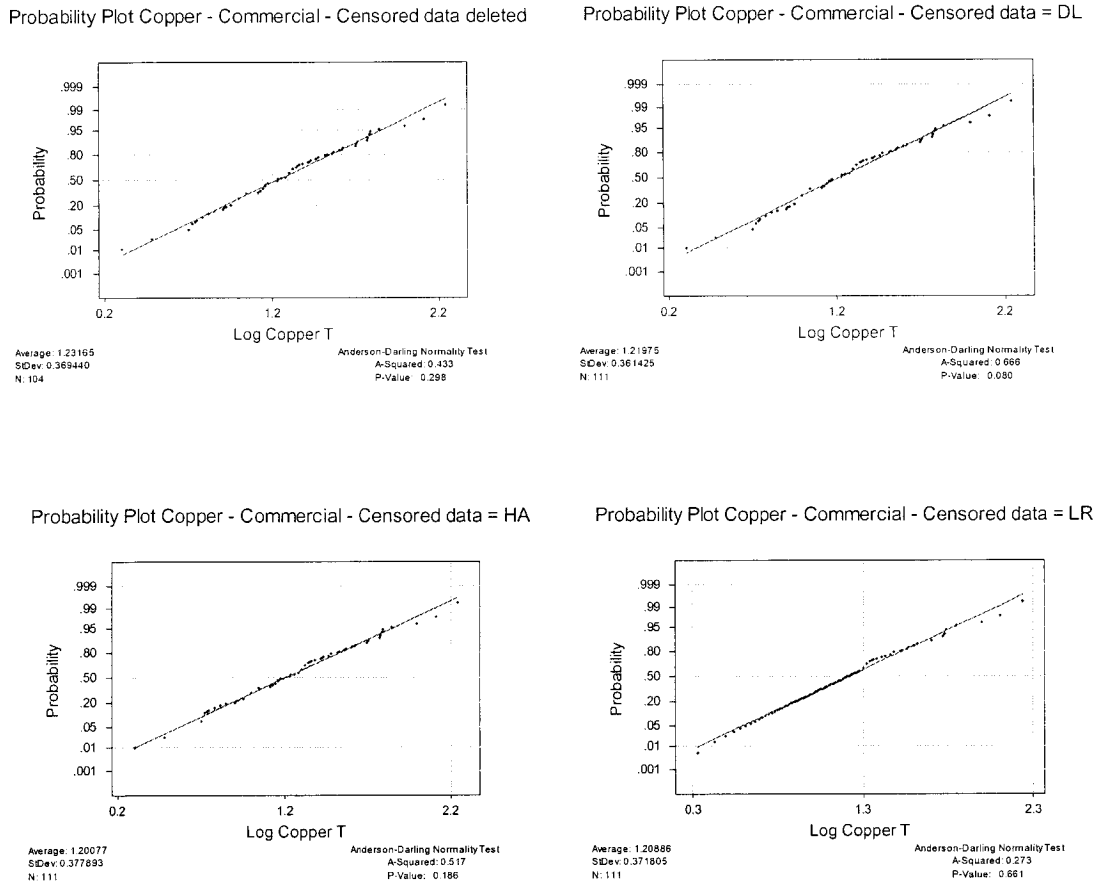
Fig. C1. Probability plot for different replacements of the censored data

The MR method uses the different detection limits that are in the sample set. The method was proposed by Hirsch and Stedinger (1987). They define a variable A_j as the number of uncensored data greater than the detection limit j and below the next detection limit. In the copper data there are 47 uncensored observations above the highest detection limit (A_3=47, DL=20). Between the detection limits 20 and 10 there are 34 observations (A_2=34, DL=10). Finally there are 23 uncensored observations above the minimum detection limit that in this case is zero (A_1=23, DL=0). In the case that the small value in the set is not a censored value this assumption must be made. It is also defined the parameter B_j as the number of censored and uncensored observations below the j

detection limit. In the example case there are 64 observations below the second detection limit (B_3=64), 29 observations below the first detection limit (B_2=29) and zero observations below the detection limit zero (B_1=0).

The method uses the probability of exceeding the jth detection limit $p_{e,j}$ to calculate the probability position of each observation.

$$(C.1) \qquad p_{e,j} = p_{e,j+1} + \frac{A_j}{A_j + B_j}\left(1 - p_{e,j+1}\right)$$

The calculations are easier from higher to lower. In the copper example there are three detection limits; by definition the probability of exceeding a fourth detection limit is zero. The probability of exceeding the third, second and first detection limit are 0.423, 0.735 and 1 respectively.

To calculate the plotting position of the censored data the Weibull formula was used in the range between the probabilities of exceeding boundaries.

$$(C.2) \qquad p(i) = (1 - p_{e,j}) + (p_{e,j} - p_{e,j+1})\frac{i}{A_j + 1}$$

This formula indicates that the values observed between the j and j+1 range are distributed using the Weibull formula. The plotting position for the censored data follows the same concept; distribute the censored data between the limits using the Weibull formula. For the censored observation the plotting positions can be calculated as:

$$(C.3) \qquad pc(i) = (1 - p_{e,j})\frac{i}{C_j + 1}$$

The formula calculates the position of the ith censored observation among the C

tied observations in the jth detection limit. The probability plot is shown in Figure C2.

Probability Plot Copper - Commercial - Censored data = MR
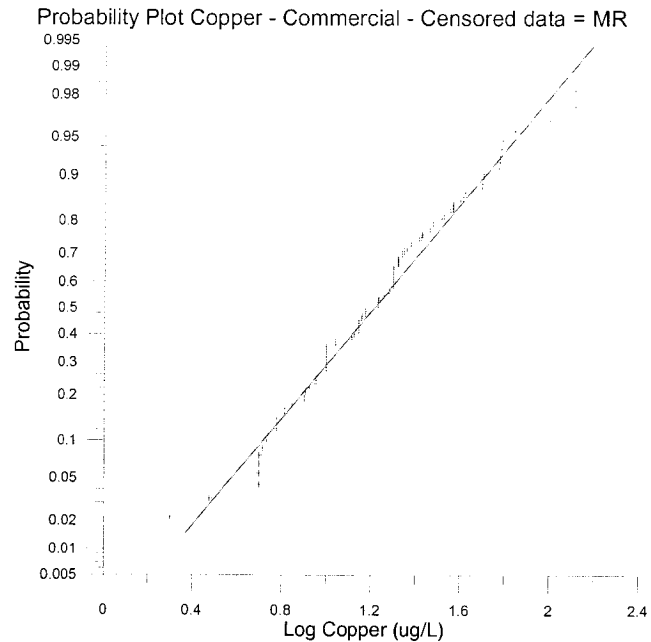
Fig. C2. Probability plot using the MR method.

The MM method presented by Cohen (1976) uses the maximum likelihood

method for the three parameter lognormal distribution.

(C.4)
$$f(x;\mu,\sigma,\gamma) = \frac{1}{\sigma\sqrt{2\pi}(x-\gamma)}\exp\left[-\frac{(\ln(x-\gamma)-\mu)^2}{2\sigma^2}\right]$$

$$\gamma < x < \infty$$

In this distribution X is a random variable lognormal (μ,σ^2,γ) and the

transformation Y = ln(X-γ) is normal (μ,σ^2). The method assumes that there are k

censored observations and n noncensored. For a censored level T_j the transformed y_j will be:

(C.5)
$$y_i = \ln(T_j - \gamma)$$

and,

(C.6)
$$\xi_j = \frac{y_i - \mu}{\sigma}$$

Cohen create a new variable Z_j that is used for solve the maximum likelihood estimators.

(C.7)
$$Z_j = Z(\xi_j) = \frac{\varphi(\xi_j)}{1 - \Phi(\xi_j)}$$

Where:

$\phi(\xi_j)$ = normal density function N(0,1)

$\Phi(\xi_j)$ = normal cumulative density function N(0,1)

Three simultaneous equations can be solved to estimate the parameters μ, σ, and γ.

(C.8)
$$\sum_{i=1}^{n} [\ln(x_i - \gamma) - \mu] + \sigma \sum_{j=1}^{k} Z_j = 0$$

(C.9)
$$\sum_{i=1}^{n}\left[\ln(x_i - \gamma) - \mu\right]^2 + \sigma^2\left[\sum_{j=1}^{k}\xi_j Z_j - n\right] = 0$$

Cohn suggested assuming a γ value and solve for μ and σ from the first two equations. After that the γ parameter can be recalculated using the third equation. In some cases the parameter g does not converge. In that case the following approximation must be used.

(C.10)
$$\gamma = x_k - \exp\left(\mu + \sigma\xi_k\right)$$

Where x_k is the kth order statistic in the sample.

The AM method is the best alternative in the case that the distribution is lognormal. In any other case the authors found that elevated bias and rmse are obtained in the mean and the variance (>1000%) if the distribution is different.

The National Council of the Paper Industry for Air and Stream Improvement had created seven technical bulletins about statistical methods used with environmental data sets (NCASI 1995). One of the reports presents a decision tree to select the appropriate statistical method and a description of the Cohen's multilevel MLE procedure. This method was recommended after compare it with other methods as replacement/deletion, D-log procedure, regression of normal order statistics balancing techniques and graphical techniques. At the end of this chapter it is shown the suggested decision tree in the case of multiple censored data.

The diagram indicates that in some cases of multiple detection limits, the problem can be solved using single censoring point (SCP) methods, for example when all the non-detected values are smaller than the detected values. In other situations the simplification cannot be done and multiple censoring point (MCP) methods must be used.

The Cohen's maximum likelihood method obtained the mean and variance estimates from the logarithm of the likelihood function and obtaining the partial derivate respect to the mean and the variance. The following equation defines the log likelihood function:

(C.11) $$L(S) = -n\ln(\sigma) - \frac{1}{2}\sum_{1}^{n}\left(\frac{x_i - \mu}{\sigma}\right)^2 + \sum_{1}^{k} r_i \ln[F_i] - n\ln[\sqrt{2\pi}]$$

Where:

S = sample set containing a total of N censored observations and fully quantified values

x_i = i^{th} fully quantified value

μ = population mean

σ = population standard deviation

k = number of censored levels

r_i = number of censored values at each censored level i

n = number of noncensored observations

F_i = F (ξ_i) = area under standard normal curve at f

ξ_i = $(T_i - \mu)/\sigma$, standard normal variate for the ith censoring level

$\phi(t)$ = $(2\pi)^{-1}$ exp $[-t^2/2]$, ordinate value of normal variate , f/(ξ_i)

T_i = the limit of detection of the ith level of censoring.

The derivates are:

(C.12)
$$\frac{\partial L}{\partial \mu} = \frac{n}{\sigma}\left[\frac{(\bar{x}-\mu)}{\sigma} - \sum_1^k \frac{r_i}{n}Z_i\right] = 0$$

(C.13)
$$\frac{\partial L}{\partial \sigma} = \frac{n}{\sigma}\left[\frac{s^2 + (\bar{x}-\mu)^2}{\sigma^2} - 1 - \sum_1^k \frac{r_i}{n}\xi_i Z_i\right] = 0$$

Z_i = the hazard function ϕ_i/F_i.

s^2 = Sample variance.

The report includes the program source code for use the Cohen's method in FORTRAN and SAS. The procedure and the code presented in the technical bulletin No. 703 were used to estimate the censored observations.

Appendix D
Unusual Sites Identified Using Xbar Plots

D1. Description

This appendix describes sites having unusual stormwater concentrations for all land uses, besides the residential areas that were described in Chapter 3.

D2. Evaluation of the Methods Selected to Estimate Non-Detected Observations

Three methods were used to estimate appropriate substitution values for the non-detected observations: delete them, replace them by half of the detection limit or estimate them using the Cohen's maximum likelihood method (an extrapolation of the probability plot of the data). The following discusses the analyses for each constituent for each land use category.

D.2.1 Hardness

Total hardness was observed for all samples, except in industrial land use areas where less than 2% of the samples were not detected. Changes in the average, median, standard deviation and coefficient of variation were not significant if the non-detected values were ignored, estimated, or replaced by half of the detection limit. Table D1 shows that there are no important differences in the industrial land descriptions use using any of these three methods.

287

Table D1. Summary Statistics for Estimated Observations for Total Hardness (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	250	250	250	139	139	139	138	138	138
% Detected	100.00			100.00			96.38		
Minimum	3.00	3.00	3.00	1.90	1.90	1.90	5.50	5.00	5.00
Maximum	401.00	401.00	401.00	356.00	356.00	356.00	888.00	888.00	888.00
Average	43.32	43.32	43.32	62.03	62.03	62.03	68.83	66.52	66.52
Median	32.00	32.00	32.00	38.90	38.90	38.90	39.00	38.50	38.50
Standard Dev.	44.87	44.87	44.87	65.17	65.17	65.17	104.55	103.32	103.32
Coeff. of Var.	1.04	1.04	1.04	1.05	1.05	1.05	1.52	1.55	1.55

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	8	8	8	127	127	127
% Detected	100.00			100.00		
Minimum	11.00	11.00	11.00	5.00	5.00	5.00
Maximum	270.00	270.00	270.00	1000.00	1000.00	1000.00
Average	145.25	145.25	145.25	57.19	57.19	57.19
Median	150.00	150.00	150.00	34.00	34.00	34.00
Standard Dev.	85.12	85.12	85.12	105.95	105.95	105.95
Coeff. of Var.	0.59	0.59	0.59	1.85	1.85	1.85

Figure D1 shows probability plots for industrial land use hardness values. The plot indicates that the mean value is smaller when the non-detected values are either estimated or replaced by half of the detection limit. The lower 40% of the distribution is displaced to the left. All the non-detected values were observed at 10 mg/L. The upper 60% of the distribution is not affected by the non-detected values. Oil and grease had censored data for 37% and 72% of the observations. Table D2 shows the differences in the descriptive statistics using the three methods. The greatest change occurred in the coefficient of variation values for freeway sites. The mean oil and grease values increased in a range of 30% to 60% when the censored observations were ignored. The difference was below 4% when the censored observations were replaced using Cohen's maximum likelihood method, or replaced by half of the detection limit.

Lognormal Probability Plot for Hardness in Industrial Land Use
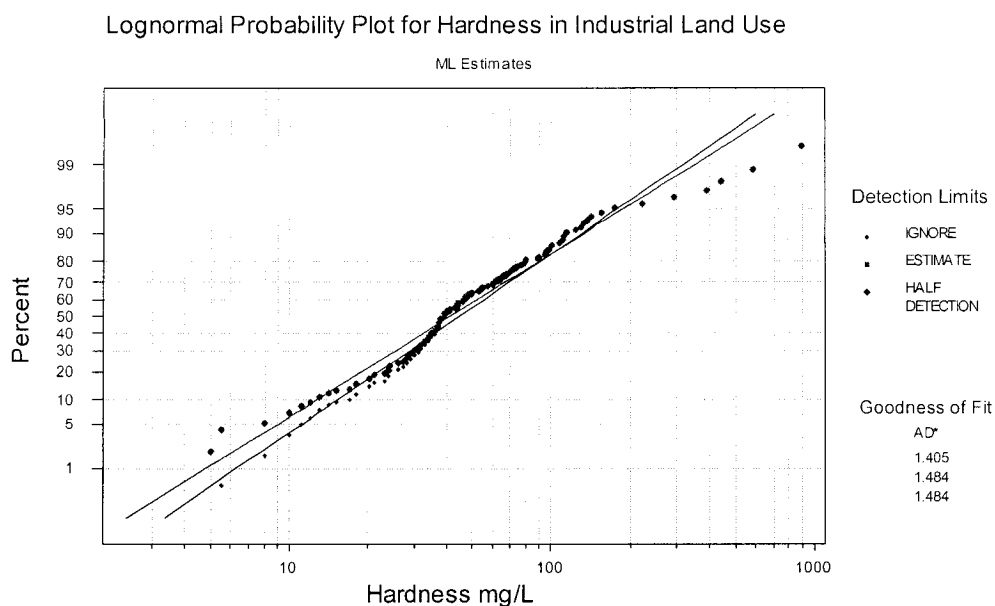
ML Estimates

Fig. D1. Estimated hardness distributions in industrial land use areas

Table D2. Summary Statistics for Estimated Observations for Oil and Grease (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	533	533	533	308	308	308	327	327	327
% Detected	57.79			70.78			65.14		
Minimum	0.20	0.02	0.20	0.80	0.03	0.25	0.50	0.00	0.25
Maximum	2980	2980	2980	359	359	359	11000	11000	11000
Average	22.85	13.87	13.89	12.63	9.42	9.39	62.87	41.40	41.39
Median	3.85	2.50	2.50	4.70	3.00	3.00	5.00	2.50	2.60
Standard Dev.	175.53	133.76	133.76	39.75	33.80	33.81	753.77	608.56	608.56
Coeff. of Var.	7.68	9.65	9.63	3.15	3.59	3.60	11.99	14.70	14.70

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	19	19	19	60	60	60
% Detected	36.84			71.67		
Minimum	0.50	0.50	0.50	3.00	0.50	0.25
Maximum	3.70	3.70	3.70	30.00	30.00	30.00
Average	1.53	1.09	1.09	8.49	6.57	6.45
Median	1.30	0.50	0.50	8.00	4.65	4.65
Standard Dev.	1.07	0.93	0.93	5.28	5.42	5.52
Coeff. of Var.	0.70	0.85	0.85	0.62	0.83	0.86

The probability plot in residential land use areas indicates that the lower tail is better described with the Cohen estimated method (Figure D2). The upper tail was the same for the estimated and the half detection limit method. About 40% of the non-detected values were at the < 1 mg/L level, and another 40% were at the < 5 mg/L level. The estimated values better describe the lower tail, however there was no significant differences in the means, standard deviations and coefficients of variation. This case is very important because the level of censoring was large (42.2%). Ignoring the non-detected values increased the mean value by more than 64% and the standard deviation by more than 30%, and reduces the coefficient of variation in 20%.

The analyses for commercial land use data resulted in a similar trend as observed for the residential land use areas (Figure D3). There is a better description of the lower tail, but the mean, standard deviation and coefficient of variation values are almost the same if the censored data are replaced by half of the detection limit or if they are estimated. The most frequent reported level of non-detected values was < 5 mg/L, followed by < 1 mg/L. The average was increased by 34%, and the standard deviation by 18%, when the censored data was ignored, and the coefficient of variation was reduced about 12% when the non-detected values were ignored.

