# The National Stormwater Quality Database, Version 1.1 

# A Compilation and Analysis of NPDES Stormwater Monitoring Information 

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#### Abstract

The National Stormwater Quality Database v. 1.1 (NSQD) contains selected water quality information from the monitoring carried out as part of the U.S. EPA's National Pollutant Discharge Elimination System (NPDES) Phase 1 stormwater permit applications and subsequent permits, during the period of 1992 to 2002. This database contains about 3,765 events from 360 sites in 65 communities from throughout the U.S. For each site, more additional data, including the percentage of each land use in the catchment, the total area, the percentage of impervious cover, the geographical location, and the season, has been included in the database. Information about the characteristics of each event is also included. Total precipitation, precipitation intensity, total runoff and antecedent dry period are also included, if collected. The database only contains information for samples collected at drainage system outfalls; in-stream samples (which were a component of some state programs) were not included in the database, although some outfalls were located in open channel conveyances.

The first phase requirements of the federal stormwater permit program were first published in the Federal Register by the EPA in 1987 and was initially applied to large cities ( $>100,000$ in population), while Phase II of the stormwater permit program was applied to all urban areas as of early 2003. This program requires significant changes in how stormwater is to be managed. Historical approaches only examined drainage issues, while the new regulations also require consideration of water quality issues.

There are a number of commonly accepted notions that are used by stormwater managers and regulators that can have major impacts on local costs and program effectiveness. This research report examines a number of these potential misconceptions to see how well they hold up under a comprehensive set of actual monitoring data collected throughout the U.S. as part of the Phase I stormwater permit program. This research report is mostly comprised of the major sections of the Ph.D. dissertation prepared by Alex Maestre in partial fulfillment of his degree requirements in the Department of Civil and Environmental Engineering at the University of Alabama.


## Contents

ABSTRACT ..... II
CONTENTS ..... III
LIST OF TABLES ..... IX
LIST OF FIGURES ..... XIII
ACKNOWLEDGEMENTS ..... XVI
Report Organization ..... 2
CHAPTER 2: THE NATIONAL STORMWATER QUALITY DATABASE (NSQD) DESCRIPTION ..... 3
Introduction ..... 3
Data Collection ..... 4
Summary of U.S. NPDES Phase I Stormwater Data in the NSQD ..... 6
Database Structure ..... 6
Site Description [Columns A through Y] ..... 13
Hydrologic Information [Columns Z through AN] ..... 16
Conventional Constituents [Columns AO through BS] ..... 19
Nutrients [Columns BU through CG] ..... 21
Metals [Columns CK through EK] ..... 22
Additional Constituents [Columns EM through HW] ..... 24
Site Descriptions and Additional Supporting Information ..... 24
Problems Encountered during NPDES Stormwater Monitoring ..... 25
Comparison of NSQD with Existing Stormwater Databases ..... 26
Land use effects ..... 31
Other Factors ..... 31
Chapter Summary ..... 32
CHAPTER 3: QA/QC PROCEDURES ..... 33
Introduction ..... 33
Quality Control/ Quality Assurance ..... 33
Unusual Monitoring Locations ..... 34
Non-Detected Analyses ..... 42
Censored Data Distribution ..... 43
Expected Percentages of Observations at Different Levels of Detection ..... 52
Effects of Non-detected Observations on Calculating Mean and Standard Deviation Values ..... 55
Effects on Mean, Median and Coefficient of Variation Values at Different Percentages of Censored Observations56
Total Suspended Solids Analyses at Different Levels of Censoring ..... 61
Summary ..... 66
CHAPTER 4: STORMWATER QUALITY DESCRIPTIONS USING THE THREE PARAMETER LOGNORMAL DISTRIBUTION ..... 67
Introduction ..... 67
The Effects of Unusual High and Low Values on Probability Distribution Parameters ..... 69
Analysis of Lognormality of Stormwater Constituents Parameters ..... 72
Three Parameter Lognormal Calculations ..... 80
Summary ..... 83
CHAPTER 5: IDENTIFICATION OF SIGNIFICANT FACTORS AFFECTING STORMWATER QUALITY USING THE NSQD ..... 88
Introduction ..... 88
Main Factors Affecting Stormwater Quality ..... 88
Effects of Stormwater Controls on Stormwater Quality ..... 90
Sampling Method Effects on Stormwater Concentrations ..... 96
Sample Compositing Procedures ..... 101
Sampling Period during Runoff Event and Selection of Events to Sample ..... 106
Type of Conveyance ..... 108
Concentration Effects Associated with Varying Amounts of Impervious Cover ..... 113
Seasonal Effects on Stormwater Quality ..... 117
Precipitation Depth Effects on Stormwater Quality ..... 118
Antecedent Period without Rain before Monitored Event ..... 120
Trends in Stormwater Quality with Time ..... 124
Summary ..... 125
CHAPTER 6: COMPARISONS OF FIRST 30-MINUTE SAMPLES TO 3-HOUR COMPOSITE SAMPLES ..... 127
Introduction ..... 127
First Flush ..... 127
Methodology ..... 127
Initial Analyses ..... 128
Nonparametric Analyses ..... 130
Results ..... 132
Summary ..... 133
Conclusion ..... 138
CHAPTER 7: EFFECTS OF LAND USE AND GEOGRAPHICAL LOCATION ON STORMWATER QUALITY ..... 139
Model Building using the NSQD ..... 139
ANOVA Evaluation of Suspended Solids Data ..... 139
Descriptive TSS Statistics ..... 140
Land Use and Geographical Area Effects for All Constituents ..... 151
Significant Land Use and Geographical Interactions Affecting MS4 Stormwater Quality ..... 152
CHAPTER 8: EXAMPLE APPLICATION OF THE NATIONAL STORMWATER QUALITY DATABASE (TSS AND NUTRIENT EXPORT CALCULATIONS FOR CHESAPEAKE BAY WATERSHEDS) ..... 170
Overview ..... 170
Data Availability ..... 170
Urban Data ..... 170
Rural Data ..... 171
Summary of Data and Load Calculations ..... 172
Statistical Analyses Performed ..... 182
Residential Area Total Suspended Solids Analyses ..... 182
Residential Area Total Phosphorus Data Analyses ..... 184
Residential Area Total Nitrogen Data Analyses ..... 188
Commercial Area Total Suspended Solids Analyses ..... 190
Commercial Area Total Phosphorus Analyses ..... 190
Commercial Area Total Nitrogen Analyses ..... 190
Industrial Area Total Suspended Solids Analyses ..... 190
Industrial Area Total Phosphorus Analyses ..... 190
Industrial Area Total Nitrogen Analyses ..... 190
Summary ..... 191
CHAPTER 9: FINDINGS AND CONCLUSIONS ..... 192
Introduction ..... 192
Major Findings, as Reported in Report Chapters ..... 192
Findings from Chapter 2: The National Stormwater Quality Database (NSQD) Description ..... 192
Findings from Chapter 3: QA/QC Procedures ..... 197
Findings from Chapter 4: Stormwater Quality Descriptions Using the Three Parameter Lognormal Distribution ..... 198
Findings from Chapter 5: Identification of Significant Factors Affecting Stormwater Quality Using the NSQD 200
Findings from Chapter 6: Comparisons of First 30-minute Samples to 3-hour Composite Samples ..... 207
Findings from Chapter 7: Effects of Land Use and Geographical Location on Stormwater Quality ..... 208
Findings from Chapter 8: Example Application of the National Stormwater Quality Database (TSS and NutrientExport Calculations for Chesapeake Bay Watersheds)211
Research Hypotheses ..... 214
Research Hypothesis 1. Lognormal distributions are robust descriptions of stormwater quality data and a few unusual values have little effect on dataset summary statistical descriptions. ..... 214
Research Hypothesis 2. Censored data can be adequately adjusted by substituting half of the detection limit, withlittle resulting effects on the mean and variance of stormwater datasets.215
Research Hypothesis 3. Different levels of imperviousness are more important than differences in land use categories when predicting stormwater constituent concentrations. ..... 215
Research Hypothesis 4. Antecedent dry periods have a significant effect on stormwater constituent concentrations.215
Research Hypothesis 5. Outfall samples collected during the "first flush" periods of storms have significantly greater concentrations than total storm composite samples. ..... 216
Recommendations for Future Stormwater Permit Monitoring Activities ..... 216
REFERENCES ..... 217
APPENDIX A: SITES INCLUDED IN THE DATABASE ..... 220
APPENDIX B: MODIFIED VALUES IN THE DATABASE ..... 239
Description ..... 239
APPENDIX C: METHODS TO ESTIMATE NON-DETECTED VALUES IN STORMWATER DATASETS ..... 249
Introduction ..... 249
Analysis of Multiple Censored Data ..... 249
APPENDIX D: UNUSUAL SITES IDENTIFIED USING XBAR PLOTS ..... 256
Evaluation of the Methods Selected to Estimate Non-Detected Observations ..... 256
Hardness ..... 256
Oil and Grease ..... 257
Total Dissolved Solids (TDS) ..... 261
Total Suspended Solids (TSS) ..... 263
Biochemical Oxygen Demand $\left(\mathrm{BOD}_{5}\right)$ ..... 264
Chemical Oxygen Demand (COD) ..... 266
Ammonia $\left(\mathrm{NH}_{3}\right)$ ..... 268
Nitrite and Nitrate $\left(\mathrm{NO}_{2}+\mathrm{NO}_{3}\right)$ ..... 270
Total Kjeldahl Nitrogen (TKN) ..... 272
Total Phosphorus ..... 274
Dissolved Phosphorus ..... 275
Total Cooper (Cu) ..... 277
Total Lead ..... 279
Total Zinc ..... 280
Sites with Unusual TSS Concentrations for Different Land Uses ..... 283
Residential and Mixed Residential Locations ..... 283
Commercial and Mixed Commercial Locations ..... 285
Industrial and Mixed Industrial Locations ..... 289
APPENDIX E: FIRST FLUSH TABLES ..... 333
Description ..... 333
APPENDIX F: DETAILED STATISTICAL TEST RESULTS TO IDENTIFY SIGNIFICANT LAND USE AND GEOGRAPHICAL INTERACTIONS ..... 347
pH ..... 347
Temperature ..... 359
Hardness ..... 364
Oil and Grease ..... 373
Total Dissolved Solids ..... 380
Total Suspended Solids ..... 388
Biochemical Oxygen Demand, 5 day ( $\mathrm{BOD}_{5}$ ) ..... 396
Chemical Oxygen Demand (COD) ..... 404
Ammonia ( $\mathbf{N H}_{3}$ ) ..... 411
Nitrite plus Nitrate $\left(\mathbf{N O}_{\mathbf{2}}+\mathbf{N O}_{\mathbf{3}}\right)$ ..... 417
Total Kjeldahl Nitrogen (TKN) ..... 424
Total Phosphorus (P) ..... 431
Dissolved Phosphorus (dissolved - P) ..... 438
Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) (Cu) ..... 445

Total Lead $(\mu \mathrm{g} / \mathrm{L})(\mathrm{Pb}) \quad 453$
Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) (Zn)