Figure D4 shows the probability plot for oil and grease data at industrial land use areas and illustrates the case when an unusual value was present in the dataset. The maximum observation was larger by a factor of 2,200 compared with the median value of the distribution. This generates a coefficient of variation of 12 when the censored data are ignored, or 14.7 in the case when they are estimated or replaced by half of the detection limit.
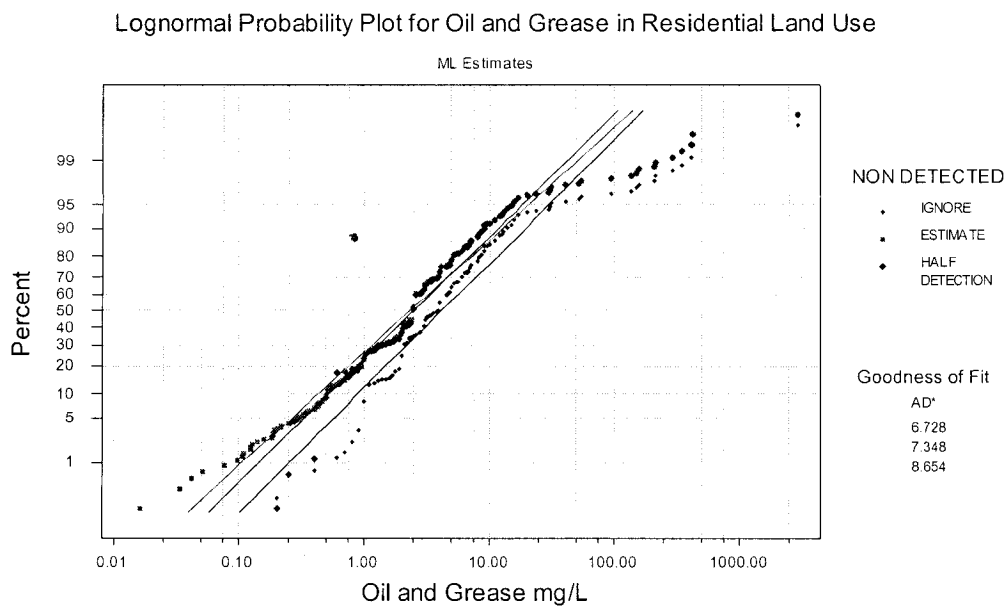
Lognormal Probability Plot for Oil and Grease in Residential Land Use

ML Estimates

Fig. D2. Estimated oil and grease distributions in residential land use areas

Lognormal Probability Plot for Oil and Grease in Commercial Land Use
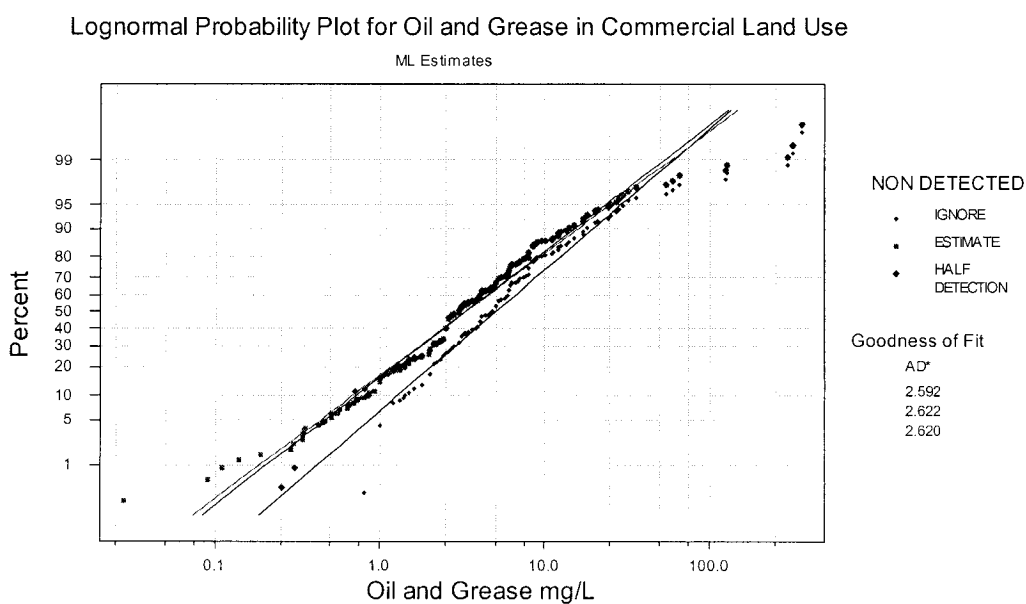
ML Estimates

Fig. D3. Estimated oil and grease distributions in commercial land use areas

The percentage of detected values for oil and grease in open space areas was very low (only 7 of 19 observations were detected) (Figure D5). Almost all of the non-

detected values were at < 1 mg/L. It was not possible to use the Cohen's maximum

likelihood method in this case because the percentage of non-detected values was too

high. Ignoring the non-detected values will increase the mean value by almost 40%

compared when the non-detected values were replaced with half of the detection limit.

The probability plot for freeway oil and grease values indicate that estimating or

replacing the censored observations for half of the detection limit does not cause a

significant difference in the coefficient of variation (Figure D6). The coefficient of

variation was 3% larger when half of the detection limit was used instead of Cohen's

method. A different situation occurs when the non-detected values were ignored. In this

case, the coefficient of variation was reduced by 30% compared with the estimated

method.
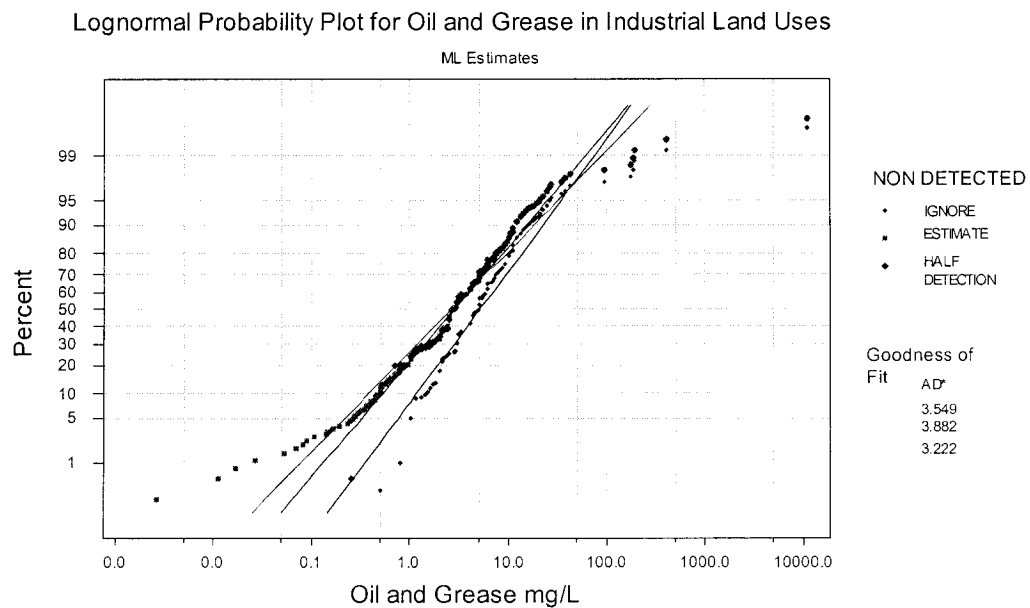
Fig. D4. Estimated oil and grease distributions in industrial land use areas
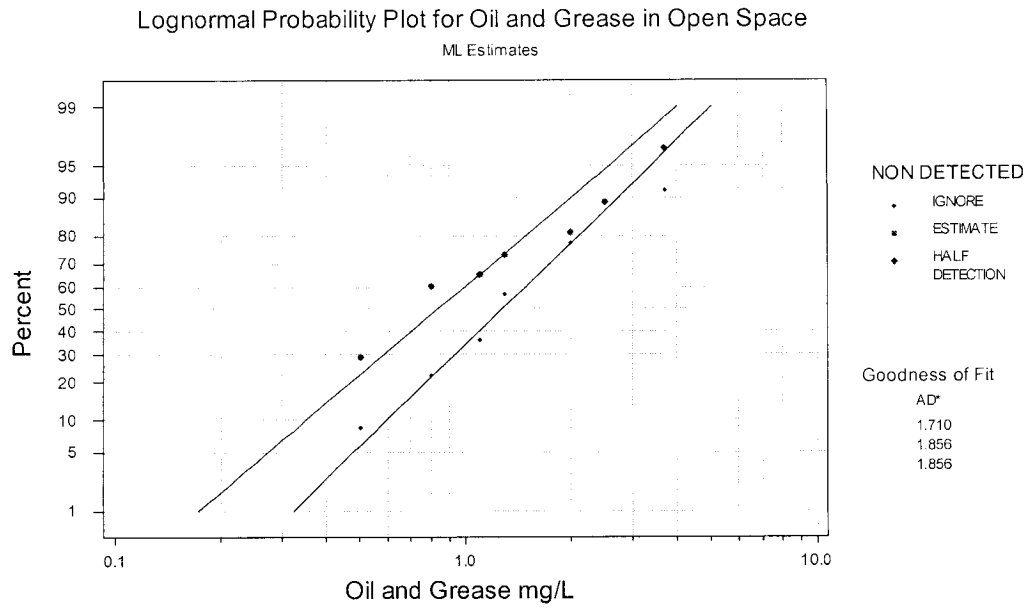
Lognormal Probability Plot for Oil and Grease in Open Space

Fig. D5. Estimated oil and grease distributions in industrial land use areas

Lognormal Probability Plot for Oil and Grease in Freeways Land Use
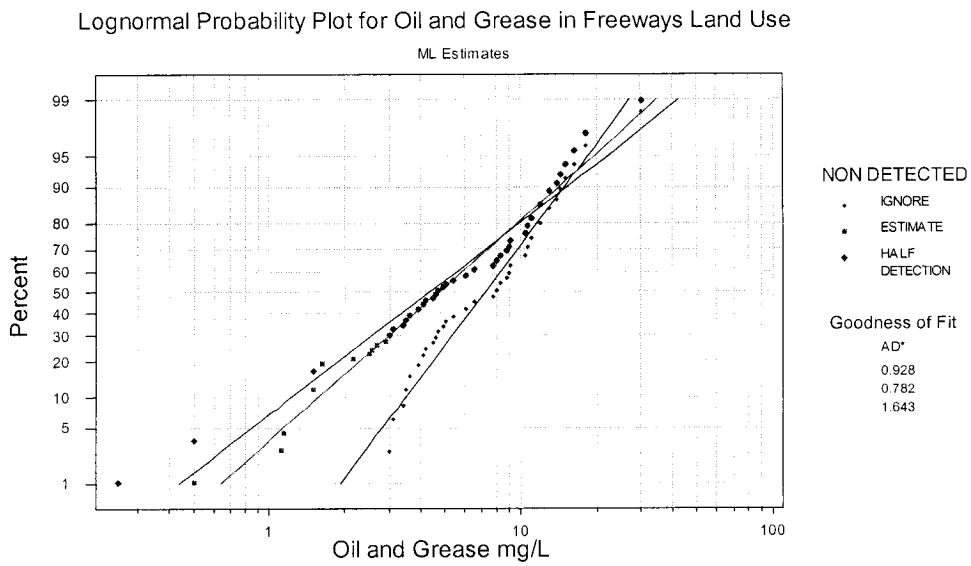
Fig. D6. Estimated oil and grease distributions in freeway land use areas

D.2.2 Total Dissolved Solids (TDS)

In all the land use categories, the percentages of non-detected TDS values were very low. The lowest percentage was observed in open space areas, with 2% not detected. No important differences were observed in the means, standard deviations and coefficients of variation when the non-detected values were ignored or estimated using with the Cohen method or substituting with half the detection limit. Descriptive statistics for each of the three methods are shown in Table D3.

Table D3. Summary Statistics for Estimated Observations for TDS (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	861	861	861	399	399	399	412	412	412
% Detected	99.19			99.50			99.51		
Minimum	3.00	3.00	0.50	4.00	4.00	4.00	4.50	1.78	2.50
Maximum	1700	1700	1700	3860	3860	3860	11200	11200	11200
Average	96.26	95.54	95.50	109.94	109.44	109.42	161.99	161.23	161.22
Median	72.00	70.50	70.50	74.00	74.00	74.00	91.00	89.50	89.50
Standard Dev.	102.45	102.35	102.38	208.76	208.36	208.37	582.40	581.09	581.09
Coeff. of Var.	1.06	1.07	1.07	1.90	1.90	1.90	3.60	3.60	3.60

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	45	45	45	97	97	97
% Detected	97.78			98.97		
Minimum	32.00	10.79	2.50	12.00	5.85	0.50
Maximum	542	542	542	470	470	470
Average	151.41	148.28	148.10	95.31	94.39	94.34
Median	124.50	119.00	119.00	77.50	77.00	77.00
Standard Dev.	109.83	110.58	110.82	76.38	76.52	76.59
Coeff. of Var.	0.73	0.75	0.75	0.80	0.81	0.81

Figure D7 shows the probability plots for residential land use TDS concentrations. The plot indicates that using half of the detection limit estimate lowers values compared to the Cohen's maximum likelihood method. The upper 95% of the distributions are

identical for the three cases. The probability plots don't indicate significant differences among the three methods for the remaining land uses. For example, Figure D8 shows the probability plots for commercial areas. The three lines overlap, except for a small fraction in the lower tail of the distribution.

Lognormal Probability Plot for TDS in Residential Land Use
ML Estimates

Fig. D7. Estimated TDS distributions in residential land use areas
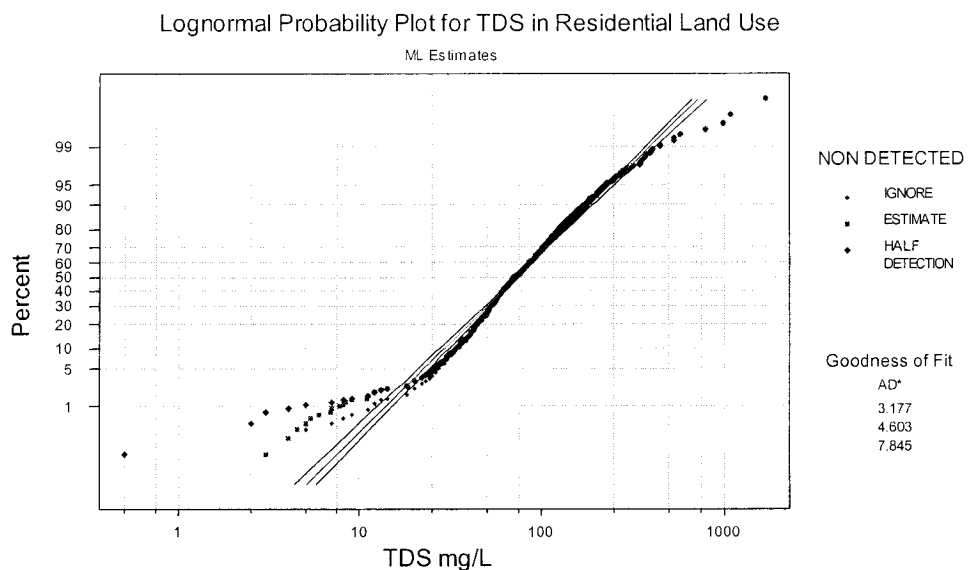
Lognormal Probability Plot for TDS in Commercial Land Use
ML Estimates
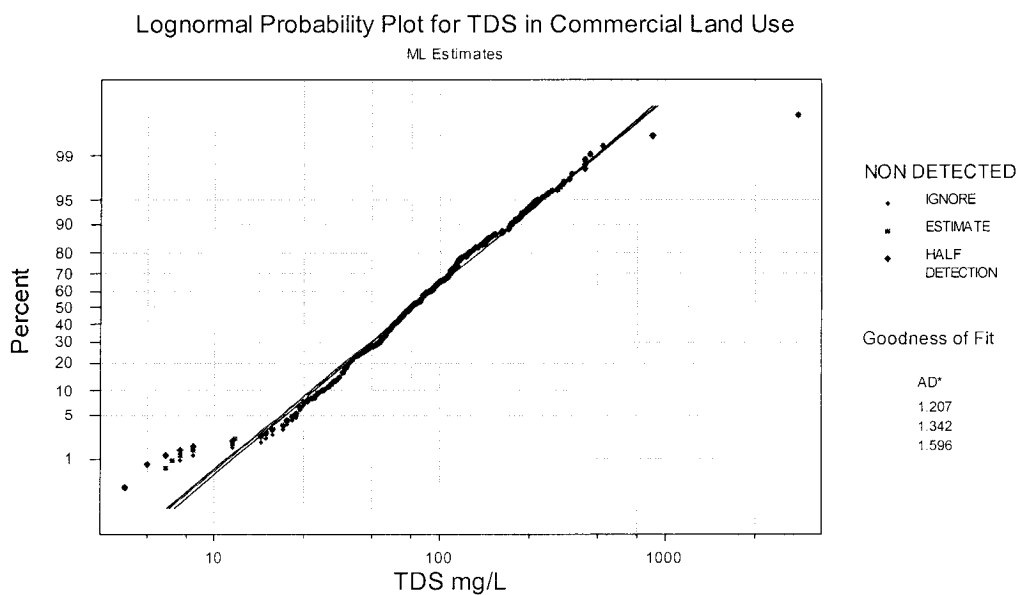
Fig. D8. Estimated TDS distributions in commercial land use

D.2.3 Total Suspended Solids (TSS)

The results for TSS were similar to above described results for TDS, the

maximum level of non-detected values was observed in open space areas, where about

5% of the observations were censored. Table D4 indicates that there are not any relevant

differences in means, standard deviations or coefficients of variation for any of the three

methods.

Table D4. Summary Statistics for Estimated Observations for TSS (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	991	990	991	458	457	458	427	426	427
% Detected	98.59			98.25			99.06		
Minimum	3.00	0.63	0.25	3.00	1.56	0.25	3.00	0.43	0.50
Maximum	2462	2462	2462	2385	2385	2385	2490	2490	2490
Average	99.84	98.53	98.46	110.06	108.45	108.18	142.44	141.36	141.12
Median	49.00	48.00	48.00	42.00	41.00	41.00	78.00	76.36	76.00
Standard Dev.	179.12	178.29	178.22	218.51	217.22	217.05	218.76	218.35	218.15
Coeff. of Var.	1.79	1.81	1.81	1.99	2.00	2.01	1.54	1.54	1.55

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	44	44	44	134	134	134
% Detected	95.45			99.25		
Minimum	3.00	1.22	0.50	3.00	3.00	0.50
Maximum	980	980	980	4800	4800	4800
Average	176.88	168.98	168.91	173.39	172.13	172.10
Median	48.50	39.00	39.00	99.00	98.50	98.50
Standard Dev.	263.04	259.44	259.49	448.85	447.39	447.41
Coeff. of Var.	1.49	1.54	1.54	2.59	2.60	2.60

The probability plots indicate that the lower values were better estimated using

half of the detection limit, rather than the Cohen's method. This indicate that with large

numbers observations and small percentage of non-detected values, replacing the missing

data by half of the detection limit will produce similar means compared to those obtained

when using the maximum likelihood method. Figure D9 shows the probability plot for

TSS concentrations for residential land use areas. The three curves overlap, indicating

than the three methods will produce practically the same result.

The probability plot for open space has the lower number of observations among

the five land uses. In this case, the pattern observed in the three methods was almost the

same. The coefficient of variation increases only 3% when the censored data was

estimated with the Cohen method, or replaced by half of the detection limit.
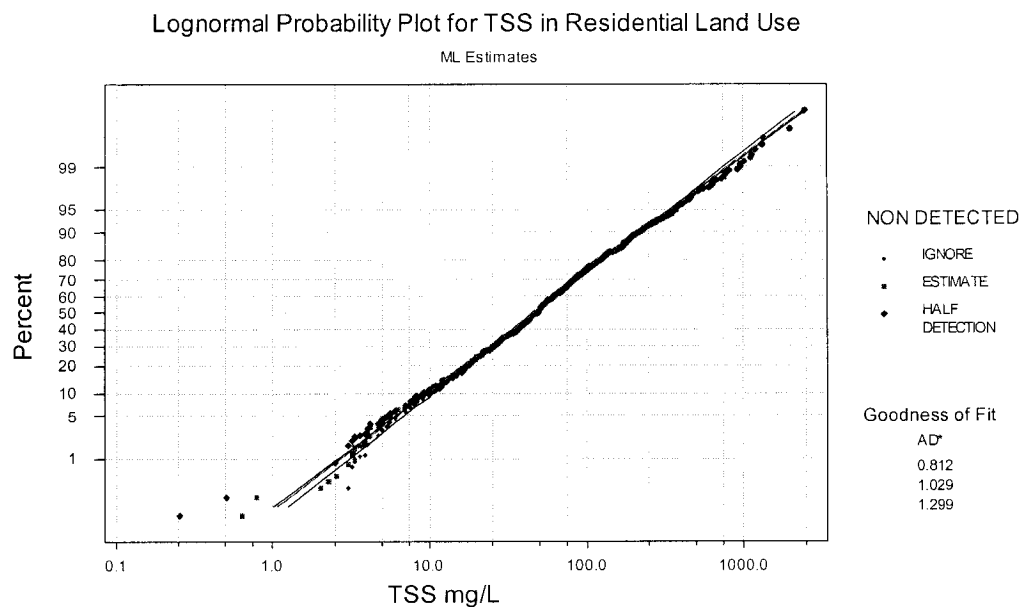
Fig. D9. Estimated TSS distributions in residential land use areas

D.2.4 Biochemical Oxygen Demand (BOD₅)

The percentage of non-detected values for BOD$_5$ was higher in open space and

freeway areas compared with the other land uses (Table D5). The lowest concentrations

were observed in open space areas with a median BOD$_5$ value of 4 mg/L. Freeways,

commercial and residential land use areas have similar concentrations, with 15 mg/L

average BOD$_5$ values. The highest BOD$_5$ concentration was observed at an industrial

land use site, however a single unusual BOD$_5$ observation of 6,920 mg/L had a large

effect on the mean, standard deviation and coefficient of variation values.

The lognormal probability plot for industrial land use areas showed one unusual

BOD$_5$ observation. This BOD$_5$ concentration was 35 times larger than the second highest

observation. This unusual value increased the standard deviation almost 18 times

compared with the other land uses. Figure D10 shows the probability plot for BOD$_5$

concentrations at industrial land use areas.

Table D5. Summary Statistics for Estimated Observations for BOD$_5$ (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	941	941	941	432	432	432	406	406	406
% Detected	97.56			97.45			95.32		
Minimum	1.00	1.00	0.50	2.00	0.75	0.50	1.00	0.55	0.50
Maximum	350	350	350	150	220	150	6920	6920	6920
Average	15.05	14.97	14.84	18.16	18.58	18.14	35.92	34.65	34.47
Median	9.00	9.00	9.00	11.00	11.00	11.00	9.00	9.00	9.00
Standard Dev.	22.25	22.34	22.11	20.25	22.59	20.63	351.89	343.62	343.61
Coeff. of Var.	1.48	1.49	1.49	1.12	1.22	1.14	9.80	9.92	9.97

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	44	43	44	26	26	26
% Detected	86.36			84.62		
Minimum	1.00	0.62	0.50	2.0	1.5	1.5
Maximum	20	20	20	89	89	89
Average	6.25	5.68	5.74	14.86	13.06	12.88
Median	5.40	4.00	4.00	8.0	6.5	6.5
Standard Dev.	4.30	4.34	4.38	18.68	17.67	17.76
Coeff. of Var.	0.69	0.76	0.76	1.26	1.35	1.38

Open space and freeway areas had the largest level of non-detected BOD$_5$ values.

The mean value for open space areas increased by 10% when the censored data were

ignored. No significance difference was observed for the variance values (Figure D11).

Estimating the non-detected value using Cohen's method, or replacing the non-detected

values by half of the detection limit results in almost the same means, standard deviations

and coefficients of variation values.

Fig. D10. Estimated BOD$_5$ distributions in industrial land use areas
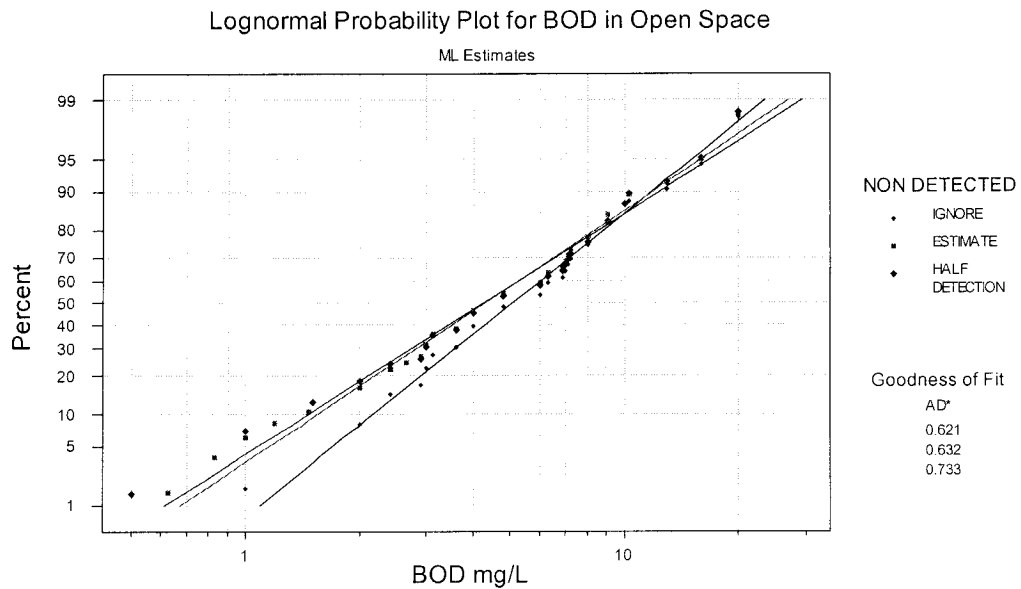
Fig. D11. Estimated BOD$_5$ distributions for open space land use areas

D.2.5 Chemical Oxygen Demand (COD)

Differences in the means, averages and coefficients of variation for COD

concentrations between the different methods for replacing the censored data were not

important, except for the open space land use area where the level of non-detected

observations was high (close to 25%) (Table D6). In the remaining land use areas, the

frequency of non-detected values was smaller than 2%.

Table D6. Summary Statistics for Estimated Observations for COD (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	796	796	796	373	373	373	361	361	361
% Detected	98.87			98.39			98.89		
Minimum	5.00	1.74	0.50	4.00	1.96	0.50	2.00	2.00	2.00
Maximum	620	620	620	635	635	635	1260	1260	1260
Average	74.34	73.55	73.52	94.11	92.70	92.63	103.23	102.26	102.17
Median	55.00	53.60	53.60	60.00	59.00	59.00	60.00	59.00	59.00
Standard Dev.	69.12	69.12	69.15	94.39	94.28	94.34	127.35	126.97	127.03
Coeff. of Var.	0.93	0.94	0.94	1.00	1.02	1.02	1.23	1.24	1.24

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	44	44	44	67	67	67
% Detected	75.00			98.51		
Minimum	8.00	3.70	5.00	2.44	2.44	2.44
Maximum	476	476	476	1012.82	1012.82	1012.82
Average	51.47	40.93	40.76	140.99	139.10	138.96
Median	42.10	24.85	24.85	100.00	100.00	100.00
Standard Dev.	79.11	70.73	70.78	148.89	148.56	148.69
Coeff. of Var.	1.54	1.73	1.74	1.06	1.07	1.07

One characteristic of the COD probability plot is that the lower tail does not

follow the trend showed by the rest of the distribution. Figure D12 shows an example

COD distribution for residential land use areas. This effect is increased when the

censored data is estimated or replaced by half of the detection limit.

The mean value in open space land use areas was increased by 25% when the censored data was ignored (Figure D13). In contrast, the coefficient of variation was reduced by almost 12 % when the non-detected values were ignored. No significant differences can be observed when the censored data was estimated using Cohen's method or replaced with half of the detection limit.
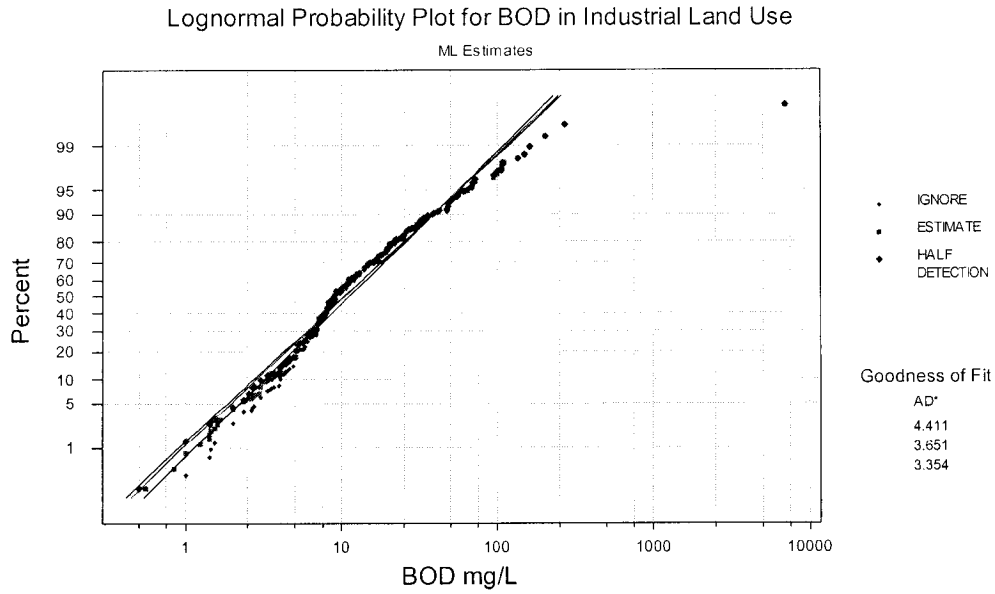
Fig. D12. Estimated COD distributions in residential land use areas
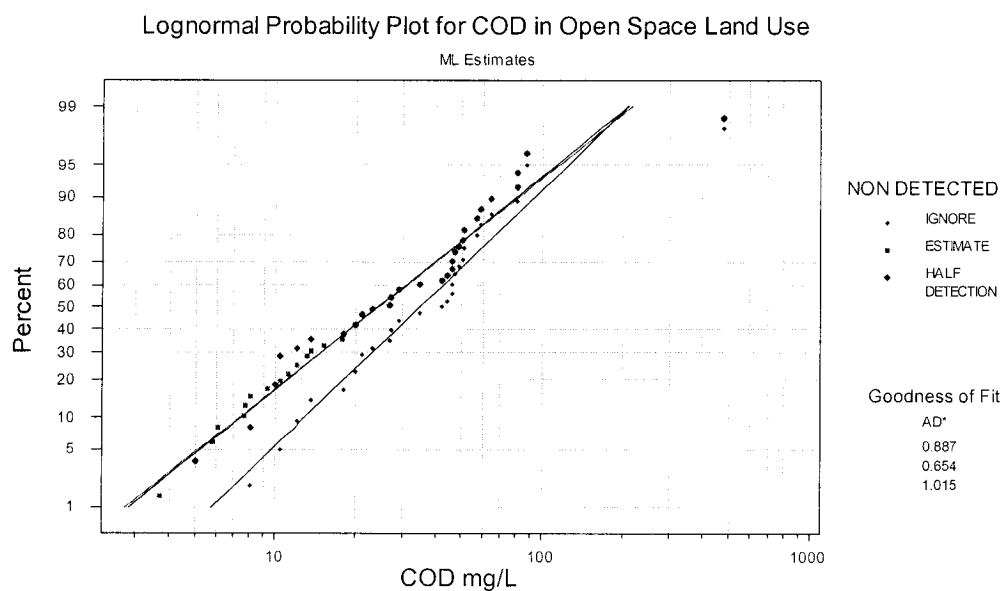
Fig. D13. Estimated COD distributions in open space land use areas

D.2.6 Ammonia (NH₃)

Ammonia had one of the largest levels of censored observations of the common stormwater constituents examined in detail (Table 3.16). The percentage of non-detected observations was about 20%, except for open space areas where it is more than 80%. The highest ammonia concentrations were observed at the freeway sites. Ignoring the censored observations increased the mean values by about 15%, while ignoring the non-detected values increased the coefficients of variation by almost 15%.

Table D7. Summary Statistics for Estimated Observations for Ammonia (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	595	595	595	299	299	299	253	252	253
% Detected	81.51			83.28			83.40		
Minimum	0.01	0.00	0.01	0.02	0.00	0.01	0.03	0.00	0.01
Maximum	5.60	5.60	5.60	7.80	7.80	7.80	9.84	9.84	9.84
Average	0.47	0.40	0.39	0.85	0.73	0.73	0.78	0.68	0.68
Median	0.32	0.27	0.25	0.50	0.41	0.40	0.47	0.38	0.36
Standard Dev.	0.51	0.48	0.48	1.02	0.97	0.97	0.96	0.91	0.91
Coeff. of Var.	1.09	1.20	1.22	1.20	1.32	1.33	1.23	1.35	1.35

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	32	32	32	79	79	79
% Detected	18.75			87.34		
Minimum	0.07	0.02	0.01	0.08	0.08	0.08
Maximum	1.80	1.80	1.80	11.87	11.87	11.87
Average	0.64	0.27	0.26	1.73	1.53	1.52
Median	0.18	0.25	0.25	1.07	0.90	0.90
Standard Dev.	0.79	0.38	0.38	2.24	2.16	2.16
Coeff. of Var.	1.24	1.43	1.44	1.30	1.41	1.42

The probability plots showed that replacing the non-detected values by half of the detection limit resulted in lower values than if Cohen's method was used. The Anderson Darling statistic for normality increased when the censored data was estimated. Figure D14 shows the probability plot for ammonia for commercial land use areas. In open

space areas, the estimated values don't seem to fit the log normal distribution (Figure D15). Estimating the censored observations using Cohen's method when more than 80% of observations were below the detection limit is certainly not recommended.
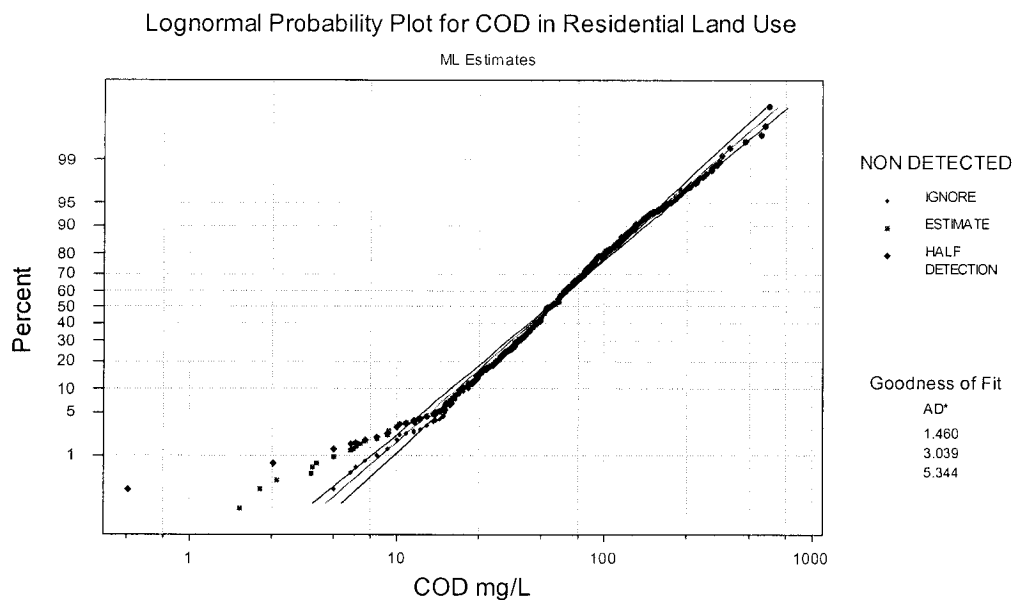

Fig. D14. Estimated ammonia distributions in commercial land use areas
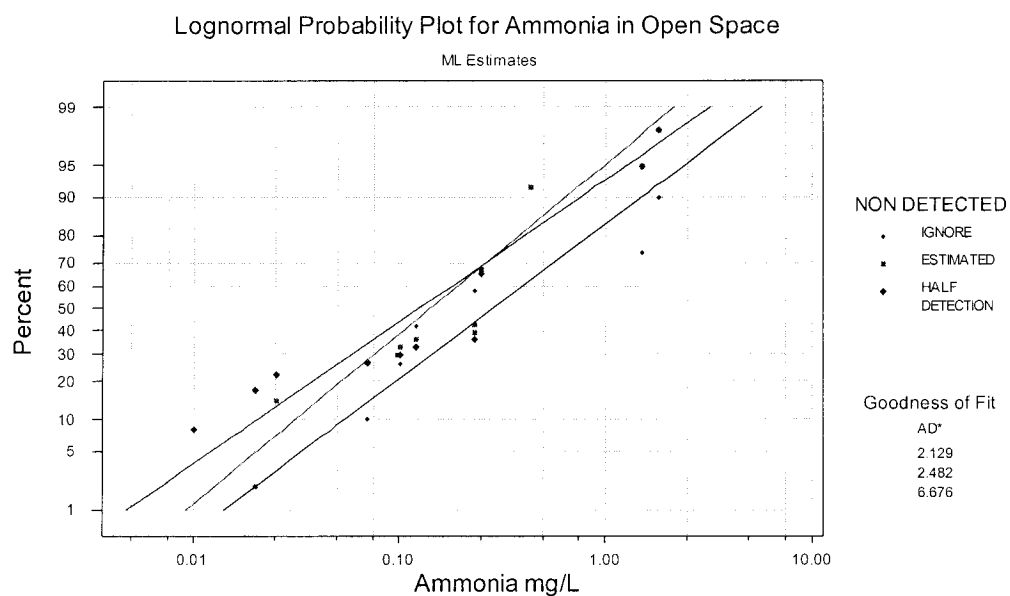
Fig. D15. Estimated Ammonia distributions in open space land use areas

D.2.7 Nitrite and Nitrate (NO₂ + NO₃)

D.2.7 Nitrite and Nitrate (NO$_2$ + NO$_3$)

The percentages of non-detected values was smaller than 5% in all the land uses

for nitrites and nitrates, except for open space areas where the level of censored values

was higher than 15%. There were no significant differences in the means, standard

deviations and coefficients of variation, except for the open space data set, when the

alternative substitution methods were used.

Table D8. Summary Statistics for Estimated Observations for NO$_2$ + NO$_3$ (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	927	927	927	425	425	425	417	417	417
% Detected	97.41			98.12			96.16		
Minimum	0.01	0.01	0.01	0.03	0.02	0.01	0.02	0.02	0.01
Maximum	18.00	18.00	18.00	8.21	8.21	8.21	8.40	8.40	8.40
Average	0.76	0.75	0.74	0.86	0.85	0.85	0.98	0.95	0.94
Median	0.59	0.58	0.58	0.61	0.60	0.60	0.73	0.72	0.70
Standard Dev.	0.87	0.86	0.86	0.91	0.91	0.91	0.87	0.86	0.86
Coeff. of Var.	1.14	1.15	1.16	1.06	1.08	1.08	0.89	0.91	0.91

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	44	44	44	25	25	25
% Detected	84.09			96.00		
Minimum	0.09	0.02	0.05	0.10	0.10	0.10
Maximum	3.33	3.33	3.33	3.00	3.00	3.00
Average	0.99	0.84	0.84	0.51	0.50	0.50
Median	0.59	0.50	0.50	0.28	0.28	0.26
Standard Dev.	0.88	0.88	0.88	0.63	0.62	0.62
Coeff. of Var.	0.89	1.04	1.04	1.23	1.23	1.25