## List of Tables

1. Samples and Sites by EPA Rain Zone 12
2. Total Samples and Sites by Land Use 13
3. Summary of Available Stormwater Data Included in NSQD, Version 1.1
4. Conventional Constituents Summary 33
5. Nutrients Summary 36
6. Summary of Metal Concentrations 38
7. Summary of Additional Constituents 40
8. Additional Site Information 41
9. Comparison of Stormwater Databases 48
10. Total Events Monitored during NURP by EPA Rain Zones 49
11. Percentages of Detected Values by Land Use Category and for the Complete Database 59
12. Percentages of Non-Detected Values for Different Reported Detection Limits by Land Use and for the Total Database 63
13. Percentage of Observations below Specific Concentrations 68
14. Descriptive Statistics for TSS Truncated at Different Levels 81
15. Sites Failing Xbar and R Chart in Residential Land Use 90
16. Sites Failing Xbar Chart in Residential Land Use 94
17. Sites Failing Xbar Chart in Commercial Land Use 95
18. Sites Failing Xbar Chart in Industrial Land Use 97
19. Power of the Test when applied to Selected Constituents in Residential Land Uses 113
20. Comparison of Goodness of Fit for Gamma, Exponential and Lognormal Distribution using the NSQD v. 1.1116
21. Goodness of Fit for Different Land Uses 123
22. Numbers and Percentage of Samples by Discrete Site Variable Category 130
23. One-Way ANOVA Results by Control Type in Residential Land Use, Rain Zone 2134
24. One-Way ANOVA Results by Control Type in Commercial Land Use, Rain Zone 2135
25. One-Way ANOVA Results by Control Type in Industrial Land Use, Rain Zone 2136
26. One-Way ANOVA Results by Control Type in Industrial Land Use, Rain Zone 3137
27. One-Way ANOVA Results by Type of Sampler in Residential Land Use, Rain Zone 2140
28. One-Way ANOVA Results by Type of Sampler in Commercial Land Use, Rain Zone 2141
29. One-Way ANOVA Results by Type of Sampler in Industrial Land Use, Rain Zone 2142
30. One-Way ANOVA Results by Sample Compositing Scheme in Residential Land Use, Rain Zone 2147
31. One-Way ANOVA Results by Sample Compositing Scheme in Commercial Land Use, Rain Zone 2148
32. One-Way ANOVA Results by Sample Compositing Scheme in Industrial Land Use, Rain Zone 2149
33. One-Way ANOVA Results by Sample Compositing Scheme in Industrial Land Use, Rain Zone 3150
34. One-Way ANOVA Results by Type of Conveyance in Residential Land Use, Rain Zone 2155
35. One-Way ANOVA Results by Type of Conveyance in Industrial Land Use, Rain Zone 2156
36. One-Way ANOVA Results by Type of Conveyance in Residential Land Use, Rain Zone 3157
37. One-Way ANOVA Results by Type of Conveyance in Industrial Land Use, Rain Zone 3158
38. Regression of Median Concentrations by Percentage of Impervious in Residential Land Use, EPA Rain Zone 2164
39. Regression of Median Concentrations by Percentage of Impervious in Commercial and Industrial Land Use, EPA Rain Zone 2165
40. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period for Residential Land Use, EPA Rain Zone 2170
41. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period for Commercial Land Use, EPA Rain Zone 2172
42. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period for Industrial Land Use, EPA Rain Zone 2
43. Comparison of Commercial and Residential Stormwater Runoff Quality from 1980/81 to 1992/93 176
44. Preliminary Number of Storm Events Selected
45. Initial Analyses 187
46. Significant First Flushes Ratios 190
47. Ranking by Methods of Sampling 198
48. Urban Monitoring Locations in the Chesapeake Bay Watershed Represented in the National Stormwater Quality Database 201
49. Land Uses in the Chesapeake Bay Watershed 203
50. Reported Mean Annual Yields and Land Use 204
51. Commercial TSS (mg/L), Mean and COV 207
52. Commercial Total Nitrogen (mg/L), Mean and COV 207
53. Average Concentration By Land Use 208
54. Fraction of Annual Flow Associated with Season and Rain Depth Categories (based on 50 years of rain at Baltimore, BWI) 209
55. Total Suspended Solids Concentrations for Land Use Categories in Anne Arundel County, Maryland 210
56. Total Phosphorus Concentrations for Land Use Categories in Anne Arundel County, Maryland 211
57. Urban Areas Concentrations 212
58. Total Nitrogen Calculated Concentrations for Land Use Categories in Anne Arundel County, Maryland 213
59. Discharge by Major Land Use Categories in Anne Arundel County, Maryland 214
60. Urban Land Use Categories in Anne Arundel County, Maryland 215
61. Other Land Use Categories Used in Anne Arundel County Calculations 216
62. Results for Residential TSS (mg/L) 222
63. Results for Residential Total Phosphorus (Impervious $<27 \%$ ) 224
64. Results for Residential Total Phosphorus (Impervious > 27 \%) 226
65. Results for Residential Total Nitrogen

## List of Figures

1. Communities included in the NSQD version 1 by Rain Zones 12
2. Drainage area by land use 23
3. Percentage of impervious are by land use 25
4. Scatter plot of percentage of impervious and Rv 26
5. Precipitation and runoff depth by land use 30
6. Box and whiskers plots for conventional constituents by single land use 34
7. Water temperature in EPA Rain Zone 5 and 6
8. Box and whiskers plots for nutrients by single land use 37
9. Box and whiskers plots for metals by single land use 39
10. Distribution of collected events using the NSQD database 50
11. Distribution of collected events using the NURP database 50
12. Example of constituents collected in residential land use by EPA Rain Zone 51
13. Distribution of collected events using the NURP database 53
14. Example scatter plots of stormwater data 56
15. Effects on the mean when using random estimated values versus ignoring the non- detected observations, at different percentage of detected values
16. Effect of ignoring the non-detected observations on the median 76
17. Effect of ignoring the non-detected observations in the median 78
18. Effect of ignoring the non-detected observations on the coefficient of variation 80
19. Effect on the mean when TSS observation are truncated 82
20. Effect on the median when the TSS dataset is truncated 83
21. Effect on the standard deviation when TSS dataset is truncated 84
22. Effect on the coefficient of variation when the TSS dataset is truncated 85
23. Box and whiskers plots for TSS, total phosphorus and total copper by EPA Rain Zone and land use 86
24. TSS box and whiskers plots in residential land use by EPA Rain Zone and location.......... 89
25. Xbar S chart for residential land use in EPA Rain Zone $2 \quad 90$
26. Probability plot of total dissolved solids in residential land use 102
27. Effect of unusual values on the coefficient of variation 104
28. Effect of unusual values on the standard deviation 106
29. Effect of unusual values on the mean 108
30. Cumulative and empirical distributions of total copper for residential land use data 109
31. Lower bounds for the power of the $D$ test for $\alpha=1,5$, and 10 percent 112
32. Normality test for total phosphorus in residential land use using the NSQD
33. $D_{\max }$ was located in the middle of the distribution 115
34. Estimated models for TSS in commercial land use 119
35. Estimated models for TDS in industrial land use 120
36. TSS distribution by controls in residential areas and EPA Rain Zone 2132
37. Comparison of reported concentrations in residential land use and EPA Rain Zone 2 for automatic vs. manual sampling methods 144
38. Comparison between time- and flow-composite options for TSS
39. Histogram of possible TSS concentrations in Flagstaff Street based on collecting three samples per year for two years 153
40. TSS concentration by type of conveyance 160
41. Plot of COD concentrations against watershed area percent imperviousness values for different land uses 161
42. TSS concentrations by impervious cover and single land use 162
43. Total nitrates regression at different percentages of impervious 164
44. Example residential area stormwater pollutant concentrations sorted by season 166
45. Example of scatter plots by precipitation depth ..... 167
46. Precipitation depth and runoff depth by land use ..... 168
47. Box and whiskers plot of days since preceding event by Rain Zone ..... 170
48. Nutrient concentration affected by dry periods since last rain inresidential land use171
49. Total phosphorus and total lead as function of antecedent dry periodin commercial land use 173
50. TSS concentrations for days since preceding event in industrial land use17451. Comparison of pollutant concentrations collected during NURP (1981)to MS4 application data (1990) at the same location 175
51. Residential lead and COD concentrations with time ..... 176
52. Cumulative probability and box and whiskers plots ..... 185
53. Analysis flow chart ..... 189
54. Hydrograph for a storm event ..... 193
55. Contributing areas in urban watersheds (Pitt, 1999) ..... 194
56. Contributing areas in urban watersheds (Pitt, 1995) ..... 195
57. Probability and box and whiskers plot of selected rain events ..... 196
58. Example of an event with peaks after the sampling period ..... 197
59. Sources of runoff, TSS and nutrients for different sources in Anne Arundel ..... 218
60. Residential TSS distribution by groups ..... 221
61. Residential TSS box and whiskers plot distribution by groups ..... 22263. Residential total phosphorus concentrations for siteshaving $<27$ \% impervious surfaces 223
62. Box and whiskers plot for residential total phosphorus concentrations for siteshaving $<27 \%$ impervious surfaces 224
63. Residential total phosphorus concentrations for siteshaving >27 \% impervious surfaces 225
64. Box and whiskers plot for residential total phosphorus concentrations for siteshaving >27 \% impervious surfaces 225
65. Residential total nitrogen concentration groups ..... 227
66. Box and whiskers plot for residential total nitrogen ..... 227

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## Chapter 1: Introduction

The first phase of the federal stormwater permit program was first published in the Federal Register by the EPA in 1987 and was initially applied to large cities ( $>100,000$ in population), while Phase II of the stormwater permit program was applied to all urban areas as of early 2003. This program requires significant changes in how stormwater is to be managed. Historical approaches only examined drainage issues, while the new regulations also required consideration of water quality issues. Unfortunately, some professionals involved with stormwater management may not have an adequate understanding of stormwater characteristics, including its effects, and treatability. As an example, there are a number of commonly accepted notions that are used by stormwater managers and regulators that can have major impacts on local costs and program effectiveness. This research report examines a number of these notions to see how well they hold up under a comprehensive set of actual monitoring data collected throughout the U.S. as part of the Phase I stormwater permit program. This research report also includes a predictive tool that can assist stormwater managers in predicting expected stormwater conditions for local areas.

Researchers from the University of Alabama and the Center for Watershed Protection assembled a large database of stormwater characteristics, the National Stormwater Quality Database (NSQD), as part of an EPA-funded section 104(b)3 project from the Office of Water. This is the largest collection of information on stormwater characteristics ever assembled for US conditions. The research described in this report used this information to test the validity of several commonly accepted notions concerning stormwater, and produced a statistical tool that hopefully can assist stormwater managers and regulators. In addition, many suggestions concerning monitoring strategies for stormwater are summarized, based on the experiences of many of the Phase I permitted communities. The cumulative value of the monitoring data collected over nearly a ten-year period from more than 200 municipalities throughout the country has a great potential in characterizing the quality of stormwater runoff and comparing it against historical benchmarks.

The data set received a comprehensive quality assurance/quality control review, based on reasonableness of data, extreme values, relationships among parameters, sampling methods, and a review of the analytical methods. The statistical analyses were conducted at several levels. Probability plots were used to identify range, randomness and normality. Multivariate analyses were also utilized to characterize significant factors affecting the data patterns. The master data set was also evaluated to develop descriptive statistics, such as measures of central tendency and standard errors. Testing was done for regional and climatic differences, the influences of land use, and the effects of storm size, drainage area and season, among other factors.

This National Stormwater Quality Database (NSQD), in its first version presented here, is not intended for comprehensive characterization purposes for all conceivable situations and to replace the need for all characterization monitoring. Some communities may have obvious unusual conditions, or adequate data may not be available in the database for their region. In these conditions, site specific local outfall monitoring may be needed. In addition, stormwater monitoring will continue to be needed for other purposes in many areas having, or anticipating, active stormwater management programs (especially when supplemented with other biological, physical, and hydrologic monitoring components). These new monitoring programs should be designed specifically for additional objectives, beyond simple characterization. These may include receiving water assessments to understand local problems, source area monitoring to identify critical sources, treatability tests to verify performance of stormwater controls for local conditions, and assessment monitoring to verify the success of local stormwater management approaches (including model calibration and verification). In many cases, however, the resources being spent for conventional outfall monitoring could be more effectively spent to better understand many of these other aspects of an effective stormwater management program.

## Report Organization

This report is divided into nine chapters and five appendices. Chapter 2 describes the National Stormwater Quality Database (NSQD). Chapter 3 describes the QA/QC procedures used during the collection of data and creation of the database, including an evaluation of alternative methods to address the presence of non-detected values. Chapter 4 addresses the hypothesis concerning the probability distributions most appropriate for the stormwater constituents. Chapter 5 describes the results of the investigations relating constituent concentrations to main factors and interactions of parameters described in the site description and hydrologic information sections of the database. Chapter 6 presents the results from the "first flush" analysis. Chapter 7 presents detailed results of the statistical tests used to develop predictive models of stormwater characteristics affected by geographical location and land use. Chapter 8 presents an example of how the data in the NSQD can be used to estimate the concentration of stormwater constituents for Maryland and Virginia (the region best represented in the database). Chapter 9 presents the conclusions and recommendations of this research.

# Chapter 2: The National Stormwater Quality Database (NSQD) Description 

## Introduction

The National Stormwater Quality Database (NSQD) was prepared by the University of Alabama and the Center for Watershed Protection under 104(b)3 funding from the U.S. Environmental Protection Agency (EPA). The NSQD is a spreadsheet database and supporting documents describing the monitoring efforts of 65 communities from throughout the U.S. that are larger than 100,000. The monitoring period covered by the NSQD is from 1992 to 2002.

Several efforts have been performed in the past to describe the water quality characteristics of stormwater constituents at different locations. The importance of this EPA-sponsored project is based on the scarcity of nationally summarized and accessible data from the existing U.S. EPA's NPDES (National Pollutant Discharge Elimination System) stormwater permit program. There have been some local and regional data summaries, but little has been done with nationwide data. A notable exception is the Camp, Dresser, and McGee (CDM) national stormwater database (Smullen and Cave 2002) that combined historical Nationwide Urban Runoff Program (NURP) (EPA 1983) data, available urban U.S. Geological survey (USGS), and selected NPDES data. Their main effort had been to describe the probability distributions of these data (and corresponding EMCs, the event mean concentrations). They concluded that concentrations for different land uses were not significantly different, so all their data were pooled into a single urban land use category.

The Clean Water Act (CWA) of 1972 was the first major national regulation in the U.S. requiring control of conventional point source discharges of water pollutants (affecting municipal and industrial discharges). Section 208 also provided the capability to implement stormwater management plans at the regional level. In 1976, the EPA enlarged the planning initiative through the "Section 208: Areawide Assessment Procedures Manual". However, in the late 1970s, some problems arose with the 208 planning projects due to inadequate data and lack of technological development (Whipple, as quoted by Pitt, et al. 1999).

Between 1978 and 1983, the EPA conducted the Nationwide Urban Runoff Program (NURP) that examined stormwater quality from separate storm sewers in different land uses (EPA 1983). This program studied 81 outfalls in 28 communities throughout the U.S. and included the monitoring of approximately 2,300 storm events. NURP is still an important reference for water quality characteristics of urban stormwater; however, the collected data poorly represented the southern area of the country and was focused mainly in residential and mixed land use areas. Since NURP, other important studies have been conducted that characterize stormwater. The USGS created a database with more than 1,100 storms from 98 monitoring sites in 20 metropolitan areas. The Federal Highway Administration (FHWA) analyzed stormwater runoff from 31 highways in 11 states during the 1970s and 1980s. Strecker (personal communication) is also collecting information from highway monitoring as part of a current NCHRP (National Cooperative Highway Research Program) funded project. The city of Austin also developed a database having more than 1,200 events.

Other regional databases also exist for U.S. data, mostly using local NPDES data. These include the Los Angeles area database, the Santa Clara and Alameda County (California) databases, the Oregon Association of Clean Water Agencies Database, and the Dallas, Texas, area stormwater database. These regional data are included in the NSQD. However, the USGS and historical NURP data are not included in the NSQD due to lack of consistent descriptive information for the older drainage areas and because of the age of the data from those prior studies. Much of the NURP data is available in electronic form at the University of Alabama's student American Water Resources Association web page at: http://www.eng.ua.edu/~awra/download.htm.

Outside the U.S., there have been important efforts to characterize stormwater. In Toronto, Canada, the Toronto Area Watershed Management Strategy Study (TAWMS) was conducted during 1983 and 1984 and extensively
monitored industrial stormwater, along with snowmelt in the Toronto urban area, for example. Numerous other investigations in South Africa, the South Pacific, Europe and Latin America have also been conducted over the past 30 years, but no large-scale summaries of that data have been prepared. About 4,000 international references on stormwater have been reviewed and compiled since 1996 by the Urban Wet Weather Flows literature review team for publication in Water Environment Research (most recently by Clark, et al. 2001, 2002, 2003, 2004). An overall compilation of these literature reviews is available at: http://www.eng.ua.edu/~rpitt/Publications/Publications.shtml. These reviews include short summaries of the papers and are organized by major topics. Besides journal articles, many published conference proceedings are also represented (including the extensive conference proceedings from the $7^{\text {th }}$ International Conference on Urban Storm Drainage held in Germany in 1996, the $8^{\text {th }}$ International Conference on Urban Storm Drainage held in Sydney, Australia, in 1999, the $9^{\text {th }}$ International Conference on Urban Storm Drainage held in Portland, OR, in 2002, and the Urban Water Systems Modeling conference series for the Toronto meetings organized by Computational Hydraulics, Inc., amongst many other specialty conferences).

In 1987, the amendments to the CWA established a two-phase program to regulate 13 classes of stormwater discharges. Two of these classifications were discharges from large and medium-sized Municipal Separate Storm Sewer Systems. A large MS4 serves an urban population of 250,000 or more, while a medium MS4 serves communities between 100,000 and 250,000 . EPA set up a permit strategy for communities complying with NPDES requirements. Monitoring data from this program have been included in some databases. The CDM National Stormwater Runoff Pollution database included 816 NPDES storm events in a database that totals approximately 3,100 events. The Rouge River National Wet Weather Demonstration Program office in Detroit included their NPDES data in their database (Smullen and Cave 2003).

Another important effort has been the development of the National Stormwater Best Management Practices Database (http://www.bmpdatabase.com). This database was created with the purpose to evaluate the performance and effectiveness of stormwater control practices, frequently labeled "best management practices," or BMP's. Detention ponds, street cleaning, and hydrodynamic devices are examples of BMPs (ASCE/EPA 2000).

## Data Collection

Data from 3,765 storm events at 360 monitoring sites were collected and are stored in version 1.1 of the NSQD. This version contains the results of approximately one fourth of the total number of communities that participated in the Phase I NPDES stormwater permit monitoring activities.

According to the published sampling guidance (40 CFR 122.21) for the permit application, each community was required to sample at least a residential, a commercial and an industrial watershed. At least three samples should be collected every year at each location. Each storm should be at least one month apart and have at least a 3 days antecedent dry period. Only samples from rain events greater than 0.1 inches, and close to the annual mean conditions, were considered valid for the analysis. It was required to collect a composite sample with subsamples collected during the first three hours of the event. An additional grab sample was required during the first 30 minutes of the event to evaluate the "first flush" effect. "First flush" refers to the hypothesis that the concentrations of stormwater constituents are higher at the beginning of the discharge event than during the complete event. Designated states were able to modify some of these sampling requirements to better address local concerns.

Most communities were required to submit annual reports describing the sampling locations and procedures, the equipment, and the quality control and quality assurance (QA/QC) procedures used during the sampling and analysis of the samples, the analytical methods used in the laboratory, and problems encountered during the sample collection. The reports also included the results of the chemical analyses performed by the laboratories.

Figure 1 is a map showing the 65 communities and 17 states included in the first version of the NSQD. The EPAfunded project was intended to focus on the Chesapeake Bay area and parts of the southern U.S. (specifically Birmingham, AL, and Atlanta, GA) as a demonstration of the usefulness of the data. However, it was possible to obtain some data from other parts of the country during the project period and these data were incorporated in the database, allowing some regional analyses. States representing most of the samples included Virginia (24\%) and Maryland (13\%). The states with low numbers of observations included Pennsylvania, Massachusetts, and Indiana.

Figure 1 also shows the EPA Rain Zones. Each zone corresponds to a geographical region with similar climatic conditions (EPA 1986). There is at least one community per rain zone indicating some geographical representation for the entire country. However, Table 1 indicates that most of the samples were collected west, south and east of the continental part of the country, with few of the large amounts of data from EPA Rain Zone 1 included in the database. EPA Rain Zones 8 and 9 have sparse available data from the Phase I monitoring program, due to few large cities in these areas.


Figure 1. Communities included in the NSQD version 1.1 by rainfall zones

Table 1. Total Samples and Sites by EPA Rain Zone

| EPA Rain Zone | Total Samples | Percentage of <br> Samples | Number of <br> Communities | Number of <br> Sites |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 69 | 1.8 | 2 | 12 |
| 2 | 2000 | 53 | 28 | 185 |
| 3 | 266 | 7.1 | 8 | 30 |
| 4 | 212 | 5.6 | 4 | 21 |
| 5 | 485 | 13 | 9 | 33 |
| 6 | 356 | 9.5 | 4 | 30 |
| 7 | 229 | 6.1 | 6 | 28 |
| 8 | 24 | 0.63 | 1 | 4 |
| 9 | 124 | 3.3 | 3 | 17 |

Each site in the database corresponds to an outfall where the runoff produced in the watershed is discharged. During the monitored events, samples were collected to identify the characteristics of the stormwater being discharged. According to the land use of the watershed, each site was classified as residential, commercial, industrial, open space, freeway, or mixed. When a single land use was not identified for the watershed, then the site was considered mixed, with a predominant land use. Table 2 indicates the total number of sites included in the database, separated by land use.

Table 2. Total Samples and Sites by Land Use

| Land use | Number of Sites | Percentage | Number of Events | Percentage |
| :--- | :---: | :---: | :---: | :---: |
| Residential | 111 | 31 | 1042 | 28 |
| Mixed Residential | 44 | 12 | 611 | 16 |
| Commercial | 51 | 14 | 526 | 14 |
| Mixed Commercial | 29 | 8.1 | 325 | 8.6 |
| Industrial | 54 | 15 | 566 | 15 |
| Mixed Industrial | 22 | 6.1 | 249 | 6.6 |
| Institutional | 1 | 0.3 | 18 | 0.5 |
| Open Space | 10 | 2.8 | 49 | 1.3 |
| Mixed Open Space | 13 | 3.6 | 168 | 4.5 |
| Freeways | 22 | 6.1 | 185 | 4.9 |
| Mixed Freeways | 3 | 0.8 | 26 | 0.7 |

About one third of the sites included in the database correspond to residential areas, another third is shared by commercial and industrial land uses. The remaining third correspond to freeways, open space, institutional and all the mixed land uses. Several schools were identified in the sites, however only one site was considered $100 \%$ institutional.

## Summary of U.S. NPDES Phase I Stormwater Data in the NSQD

Table 3 is a summary of selected data collected and entered into the database. The data are separated into 11 land use categories: residential, commercial, industrial, institutional, freeways, and open space, plus mixtures of these land uses. Summaries are shown for the major land use areas and for the total data set combined. The full database includes all of the data. The total number of observations and the percentage of observations above the detection limits are also shown on this summary table. In general, the coefficient of variation (COV) values range from 1.0 to 2.0 for the majority of pollutants across all major land uses.

The following sections describe the structure of the full database and present some findings. The findings presented are focused on specific issues and are illustrated using small portions of the complete database to minimize the effects of other interacting factors (such as using data from a single region and land use to show the effects of sampling methods, for example). Later sections of this report present more comprehensive discussions of the data that do consider interactions of the many factors available in the database.