The probability plots for residential, commercial and industrial land use areas

show a different trend for the lower tail of the distribution up to the 10th percentile for the

different methods. The departures from normality are more evident in the case when the

censored observations are replaced by half of the detection limit (Figure D16). In open

space areas, when the censored data was estimated or replaced, the coefficient of

variation increased almost 17% due the elevated level of censoring (Figure D17). There were no observed differences in the means, standard deviations and coefficients of variation when the censored values were replaced by half of the detection limit or estimated using Cohen's method.
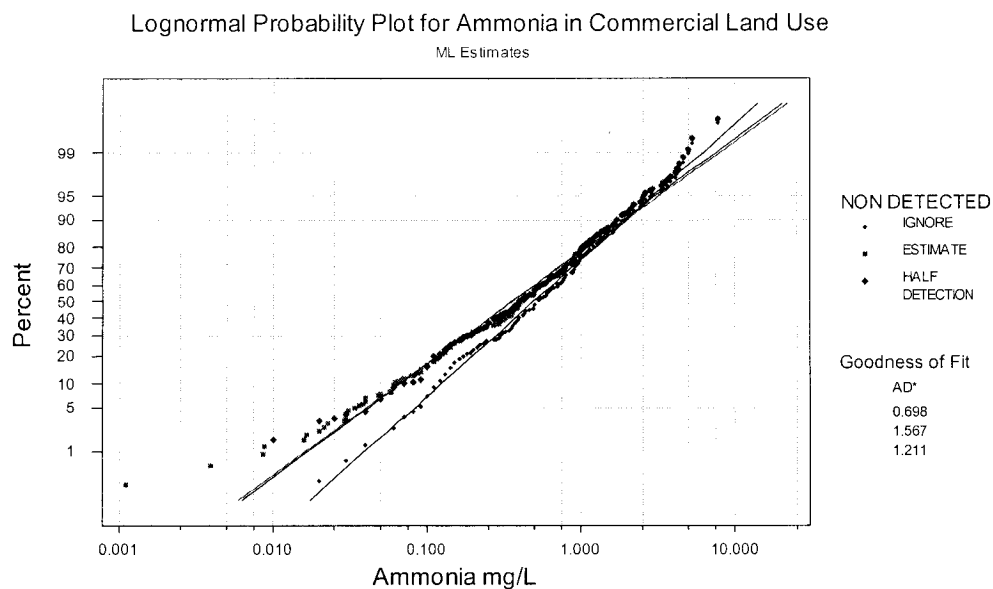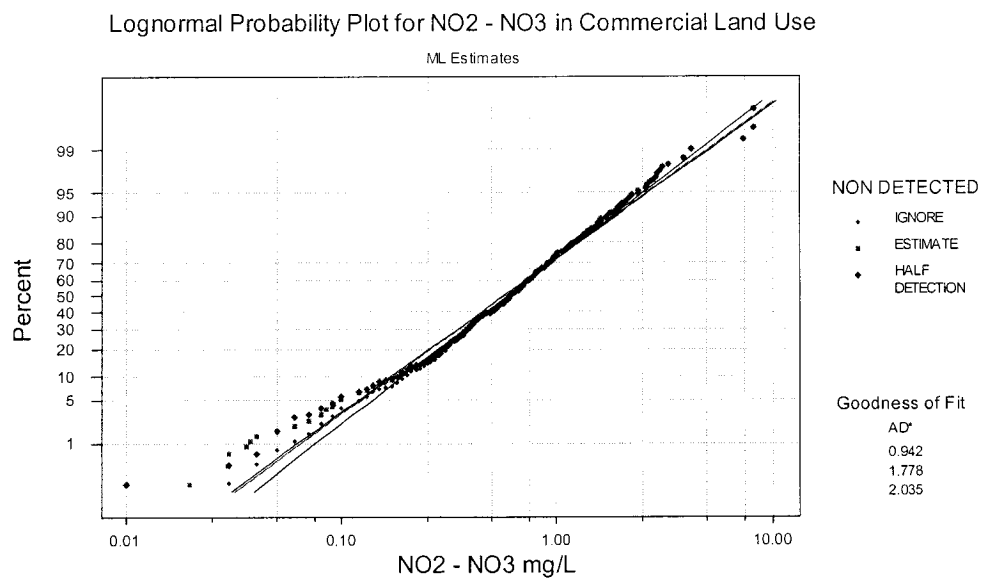

Fig. D16. Estimated nitrate - nitrite distributions in commercial land use areas
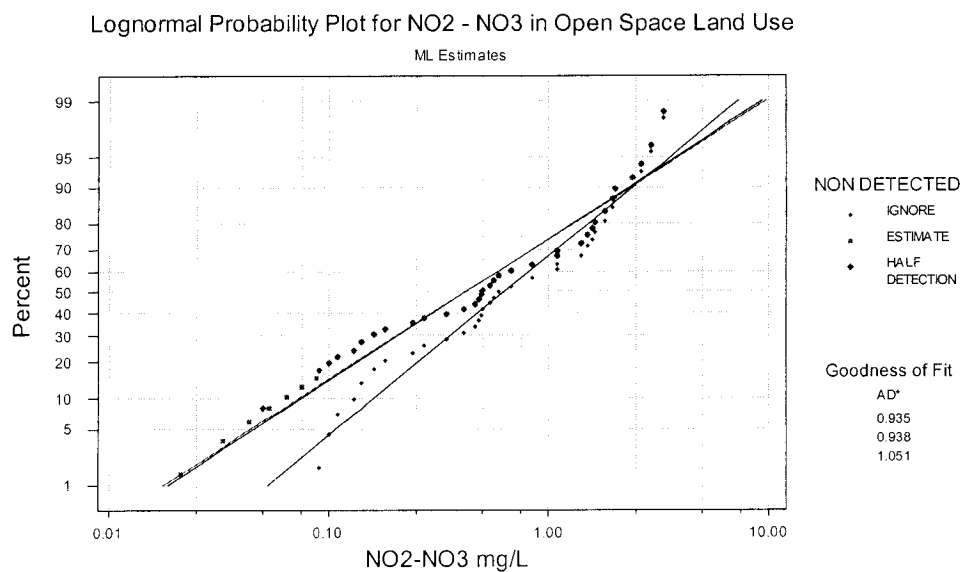
Fig. D17. Estimated nitrate - nitrite distributions in open space land use areas

D.2.8 Total Kjeldahl Nitrogen (TKN)

The level of censoring for TKN was smaller than 4% for all land use areas except

for open space areas. The highest TKN concentrations were observed in freeway areas,

and the lowest TKN concentrations were observed in open space areas (Table D9). Large

changes in the coefficient of variation were observed in open space areas when using

Cohen's method (an increase of 15%) and when replacing the censored values by half of

the detection limit (increases of 22%).

Table D9. Summary Statistics for Estimated Observations for TKN (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	957	957	957	449	449	449	439	439	439
% Detected	96.76			97.33			95.90		
Minimum	0.05	0.00	0.01	0.05	0.02	0.01	0.05	0.01	0.01
Maximum	36.00	36.00	36.00	15.00	15.00	15.00	25.00	25.00	25.00
Average	1.96	1.91	1.90	2.23	2.18	2.17	2.23	2.17	2.16
Median	1.43	1.40	1.40	1.59	1.55	1.55	1.40	1.37	1.37
Standard Dev.	2.05	2.04	2.04	2.08	2.07	2.08	2.56	2.53	2.54
Coeff. of Var.	1.05	1.07	1.07	0.93	0.95	0.96	1.15	1.17	1.18

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	45	45	45	125	125	125
% Detected	71.11			96.80		
Minimum	0.20	0.20	0.20	0.20	0.19	0.05
Maximum	4.70	4.70	4.70	36.15	36.15	36.15
Average	1.35	1.08	1.03	3.29	3.20	3.19
Median	0.74	0.50	0.50	2.00	1.93	1.93
Standard Dev.	1.20	1.10	1.13	4.49	4.44	4.45
Coeff. of Var.	0.89	1.02	1.09	1.37	1.39	1.39

The lognormal probability plot follows a straight line, except for the lower tail up

to the 5th percentile (Figure D18). The effect on the Anderson Darling statistic is

increased when the censored data is estimated. The effect is higher when the non-detected

values are replaced by half of the detection limit, instead of being estimated using

Cohen's maximum likelihood estimator. In open space areas when the level of censoring

is elevated and the number of observations is low, the Cohen's estimated method did not

follow a lognormal distribution. In Figure D19 two groups seem to exist, but it is

important to mention that more than 44% of the total TKN observations were lower than

0.5 mg/L. All the censored values in this land use were located at 0.5 mg/L TKN.
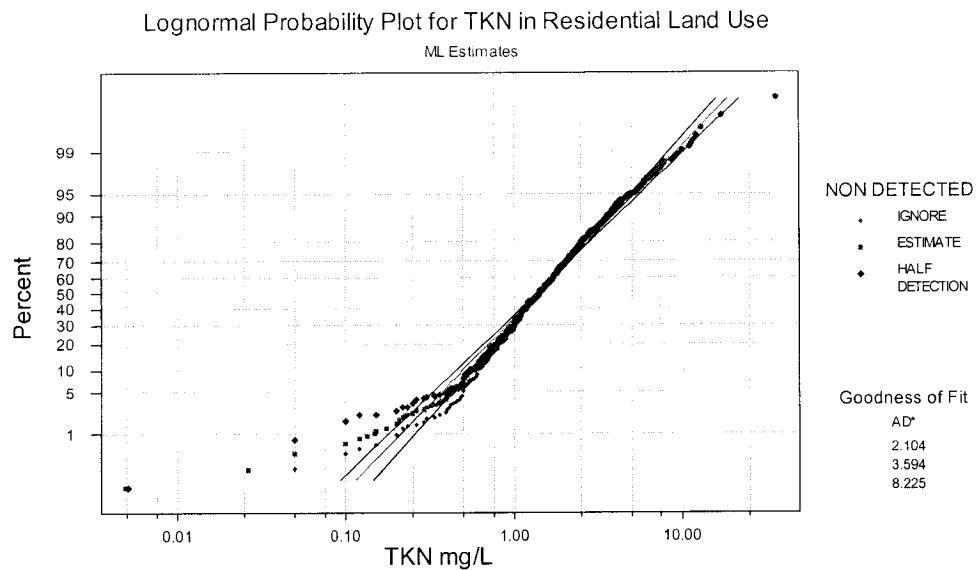
Fig. D18. Estimated TKN distributions in residential land use areas
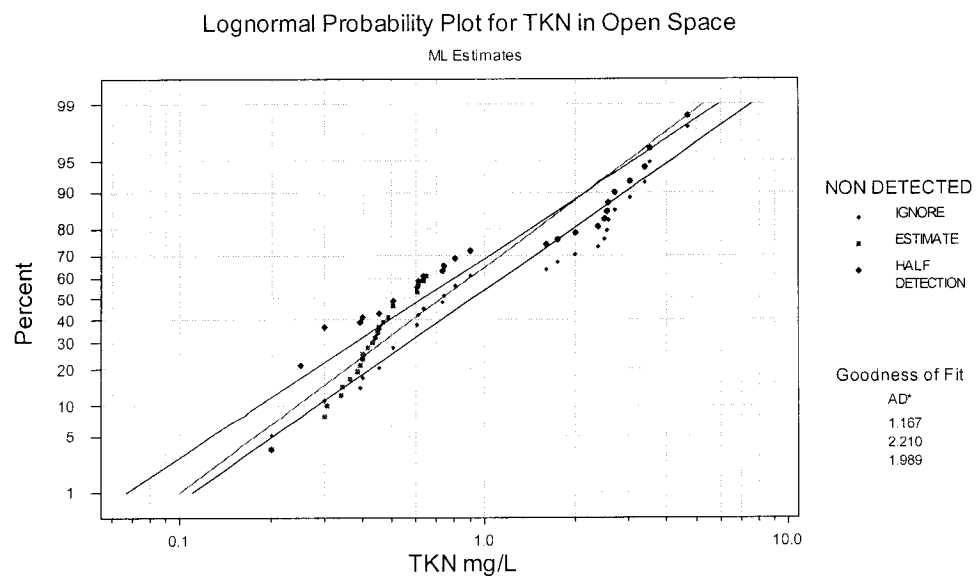
Fig. D19. Estimated TKN distributions in open space land use areas

D.2.9 Dissolved Phosphorus

Dissolved phosphorus has a large amount of non-detected values in all the land

use areas (about 13 to 20%), except for freeways were only 5% of the observations were

censored. In general, ignoring the non-detected values increased the means and standard

deviations and reduced the coefficients of variation. Table D10 shows the descriptive

statistics for dissolved phosphorus.

Table D10. Summary Statistics for Estimated Observations Dissolved Phosphorus (mg/L)

	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
Land use	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	738	738	738	323	323	323	325	325	325
% Detected	84.15			81.11			87.38		
Minimum	0.009	0.001	0.005	0.01	0.00	0.01	0.003	0.003	0.003
Maximum	1.69	1.69	1.69	1.60	1.60	1.60	1.60	1.60	1.60
Average	0.23	0.20	0.20	0.21	0.18	0.19	0.17	0.16	0.16
Median	0.17	0.14	0.14	0.11	0.09	0.09	0.11	0.10	0.10
Standard Dev.	0.21	0.21	0.21	0.27	0.25	0.25	0.20	0.19	0.19
Coeff. of Var.	0.94	1.04	1.05	1.24	1.35	1.34	1.18	1.23	1.23

	OPEN SPACE			FREEWAY		
Land use	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	44	44	44	22	22	22
% Detected	79.55			95.45		
Minimum	0.010	0.003	0.005	0.06	0.01	0.01
Maximum	0.52	0.52	0.52	6.97	6.97	6.97
Average	0.18	0.16	0.17	0.78	0.75	0.75
Median	0.13	0.09	0.14	0.20	0.20	0.20
Standard Dev.	0.16	0.15	0.15	1.66	1.63	1.63
Coeff. of Var.	0.89	0.95	0.87	2.13	2.18	2.18

As in the previous cases, ignoring the censored observations results in larger

mean values. There were no observed practical differences between the maximum

likelihood method and replacing the non-detected values with half of the detection limit

(Figure D20). , Dissolved phosphorus had the lowest level of censoring at freeway sites.

The probability plot indicates that the distribution is heavy in the tails; the slope between the 20th and 60th percentiles is higher than in the tails (Figure D21).

Lognormal Probability Plot for Dissolved Phosphorus in Industrial Land Use
ML Estimates
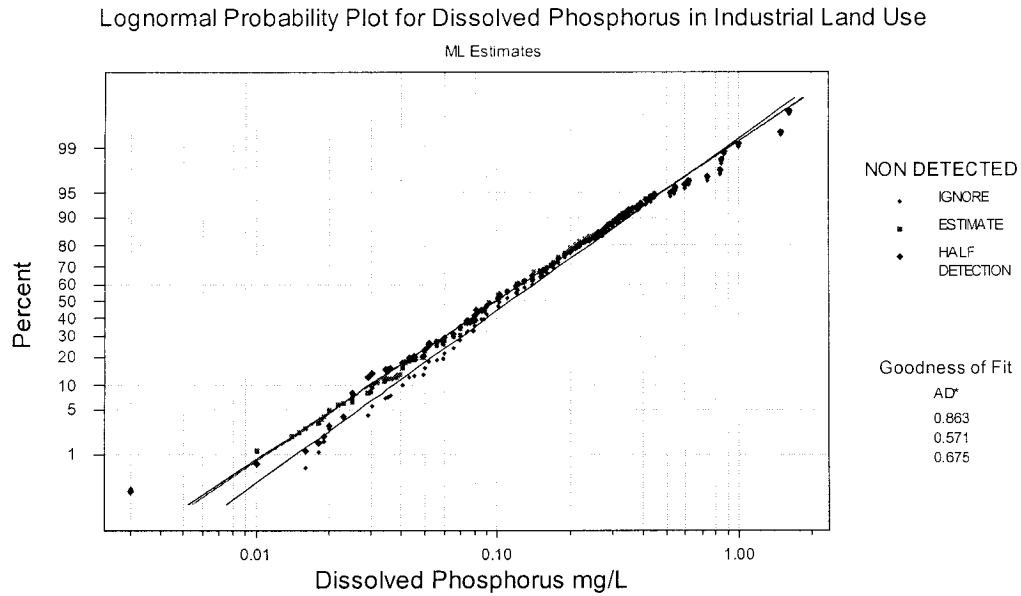
Fig. D20. Estimated dissolved phosphorus distributions in industrial land use areas

Lognormal Probability Plot for Dissolved Phosphorus in Freeways
ML Estimates
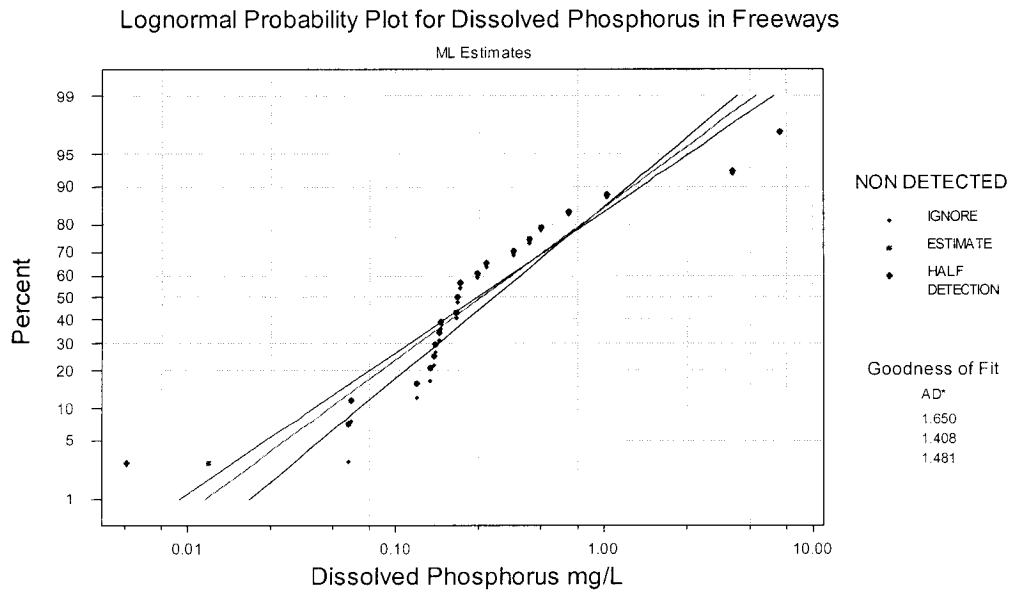
Fig. D21. Estimated dissolved phosphorus distributions in freeways land use areas

D.2.10 Total Phosphorus

Total phosphorus has low level of censored observations (less than 5%) at all land

use areas, except for open space (where it is close to 15%) (Table D11). Variations in the

coefficient of variation were not significant, except in open space areas where ignoring

the censored observations reduces the coefficient of variation by almost 7%.

Table D11. Summary Statistics for Estimated Observations for Total Phosphorus (mg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	963	963	963	446	446	446	434	434	434
% Detected	96.88			95.74			95.85		
Minimum	0.01	0.00	0.01	0.02	0.01	0.01	0.02	0.01	0.01
Maximum	6.90	6.90	6.90	3.35	3.35	3.35	7.90	7.90	7.90
Average	0.42	0.41	0.41	0.35	0.34	0.34	0.46	0.45	0.45
Median	0.30	0.30	0.30	0.22	0.22	0.22	0.26	0.25	0.25
Standard Dev.	0.47	0.47	0.47	0.40	0.39	0.39	0.64	0.63	0.63
Coeff. of Var.	1.13	1.14	1.14	1.16	1.16	1.16	1.39	1.41	1.40

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	46	46	46	128	128	128
% Detected	84.78			99.22		
Minimum	0.02	0.01	0.01	0.06	0.05	0.02
Maximum	15.40	15.40	15.40	7.19	7.19	7.19
Average	0.68	0.59	0.60	0.43	0.43	0.43
Median	0.31	0.22	0.25	0.25	0.25	0.25
Standard Dev.	2.43	2.24	2.24	0.76	0.76	0.76
Coeff. of Var.	3.54	3.77	3.74	1.76	1.77	1.77

When the censored data is ignored, the observations followed a lognormal

distribution. However, if the non-detected values are replaced by half of the detection

limit or estimated using the Cohen method, the lower tail has lower values than expected.

There is an unusual observation 20 times higher than the second highest

observation for the open space data (Figure D22). The most frequent non-detected

observation was <0.5 mg/L. Replacing the censored observations by half of the detection

limit produces values smaller than those estimated by Cohen's method. In the freeway

plot, it was observed that the higher observations are higher than the lognormal trend.

The upper 20th percentile has a different slope than the remaining observations shown on

the distribution.

Lognormal Probability Plot for Total Phosphorus in Open Space Land Use
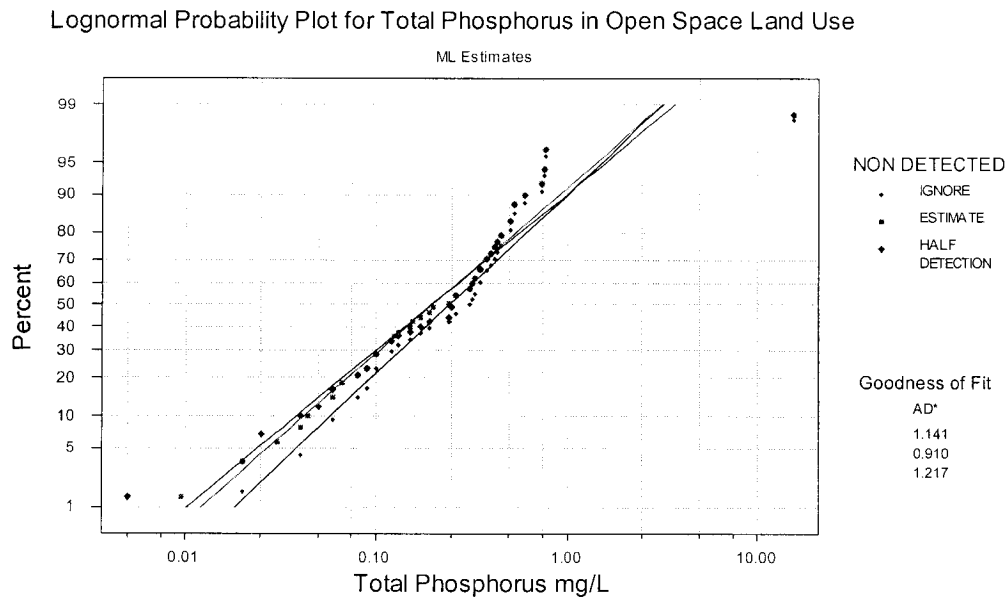
Fig. D22. Estimated total phosphorus distributions in open space land use areas

D.2.11 Total Cooper (Cu)

The levels of censoring for copper vary from 1 to 15% among the different land

uses. When the non-detected values are estimated or replaced by half of the detection

limit, the coefficients of variation increased between 1% and 6%, in addition there is a

reduction in the means and standard deviations. Table D12 shows the descriptive

statistics for each method by land use.

Table D12. Summary Statistics for Estimated Observations for Total Cooper (µg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	799	799	799	387	387	387	415	415	415
% Detected	83.60			92.76			89.64		
Minimum	1.00	0.25	0.23	1.50	1.50	1.00	1.97	1.77	1.00
Maximum	590	590	590	384	384	384	1360	1360	1360
Average	21.06	18.54	18.51	29.02	27.47	27.30	47.00	43.37	42.98
Median	12.00	10.00	10.00	17.00	15.60	15.00	21.88	20.00	20.00
Standard Dev.	38.51	35.70	35.69	42.92	41.73	41.79	93.81	89.47	89.60
Coeff. of Var.	1.83	1.93	1.93	1.48	1.52	1.53	2.00	2.06	2.08

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	39	39	39	97	97	97
% Detected	74.36			98.97		
Minimum	2.00	2.00	2.00	5.00	5.00	5.00
Maximum	210	210	210	244	244	244
Average	19.15	15.79	15.65	48.29	47.86	47.85
Median	10.00	5.30	5.00	34.70	33.40	33.40
Standard Dev.	38.97	33.98	34.00	45.91	45.87	45.89
Coeff. of Var.	2.04	2.15	2.17	0.95	0.96	0.96

The lognormal probability plots for residential and commercial land use areas indicate that the upper 5th percentile of the copper concentrations have higher values than expected if the distribution was lognormal. This observation is important because the upper tail of the distribution has an important effect in the mean and standard deviation values of the dataset.

In open space areas, replacing the non-detected values by the Cohen's method or replace the non-detected values reduce the means and standard deviations of the distribution by 18% and 13%, respectively. The probability plot for freeway areas is almost a perfect lognormal trend (Figure D23). In this case, the level of non-detected values was only 1%, and the difference in the coefficients of variations was also 1%.

Lognormal Probability Plot for Total Cooper in Open Space
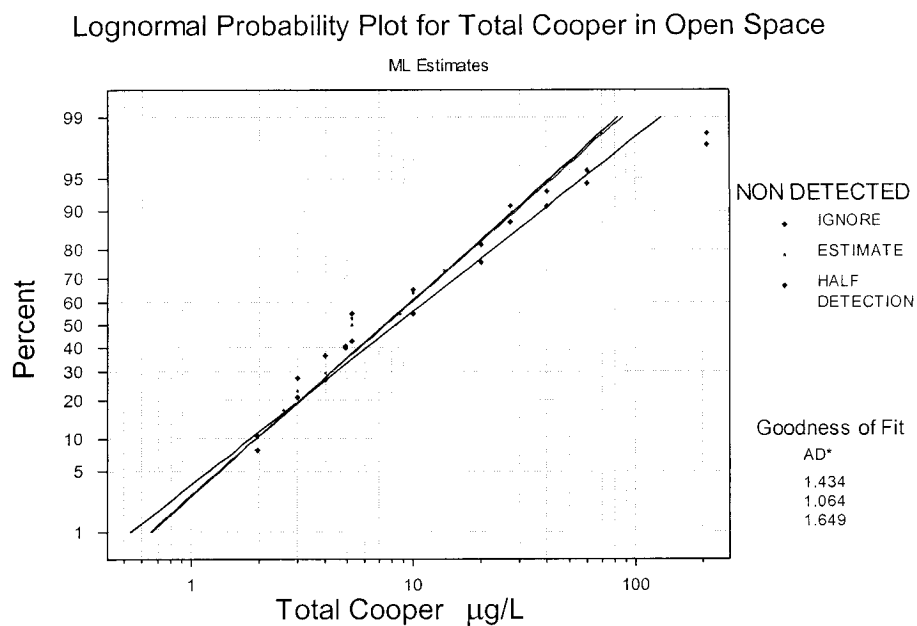
Fig. D23. Estimated total cooper distributions in open space land use areas

D.2.12 Total Lead

The level of non-detected values for lead varied from 0 to 58%. All the observations at the freeway sites indicate a presence of lead, in addition to the highest concentration among the land uses. Open land use areas had the highest level of non-detected lead values. There was about a 10% reduction in the coefficient of variation when the censored data was ignored. Table D13 shows the descriptive statistics for each method.

The probability plots indicate that when replacing the censored data by half of the detection limit, the values are smaller than when using Cohen's method (Figure D24). Estimating the censored values reduces the Anderson Darling statistic, providing a better fit with a lognormal distribution. In open space areas, most of the censored values were observed at < 40 mg/L, < 50 mg/L and < 100 mg/L. In all land use areas, almost 80% of

the lead observations were smaller than 50 mg/L. In open space areas, the estimated

means, standard deviations and coefficients of variation are dubious because most of the

censored observations were located in the upper part of the distribution (the frequency of

non-detectable observations was quite high, at about 58%).

Table D13. Summary Statistics for Estimated Observations for Total Lead (μg/L)

	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
Land use	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	788	723	788	377	355	377	411	377	411
% Detected	71.32			85.41			76.40		
Minimum	0.50	0.03	0.10	1.00	0.21	0.35	1.00	0.21	0.50
Maximum	585	585	585	689.07	689.07	689.07	1200	1200	1200
Average	26.00	21.03	22.08	37.42	34.27	33.84	70.10	59.52	57.49
Median	12.00	8.20	10.00	18.00	17.00	17.00	25.00	20.00	20.00
Standard Dev.	48.98	44.21	43.17	59.53	57.56	56.07	128.57	119.79	115.57
Coeff. of Var.	1.88	2.10	1.96	1.59	1.68	1.66	1.83	2.01	2.01

	OPEN SPACE			FREEWAY		
Land use	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	45	29	45	107	107	107
% Detected	42.22			100.00		
Minimum	0.20	0.08	0.10	1.60	1.60	1.60
Maximum	150	150	150	450	450	450
Average	28.39	19.21	23.98	48.77	48.77	48.77
Median	10.00	3.16	10.00	25.00	25.00	25.00
Standard Dev.	47.36	40.10	33.70	70.74	70.74	70.74
Coeff. of Var.	1.67	2.09	1.41	1.45	1.45	1.45

D.2.13 Total Zinc

The percentage of non-detected zinc values was smaller than 4%, except for open

space areas where it was close to 30% (Table D14). No important changes in the

coefficient of variations were observed, except for open space areas where ignoring the

censored values reduced the coefficients of variation by 13%.

Lognormal Probability Plot for Total Lead in Industrial Land Use
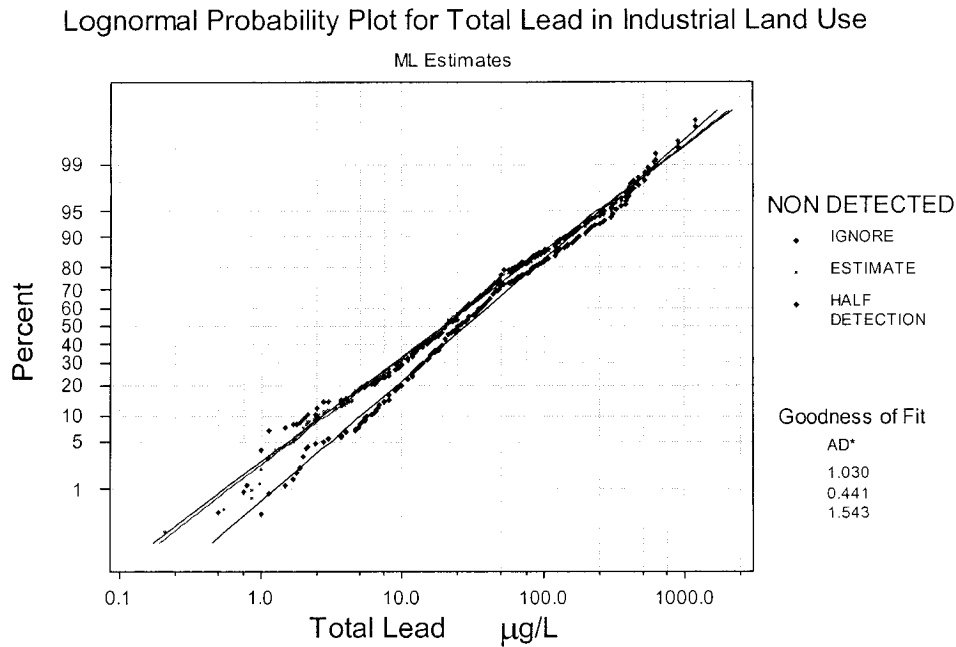
ML Estimates

Fig. D24. Estimated total lead distributions in industrial land use areas

Table D14. Summary Statistics for Estimated Observations for Total Zinc (µg/L)

Land use	RESIDENTIAL			COMMERCIAL			INDUSTRIAL		
	Ignore	Estimate	HD	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	810	810	810	392	392	392	432	432	432
% Detected	96.42			98.98			98.61		
Minimum	3.00	0.48	0.30	5.00	5.00	5.00	5.77	3.05	2.00
Maximum	1580	1580	1580	3050	3050	3050	8100	8100	8100
Average	116.70	113.53	113.23	225.32	224.06	223.55	318.25	315.02	314.34
Median	73.00	70.00	70.00	150.00	150.00	150.00	209.50	204.50	201.00
Standard Dev.	151.81	150.25	150.24	275.81	274.74	274.96	474.36	471.89	472.21
Coeff. of Var.	1.30	1.32	1.33	1.22	1.23	1.23	1.49	1.50	1.50

Land use	OPEN SPACE			FREEWAY		
	Ignore	Estimate	HD	Ignore	Estimate	HD
Observations	45	45	45	93	93	93
% Detected	71.11			96.77		
Minimum	5.00	2.00	2.50	6.00	6.00	2.50
Maximum	390	390	390	1829	1829	1829
Average	72.44	55.90	55.62	279.43	271.63	271.52
Median	40.00	20.00	20.00	200.00	194.49	194.49
Standard Dev.	96.88	85.85	85.99	281.16	279.87	279.98
Coeff. of Var.	1.34	1.54	1.55	1.01	1.03	1.03

The probability plot indicates that in the lower tail, replacing the non-detected observations by half of the detection limit will create smaller values than when estimate them using the Cohen's method (Figure D25). In open space areas, if the censored data are estimated using Cohen's method, there is a reduction in the mean and variance of the dataset of 23% and 12%, respectively, however the coefficients of variation increased by 15% (Figure D26).

Lognormal Probability Plot for Total Zinc in Residential Land Use
ML Estimates
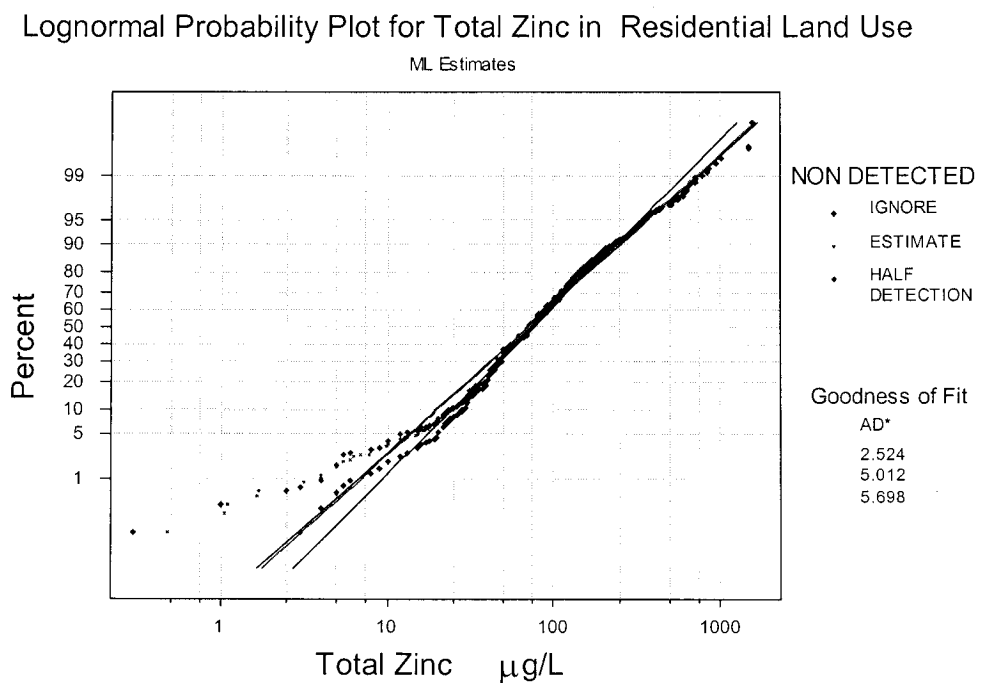
Fig. D25. Estimated total zinc distributions in residential land use areas

Lognormal Probability Plot for Total Zinc in Open Space Land Use
ML Estimates
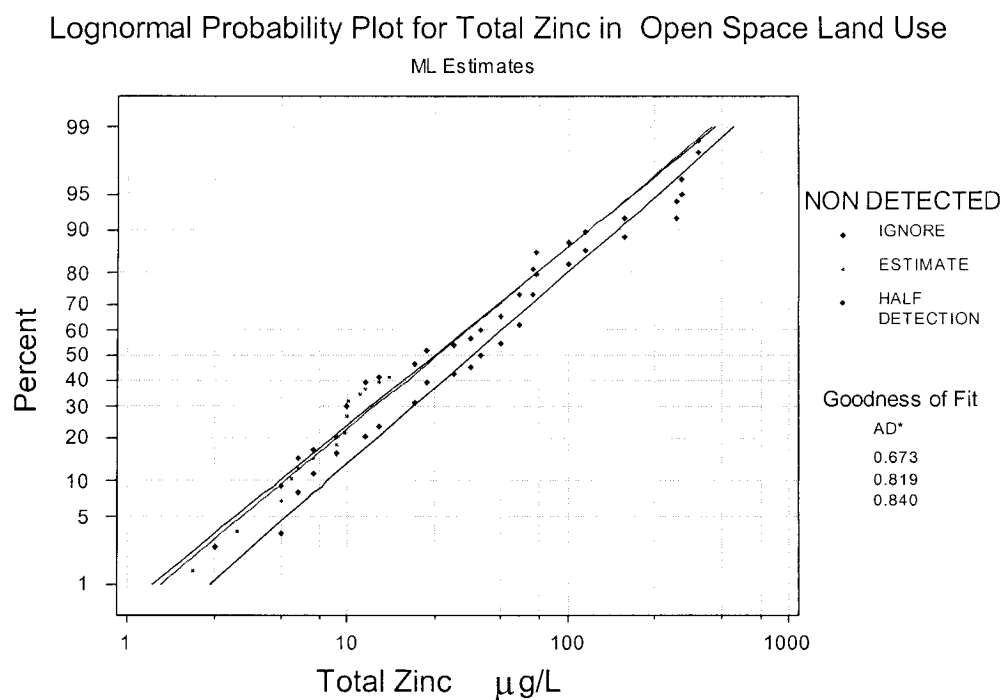
Fig. D26. Estimated total zinc distributions in open space land use areas

D.3 Sites with Unusual TSS Concentrations for Different Land Uses

This section presents the continuation of the example presented in Chapter 3,

section 3.4.1, where sites having unusual conditions were identified and examined more

carefully to try to understand the reasons for these values. A similar procedure was

followed in this appendix for the commercial, industrial and mixed land use areas to

complement the Chapter 3 analyses, which were conducted for residential areas only.

D.3.1 Residential and Mixed Residential Locations

The box and whisker plot (Figure D27 shows TSS concentrations by rain zone

and location) indicates that there is only one site that seems to have a different TSS

concentration probability distribution compared to the remaining sites in this group. The

site of interest is located in a residential-commercial area in Wooden Bridge Run, Philadelphia (PAPH1051), and has much lower concentrations that the other sites. Only two samples were collected at this site, and both were below 15 mg/L. The few samples available reduce the significance of this finding, however.

The results from the Xbar S chart analyses for mixed residential land uses are presented in Table D15. These analyses consider the numbers of samples and the variability of the data from each site, compared to the complete data set in the category being examined.

Table D15. Sites failing Xbar and S Chart Tests in Mixed Residential Land Use Areas

EPA Rain Zone	Sites Failing Xbar Chart Test	Sites Failing S Chart Test
ALL	9COCSA004(H) 2NCFVROSE(L) 7ORPOA005(H) 2TNKXTYGV(H) 5TXFWA005(H) 2VAVBTYV5(L)	GAFUCOS3(H)
1	None	None
2	TNKXTYGV(H) VAVBTYV5(L)	None
3	None	None
4	None	None
5	None	None
6	None	None
7	None	None
8	None	None
9	None	None

Boxplot TSS in Mixed Residential Landuses

Fig. D27. Box and whiskers plots for TSS concentrations at mixed residential land use areas

When 40 mixed residential sites were examined, only six sites were designated as being "out of control." These sites had unusual concentrations that were outside a band described by three standard deviations from the mean values. Two sites with means (log) below the lower control limit were located in EPA Rain Zone 2 (North Carolina and Virginia). The site located at Long Island Creek in Fulton County, Georgia has the largest standard deviation among the mixed residential sites examined. However, the S chart indicates that this site is in control compared to other sites in EPA Rain Zone 3.

There are 21 sites located in mixed residential land use and EPA Rain Zone 2, 17 sites have more than one observation each. Two sites, one above the upper control limit and one below the control limit, were observed in the Xbar chart. The site with the high median (log) value is located in Gallaher view, Knoxville, Tennessee (TNKXTYGV, 38 observations, median TSS = 105 mg/L). This site information included construction activity in the north part of the watershed, and a self-storage business, north and east of Cedar Hills apartments. The site located in Holland road, Virginia Beach (VAVBTYV5, 26 observations, median TSS = 32 mg/L) has wet ponds in the watershed that seem to control high concentrations, but the average value is the same as the other mixed residential sites.

The ANOVA analyses indicate that there is at least one EPA Rain Zone with TSS concentrations different than the other EPA Rain Zone with a p-value smaller than 1%. The Dunnett's comparison test at a family error rate of 5% indicates that EPA Rain Zones 5 and 7 have higher concentrations than those observed in EPA Rain Zone 2.

In summary, at a family error rate of 5%, higher concentrations occurred in EPA Rain Zones 5 (six sites, median TSS = 108 mg/L) and 7 (two sites, TSS = 175 mg/L)

compared with EPA Rain Zone 2 (21 sites, TSS = 59 mg/L). The Kruskall-Wallis test

indicates that there is a significant difference in the TSS median concentrations (with a p-

value close to zero). Site TNKXTYGV has higher characteristics than the other

residential mixed sites, most likely due to the noted construction activity close to the

outfall location.

D.3.2 Commercial and Mixed Commercial Locations

Box plots Xbar and S charts and ANOVA tests were used for commercial land

use data. Figure D28 identifies a site with high TSS concentration in EPA Rain Zone 4

(KATOJACK, 15 observations, median TSS = 603 mg/L). In general it seems that sites in

EPA Rain Zone 7 and 9 have higher concentrations than the other EPA Rain Zones. No

other trend or variation among EPA Rain Zone was identified from the box plot.