## Database Structure

The database has five major sections: General Information, Items Description, Constituents and Parameters, and the Database itself. In addition, detailed site information along with aerial photographs and topographic maps is provided for each municipality and monitoring location. Each of the sections is a tab in the bottom part of the spreadsheet.
Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1

|  | Area (acres) | \% Impervious | Precipitation Depth (in) | Runoff Depth (in) | Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ <br> $@ 25^{\circ} \mathrm{C}$ ) | $\begin{gathered} \text { Hardness } \\ (\mathrm{mg} / \mathrm{L} \\ \text { CaCO3) } \\ \hline \end{gathered}$ | Oil and Grease (mg/L) | pH | Temperature (C) | $\begin{aligned} & \text { TDS } \\ & \text { (mg/L) } \end{aligned}$ | $\begin{aligned} & \text { TSS } \\ & \text { (mg/L) } \end{aligned}$ | $\begin{gathered} \mathrm{BOD}_{5} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \text { COD } \\ & \text { (mg/L) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Overall Summary (3765) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 3765 | 2209 | 3316 | 1495 | 685 | 1082 | 1834 | 1665 | 861 | 2956 | 3493 | 3105 | 2750 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 98.7 | 66.1 | 100 | 100 | 99.0 | 97.9 | 96.2 | 98.4 |
| Median | 57.3 | 50.0 | 0.48 | 0.15 | 121 | 38.0 | 4.3 | 7.5 | 16.5 | 80 | 59 | 8.6 | 53 |
| Coefficient of variation | 3.7 | 0.4 | 1.0 | 1.9 | 1.6 | 1.4 | 9.7 | 0.1 | 0.4 | 3.4 | 1.8 | 7.4 | 1.1 |
| Residential (1042) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 1042 | 614 | 919 | 372 | 104 | 215 | 483 | 286 | 181 | 814 | 978 | 908 | 748 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 100 | 54.9 | 100 | 100 | 99.1 | 98.3 | 97.1 | 98.7 |
| Median | 57.3 | 37.0 | 0.48 | 0.10 | 102 | 32.0 | 4.0 | 7.2 | 17.0 | 72.0 | 49 | 9.0 | 54.5 |
| Coefficient of variation | 4.8 | 0.4 | 1.0 | 1.5 | 1.6 | 1.1 | 7.8 | 0.1 | 0.4 | 1.1 | 1.8 | 1.5 | 0.93 |
| Mixed Residential (611) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 611 | 278 | 491 | 262 | 105 | 168 | 283 | 333 | 137 | 491 | 582 | 549 | 465 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 98.2 | 70.3 | 100 | 100 | 99.2 | 98.3 | 94.2 | 99.6 |
| Median | 150.8 | 44.9 | 0.53 | 0.12 | 112 | 40.0 | 4.0 | 7.50 | 15.5 | 86 | 66 | 7.8 | 43 |
| Coefficient of variation | 2.1 | 0.3 | 0.8 | 1.3 | 1.2 | 1.1 | 2.6 | 0.1 | 0.3 | 5.2 | 1.6 | 1.3 | 1.2 |
| Commercial (527) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 527 | 284 | 462 | 146 | 78 | 156 | 331 | 191 | 98 | 418 | 503 | 452 | 393 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 100 | 71.9 | 100 | 100 | 99.5 | 95.2 | 97.6 | 98.5 |
| Median | 38.8 | 84.5 | 0.42 | 0.29 | 107 | 36.5 | 4.6 | 7.4 | 16.0 | 72 | 43 | 11.0 | 58 |
| Coefficient of variation | 1.2 | 0.1 | 1.0 | 1.0 | 1.0 | 1.1 | 3.0 | 0.1 | 0.4 | 1.9 | 2.0 | 1.1 | 1.0 |
| Mixed Commercial (324) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 324 | 237 | 305 | 118 | 59 | 98 | 134 | 156 | 98 | 265 | 297 | 277 | 267 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 99.0 | 79.9 | 100 | 100 | 99.6 | 99.7 | 98.9 | 99.6 |
| Median | 75.0 | 60.0 | 0.47 | 0.28 | 100 | 36.0 | 5.0 | 7.60 | 14.5 | 69.5 | 54.5 | 9.0 | 60 |
| Coefficient of variation | 1.4 | 0.3 | 1.0 | 0.9 | 0.8 | 1.8 | 2.9 | 0.1 | 0.4 | 1.9 | 1.3 | 1.7 | 1.0 |
| Industrial (566) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 566 | 292 | 482 | 215 | 102 | 132 | 315 | 248 | 140 | 431 | 521 | 455 | 386 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 96.2 | 64.8 | 100 | 100 | 99.5 | 97.7 | 95.4 | 99.0 |
| Median | 39.5 | 75.0 | 0.50 | 0.16 | 139 | 39.0 | 4.8 | 7.50 | 17.9 | 86 | 81 | 9.0 | 58.6 |
| Coefficient of variation | 1.1 | 0.3 | 0.9 | 1.2 | 1.3 | 1.5 | 11.8 | 0.1 | 0.3 | 3.6 | 1.6 | 10.0 | 1.2 |

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 - Continued

|  | $\begin{gathered} \text { Area } \\ \text { (acres) } \end{gathered}$ | $\begin{gathered} \% \\ \text { Impervious } \\ \hline \end{gathered}$ | Precipitation Depth (in) | Runoff Depth (in) | Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ @25응 | Hardness (mg/L CaCO3) | Oil and Grease (mg/L) | pH | Temperature (C) | $\begin{gathered} \text { TDS } \\ (\mathrm{mg} / \mathrm{L}) \\ \hline \end{gathered}$ | $\begin{gathered} \text { TSS } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\mathrm{BOD}_{5}$ (mg/L) | $\begin{gathered} \operatorname{COD} \\ (\mathrm{mg} / \mathrm{L}) \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mixed Industrial (218) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 218 | 118 | 193 | 117 | 56 | 75 | 72 | 152 | 57 | 186 | 207 | 178 | 175 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 93.3 | 80.6 | 100 | 100 | 99.5 | 100 | 95.5 | 98.9 |
| Median | 168.0 | 44.0 | 0.45 | 0.29 | 126 | 29.3 | 9.0 | 7.70 | 18.0 | 90 | 82 | 7.5 | 39.9 |
| Coefficient of variation | 1.8 | 0.3 | 0.9 | 1.2 | 0.8 | 0.6 | 1.8 | 0.1 | 0.3 | 0.8 | 1.4 | 1.8 | 1.2 |
| Institutional (18) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 18 | 18 | 17 | 14 |  |  |  |  |  | 18 | 18 | 18 | 18 |
| $\%$ of samples above detection | 100 | 100 | 100 | 100 |  |  |  |  |  | 100 | 94.4 | 88.9 | 88.9 |
| Median | 36.0 | 45.0 | 0.18 | 0.00 |  |  |  |  |  | 52.5 | 17 | 8.5 | 50 |
| Coefficient of variation | 0 | 0 | 0.9 | 2.1 |  |  |  |  |  | 0.7 | 0.83 | 0.7 | 0.9 |
| Freeways (185) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 185 | 154 | 182 | 144 | 86 | 127 | 60 | 111 | 31 | 97 | 134 | 26 | 67 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 100 | 71.7 | 100 | 100 | 99.0 | 99.3 | 84.6 | 98.5 |
| Median | 1.6 | 80.0 | 0.54 | 0.41 | 99 | 34.0 | 8.0 | 7.10 | 14.0 | 77.5 | 99 | 8 | 100 |
| Coefficient of variation | 1.4 | 0.13 | 1.1 | 1.7 | 1.0 | 1.9 | 0.6 | 0.1 | 0.4 | 0.8 | 2.6 | 1.3 | 1.1 |
| Mixed Freeways (26) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 26 |  | 26 |  | 21 | 12 | 20 | 17 | 17 | 15 | 23 | 23 | 15 |
| $\%$ of samples above detection | 100 |  | 100 |  | 100 | 100 | 100 | 100 | 100 | 100 | 100 | 100.0 | 100.0 |
| Median | 63.1 |  | 0.47 |  | 353 | 83 | 4.5 | 7.7 | 16.0 | 177 | 88 | 8.2 | 47 |
| Coefficient of variation | 0.7 |  | 0.8 |  | 0.6 | 0.3 | 1.8 | 0.1 | 0.3 | 0.4 | 1.1 | 1.2 | 0.5 |
| Open Space (49) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 49 | 37 | 41 | 11 | 2 | 8 | 19 | 19 | 2 | 45 | 44 | 44 | 43 |
| $\%$ of samples above detection | 100 | 100 | 100 | 100 | 100 | 100 | 36.8 | 100 | 100 | 97.8 | 95.5 | 86.4 | 76.74 |
| Median | 85 | 2.0 | 0.52 | 0.05 | 113 | 150 | 1.3 | 7.70 | 14.6 | 125 | 48.5 | 5.4 | 42.1 |
| Coefficient of variation | 1.5 | 1.0 | 1.2 | 1.4 | 0.5 | 0.6 | 0.7 | 0.08 | 0.7 | 0.7 | 1.5 | 0.7 | 1.5 |
| Mixed Open Space (168) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 168 | 131 | 167 | 93 | 65 | 70 | 90 | 128 | 76 | 148 | 153 | 145 | 145 |
| \% of samples above detection | 100 | 100 | 100 | 100 | 100 | 100 | 60.0 | 100 | 100 | 99.3 | 97.4 | 96.6 | 96.6 |
| Median | 115.4 | 33.0 | 0.51 | 0.10 | 215 | 64.2 | 8.5 | 7.9 | 16.0 | 109 | 78.0 | 6.0 | 34 |
| Coefficient of variation | 0.8 | 0.4 | 0.8 | 1.2 | 1.7 | 1.3 | 1.5 | 0.1 | 0.3 | 2.2 | 1.6 | 2.7 | 1.6 |

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 - Continued

|  | Fecal Coliform (mpn/100 mL ) | Fecal Streptococcus (mpn/100 mL ) |  | $\begin{gathered} \text { Total E. } \\ \text { Coli } \\ (\mathrm{mpn} / 100 \\ \mathrm{mL}) \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{NH} 3 \\ (\mathrm{mg} / \mathrm{L}) \\ \hline \end{gathered}$ | $\begin{gathered} \mathrm{N} 02+\mathrm{NO} 3 \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Nitrogen, Total Kjeldahl (mg/L) | Phosphorus, filtered (mg/L) | Phosphorus, total (mg/L) | Sb, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | As, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | As, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Be, total ( $\mu \mathrm{g} / \mathrm{L}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Overall Summary (3765) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 1704 | 1141 | 83 | 67 | 1908 | 3075 | 3191 | 2477 | 3285 | 874 | 1507 | 210 | 947 |
| \% of samples above detection | 91.2 | 94.0 | 90.4 | 95.5 | 71.3 | 97.3 | 95.6 | 85.1 | 96.5 | 7.2 | 49.9 | 27.1 | 7.7 |
| Median | 5091 | 17000 | 12000 | 1750 | 0.44 | 0.60 | 1.4 | 0.13 | 0.27 | 3.0 | 3.0 | 1.5 | 0.4 |
| Coefficient of variation | 4.6 | 3.8 | 2.4 | 2.3 | 1.4 | 0.97 | 1.2 | 1.6 | 1.5 | 1.7 | 2.6 | 1.0 | 2.5 |
| Residential (1042) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 402 | 257 |  | 14 | 572 | 889 | 922 | 690 | 926 |  | 395 |  | 282 |
| \% of samples above detection | 87.8 | 87.9 |  | 100 | 82.2 | 97.6 | 96.5 | 83.5 | 96.8 |  | 40.8 |  | 7.8 |
| Median | 7000 | 24300 |  | 700 | 0.31 | 0.60 | 1.5 | 0.18 | 0.31 |  | 3.0 |  | 0.5 |
| Coefficient of variation | 5.2 | 1.7 |  | 1.6 | 1.1 | 1.1 | 1.1 | 0.9 | 1.1 |  | 2.2 |  | 2.5 |
| Mixed Residential (611) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 336 | 178 | 26 | 11 | 282 | 531 | 517 | 430 | 552 |  | 158 |  | 97 |
| \% of samples above detection | 94.3 | 97.8 | 84.6 | 90.9 | 58.5 | 97.9 | 95.0 | 83.3 | 96.2 |  | 65.9 |  | 11.3 |
| Median | 11210 | 27500 | 5667 | 1050 | 0.39 | 0.57 | 1.4 | 0.13 | 0.28 |  | 3.0 |  | 0.3 |
| Coefficient of variation | 3.2 | 2.1 | 1.3 | 2.1 | 1.6 | 0.78 | 1.7 | 1.1 | 1.7 |  | 3.9 |  | 2.7 |
| Commercial (527) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 253 | 201 |  |  | 300 | 445 | 469 | 343 | 466 |  | 235 |  |  |
| \% of samples above detection | 88.9 | 92.5 |  |  | 83.3 | 98.0 | 97.4 | 81.0 | 95.9 |  | 33.6 |  |  |
| Median | 4600 | 12000 |  |  | 0.50 | 0.6 | 1.5 | 0.11 | 0.22 |  | 2.3 |  |  |
| Coefficient of variation | 3.0 | 2.7 |  |  | 1.2 | 1.1 | 0.9 | 1.3 | 1.2 |  | 2.9 |  |  |
| Mixed Commercial (324) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 116 | 95 |  |  | 173 | 284 | 276 | 221 | 290 | 89 | 139 |  |  |
| \% of samples above detection | 94.8 | 98.9 |  |  | 67.1 | 96.8 | 96.0 | 93.7 | 98.6 | 11.9 | 45.5 |  |  |
| Median | 5400 | 11900 |  |  | 0.60 | 0.58 | 1.4 | 0.12 | 0.26 | 15.0 | 2.0 |  |  |
| Coefficient of variation | 3.0 | 2.6 |  |  | 1.0 | 0.7 | 0.9 | 2.1 | 1.5 | 1.0 | 1.0 |  |  |
| Industrial (566) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 315 | 189 |  |  | 272 | 461 | 483 | 344 | 478 | 152 | 255 |  | 197 |
| \% of samples above detection | 87.3 | 93.7 |  |  | 78.3 | 96.3 | 96.3 | 88.1 | 96.2 | 14.5 | 52.9 |  | 10.7 |
| Median | 2400 | 12000 |  |  | 0.42 | 0.69 | 1.4 | 0.10 | 0.25 | 3.7 | 4.0 |  | 0.38 |
| Coefficient of variation | 5.7 | 7.0 |  |  | 1.3 | 0.92 | 1.1 | 1.2 | 1.4 | 1.4 | 1.4 |  | 2.5 |

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 - Continued

|  | Fecal Coliform (mpn/100 mL) | $\qquad$ | $\begin{gathered} \text { Total } \\ \text { Coliform } \\ \text { (mpn/10 } \\ 0 \mathrm{~mL} \text { ) } \\ \hline \end{gathered}$ |  | NH 3 (mg/L) | $\begin{gathered} \text { N02+NO3 } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Nitrogen, Total Kjeldah (mg/L) | Phosphorus, filtered (mg/L) | $\begin{gathered} \text { Phospho } \\ \text {-rus, } \\ \text { total } \\ \text { (mg/L) } \\ \hline \end{gathered}$ | $\begin{gathered} \text { Sb, total } \\ (\mu \mathrm{g} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \text { As, total } \\ (\mu \mathrm{g} / \mathrm{L}) \\ \hline \end{gathered}$ | As, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | $\begin{gathered} \text { Be, total } \\ (\mu \mathrm{g} / \mathrm{L}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mixed Industrial (218) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 79 | 59 | 14 |  | 99 | 173 | 160 | 179 | 177 |  | 93 |  |  |
| \% of samples above detection | 98.7 | 96.9 | 71.4 |  | 30.3 | 98.8 | 92.5 | 84.4 | 95.5 |  | 88.2 |  |  |
| Median | 3033 | 11000 | 2467 |  | 0.58 | 0.59 | 1.1 | 0.08 | 0.20 |  | 3.5 |  |  |
| Coefficient of variation | 2.5 | 2.5 | 1.5 |  | 0.8 | 0.7 | 1.5 | 2.3 | 1.6 |  | 0.9 |  |  |
| Institutional (18) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations |  |  |  |  | 18 | 18 | 18 | 17 | 17 |  |  |  |  |
| \% of samples above detection |  |  |  |  | 88.9 | 100 | 100 | 82.4 | 94.1 |  |  |  |  |
| Median |  |  |  |  | 0.31 | 0.6 | 1.35 | 0.13 | 0.18 |  |  |  |  |
| Coefficient of variation |  |  |  |  | 0.5 | 0.6 | 0.5 | 0.5 | 1.0 |  |  |  |  |
| Freeways (185) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 49 | 25 | 16 | 13 | 79 | 25 | 125 | 22 | 128 |  | 61 | 72 |  |
| \% of samples above detection | 100 | 100 | 100 | 100 | 87.3 | 96.0 | 96.8 | 95.5 | 99.2 |  | 55.7 | 50.0 |  |
| Median | 1700 | 17000 | 50000 | 1900 | 1.07 | 0.28 | 2.0 | 0.20 | 0.25 |  | 2.4 | 1.4 |  |
| Coefficient of variation | 2.0 | 1.2 | 1.5 | 2.2 | 1.3 | 1.2 | 1.4 | 2.1 | 1.8 |  | 0.7 | 2.0 |  |
| Mixed Freeways (26) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 20 | 16 |  |  |  | 22 | 22 | 11 | 22 |  | 15 |  |  |
| \% of samples above detection | 85.0 | 93.8 |  |  |  | 100 | 100 | 100 | 100 |  | 80 |  |  |
| Median | 2600 | 19000 |  |  |  | 0.9 | 2.3 | 0.03 | 0.34 |  | 3.0 |  |  |
| Coefficient of variation | 2.3 | 1.1 |  |  |  | 0.7 | 1.3 | 0.9 | 0.7 |  | 0.7 |  |  |
| Open Space (68) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 23 | 22 |  |  | 32 | 44 | 45 | 44 | 46 |  | 19 |  |  |
| $\%$ of samples above detection | 91.3 | 90.9 |  |  | 18.8 | 84.1 | 71.1 | 79.6 | 84.8 |  | 31.6 |  |  |
| Median | 7200 | 24900 |  |  | 0.18 | 0.59 | 0.74 | 0.13 | 0.31 |  | 4.0 |  |  |
| Coefficient of variation | 1.1 | 1.0 |  |  | 1.24 | 0.9 | 0.9 | 0.9 | 3.5 |  | 0.4 |  |  |
| Mixed Open Space (168) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 86 | 75 |  |  | 71 | 152 | 123 | 148 | 152 |  | 88 |  |  |
| $\%$ of samples above detection | 97.7 | 100 |  |  | 22.5 | 97.4 | 90.2 | 85.8 | 96.1 |  | 44.3 |  |  |
| Median | 3000 | 21000 |  |  | 0.51 | 0.7 | 1.1 | 0.09 | 0.25 |  | 3.0 |  |  |
| Coefficient of variation | 2.3 | 2.4 |  |  | 1.2 | 0.8 | 0.9 | 1.1 | 1.1 |  | 0.9 |  |  |

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 - Continued

|  | Cd, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cd, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cr, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cr , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cu , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cu , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Pb , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Pb , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Hg , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Ni , total ( $\mu \mathrm{g} / \mathrm{I}$ ) | Ni, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Zn, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Zn, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Overall Summary (3765) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 2574 | 389 | 1598 | 261 | 2722 | 411 | 2949 | 446 | 1014 | 1430 | 246 | 3007 | 381 |
| \% of samples above detection | 40.6 | 30.3 | 70.2 | 60.5 | 87.4 | 83 | 77.7 | 49.8 | 10.2 | 59.8 | 64.2 | 96.6 | 96.3 |
| Median | 1.0 | 0.50 | 7.0 | 2.1 | 16 | 8.0 | 17.0 | 3.0 | 0.20 | 8.0 | 4.0 | 116 | 52 |
| Coefficient of variation | 3.7 | 1.1 | 1.5 | 0.7 | 2.2 | 1.6 | 1.8 | 2.0 | 2.5 | 1.2 | 1.5 | 3.3 | 3.9 |
| Residential (1042) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 695 |  | 404 |  | 771 | 90 | 762 | 108 | 275 | 392 | 25 | 784 | 87 |
| \% of samples above detection | 31.1 |  | 53.2 |  | 83.1 | 63.3 | 69.4 | 33.3 | 6.9 | 44.1 | 44.0 | 96.2 | 89.7 |
| Median | 0.5 |  | 4.5 |  | 12 | 7.0 | 12.0 | 3.0 | 0.20 | 5.6 | 2.0 | 73 | 31.5 |
| Coefficient of variation | 3.4 |  | 1.2 |  | 1.8 | 2.0 | 1.9 | 1.9 | 0.9 | 1.2 | 0.5 | 1.3 | 0.8 |
| Mixed Residential (611) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 420 | 30 | 193 | 21 | 432 | 29 | 500 | 30 | 115 | 150 | 25 | 515 | 28 |
| \% of samples above detection | 34.5 | 40.0 | 81.3 | 52.4 | 83.8 | 72.4 | 78.4 | 46.7 | 15.7 | 60 | 72.0 | 92.6 | 100 |
| Median | 0.9 | 0.30 | 7.0 | 2.0 | 16 | 5.5 | 16 | 3.0 | 0.20 | 7.8 | 5.5 | 95 | 48 |
| Coefficient of variation | 3.6 | 0.6 | 1.5 | 0.8 | 1.2 | 0.9 | 1.4 | 0.7 | 0.8 | 0.8 | 0.9 | 0.9 | 0.9 |
| Commercial (527) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 379 | 47 | 257 | 27 | 408 | 48 | 399 | 59 | 170 | 242 | 23 | 414 | 49 |
| \% of samples above detection | 41.7 | 23.4 | 60.7 | 40.7 | 92.9 | 79.2 | 85.5 | 52.5 | 6.5 | 60.3 | 47.8 | 99.0 | 100 |
| Median | 0.96 | 0.30 | 6.0 | 2.0 | 17 | 7.57 | 18.0 | 5.0 | 0.20 | 7.0 | 3.0 | 150 | 59 |
| Coefficient of variation | 2.7 | 1.3 | 1.3 | 0.6 | 1.5 | 0.8 | 1.6 | 1.6 | 0.8 | 1.2 | 0.8 | 1.2 | 1.4 |
| Mixed Commercial (324) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 188 | 41 | 128 | 27 | 191 | 41 | 244 | 41 |  | 102 | 26 | 243 | 39 |
| \% of samples above detection | 49.5 | 34.1 | 88.3 | 66.7 | 93.2 | 80.5 | 88.1 | 63.4 |  | 78.4 | 69.2 | 98.8 | 100 |
| Median | 0.9 | 0.35 | 5.0 | 2.5 | 17.5 | 10 | 17.0 | 3.5 |  | 5.1 | 3.5 | 131.4 | 73 |
| Coefficient of variation | 1.1 | 0.8 | 1.1 | 0.7 | 3.0 | 0.6 | 1.4 | 0.8 |  | 1.3 | 0.6 | 1.7 | 0.8 |
| Industrial (566) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 435 | 42 | 250 | 36 | 455 | 42 | 452 | 51 | 199 | 237 | 36 | 473 | 42 |
| \% of samples above detection | 49.0 | 54.8 | 72.0 | 55.6 | 88.6 | 90.5 | 75.0 | 52.9 | 13.9 | 61.6 | 58.3 | 98.9 | 95.2 |
| Median | 2.0 | 0.60 | 12.0 | 3.0 | 20.8 | 8.0 | 24.9 | 5.0 | 0.20 | 14.0 | 5.0 | 199 | 112 |
| Coefficient of variation | 2.2 | 1.1 | 1.2 | 0.7 | 2.0 | 0.7 | 1.9 | 1.6 | 2.7 | 1.0 | 1.4 | 1.5 | 3.6 |