Boxplot TSS in Commercial Land Uses

EPA Rain Zone and Location_ID

Total Suspended Solids mg/L
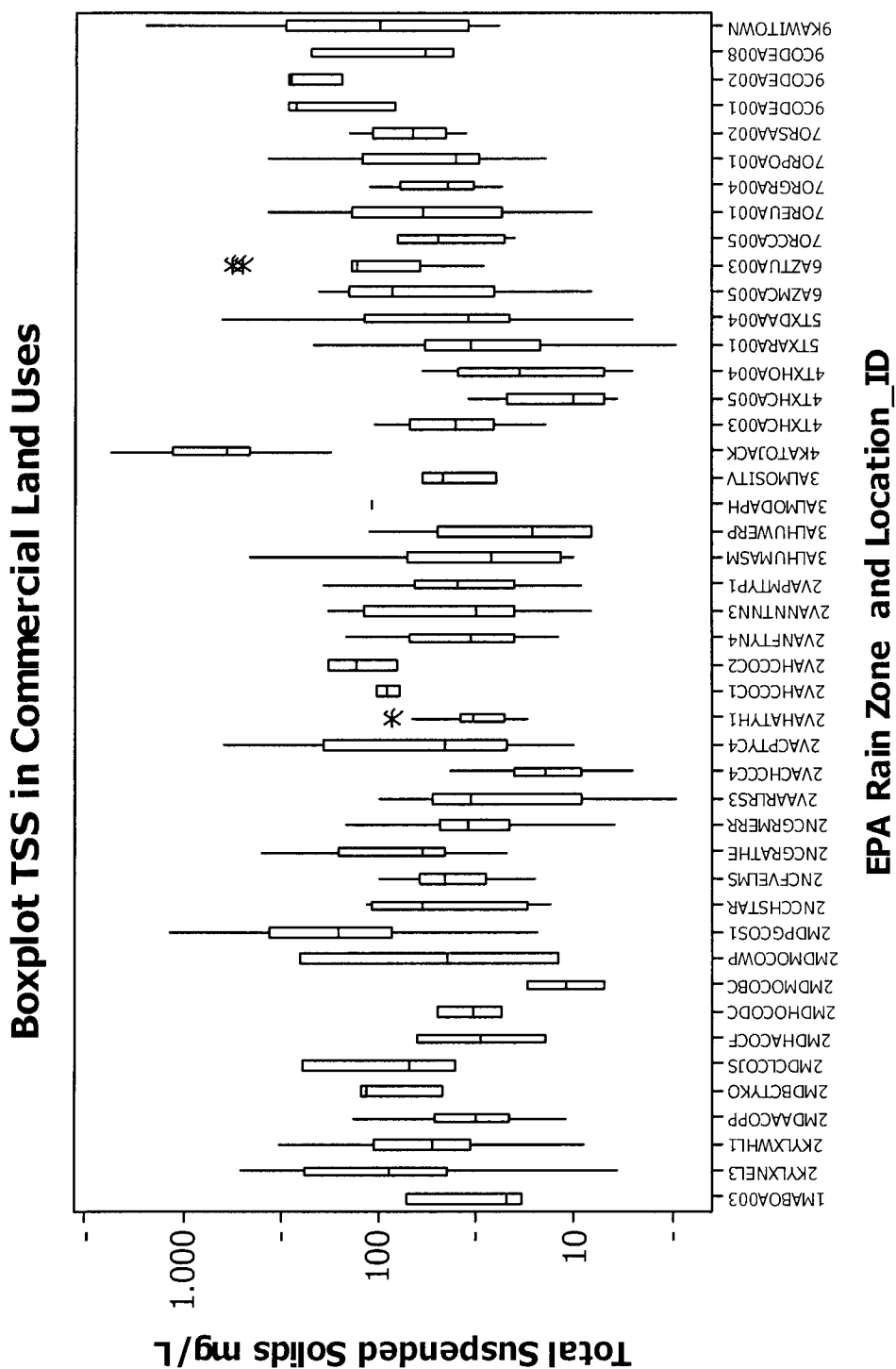
Fig. D28. Box and whiskers plots for TSS concentrations at commercial land use areas

The second approach was to identify unusual sites by EPA Rain Zone using Xbar

S charts. The results for commercial land uses are presented in Table D16.

Table D16. Sites failing Xbar and S Chart Tests in Commercial Land Use Areas

EPA Rain Zone	Sites Failing Xbar Chart Test	Sites Failing S Chart Test
ALL	2MDPGCOS1(H) 2VACHCCC4(L) 3ALHUWERP(L) 4KATOJACK(H) 4TXHCA005(L) 4TXHOA004(L) 9KAWITOWN(H)	None
1	None	None
2	2MDPGCOS1(H) 2VACHCCC4(L)	None
3	None	None
4	4KATOJACK(H)	None
5	None	None
6	None	None
7	None	None
9	None	None

The Xbar S plot did not indicate any trend by geographical region for the 45 sites.

Sites with low concentrations were observed in EPA Rain Zones 2, 3 and 4. There were

three sites identified with concentrations above the control limit, one in EPA Rain Zone

2, another in EPA Rain Zone 9, and a site identified by the box plot in EPA Rain Zone 4.

In EPA Rain Zone 2, two sites were found outside the control limits.

MDPGCOS1 is located in a shopping center in Arena Plaza, Price Georges County,

Maryland. 26 samples were collected at this location. The median TSS concentration for

this site is 158 mg/L. No reason was given for the high observed TSS concentrations. The

second site is located at Clover Leaf Mall in Chesterfield County, Virginia (VACHCCC4,

12 observations, 60 acres, median TSS = 14 mg/L). There is no clear reason that explains

the low concentrations found at this location. No sites outside the control limits were

found in other EPA Rain Zones except for EPA Rain Zone 4. This outfall is located in

Jackson Street in Topeka, Kansas. The high TSS concentrations may have been affected by tracking of sediment from a sand quarry close to the watershed. There were collected 16 samples collected between April 1998 and Septembers 2002.

The ANOVA test indicated that there was a significant difference among EPA Rain Zones (P-value = 0). The Dunnett's comparison test, with a family error of 5%, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS = 48 mg/L) are larger in EPA Rain Zones 4 (median TSS= 82 mg/L) and 9 (median TSS = 128 mg/L). The median TSS concentrations at the remaining EPA Rain Zones are not statistically different than the observed in EPA Rain Zone 2.

There are 24 sites located in mixed commercial land use with more than one observation. EPA Rain Zone 2 has the largest number of sites (10 sites), followed by EPA Rain Zone 5 (5 sites). Figure D29 shows the box plots for mixed commercial land uses by EPA Rain Zone.

Boxplot of TSS in Mixed Commercial Land Use
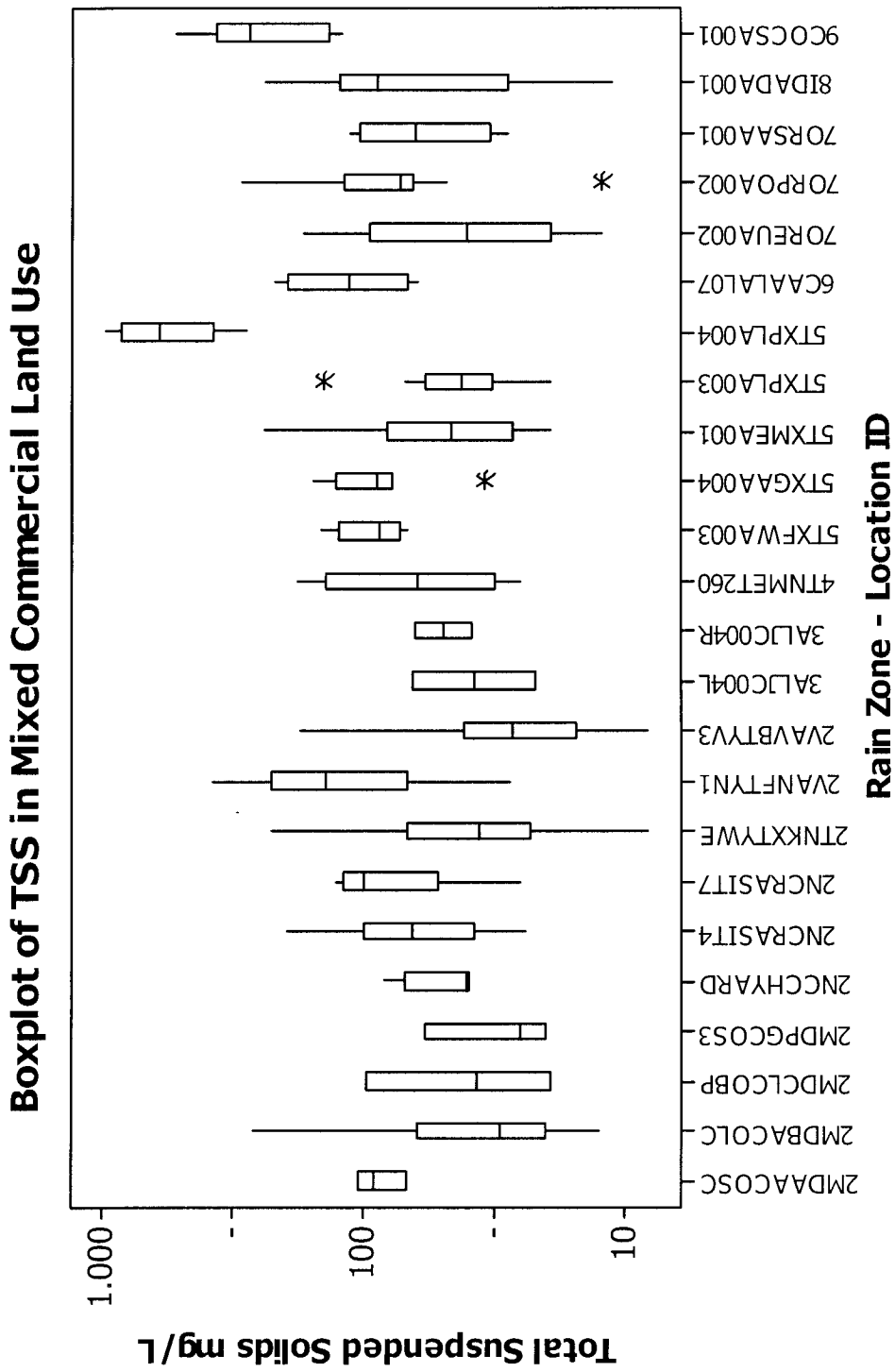
Fig. D29. Box and whiskers plots for TSS concentrations at mixed commercial land use areas

The box plot indicates that there is a mixed commercial site located in Plano, Texas and a site in Colorado with higher concentrations than the other sites in this category. Because the low number of sites sampled by geographical region, it is not possible to identify any trend by EPA Rain Zone. Table D17 lists those sites outside the control limits by EPA Rain Zone for all of the mixed commercial sites.

Table D17. Sites failing Xbar and S Chart Tests in Mixed Commercial Land Use Areas

EPA Rain Zone	Sites Failing Xbar Chart Test	Sites Failing S Chart Test
ALL	2TNKXTYWE(L) 2VANFTYN1(H) 2VAVBTYV3(L) 5TXPLA004(H) 9COCSA001(H)	None
1	None	None
2	2VANFTYN1(H) 2VAVBTYV3(L)	None
3	None	None
4	None	None
5	5TXPLA004(H)	None
6	None	None
7	None	None
9	None	None

The Xbar chart for all mixed commercial observations indicates that sites with high TSS concentrations occurred in EPA Rain Zones 5 and 9. In EPA Rain Zone 2, three sites were outside of the control limits, two below the lower control limit and one above the upper control limit. As in the commercial site analyses, EPA Rain Zone 9 seems to have higher TSS concentrations than the other EPA Rain Zones.

The analysis by EPA Rain Zone indicates that only EPA Rain Zones 2 and 5 have sites outside the control limits. In EPA Rain Zone 2, the site with high concentrations (VANFTYN1) is located at Armisted Avenue in Norfolk, Virginia. A total of 28 observations were collected at this site. The median TSS for this location was 117 mg/L. The site having unusually low median TSS concentration was at Haygood, Virginia

Beach, Virginia (VAVBTYV3). A total of 33 storms were sampled at this site. The median TSS concentration at this location was 26 mg/L. This site is 79% commercial and 13% open space. The site having unusually high TSS concentrations in EPA Rain Zone 5 is located at Spring Creek, Plano, Texas (TXPLA004). There are 7 events from this site in the database. The median TSS concentration is 575 mg/L. No information was found to explain the elevated concentrations. Another site that appear to be outside the control limits compared to all the sites, but not in its group. It is located in Sixteenth Hole Valley, Colorado Springs, Colorado. The median concentration for this site was 251 mg/L. This site has two automobile dealerships and a gas station, along with evidence of erosion observed in the aerial photograph.

The ANOVA analysis indicates that there are significant differences among EPA Rain Zones (P-value = 0) in mixed commercial land uses. The Dunnett's comparison test, with a family error of 5%, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS = 46 mg/L) are larger in EPA Rain Zones 5 (median TSS = 72 mg/L) and 9 (median TSS = 254 mg/L). The median TSS in the remaining EPA Rain Zones is not statistically different than those observed in EPA Rain Zone 2.

D.3.3 Industrial and Mixed Industrial Locations

Box plots Xbar and S charts and ANOVA tests were used to examine the observations from sites located in industrial land use areas. Figure D30 shows the box plots by EPA Rain Zone and location. Sites located in EPA Rain Zones 6 and 9 seem to have higher concentrations than the remaining industrial sites. A site with two unusually low concentrations was located in Boston, Massachusetts.
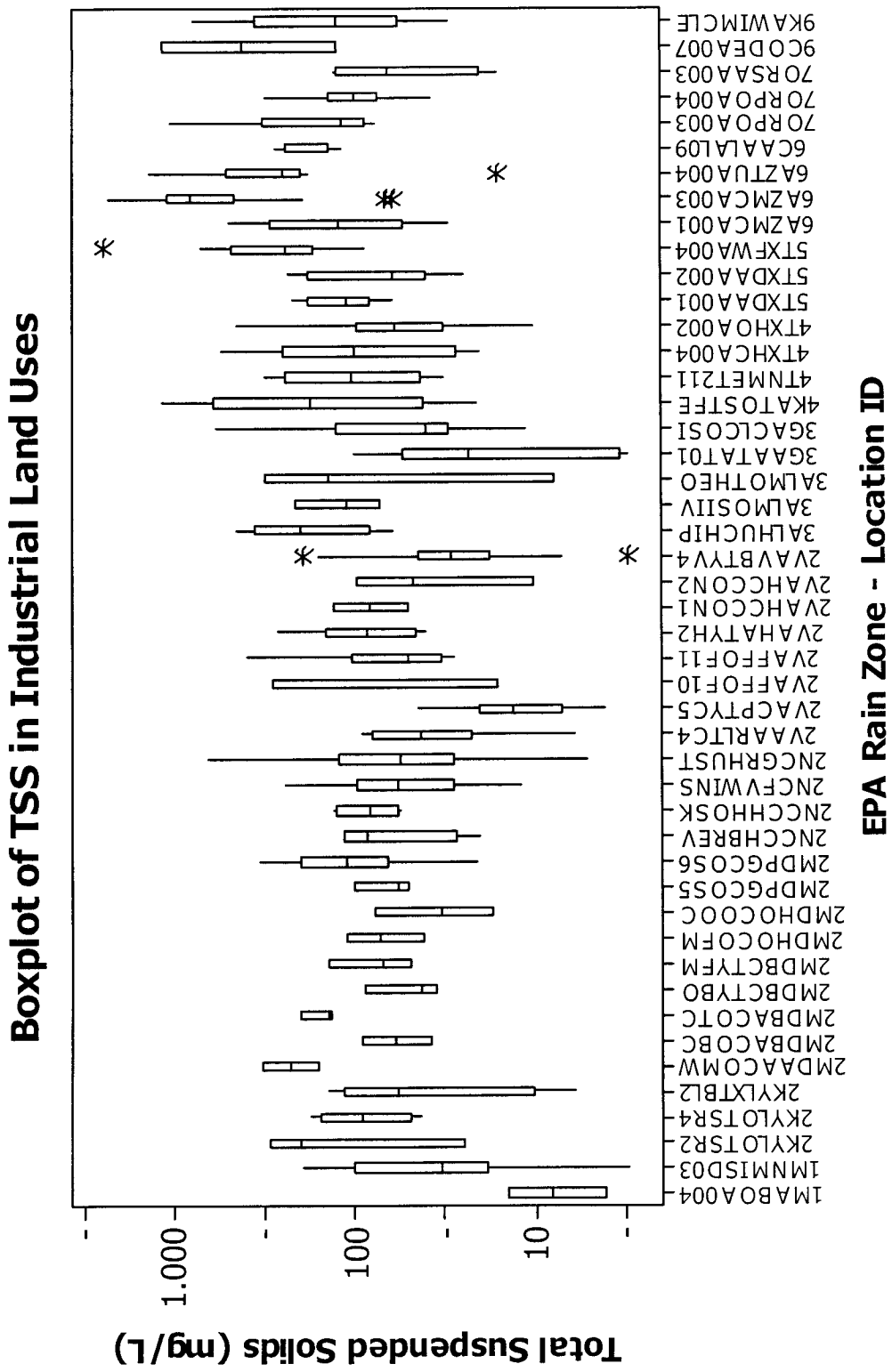
Fig. D30. Box and whiskers plots for TSS concentrations at industrial land use areas

Table D18 shows those industrial sites that are outside the control limits by the

pooled dataset and by each EPA Rain Zone.

Table D18. Sites failing Xbar and S Chart Tests in Industrial Land Use Areas

EPA Rain Zone	Sites Failing Xbar Chart Test	Sites Failing S Chart Test
ALL	1MABOA004(L) 2VACPTYC5(L) 2VAVBTYV4(L) 3GAATAT01(L) 5TXFWA004(H) 6AZMCA003(H) 6AZTUA004(H)	None
1	None	None
2	MDPGCOS6(H) VACPTYC5(L) VAVBTYV4(L)	None
3	None	None
4	None	None
5	TXFWA004(H)	None
6	AZMCA003(H)	None
7	None	None
9	None	None

As in the other land uses, sites with concentrations below the control limit were

observed in EPA Rain Zones 1, 2 and 3. Sites with median concentrations larger than the

upper control limit were located in EPA Rain Zones 5 and 6. Three sites were outside the

control limits in EPA Rain Zone 2, one in EPA Rain Zone 5, and one in EPA Rain Zone

6. The two sites in EPA Rain Zone 2 with low concentrations were located in Virginia,

and the site with high concentrations was located in Maryland. One of the sites located in

Virginia was located in Cavalier Industrial Park in the city of Chesapeake (VACPTYC5).

This 16 acres site is 92% industrial, with the remaining 8% open space. A total of 15

samples were collected from this site during the period 1993 to 1999. The median TSS

concentration for this site is 13 mg/L. No additional information was observed in the

aerial photos that might explain the low concentrations.

The second site was located in Viking Drive, Virginia Beach (VAVBTYV5). This 29-acre site had 55 percent of impervious surfaces. There are 30 samples from this site in the database. The samples were collected between 1992 and 1999. The median TSS concentration is 29 mg/L.

The site with elevated concentrations in EPA Rain Zone 2 is located in Pennsy Drive in Riverdale, Prince George County, Maryland (MDPGCOS6). This 42.4-acre size site has a grass swale drainage system. There are 30 samples in the database from this location. The samples were collected between 1994 and 1997. The median TSS concentration is 98 mg/L. The site is located next to Glenridge Elementary School. The aerial photo shows construction activity in the northwest part of the watershed.

The site with high concentration in EPA Rain Zone 5 is located at Dry Branch, in Fort Worth, Texas (TXFWA004). A total of 21 samples were obtained at this site. The median TSS for this location is 288 mg/L. Several bare ground open space areas were observed in the aerial photograph. The site located in EPA Rain Zone 6 is at 27th Avenue at Salt River in Maricopa County, Arizona (AZMCA003). There are 27 samples from this location. The median TSS concentration is 660 mg/L. The scarce vegetation and the type of soils may be the reason of this elevated median value.