Table 3. Summary of Available Stormwater Data Included in NSQD, version 1.1 - Continued

|  | Cd, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cd, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cr, total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cr , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cu , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Cu , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Pb , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Pb , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Hg , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Ni , total ( $\mu \mathrm{g} / \mathrm{l}$ ) | Ni , filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) | Zn , total ( $\mu \mathrm{g} / \mathrm{L}$ ) | Zn, filtered ( $\mu \mathrm{g} / \mathrm{L}$ ) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mixed Industrial (218) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 145 | 25 | 109 | 15 | 150 | 24 | 213 | 25 | 58 | 74 | 15 | 212 | 24 |
| \% of samples above detection | 60.7 | 92.0 | 92.7 | 66.7 | 90.0 | 100.0 | 82.6 | 92.0 | 22.4 | 83.8 | 100.0 | 98.6 | 95.8 |
| Median | 1.6 | 0.60 | 8.0 | 2.0 | 23 | 6.0 | 20.0 | 5.0 | 0.3 | 12 | 5.0 | 172 | 2100 |
| Coefficient of variation | 1.9 | 0.6 | 1.7 | 0.7 | 0.8 | 0.6 | 1.4 | 1.0 | 0.6 | 0.8 | 0.6 | 3.1 | 1.2 |
| Institutional (18) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations |  |  |  |  |  |  | 18 |  |  |  |  | 18 |  |
| \% of samples above detection |  |  |  |  |  |  | 77.8 |  |  |  |  | 100 |  |
| Median |  |  |  |  |  |  | 5.75 |  |  |  |  | 305 |  |
| Coefficient of variation |  |  |  |  |  |  | 0.8 |  |  |  |  | 0.8 |  |
| Freeways (185) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 95 | 114 | 76 | 101 | 97 | 130 | 107 | 126 |  | 99 | 95 | 93 | 105 |
| \% of samples above detection | 71.6 | 26.3 | 98.7 | 78.2 | 99.0 | 99.2 | 100 | 50.0 |  | 89.9 | 67.4 | 96.8 | 99.1 |
| Median | 1.0 | 0.68 | 8.3 | 2.3 | 34.7 | 10.9 | 25 | 1.8 |  | 9.0 | 4.0 | 200 | 51 |
| Coefficient of variation | 0.9 | 1.0 | 0.7 | 0.7 | 1.0 | 1.5 | 1.5 | 1.7 |  | 0.9 | 1.4 | 1.0 | 1.9 |
| Mixed Freeways (26) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 23 |  | 15 |  | 23 |  | 23 |  |  |  |  | 23 |  |
| \% of samples above detection | 56.5 |  | 100 |  | 100 |  | 56.5 |  |  |  |  | 100 |  |
| Median | 0.5 |  | 6.0 |  | 14 |  | 10.0 |  |  |  |  | 130 |  |
| Coefficient of variation | 2.2 |  | 1.0 |  | 1.0 |  | 1.3 |  |  |  |  | 0.9 |  |
| Open Space (68) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 38 |  | 36 |  | 39 |  | 45 |  |  |  |  | 45 |  |
| \% of samples above detection | 55.3 |  | 36.1 |  | 74.4 |  | 42.2 |  |  |  |  | 71.1 |  |
| Median | 0.38 |  | 5.4 |  | 10 |  | 10.0 |  |  |  |  | 40 |  |
| Coefficient of variation | 1.9 |  | 1.7 |  | 2.0 |  | 1.7 |  |  |  |  | 1.3 |  |
| Mixed Open Space (168) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Number of observations | 107 |  | 88 |  | 108 |  | 155 |  | 27 | 51 |  | 156 |  |
| \% of samples above detection | 18.7 |  | 81.8 |  | 89.8 |  | 74.2 |  | 14.8 | 72.5 |  | 98.1 |  |
| Median | 2.0 |  | 6.0 |  | 9.0 |  | 10 |  | 0.15 | 8.0 |  | 80 |  |
| Coefficient of variation | 1.4 |  | 1.3 |  | 1.0 |  | 2.3 |  | 0.4 | 1.1 |  | 1.1 |  |

In the General Information tab, the spreadsheet lists the states and municipalities included in the current version of the database. The second tab describes the two main sections of the database: site descriptions and event descriptions. In the items description section, each column in the database is described. The last column in this table shows an example of the value expected in each column. The third tab describes the constituents and parameters included in the database, the number of observations, and the percentage of samples having detected observations. This table is useful to identify those constituents with high percentages of detected values.

The last tab in the database contains the data itself; a matrix of 232 columns by 3,765 rows containing all the data collected and reviewed. Each row represents a storm event for each monitoring location. This part of the table is divided in seven subsections describing the site location, the hydrology of the event and equipment used, and the constituent classifications. Each section of the database is described in the following discussion, with detailed analyses presented in Chapters 4 through 8 of this report.

The following discussion will require a copy of the database for reference. This is available at: http://unix.eng.ua.edu/ rpitt/Research/ms4/mainms4.shtml. Each of the sections and columns included in the spreadsheet will be explained in detail. Summary statistics, probability plots and box and whiskers plots will be used to describe the most important parameters.

## Site Description [Columns A through Y]

Column A is an identifier of each storm event stored in the database. It is the table key. Column B describes the site main land use or activity: residential (RE), commercial (CO), industrial (ID), institutional (IS), open space (OS), and freeways (FW). In the case when more than one land use is present, a combination code is used beginning with the land use with the most area in the watershed. For example, if a site was $70 \%$ residential and $30 \%$ commercial, the site was coded as RE_CO. The percentage of each land use is indicated in the columns J through O .

Column C describes the month of the year when the sample was collected as follows: winter (WI) if the sample was collected in November, December or January; spring (SP) if the sample was collected in February, March, or April; summer (SU) if the sample was collected in May, June or July; and fall (FA) if the sample was collected in August, September, or October. A reasonably uniform number of samples were collected during each of the four periods: about $29 \%$ of the samples were collected in the winter, $30 \%$ in the spring, $19 \%$ in the summer, and $23 \%$ in the fall.

Columns D through F indicate the location of the site. LOCATION_ID is the key for sorting the sites, and is a code of eight characters: the first two letters indicate the name of the state, the next four letters is a code for the community, and the last two letters represent the site name. Columns E and F are the name of the community and the name of the site. Column $G$ is the contact information of the person in that community that supplied the database information. Columns H through M are the percentages of the separate land uses in the drainage area, as described in column B.

Column N indicates the total watershed drainage area in acres. Figure 2 shows the distribution of the area by land use. The distribution of the watersheds areas can be considered approximately lognormal. Commercial, industrial, open space, and residential land uses have approximately the same distribution of drainage areas for the monitored outfalls, with a range between ten and one thousand acres. The median monitored watershed area for commercial and industrial sites was about 43 acres, while the median watershed area in residential and open space areas was about 65 acres. Freeways had smaller areas than the other land uses, with median areas being about 2 acres, with a range varying between one and one hundred acres.

Columns O and P list the approximate latitude and longitude of the outfall location in degrees, minutes, and seconds. Most of these coordinates were obtained using the Teraserver website. Column S indicates the EPA Rain Zone location of each site (Figure 1 and Table 1). About $52 \%$ of the sites are located in the EPA Rain Zone 2, which contains the Chesapeake Bay region, the main targeted area for this database. Each of the Rain Zones 3 through 7 has about $8 \%$ of the total sites. Rain Zones 1 and 9 have each about $3 \%$ of the sites. Rain Zone 8 has only one community with four locations, or about $1 \%$ of the total number of sites.


Figure 2. Drainage areas by land use

Column R indicates the total percentage of impervious surfaces reported for each site. Only Newport News, Virginia, contained information describing how the impervious areas were hydraulically connected to the drainage systems. It is expected that a watershed with high levels of impervious (a parking lot for example), is mostly directly connected due to little opportunity for draining to pervious areas. Less water is therefore infiltrated and the stormwater rapidly moves to the connected outfall. About 169 sites (about $47 \%$ of the total number of sites) included percentage of impervious surfaces in their annual reports or permit applications. Of this response, about 69 sites were for single or mixed residential areas, 34 sites were single or mixed industrial areas, 34 sites were single or mixed commercial area, 17 sites were single or mixed freeway areas, and 15 sites were single or mixed open space areas.

Figure 3 shows a box and whiskers plot of the reported impervious surface values for the predominant land uses. As expected, the open space sites have the lowest percentage of impervious surfaces (mean about $3.3 \%$ ), while the mean impervious surface value for the freeway sites is $92 \%$. Industrial and commercial area impervious surface values are higher, with means of $67 \%$ and $81 \%$ respectively. Residential areas cover almost the complete range, from about 7 to $89 \%$. The impervious surfaces for residential areas are intermediate between the values for open space and the industrial/commercial values, as expected. The mean percentage of impervious areas in residential areas is approximately $41 \%$.


Figure 3. Percentage of impervious surfaces by land use

Column S is a qualifier for the total percentage of impervious surface area in the test watershed, indicating if there was an apparent increase in the percentage of imperviousness during the monitoring period, based on examinations of aerial photographs. Only one site (Pylon Street in Forth Worth, TX) had an apparent increase in the percentage of impervious area during the monitoring period. Column $T$ indicates the volumetric runoff coefficient (Rv), or the ratio between the total runoff depth divided by the precipitation depth for each event. Figure 4 is a scatter plot of the reported percentage of impervious areas and reported Rv. As expected, higher volumetric runoff coefficients are reported for heavily paved areas, such as parking lots or freeways, compared to areas having much more landscaped areas, such as residential areas or parks. However, it is possible that some of the reported Rv values are simply calculated from the percent impervious cover values, and not from monitored rainfall and monitored runoff values.


Figure 4. Scatter plot of percentage of impervious and Rv

None of the monitoring agencies reported the TR-55 curve number for the sites. This value is used to estimate the runoff volume using the Soil Conservation Service, SCS (now Natural Resources Conservation Service, NRCS) TR55 method. Curve numbers (column U) were therefore not examined during this analysis. Only eight sites indicated the year when the land was developed, and these are shown in column V . Because of the low number of observations, this factor also could not be used in the data analyses.

Column W indicates the type of stormwater conveyance reported for the monitored area. This parameter indicates if the site is drained with "curb and gutter" systems typical of areas with high percentages of imperviousness, or if the water is transported beside the road through a grass-lined drainage channel (swales), more common in lower density areas. About $26 \%$ of the sites did not report the type of conveyance or it was not possible to identify them using the aerial photographs. Curb and gutter systems were reported for $65 \%$ of the sites, while grass swales were reported for $9 \%$ of the sites. Grass swales are usually considered a stormwater control, or "BMP," due to their ability to infiltrate large fractions of the runoff before discharge. They may provide some limited concentration reductions of particulate pollutants, but only for the shallowest flows. Detailed analyses are presented in Chapter 5 of this report.

The next column indicates if the site has wet detention ponds. About seven sites (out of the 360 total sites) have a wet pond at the outfall, nine sites have ponds in the watershed, and three sites have ponds in series, all upstream of the monitoring location. Other reported stormwater controls included: dry detention ponds (4 sites), small underground detention storage tanks ( 2 sites), besides the 32 sites having grass swales as noted above.

The final column in this section $(\mathrm{Y})$ includes important comments that were not assigned to any of the other columns. Typical information in this column is the size of the pipe; if the outlet is a circular (pipe), or a square (box culvert); the number of pipes discharging from the watershed; or if there is a USGS monitoring station at the outfall that reported the data in the NSQD.

## Hydrologic Information [Columns Z through AN]

Column Z is the identifier of each storm event stored in the database. It is used as a table sorting "key." Generally, it contains information about the location and the sampling date. Column AA indicates the precipitation depth recorded during the event, in inches. About 3,300 events included this parameter. Precipitation depth, flow volume
and similar hydrologic parameters were included in the annual reports or permit applications usually as appendices. During the data collection process, some of these appendices were not copied or located. The highest percentage of events with precipitation by land use was observed in single and mixed freeways (about 99\%). The lowest percentage of events with precipitation data was observed in single and mixed residential areas, with $85 \%$ of the sites reporting this information. The percentage for the other land uses were: $87 \%$ for single and mixed industrial, $90 \%$ in single and mixed commercial, $96 \%$ in single and mixed open space.

Figure 5 shows the distribution of the available precipitation depth data by land use. The range of precipitation depth varies between 0.01 and 6 inches, indicating that some of the reported events were outside of the range specified by the general monitoring guidance (minimum of 0.1 inches and close to annual average characteristics). The distribution of the rainfall depth data is approximately lognormal, with a median between 0.4 to 0.6 inches. All the land uses have a similar pattern, with approximately the same variance. The mixed freeway category seems to have a narrower range, but they only represent $0.5 \%$ of the total events that have precipitation data. Column AB is a qualifier for the precipitation depth data. Some communities collected the data on site, while others used rain gauge data collected from a local airport. Rain gauges located on site are preferred as they are expected to better represent the rainfall conditions that occurred on the monitored site for the monitored event. Twelve percent of the total database events did not include precipitation depth data, $42 \%$ of the events were associated with rain data collected on site, $23 \%$ of the events did not indicate how the reported rain data was obtained, $7 \%$ of the events are associated with rain data from the local airport rain gauge, and the remaining $16 \%$ used other methods to determine the event rainfall data, such as regional rain gauges associated with flood monitoring systems.

Columns AC through AF indicate the starting and ending date and time of the event. Column AG indicates the maximum reported 15 -minute rain intensity for each event. Events having high rain intensities have high kinetic energies, and it is hypothesized that these events will have increased washoff or erosion of particulate pollutants from watershed surfaces. However, only $1 \%$ of the database events reported this parameter. Column AG information was therefore not included in any of the data analyses.