The ANOVA analysis indicates that there are significant differences among EPA Rain Zones (P-value = 0) for industrial land uses. The Dunnett's comparison test with a family error of 5%, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS = 53 mg/L) are larger for EPA Rain Zones 4 (median TSS = 92 mg/L), 5 (median TSS = 147 mg/L), 6 (median TSS = 288 mg/L), 7 (median TSS = 120 mg/L),

and 9 (median TSS = 170 mg/L). The median TSS concentrations in the EPA Rain Zone

1 and 3 are not statistically different than those observed in EPA Rain Zone 2.

The box plots in mixed industrial land uses are shown in Figure D31. Most of the

box plots have the same median except for those located in EPA Rain Zone 9. The sites

that fail the quality control charts are shown in Table D19. Three sites are outside the

control limits for mixed industrial land uses. Two sites in Colorado and one site in North

Carolina are out of control. This result is similar to those observed in the other land uses.

When each EPA Rain Zone was analyzed individually, no sites were found to be out of

control.

Boxplot TSS in Mixed Indsutrial Land Uses
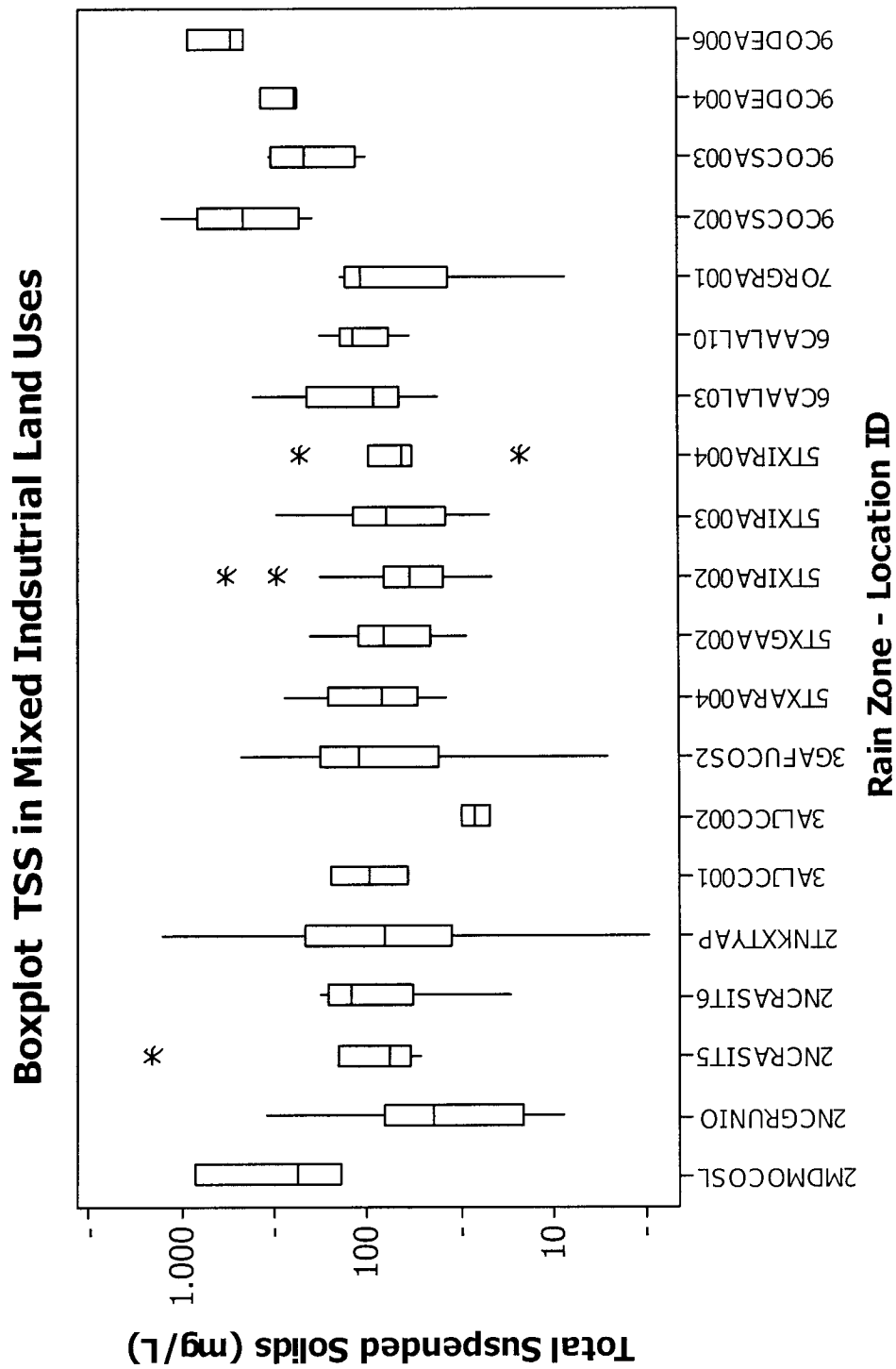
Fig. D31. Box and whiskers plots for TSS concentrations at industrial land use areas

Table D19. Sites failing Xbar and S Chart Tests in Mixed Industrial Land Use Areas

EPA Rain Zone	Sites Failing Xbar Chart Test	Sites Failing S Chart Test
ALL	9COCSA002(H) 9CODEA006(H) 2NCGRUNIO(L)	None
2	None	None
3	None	None
5	None	None
6	None	None
7	None	None
9	None	None

The ANOVA analysis indicates that there are significant differences among EPA Rain Zones (P-value = 0) at mixed industrial land uses. The Dunnett's comparison test with a family error of 5%, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS = 82 mg/L) are larger only for EPA Rain Zone 9 (median TSS = 341 mg/L). The median TSS concentrations in EPA Rain Zones 3, 5, 6, and 7 are not statistically different from the median TSS concentrations found in EPA Rain Zone 2.

Appendix E
First Flush Tables

E.1 Description

The following table shows the summary statistic for each constituent included in the database.

Table E1. Results of Preliminary Statistical Analysis for Total Suspended Solids (TSS)

TSS (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	90	90	54.95	1.740	0.106	0.168	0.254	0.730	1.85
Commercial First Flush	90	90	101.86	2.008	0.200	-0.508	0.254	0.016	
Industrial Composite	83	83	66.07	1.820	0.186	-0.021	0.264	0.336	0.97
Industrial First Flush	83	83	63.97	1.806	0.374	-0.157	0.264	0.055	
Institutional Composite	18	18	16.48	1.217	0.110	-0.176	0.536	0.122	2.12
Institutional First Flush	18	18	34.99	1.544	0.145	-0.164	0.536	0.846	
Open Space Composite	32	32	21.98	1.342	0.424	-0.526	0.414	0.511	0.95
Open Space First Flush	32	32	20.89	1.320	0.563	-0.126	0.414	0.847	
Residential Composite	144	144	37.50	1.574	0.217	-0.033	0.202	0.282	1.84
Residential First Flush	144	144	69.02	1.839	0.302	-0.267	0.202	0.533	
All Land Uses Composite	372	372	44.36	1.647	0.226	-0.381	0.126	0.008	1.60
All Land Uses First Flush	372	372	70.96	1.851	0.335	0.457	0.126	0	

Table E2. Results of Preliminary Test Analysis for Total Suspended Solids (TSS)

TSS (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t-Test	Result
Commercial	N/A	U = 5.345 ; P = 0	0.014	N/A	Different (first flush)
Industrial	0.627	U = 0.483 ; P = 0.31	0.222	0.432	Same (no first flush)
Institutional	0.007	U = 3.095 ; P = 0	0.309	0.001	Different
Open Space	0.706	U = 0.39 ; P = 0.35	0.183	0.614	Same
Residential	N/A	U = 4.89 ; P = 0	0	N/A	Different
All Land Uses	N/A	U = 6.65 ; P = 0	0	N/A	Different

Table E3. Results of Preliminary Statistical Analysis for Turbidity (NTU)

Turbidity (NTU)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	11	11	19.68	1.294	0.062	-0.984	0.661	0.186	1.32
Commercial First Flush	11	11	26.00	1.415	0.078	0.523	0.661	0.564	
Residential Composite	12	12	23.44	1.370	0.163	0.213	0.637	0.721	1.24
Residential First Flush	12	12	28.97	1.462	0.148	1.407	0.637	0.168	
All Land Uses Composite	26	26	21.73	1.337	0.109	0.204	0.456	0.406	1.26
All Land Uses First Flush	26	26	27.48	1.439	0.105	1.197	0.456	0.108	

Table E4. Results of Preliminary Test Analysis for Turbidity (NTU)

Turbidity (NTU)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t-Test	Result
Commercial	0.224	U = 1.26 ; P > 0.1	0.652	0.219	Same (no first flush)
Residential	0.418	U = 0.853 ; P > 0.1	0.240	0.021	Same
All Land Uses	0.124	U = 0.673 ; P = 0.25	0.134	0	Same

Table E5. Results of Preliminary Statistical Analysis for pH

pH	Total Events	Selected Cases	Median	Var (Log)	Skew	SE Skew	Test Norm. p-value	Median Ratio
Commercial Composite	17	17	7.4	0.368	-0.299	0.550	0.527	1.03
Commercial First Flush	17	17	7.6	0.509	0.788	0.550	0.351	
Industrial Composite	16	16	6.755	0.194	0.482	0.564	0.179	1.00
Industrial First Flush	16	16	6.750	0.388	-0.854	0.564	0.307	
Residential Composite	26	26	7.213	0.195	-0.520	0.456	0.447	1.01
Residential First Flush	26	26	7.250	0.212	-0.283	0.456	0.408	
All Composite	63	63	7.2	0.302	0.102	0.302	0.562	1.01
All First Flush	63	63	7.3	0.437	0.036	0.302	0.110	

Table E6. Results of Preliminary Test Analysis for pH

pH	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	0.208	U = 1.28 ; P = 0.10	0.007	N/A	Same
Industrial	N/A	U = 0.428 ; P = 0.33	0.341	0.828	Same
Residential	0.308	U = 1.32 ; P = 0.09	0	N/A	Same
All Land Uses	0.219	U = 1.68 ; P = 0.05	0	N/A	Same

Table E7. Results of Preliminary Statistical Analysis for BOD_5

BOD₅ (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	83	83	15.21	1.182	0.125	0.263	0.264	0.513	1.77
Commercial First Flush	83	83	26.98	1.431	0.153	-0.241	0.264	0.390	
Industrial Composite	80	80	15.14	1.18	0.188	0.190	0.269	0.013	1.58
Industrial First Flush	80	80	23.99	1.38	0.180	-0.502	0.269	0.044	
Institutional Composite	18	18	7.48	0.874	0.151	-0.737	0.536	0.247	1.67
Institutional First Flush	18	18	12.47	1.096	0.173	-0.732	0.536	0.281	
Open Space Composite	28	28	3.79	0.579	0.148	0.523	0.441	0.242	1.07
Open Space First Flush	28	28	4.05	0.607	0.197	0.449	0.441	0.077	
Residential Composite	133	133	12.59	1.100	0.154	0.314	0.210	0.137	1.67
Residential First Flush	133	133	20.99	1.322	0.220	-0.150	0.210	0.010	
All Land uses Composite	344	344	12.53	1.098	0.184	0.073	0.131	0.003	1.67
All Land Uses First Flush	344	344	20.89	1.320	0.233	-0.385	0.131	0	

Table E8. Results of Preliminary Test Analysis for BOD_5

BOD₅	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t - Test	Result
Commercial	0	U = 4.85 ; P = 0	0.013	N/A	Different (first flush)
Industrial	0.007	U = 2.76 ; P = 0	0.434	0.012	Different
Institutional	0.027	U = 2.46 ; P = 0.01	0.056	0.001	Different
Open Space	0.706	U = 0.39 ; P = 0.35	0.183	0.614	Same (no first flush)
Residential	N/A	U = 4.89; P = 0	0	N/A	Different
All Land Uses	N/A	U = 6.65 ; P = 0	0	N/A	Different

Table E9. Results of Preliminary Statistical Analysis for COD

COD (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite)	91	91	71.94	1.857	0.075	0.261	0.253	0.022	2.29
Commercial First Flush	91	91	164.82	2.217	0.119	-0.201	0.253	0.877	
Industrial Composite	84	84	75.34	1.877	0.100	0.167	0.263	0.014	1.43
Industrial First Flush	84	84	107.40	2.031	0.151	-0.141	0.263	0.804	
Institutional Composite	18	18	43.85	1.642	0.220	-0.456	0.536	0.567	2.73
Institutional First Flush	18	18	119.67	2.078	0.151	-0.969	0.536	0.105	
Open Space Composite	28	28	20.00	1.301	0.130	0.441	0.441	0.084	0.67
Open Space First Flush	28	28	13.43	1.128	0.211	0.731	0.441	0.013	
Residential Composite	140	140	67.92	1.832	0.095	0.271	0.205	0.008	1.63
Residential First Flush	140	140	110.41	2.043	0.138	-0.831	0.205	0.005	
All Land Uses Composite	363	363	65.92	1.819	0.123	-0.293	0.128	0	1.71
All Land Uses First Flush	363	363	112.98	2.053	0.194	-0.710	0.128	0	

Table E10. Results of Preliminary Test Analysis for COD

COD	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 4.83 ; P = 0	0.269	0	Different (first flush)
Industrial	N/A	U = 1.67 ; P = 0.05	0.691	0.01	Different
Institutional	0.01	U = 2.94 ; P = 0	0.677	0	Different
Open Space	N/A	U = 0.269 ; P = 0.39	0.004	N/A	Same (no first flush)
Residential	N/A	U = 6.715 ; P = 0	0	N/A	Different
All Land Uses	N/A	U =9.19 ; P = 0	0	N/A	Different

Table E11. Results of Preliminary Statistical Analysis for Total Dissolved Solids (TDS)

TDS (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	82	82	73.28	1.865	0.064	-0.338	0.266	0.263	1.83
Commercial First Flush	82	82	133.97	2.127	0.065	-0.219	0.266	0.115	
Industrial Composite	82	81	97.72	1.990	0.093	-0.482	0.267	0.341	1.32
Industrial First Flush	82	81	128.82	2.110	0.126	-0.513	0.267	0.109	
Institutional Composite	18	18	52.48	1.720	0.068	-0.034	0.536	0.360	2.66
Institutional First Flush	18	18	139.64	2.145	0.090	-0.303	0.536	0.158	
Open Space Composite	31	30	69.98	1.845	0.051	0.617	0.427	0.376	1.07
Open Space First Flush	31	30	74.99	1.875	0.104	-1.483	0.427	0.005	
Residential Composite	137	133	70.31	1.870	0.119	-0.245	0.210	0.041	1.52
Residential First Flush	137	133	107.15	2.030	0.125	0.500	0.210	0.167	
All Land Uses Composite	354	342	77.62	1.890	0.083	0.188	0.132	0.334	1.55
All Land Uses First Flush	354	342	120.23	2.080	0.104	0.225	0.132	0.126	

Table E12. Results of Preliminary Test Analysis for Total Dissolved Solids (TDS)

TDS (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	0	U = 7.33 ; P = 0	0.160	0	Different (first flush)
Industrial	0.0245	U = 2.28 ; P = 0.01	0.070	0.003	Different
Institutional	0.0118	U = 2.945 ; P = 0	0.544	0	Different
Open Space	N/A	U = 0.161 ; P = 0.44	0	N/A	Same (no first flush)
Residential	N/A	U = 4.89 ; P = 0	0	N/A	Different
All Land Uses	0	U = 7.58 ; P = 0	0	N/A	Different

Table E13. Results of Preliminary Statistical Analysis for O&G

O&G (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	10	10	5.19	0.715	0.068	-0.976	0.687	0.016	1.54
Commercial First Flush	10	10	8.00	0.027	0.903	1.641	0.687	0.019	
Residential Composite	8	4	5.00	0.699	0.066	1.985	1.014	0.013	2.05
Residential First Flush	8	4	10.23	1.010	0.134	0.003	1.014	0.056	
All Land Uses Composite	18	14	5.00	0.699	0.073	-0.370	0.597	0.015	1.60
All Land Uses First Flush	18	14	8.00	0.903	0.051	0.890	0.597	0.011	

Table E14. Results of Preliminary Test Analysis for O&G

O&G (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t-Test	Result
Commercial	N/A	U = 6.198 ; P < 0.01	0.222	0.004	Different
Residential	N/A	U = 1.069 ; P > 0.1	0.049	0.306	Same
All Land Uses	N/A	U = 4.072 ; P = 0	0.036	N/A	Different

Table E15. Results of Preliminary Statistical Analysis for Fecal Coliforms

Fecal Coliforms (mpn/100 mL)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	12	12	67764	4.831	1.099	-0.691	0.637	0.627	0.87
Commercial First Flush	12	12	58884	4.770	1.732	-0.388	0.637	0.228	0.87
Residential Composite	10	9	41976	4.623	0.292	0.485	0.717	0.276	0.98
Residential First Flush	10	9	41020	4.643	0.685	0.247	0.717	0.799	0.98
All Land Uses Composite	22	21	46238	4.665	0.745	-0.886	-0.515	0.511	1.21
All Land Uses First Flush	22	21	55976	4.748	1.269	0.501	0.501	0.391	1.21

Table E16. Results of Preliminary Test Analysis for Fecal Coliforms

Fecal Coliforms (mpn/100 mL)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 0 ; P > 0.10	0.833	0.583	Same
Residential	N/A	U = 0.289 ; P > 0.1	0.016	0.973	Same
All Land Uses	N/A	U = 0.181 ; P = 0.43	0.086	0.665	Same

Table E17. Results of Preliminary Statistical Analysis for Fecal Streptococcus

Fecal Streptococcus (mpn/100 mL)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	12	11	37153	4.570	0.780	-0.255	0.661	0.948	1.05
Commercial First Flush	12	11	38904	4.590	1.094	0.009	0.661	0.722	
Residential Composite	11	8	77625	4.890	0.231	-0.223	0.752	0.426	1.30
Residential First Flush	11	8	101158	5.005	0.327	-0.659	0.752	0.319	
All Land Uses Composite	26	22	43651	4.640	0.536	-0.513	0.491	0.713	1.11
All Land Uses First Flush	26	22	48417	4.685	0.705	-0.188	0.491	0.802	

Table E18. Results of Preliminary Test Analysis for Fecal Streptococcus

Fecal Streptococcus (mpn/100mL)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 0.281 ; P > 0.10	0.027	N/A	Same (no first flush)
Residential	N/A	U = 0.344 ; P > 0.10	0.109	0.905	Same
All Land Uses	N/A	U = 0.309 ; P = 0.38	0.033	N/A	Same

Table E19. Results of Preliminary Statistical Analysis for Ammonia

Ammonia (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	70	52	0.76	-0.122	0.147	-0.245	0.330	0.237	2.11
Commercial First Flush	70	52	1.60	0.204	0.117	-0.718	0.330	0.027	
Industrial Composite	40	33	0.62	-0.208	0.166	-0.399	0.409	0.284	1.08
Industrial First Flush	40	33	0.67	-0.174	0.201	-0.535	0.409	0.046	
Institutional Composite	18	16	0.31	-0.509	0.058	-0.038	0.564	0.273	1.66
Institutional First Flush	18	16	0.51	-0.290	0.077	0.284	0.564	0.384	
Residential Composite	119	86	0.50	-0.301	0.370	0.779	0.260	0.001	1.36
Residential First Flush	119	86	0.68	-0.168	0.172	0.195	0.260	0.519	
All Land Uses Composite	269	190	0.52	-0.284	0.251	0.501	0.176	0.002	1.54
All Land Uses First Flush	269	190	0.80	-0.097	0.176	-0.197	0.176	0.713	

* Ammonia in Open Space was found in 22 events. Only 3 events had values above the detection limit

Table E20. Results of Preliminary Test Analysis for Ammonia

Ammonia (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 4.467 ; P = 0	0.028	N/A	Different
Industrial	N/A	U = 0.113 ; P = 0.46	0.262	0.985	Same
Institutional	0.0287	U = 2.484 ; P = 0.01	0.254	0	Different
Residential	N/A	U = 2.283 ; P = 0.01	0	N/A	Different
All Land Uses	N/A	U = 4.092 ; P = 0	0	N/A	Different

Table E21. Results of Preliminary Statistical Analysis for $NO_2 + NO_3$

$NO_2 + NO_3$ (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	84	82	0.75	-0.125	0.095	-0.092	0.266	0.188	1.73
Commercial First Flush	84	82	1.30	0.114	0.166	-0.790	0.266	0.007	
Industrial Composite	72	71	0.90	-0.046	0.073	-0.240	0.285	0.807	1.31
Industrial First Flush	72	71	1.18	0.072	0.116	-0.839	0.285	0.030	
Institutional	18	18	0.60	-0.222	0.122	-0.714	0.536	0.117	1.70
Institutional First Flush	18	18	1.02	0.009	0.151	0.268	0.536	0.381	
Open Space Composite	30	21	0.24	-0.620	0.290	0.468	0.501	0.141	0.96
Open Space First Flush	30	21	0.23	-0.638	0.356	0.823	0.501	0.030	
Residential Composite	121	118	0.60	-0.222	0.104	-0.196	0.223	0.504	1.66
Residential First Flush	121	118	1.00	-0.002	0.125	-0.292	0.223	0.102	
All Land Uses Composite	324	310	0.70	-0.155	0.124	-0.497	0.138	0	1.50
All Land Uses First Flush	324	310	1.05	0.021	0.162	-0.584	0.138	0	

Table E22. Results of Preliminary Test Analysis for $NO_2 + NO_3$

$NO_2 + NO_3$	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 3.286 ; P = 0	0	N/A	Different (first flush)
Industrial	N/A	U = 1.836 ; P = 0.03	0.941	0.034	Different
Institutional	0.043	U = 2.242 ; P = 0.01	0.026	N/A	Different
Open Space	N/A	U = 0.209 ; P = 0.42	0.023	N/A	Same (no first flush)
Residential	0	U = 4.769 ; P = 0	0.023	N/A	Different
All Land Uses	N/A	U = 5.834 ; P = 0	0	N/A	Different

Table E23. Results of Preliminary Statistical Analysis for Total Nitrogen

Table E23. Results of Preliminary Statistical Analysis for Total Nitrogen

Total N (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	19	19	1.42	0.152	0.180	-0.133	0.524	0.215	1.35
Commercial First Flush	19	19	1.91	0.281	0.203	-0.617	0.524	0.337	
Industrial Composite	19	16	2.01	0.303	0.286	-0.306	0.564	0.431	1.79
Industrial First Flush	19	16	3.61	0.557	0.349	-0.452	0.564	0.029	
Open Space Composite	6	6	1.39	0.142	0.112	-0.150	0.845	0.330	1.53
Open Space First Flush	6	6	2.12	0.326	0.248	-0.100	0.845	0.221	
Residential Composite	31	30	1.67	0.222	0.325	1.22	0.427	0.009	0.88
Residential First Flush	31	30	1.47	0.166	0.447	-0.587	0.427	0.367	
All Land Uses Composite	77	73	1.60	0.204	0.253	0.769	0.281	0.136	1.22
All Land Uses First Flush	77	73	1.95	0.290	0.331	0.599	0.281	0.071	

Table E24. Results of Preliminary Test Analysis for Total Nitrogen

Total N (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	0.220	U = 1.234 ; P = 0.11	0.329	0.013	Same
Industrial	N/A	U = 0.460 ; P = 0.32	0.759	0.161	Same
Open Space	N/A	U = 0 ; P > 0.104	0.339	0.703	Same
Residential	N/A	U = 0.106 ; P = 0.46	0.002	N/A	Same
All Land Uses	N/A	U = 0.919 ; P = 0.18	0	N/A	Same

Table E25. Results of Preliminary Statistical Analysis for TKN

TKN (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	93	86	1.63	0.213	0.085	-0.275	0.260	0.003	1.71
Commercial First Flush	93	86	2.80	0.447	0.120	-0.117	0.260	0.714	
Industrial Composite	77	76	1.69	0.227	0.116	1.157	0.276	0	1.35
Industrial First Flush	77	76	2.27	0.356	0.130	0.536	0.276	0.232	
Open Space Composite	32	14	0.61	-0.215	0.142	0.585	0.597	0.109	1.28
Open Space First Flush	32	14	0.78	-0.107	0.269	0.948	0.597	0.139	
Residential Composite	131	123	1.40	0.146	0.110	1.752	0.218	0	1.65
Residential First Flush	131	123	2.31	0.364	0.115	0.309	0.218	0.076	
All Land Uses Composite	335	301	1.50	0.176	0.114	0.856	0.140	0	1.60
All Land Uses First Flush	335	301	2.40	0.380	0.139	0.088	0.140	0	

Table E26. Results of Preliminary Test Analysis for TKN

TKN (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 6.499 ; P = 0	0.126	0	Different (first flush)
Industrial	N/A	U = 1.698 ; P = 0.04	0.054	0.063	Different
Open Space	N/A	U = 0.374 ; P = 0.35	0.116	0.364	Same (no first flush)
Residential	N/A	U = 6.079 ; P = 0	0	N/A	Different
All Land Uses	N/A	U = 7.68 ; P = 0	0	N/A	Different

Table E27. Results of Preliminary Statistical Analysis for Total Phosphorus

Total P (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	89	77	0.34	-0.469	0.160	-0.454	0.274	0.129	1.44
Commercial First Flush	89	77	0.49	-0.310	0.205	0.033	0.274	0.035	
Industrial Composite	84	71	0.29	-0.538	0.130	0.495	0.285	0.003	1.42
Industrial First Flush	84	71	0.41	-0.387	0.257	-0.441	0.285	0.397	
Institutional Composite	17	17	0.17	-0.770	0.203	-0.736	0.550	0.374	1.24
Institutional First Flush	17	17	0.21	-0.678	0.066	-0.177	0.550	0.704	
Open Space Composite	32	20	0.09	-1.023	0.147	0.613	0.512	0.218	1.05
Open Space First Flush	32	20	0.10	-1.000	0.381	0.833	0.512	0.288	
Residential Composite	140	128	0.28	-0.553	0.252	1.232	0.214	0	1.46
Residential First Flush	140	128	0.41	-0.389	0.188	-0.335	0.214	0.042	
All Land Uses Composite	363	313	0.28	-0.553	0.209	0.605	0.138	0	1.45
All Land Uses First Flush	363	313	0.41	-0.391	0.238	-0.258	0.138	0.003	

Table E28. Results of Preliminary Test Analysis for Total Phosphorus

Total P (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 3.089 ; P = 0	0.594	0	Different (first flush)
Industrial	N/A	U = 0.864 ; P = 0.19	0.194	0.667	Same (no first flush)
Institutional	N/A	U = 0.774 ; P = 0.22	0.044	N/A	Same
Open Space	N/A	U = 0.142 ; P = 0.44	0.091	0.527	Same
Residential	N/A	U = 2.671 ; P = 0	0	N/A	Different
All Land Uses	N/A	U = 3.641 ; P = 0	0	N/A	Different

Table E29. Results of Preliminary Statistical Analysis for Dissolved Phosphorus

Dissolved P (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	91	69	0.16	-0.788	0.152	0.467	0.289	0	1.23
Commercial First Flush	91	69	0.20	-0.699	0.212	0.904	0.289	0.005	
Industrial Composite	77	50	0.14	-0.854	0.142	1.248	0.337	0.093	1.04
Industrial First Flush	77	50	0.14	-0.839	0.160	0.406	0.337	0.043	
Institutional Composite	18	14	0.13	-0.891	0.066	-0.114	0.597	0.563	1.05
Institutional First Flush	18	14	0.13	-0.870	0.095	-0.770	0.597	0.122	
Open Space Composite	32	14	0.05	-1.301	0.111	-0.073	0.597	0.601	0.69
Open Space First Flush	32	14	0.03	-1.460	0.087	1.061	0.597	0.017	
Residential Composite	130	105	0.17	-0.770	0.117	0.152	0.236	0.458	1.24
Residential First Flush	130	105	0.21	-0.678	0.170	0.121	0.236	0.044	
All Land Uses Composite	350	254	0.15	-0.824	0.143	0.353	0.153	0.051	1.07
All Land Uses First Flush	350	254	0.16	-0.796	0.200	0.401	0.153	0.001	

Table E30. Results of Preliminary Test Analysis for Dissolved Phosphorus

Dissolved P (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 1.582 ; P = 0.06	0.046	N/A	Same
Industrial	N/A	U = 0.051 ; P = 0.48	0.063	0.881	Same
Institutional	0.549	U = 0.605 ; P = 0.27	0.015	N/A	Same
Open Space	N/A	U = 0.760 ; P = 0.22	0.018	N/A	Same
Residential	N/A	U = 1.702 ; P = 0.04	0.039	N/A	Different
All Land Uses	N/A	U = 1.657 ; P = 0.05	0	N/A	Same

Table E31. Results of Preliminary Statistical Analysis for Orthophosphate

Orthophosphate (mg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Industrial Composite	6	6	0.16	-0.797	0.287	-0.047	0.845	0.838	1.55
Industrial First Flush	6	6	0.25	-0.607	0.356	-0.106	0.845	0.720	
Residential Composite	14	14	0.19	-0.714	0.554	2.557	0.597	0.001	0.95
Residential First Flush	14	14	0.18	-0.737	0.214	0.708	0.597	0.362	
All Land Uses Composite	22	22	0.19	-0.714	0.423	2.270	0.491	0.004	1.30
All Land Uses First Flush	22	22	0.25	-0.600	0.222	0.260	0.491	0.503	

Table E32. Results of Preliminary Test Analysis for Orthophosphate

Orthophosphate (mg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Industrial	0.471	U = 0.772 ; P > 0.104	0.071	0.611	Same
Residential	N/A	U = 0.022 ; P = 0.49	0	N/A	Same
All Land Uses	N/A	U = 0.460 ; P = 0.32	0	N/A	Same

Table E33. Results of Preliminary Statistical Analysis for Total Cadmium

Total Cadmium (µg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	74	48	0.56	-0.253	0.246	-0.325	0.343	0	2.15
Commercial First Flush	74	48	1.20	0.079	0.261	0.080	0.343	0.089	
Industrial Composite	80	41	1	0	0.124	-0.015	0.369	0.008	1.00
Industrial First Flush	80	41	1	0	0.130	0.261	0.369	0.065	
Open Space Composite	30	15	0.23	-0.638	0.282	1.074	0.580	0.183	1.30
Open Space First Flush	30	15	0.30	-0.523	0.325	0.465	0.580	0.402	
Residential Composite	123	33	0.28	-0.553	0.359	0.693	0.409	0.002	2.00
Residential First Flush	123	33	0.56	-0.252	0.264	0.512	0.409	0.061	
All Land Uses Composite	325	139	0.60	-0.222	0.269	-0.065	0.206	0.071	1.62
All Land Uses First Flush	325	139	0.97	-0.013	0.249	0.041	0.206	0.241	

Table E34. Results of Preliminary Test Analysis for Total Cadmium

Total Cadmium (µg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	0.006	U = 2.797 ; P = 0	0.009	N/A	Different (first flush)
Industrial	0.922	U = 0.100 ; P = 0.46	0.118	0.529	Same (no first flush)
Open Space	0.442	U = 0.765 ; P = 0.22	0.292	0.191	Same
Residential	0.038	U = 2.131 ; P = 0.02	0.015	N/A	Different
All Land Uses	0.005	U = 2.839 ; P = 0	0	N/A	Different

Table E35. Results of Preliminary Statistical Analysis for Total Chromium

Total Chromium (µg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	47	22	6.81	0.833	0.086	-0.051	0.491	0.911	1.67
Commercial First Flush	47	22	11.40	1.057	0.134	-0.796	0.491	0.121	
Industrial Composite	54	25	8.79	0.944	0.111	0.338	0.464	0.456	1.36
Industrial First Flush	54	25	11.99	1.079	0.155	-0.307	0.464	0.784	
Open Space Composite	16	4	2.64	0.422	0.169	-0.556	1.014	0.492	1.70
Open Space First Flush	16	4	4.50	0.653	0.015	1.291	1.014	0.355	
Residential Composite	86	31	8.00	0.903	0.169	-0.077	0.421	0.612	1.24
Residential First Flush	86	31	9.91	0.996	0.137	0.326	0.421	0.904	
All Land Uses Composite	218	82	7.50	0.875	0.140	-0.104	0.266	0.591	1.47
All Land Uses First Flush	218	82	10.99	1.041	0.141	-0.056	0.266	0.803	

Table E36. Results of Preliminary Test Analysis for Total Chromium

Total Chromium (µg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t-Test	Result
Commercial	0.0513	U = 2.024 ; P = 0.02	0.283	0.036	Different
Industrial	0.3032	U = 1.023 ; P = 0.15	0.216	0.320	Same
Open Space	0.3032	U = 1.586 ; P = 0.10	0.160	0.199	Same
Residential	0.6023	U = 0.519 ; P = 0.30	0.007	N/A	Same
All Land Uses	0.0547	U = 1.939 ; P = 0.03	0.001	N/A	Different

Table E37. Results of Preliminary Statistical Analysis for Total Copper

Total Copper (µg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	92	82	16.98	1.230	0.083	-0.038	0.266	0.117	1.62
Commercial First Flush	92	82	27.48	1.439	0.120	0.343	0.266	0.035	
Industrial Composite	84	76	25.00	1.398	0.079	0.184	0.276	0.344	1.24
Industrial First Flush	84	76	30.97	1.491	0.166	-0.014	0.276	0.007	
Institutional Composite	18	7	16.98	1.230	0.083	-0.228	0.794	0.167	0.94
Institutional First Flush	18	7	16.00	1.204	0.047	0.954	0.794	0.555	
Open Space Composite	30	22	5.14	0.711	0.103	0.085	0.491	0.252	0.78
Open Space First Flush	30	22	4.00	0.602	0.120	1.005	0.491	0.015	
Residential Composite	144	108	11.99	1.079	0.082	-0.677	0.233	0	1.33
Residential First Flush	144	108	16.00	1.204	0.087	0.023	0.233	0.256	
All Land Uses Composite	368	295	15.00	1.176	0.116	-0.268	0.142	0	1.33
All Land Uses First Flush	368	295	20.00	1.301	0.167	0.009	0.142	0	

Table E38. Results of Preliminary Test Analysis for Total Copper

Total Copper (µg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 5.160 ; P = 0	0.001	N/A	Different (first flush)
Industrial	N/A	U = 1.864 ; P = 0.03	0.329	0.012	Different
Institutional	0.5224	U = 0.665 ; P > 0.099	0.318	0.029	Same (no first flush)
Open Space	N/A	U = 0.846 ; P = 0.19	0.074	0.337	Same
Residential	N/A	U = 4.029 ; P = 0	0.292	0	Different
All Land Uses	N/A	U = 5.146 ; P = 0	0	N/A	Different

Table E39. Results of Preliminary Statistical Analysis for Total Lead

Total Lead (µg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	89	83	16.98	1.230	0.062	0.075	0.264	0.824	1.65
Commercial First Flush	89	83	27.99	1.447	0.123	0.070	0.264	0.476	1.65
Industrial Composite	84	71	16.98	1.230	0.160	0.527	0.285	0.081	1.41
Industrial First Flush	84	71	23.99	1.380	0.240	0.319	0.285	0.608	1.41
Institutional Composite	18	13	7.00	0.845	0.082	0.675	0.616	0.158	2.28
Institutional First Flush	18	13	15.96	1.203	0.051	0.128	0.616	0.228	2.28
Open Space Composite	31	16	5.00	0.699	0.381	-0.303	0.564	0.199	0.90
Open Space First Flush	31	16	4.48	0.651	0.346	-0.466	0.564	0.563	0.90
Residential Composite	140	93	8.79	0.944	0.231	0.084	0.250	0.884	1.48
Residential First Flush	140	93	13.00	1.114	0.204	0.130	0.250	0.105	1.48
All Land Uses Composite	364	278	13.00	1.114	0.198	-0.365	0.146	0.006	1.50
All Land Uses First Flush	364	278	19.50	1.290	0.239	-0.307	0.146	0.401	1.50

Table E40. Results of Preliminary Test Analysis for Total Lead

Total Lead (µg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	0	U = 5.256 ; P = 0	0.794	0	Different
Industrial	0.083	U = 1.742 ; P = 0.04	0.167	0.016	Different
Institutional	0.004	U = 3.973 ; P = 0	0.680	0.000	Different
Open Space	0.771	U = 0.292 ; P = 0.39	0.008	0.578	Same
Residential	0.012	U = 2.59 ; P = 0	0.014	N/A	Different
All Land Uses	N/A	U = 4.77 ; P = 0	0	N/A	Different

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Table E41. Results of Preliminary Statistical Analysis for Total Nickel

Total Nickel (µg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	47	23	5.00	0.699	0.094	0.660	0.481	0.254	2.40
Commercial First Flush	47	23	11.99	1.079	0.134	-0.606	0.481	0.523	
Industrial Composite	51	22	7.00	0.845	0.106	-0.293	0.491	0.229	1.00
Industrial First Flush	51	22	7.00	0.845	0.197	0.605	0.491	0.228	
Residential Composite	83	18	7.48	0.874	0.094	0.152	0.536	0.814	1.20
Residential First Flush	83	18	8.99	0.954	0.115	1.551	0.536	0.048	
All Land Uses Composite	213	64	6.00	0.778	0.104	0.146	0.299	0.161	1.50
All Land Uses First Flush	213	64	8.99	0.954	0.147	0.322	0.299	0.443	

Table E42. Results of Preliminary Test Analysis for Total Nickel

Total Nickel (µg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	0.006	U = 3.005 ; P = 0	0.128	0.002	Different (first flush)
Industrial	0.715	U = 0.365 ; P = 0.36	0.203	0.484	Same (no first flush)
Residential	N/A	U = 1.143 ; P = 0.13	0.512	0.098	Same
All Land Uses	0.014	U = 2.539 ; P = 0.01	0.367	0.001	Different

Table E43. Results of Preliminary Statistical Analysis for Total Zinc

Total Zinc (µg/L)	Total Events	Selected Cases	Median	Median (Log)	Var (Log)	Skew (Log)	SE Skew (Log)	Test Norm. (Log) p-value	Median Ratio
Commercial Composite	90	90	149.97	2.176	0.089	-1.359	0.254	0	1.93
Commercial First Flush	90	90	289.07	2.461	0.139	-0.374	0.254	0.647	
Industrial Composite	83	83	225.94	2.354	0.184	0.828	0.264	0	1.54
Industrial First Flush	83	83	348.34	2.542	0.135	-0.181	0.264	0.930	
Institutional Composite	18	18	304.79	2.484	0.114	-0.227	0.536	0.878	2.48
Institutional First Flush	18	18	755.09	2.878	0.133	-0.696	0.536	0.055	
Open Space Composite	21	21	20.00	1.301	0.165	0.081	0.501	0.073	1.25
Open Space First Flush	21	21	25.00	1.398	0.075	-0.242	0.501	0.295	
Residential Composite	136	136	69.34	1.841	0.114	0.824	0.208	0.003	1.58
Residential First Flush	136	136	109.90	2.041	0.200	-0.232	0.208	0.014	
All Land Uses Composite	350	350	125.89	2.100	0.216	0.121	0.130	0.001	1.59
All Land Uses First Flush	350	350	199.99	2.301	0.268	0.437	0.130	0.020	

Table E44. Results of Preliminary Test Analysis for Total Zinc

Total Zinc(µg/L)	Mann Wittn. p-value	Fligner Policello	Normality for t-Test p-value	Paired t – Test	Result
Commercial	N/A	U = 6.156 ; P = 0	0	N/A	Different
Industrial	N/A	U = 2.087 ; P = 0.02	0.006	N/A	Different
Institutional	0.007	U = 3.1 ; P = 0	0.498	0	Different
Open Space	N/A	U = 0.023 ; P = 0.49	0.667	0.977	Same
Residential	N/A	U = 4.329 ; P = 0	0	N/A	Different
All Land Uses	N/A	U = 5.374 ; P = 0	0	N/A	Different