Runoff depth (column AH) is the total volume of stormwater that leaves the monitored watershed during the rain event. For a directly connected paved parking lot, the runoff depth (expressed in inches of runoff for the complete drainage area) is only slightly smaller than the precipitation depth. In contrast, a park having mostly pervious surfaces would record total runoff volumes much smaller than the rain depth because most of the rainwater is infiltrated before it drains from the site. About $36 \%$ of the events included runoff data.

Figure 5 also shows the probability plots of runoff depth for each land use. As expected, smaller runoff values were observed in open space and residential areas, while freeways, mixed commercial, and mixed industrial land uses have runoff distributions similar to the rain distributions observed in the precipitation panel. A different pattern was observed for runoff at freeways, which are characterized by their small area and high percentage of impervious cover.

Probability Plot of Precipitation and Runoff Depth

Lognormal



Figure 5. Precipitation and runoff depth by land use

Column AJ indicates if the runoff and precipitation were measured during the complete event or only the first three hours of the storm. The basic NPDES stormwater monitoring guidelines indicates that samples must be collected at least during the first three hours of the event. If the runoff and precipitation were not monitored for the complete event, then site hydrology confusion would occur. Most of the communities recorded the runoff for the complete event, even if monitoring only occurred for three hours. Only Greensboro, Topeka, Chesterfield County, and Fayetteville recorded runoff only for the first three hours of the events.

Column AK indicates if the events were from composite sampling, as required by the Federal Regulations guidance. First flush events were included in the first version of the database, version 1.0. After the paired first flush statistical analyses (see Chapter 6), these first-flush data were removed from the main database to eliminate confusion, leaving only the composite samples in the main database.

Column AL indicates if the composite sample was collected using automatic equipment, or if manual sampling was used. This column can be used to evaluate possible differences in the recorded concentrations due to the sampling method. About $81 \%$ of the events were collected using automatic samplers, $10.5 \%$ used manual sampling, and about $8.5 \%$ of the events did not have any reported sampling method. Detailed analyses concerning the effects of manual versus automatic sampling is discussed in Chapter 5 of this report.

Column AM describes if the collected sample was a flow-weighted or time-weighted composite sample. A flowweighted composite sample is comprised of several equal volume subsamples that were collected according to the flow rate of the runoff water. The sampler is programmed to collect a subsample for a specified constant flow increment. The total volume in the single composite bottle is therefore proportionate to the total runoff volume associated with the monitored event. A time-weighted composite sample is made up of several equal volume subsamples that were collected at constant periods of time and collected into a single large composite sample bottle. At the end of the event, the total volume of sample in the composite sample bottle is proportionate to the duration of the event. About $73 \%$ of the events in the database were collected using flow-weighted composite sampling methods, while only $8 \%$ of the events were collected using time-weighted composite sampling methods. No composite sampling method information was available for the remaining $19 \%$ of the events.

The last column in this database section describes the number of days without rain prior to the event sampling. It is usually hypothesized that an increase in the number of dry days prior to an event would cause an increase in the constituent concentration. About $38 \%$ of the events had this information available. Detailed analyses are presented in Chapter 5 of this report.

## Conventional Constituents [Columns AO through BS]

This section of the database contains measurement values for conventional stormwater constituents (conductivity, DO, hardness, oil and grease, pH , temperature, TDS, TSS, $\mathrm{BOD}_{5}, \mathrm{COD}$, fecal coliforms, and fecal streptococcus).

Table 3, presented earlier, contains a summary showing the total number of samples included in the database classified by land use, the percentages of samples detected, the medians, and the coefficients of variation. In general, the lowest concentrations were usually found at open space land uses, followed by residential areas. The highest concentrations were observed at freeway land use sites. Table 4 is a summary contrasting the land uses having the lowest and the highest concentrations of these constituents.

The Mann-Whitney test was used to determine if there is a significant difference between the land uses having the lowest and highest concentrations. As a complement, one-way ANOVA analyses were used to identify if a significant difference existed among any of the land uses. As the number of samples increase, the power of the test also increases. P-values close to zero will indicate that the concentration of at least one land use is statistically different than the other land uses (true for all constituents in Table 4, except for Dissolved Oxygen).

Table 4. Conventional Constituents Summary

| Constituent | Land use having the lowest median concentration |  |  | Land use having the highest median concentration |  |  | MannWhitney Test | 1-Way ANOVA by Land Use |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | Land Use | Median | n | Land Use | Median | p-value | p-value |
| Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ) | 106 | RE | 96.5 | 108 | ID | 135.5 | 0 | 0 |
| Dissolved Oxygen (mg/L) | 39 | ID | 7.3 | 30 | RE | 7.8 | 0.064 | 0.325 |
| Hardness (mg/L CaCO3) | 350 | RE | 32 | 139 | CO | 38.9 | 0.009 | 0 |
| Oil and Grease Total (mg/L) | 308 | RE | 3.85 | 43 | FW | 8.0 | 0 | 0.001 |
| pH (s.u.) | 111 | FW | 7.1 | 234 | ID | 7.5 | 0 | 0 |
| Temperature ( ${ }^{\circ} \mathrm{C}$ ) | 31 | FW | 14 | 140 | ID | 17.8 | 0 | 0 |
| Total Dissolved Solids (mg/L) | 854 | RE | 72 | 411 | ID | 92 | 0 | 0 |
| Total Suspended Solids (mg/L) | 977 | RE | 49 | 133 | FW | 99 | 0 | 0 |
| Biological Oxygen Demand (mg/L) | 38 | OP | 5.4 | 421 | CO | 11 | 0 | 0 |
| Chemical Oxygen Demand (mg/L) | 33 | OP | 42.1 | 66 | FW | 100 | 0 | 0 |
| Fecal Coliform (colonies/100 mL) | 261 | ID | 2500 | 21 | OP | 7200 | 0.014 | 0 |
| $\begin{array}{l}\text { Fecal Streptococcus (colonies/100 } \\ \mathrm{mL} \text { ) }\end{array}$ | 166 | CO | 10285 | 273 | RE | 24600 | 0 | 0.003 |

Figure 6 contains examples of grouped box and whiskers plots for several constituents for different major land use categories. The freeways sites had the highest reported TSS, COD and oil and grease concentrations. Statistical ANOVA analyses for all land use categories found significant differences for land use categories for all constituents except for dissolved oxygen. Turbidity, total solids, total coliform and total E-coli have not enough samples in each group to evaluate if there is a difference among all land uses. Chapter 5 presents more comprehensive analyses for specific site conditions (considering interactions of land use, geographical location, etc.).


Figure 6. Box and whiskers plots for conventional constituents by single land use

Stormwater temperature depends of many factors, including season, the time of the day, and the types of surfaces in a land use. Column C shows the season of the year when each sample was obtained, the most obvious factor affecting runoff temperature.

Figure 7 shows the water temperatures for each month for the samples collected in the EPA Rain Zones 5 and 6 combined. Similar patterns were observed in the other EPA Rain Zones. Two main periods can be identified in this plot: from February to July the water temperature rises and from August to January the water temperature decreases. Table 4 shows that for almost all conventional constituents, residential and open space land uses have the lowest concentrations, except for pathogen indicators. Industrial and freeway land uses generally have the highest concentrations.


Figure 7. Water temperature in EPA Rain Zones 5 and 6 (line links median values for each month)

## Nutrients [Columns BU through CG]

This section in the database contains the compounds associated with nitrogen and phosphorus compounds. Table 5 shows a summary of the land uses having the lowest and highest concentrations for each constituent. Again, the Mann Whitney and ANOVA tests were used to evaluate if there was a significant difference between land uses for these constituents.

In contrast to the conventional constituents, dissolved and total phosphorus have the highest concentrations in residential land uses. There was no significant difference noted for total nitrogen for the different land uses. The median ammonia concentration in freeway stormwater is almost three times the median concentration observed in residential and open space land uses, while freeways have the lowest orthophosphate and nitrite-nitrate concentrations; almost half of the concentration levels that were observed in industrial land uses. Figure 8 shows box plots for TKN, total phosphorus, and nitrite-nitrate for several land uses. It shows that even if there are differences in the median concentrations by a factor of two or three between the land uses, the extreme range of the concentrations within a single land uses can still vary by two or three orders of magnitude. Again, Chapter 5 examines many factors affecting these concentrations, in addition to land use.

Table 5. Nutrients Summary

| Constituent | Land use having the smallest median concentration |  |  | Land use having the largest median concentration |  |  | MannWhitney Test | 1-Way ANOVA by Land Use |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | Land Use | Median | n | Land Use | Median | $p$-value | p-value |
| Ammonia (mg/L) | 485 | RE | 0.31 | 69 | FW | 1.07 | 0 | 0 |
| Nitrogen Nitrite-Nitrate ( $\left.\mathrm{NO}_{2}+\mathrm{NO}_{3}\right)(\mathrm{mg} / \mathrm{L})$ | 24 | FW | 0.28 | 429 | ID | 0.71 | 0 | 0.001 |
| Nitrogen Total (mg/L) | 63 | ID | 2.03 | 81 | RE | 2.30 | 0.25 | 0.698 |
| Nitrogen Kjeldahl Total (TKN) (mg/L) | 32 | OP | 0.74 | 121 | FW | 2.00 | 0 | 0 |
| Phosphate Ortho (mg/L) | 103 | FW | 0.09 | 66 | ID | 0.23 | 0 | 0 |
| Phosphorous Dissolved (mg/L) | 283 | ID | 0.11 | 621 | RE | 0.17 | 0 | 0 |
| Phosphorous Total (mg/L) | 427 | CO | 0.22 | 933 | RE | 0.30 | 0 | 0 |



Figure 8. Box and whiskers plots for nutrients by single land use

## Metals [Columns CK through EK]

This section in the database contains the metal concentrations. Industrial land uses have higher median concentrations of heavy metals than any of the other land uses, followed by freeways. Table 6 shows the ANOVA results for metals. As expected, open space and residential land uses have the lowest median concentrations. In almost all cases, the median metal concentrations at the industrial areas were about three times the median concentrations observed in open space and residential areas. Arsenic, cadmium, chromium, copper, lead, nickel, and zinc showed significant differences between the extreme land uses at the $1 \%$ level of confidence, or less. Other constituents are also included in the database (antimony, beryllium, cyanide, mercury, selenium, silver, and thallium), along with dissolved forms of the metals. Too few observations and large fractions of undetected observations hindered statistical analyses of these other metals.

Table 6. Summary of Metals Concentration

| Constituent |  |  |  |  |  |  |  | Land use having the <br> smallest concentration |  |  | Land use having the <br> largest concentration |  | Mann- <br> Whitney <br> Test | 1-Way <br> ANOVA <br> by Land <br> Use |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | Land Use | Median | p-value | Land Use | Median | p-value | p-value |  |  |  |  |  |  |
| Arsenic Total $(\mu \mathrm{g} / \mathrm{L})$ | 70 | CO | 2.4 | 145 | ID | 4.0 | 0 | 0 |  |  |  |  |  |  |
| Cadmium Total $(\mu \mathrm{g} / \mathrm{L})$ | 219 | RE | 0.5 | 223 | ID | 1.9 | 0 | 0 |  |  |  |  |  |  |
| Chromium Total $(\mu \mathrm{g} / \mathrm{L})$ | 241 | RE | 4.6 | 186 | ID | 14.0 | 0 | 0 |  |  |  |  |  |  |
| Copper Total $(\mu \mathrm{g} / \mathrm{L})$ | 29 | OP | 10 | 96 | FW | 34.7 | 0 | 0 |  |  |  |  |  |  |
| Lead Total $(\mu \mathrm{g} / \mathrm{L})$ | 19 | OP | 10 | 343 | ID | 26 | 0 | 0 |  |  |  |  |  |  |
| Nickel Total $(\mu \mathrm{g} / \mathrm{L})$ | 190 | RE | 5.4 | 156 | ID | 16 | 0 | 0 |  |  |  |  |  |  |
| Zinc Total $(\mu \mathrm{g} / \mathrm{L})$ | 32 | OP | 40 | 455 | ID | 200 | 0 | 0 |  |  |  |  |  |  |

Figure 9 contains examples of grouped box and whiskers plots for lead, copper, and zinc constituents for different major land use categories. The highest lead and zinc concentrations were found in industrial land uses, while the highest copper concentrations were observed at freeways sites.


Figure 9. Box and whiskers plots for metals by single land use

## Additional Constituents [Columns EM through HW]

These columns contain information for additional constituents that were sampled only during the permit application period (first year of sampling). Some constituents having more than a $30 \%$ detection level included:
methylenechloride, total petroleum hydrocarbon (TPH), total organic carbon, chloride, nitrate nitrogen, nitrite nitrogen, total organic nitrogen, and iron.

Table 7 shows summaries for these additional constituents that have enough samples to identify significant differences between land uses. Only total petroleum hydrocarbon (TPH) and nitrite nitrogen showed significant differences (at the $5 \%$ significance level) between the land uses having lowest and highest median concentrations. The median stormwater TPH concentration in residential areas is almost half the median TPH stormwater concentration at freeway sites.

Table 7. Summary of Additional Constituents

| Constituent | Land use having the <br> smallest concentration |  | Land use having the <br> largest concentration |  | Mann-Whitney <br> Test |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{n}$ | Land <br> Use | Median | $\mathbf{n}$ | Land <br> Use | Median | p- <br> value | Significant $\boldsymbol{\alpha}$ <br> $\mathbf{s} \%$ |
| Total Petroleum Hydrocarbon (mg/L) | 36 | RE | 0.38 | 20 | FW | 0.78 | 0 | Yes |
| Chloride (mg/L) <br> (FW and OP not included) | 42 | ID | 7.1 | 38 | CO | 9.5 | 0.25 | No |
| Nitrogen Nitrate (mg/L) <br> (CO and OP not included) | 13 | RE | 0.69 | 98 | FW | 0.84 | 0.58 | No |
| Nitrogen Nitrite (mg/L) <br> (CO and OP not included) | 7 | ID | 0.07 | 42 | FW | 0.17 | 0.01 | Yes |
| Nitrogen Total Organic (mg/L) <br> (FW not included) | 12 | RE | 0.96 | 5 | CO | 1.97 | 0.19 | No |
| Iron (mg/L) | 6 | RE | 2.99 | 27 | FW | 3.60 | 0.27 | No |

## Site Descriptions and Additional Supporting Information

Supplemental reports were created containing additional information for each community. These site descriptions include (depending on available information) the land use and impervious surfaces for the monitored site, aerial photographs and a topographic map of the area, and descriptions of the sampling procedures and quality control (QA/QC) used during sample collection and analysis. The QA/QC description indicates if blank samples were used during the analysis to check the equipment, the protocols used during the sample collection, and in some cases, the chain of custody of the samples. These supplemental reports also contain descriptions of the sampled parameters, analytical methods, and field instrumentation used by the community.

About $38 \%$ of the aerial photographs have better than 1-meter resolution and the remaining photos have 1-meter resolution. The locations of most of the outfalls were included in the database in the Q and R columns (Latitude and Longitude). Table 8 shows the total number of sites with high-resolution aerial photos and with watershed delineations.

Table 8. Additional Site Information

| EPA Rain-Zone | Number of <br> Communities | Number <br> of Sites | Sites with high- <br> resolution aerial <br> photos (resolution <br> $\mathbf{0 . 2 5 ~ m}$ ) | Sites with <br> watershed <br> delineations |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 2 | 12 | 0 | 2 |
| 2 | 28 | 185 | 38 | 18 |
| 3 | 8 | 30 | 15 | 20 |
| 4 | 4 | 21 | 15 | 17 |
| 5 | 9 | 33 | 18 | 0 |
| 6 | 4 | 30 | 20 | 9 |
| 7 | 6 | 28 | 19 | 0 |
| 8 | 1 | 4 | 0 | 0 |
| 9 | 3 | 17 | 13 | 8 |

Watershed delineations are an important component of the site descriptions by identifying the extent of the contributing area, the different land uses located in the watershed and the sampling location. Only $20 \%$ of the sites included their watershed delineations.

Most communities followed the sampling recommendations presented in the Code of Federal Register (40 CFR 122.21), although delegated NPDES state agencies were able to modify the specific requirements to better address local concerns. Almost all communities collected samples at least during the first 3 hours of the event (or the complete event if the duration was shorter). For about $66 \%$ of the events, the communities calculated the total runoff for the duration of the total event discharge, but used the concentrations from the shorter monitoring period. Chapter 6 includes a detailed analysis of first-flush concentrations that may indicate the maximum errors that may occur with truncated sampling periods. Seven percent of the events included runoff for only the first three hours of the event. The remaining $25 \%$ of the events did not include runoff volume data, or it was not clear if the runoff volume data was obtained during the first three hours, or for the whole event.

Another important monitoring aspect described in the site descriptions is how the composite sample was created. There are two compositing options: flow-weighted and time-weighted. During the time-weighted compositing scheme, subsamples of equal volume were obtained at specific time intervals during the three hour sampling period. All the subsamples were collected in a single bottle, creating the composite sample. In the flow-weighted compositing case, the subsamples were collected for a set flow increment. About $71 \%$ of the events were collected using flow-weighted sampling, $5 \%$ of the events were collected using time-weighted sampling, and it was not clear how the remaining $24 \%$ of the samples were collected. Roa-Espinosa and Bannerman (1995) found that timeweighted composite sampling could be representative of the sampling period, if many subsamples are collected throughout the storm period. Time-weighted compositing is much simpler and less expensive than flow-weighted composite sampling, but may have a slight error in the measured concentrations, compared with the flow-weighted method.

About $62 \%$ of the 65 communities represented in the NSQD indicated that they used automatic samplers during their monitoring activities, about $34 \%$ did not indicate how they collected their samples, and $4 \%$ collected their samples manually. ISCO samplers were the most commonly used automatic sampler, with about $24 \%$ of the sites using ISCO 2700,3700 or 6700 samplers. American Sigma samplers were used at about $12 \%$ of the 65 communities. The most common American Sigma sampler models included 800SL, 900 AV and 900 MAX. About $69 \%$ of the communities did not indicate how, or if, they measured flow, and did not report any flow data. About $20 \%$ of the sites used ISCO 3230 or 4230 flow meters. The remaining $11 \%$ used other methods to estimate the stormwater discharge volumes.

## Problems Encountered during NPDES Stormwater Monitoring

About 58\% of the communities also described problems found during the monitoring process and these are summarized in the site summary reports. Some communities reported more than one problem. One of the basic sampling requirements was to collect three samples every year for each of the land use stations. These samples were to be collected at least one month apart during rains having at least 0.1 inch rains, and with at least 72 hours from the previous 0.1 -inch storm event. It was also required (when feasible), that the variance in the duration of the event and the total rainfall not exceeded the median rainfall for the area. About $47 \%$ of the communities reported problems meeting these requirements. In many areas of the country, it was difficult to have three storm events per year having these characteristics. The second most frequent problem, reported by $26 \%$ of the communities, concerned backwater tidal influences during sampling, or the outfall became submerged during the event. In other cases, it was observed that there was flow under the pipe (flowing outside of the pipe, in the backfill material, likely groundwater), or sometimes there was not flow at all. About $12 \%$ of the communities described errors related to malfunctions of the sampling equipment. Most of the communities with equipment failures did not report the reasons of the failure. When reported, the equipment failures were due to incompatibility between the software and the equipment, clogging of the rain gauges, and obstruction in the sampling or bubbler lines. Memory losses in the equipment recording data were also periodically reported. Other reported problems were associated with lighting, false starts of the automatic sampler before the runoff started, and operator error due to misinterpretation of the equipment configuration manual.

Sites located on the East coast (Hampton, VA for example) where the hurricane season produces frequent large storms, especially having a high water table, were especially susceptible. Base flows can commonly occur in separate storm drainage systems for a variety of reasons and they may be more important during some seasons than during others. In many cases, they cannot be avoided and should be included in the monitoring program, and their effects need to be recognized as an important flow phase. As an example, Pitt and McLean (1986) found dry weather base flows to be significant sources of many pollutants, even during a comprehensive research project that spent much time surveying the test watersheds to ensure they did not have any inappropriate discharges entering the storm drainage system.

Capturing runoff events within the acceptable range of rain depth was difficult for some monitoring agencies. Rain depth cannot be precisely predicted in many areas of the country. Also, if using rain gauge data from a location distant from the monitoring location, the reported rain depth may not have been representative of the depth that occurred at the site. The rain gauges need to be placed close to the monitored watersheds. This was likely one of the reasons why the runoff depths periodically exceeded the reported rain depths. Rain in urban areas can vary greatly over small distances. The ASCE/EPA (2002) recommended that rainfall gauges be located as close as possible to the monitoring station. In the NSQD, about $7 \%$ of the events had site precipitation estimated using rain gauge located at the city airport. About $16 \%$ of the events had precipitation depth estimated using their own monitoring network (Hampton Road Sanitation District, for example). Some communities had precipitation networks that were used for flood control purposes for the surrounding area. These networks can be considered better than the single airport rain gauge, but should at least be supplemented with a rain gauge located in the monitored watershed. Another factor that needs to be considered is the size of the watershed. Large watersheds cannot be represented with a single rain gauge at the monitoring station; in those cases the monitoring networks will be a better approach. Large watersheds are more difficult to represent with a single rain depth value.

Many of the monitoring stations lacked flow monitoring instrumentation, or did not properly evaluate the flow data. Accurate flow monitoring can be difficult, but it greatly adds to the value of the expensive water quality data. As noted previously, base flows also need to be properly removed from the event measurements so only direct runoff quantities are reported. It is probably unreasonable to expect to have a permanent flow monitoring station installed at a location where only manual grab samples are being obtained. However, manual flow monitoring can be conducted during manual sampling by carefully noting the flow stage in previously surveyed locations. These observations will need to be obtained during the complete duration of the event.

The three hour monitoring period that most used may have resulted in some bias in the reported water quality data. This limit was likely used to minimize the length of time personnel needed to be at a monitoring location during manual sampling activities. Also, it is unlikely that manual samplers were able to initiate sampling near the beginning of the events, unless they were deployed in anticipation of an event later in the day. A more cost-effective and reliable option would be to have semi-permanent monitoring stations located at the monitoring locations and sampling equipment installed in anticipation of a monitored event. Most monitoring agencies operated three to five land use stations at one time. This number of samplers, and flow equipment, could have been deployed in anticipation of an acceptable event and would not need to be installed in the field continuously.

Some of the site descriptions lacked important information and local personnel sometimes did not have the needed information. This was especially critical for watershed delineations on maps of the area. Also, few of the watershed descriptions adequately described how the impervious areas were connected to the drainage system, one of the most important factors affecting urban hydrologic analyses. In most cases, information concerning local stormwater controls was able to be determined from a variety of sources, but it was not clearly described in the annual reports.

## Comparison of NSQD with Existing Stormwater Databases

The NSQD, with 3,765 events (from the 1992-2002 period) represented sites throughout much of the US for most land uses, and for many constituents. It is therefore the most comprehensive stormwater quality database currently available for US stormwater conditions. The historical NURP database (sampling period in the late 1970s and early 1980s) contains the results from 2,300 national stormwater events, while the CDM National Urban Stormwater Quality Database includes the results of approximately 3,100 events (including the NURP data, plus additional data collected by the USGS and about 30 NPDES permits; Smullen and Cave, 2002). Table 9 compares the results of the pooled EMC's from the NURP (calculated by Smullen and Cave 2002), CDM, and NSQD databases.

The NURP means and medians were computed by Smullen and Cave (2002) using the EPA (1983) data. The CDM and the NSQD results are similar for all constituents, except for lead and zinc. All three databases have similar reported median and mean concentrations for COD and BOD and the nutrients, but are apparently different for TSS and the heavy metals. The pooled mean event mean concentration (EMC) for TSS was 2.3 times larger in the NURP database compared to the NSQD. The largest reduction in mean EMCs was found for lead ( 7.9 times larger for NURP) followed by copper ( 7.9 times larger for NURP) and zinc (1.6 times large for NURP).

Table 9. Comparison of Stormwater Databases

| Constituent | Units | Source | Event Mean Concentrations |  | Number of events |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Mean | Median |  |
| Total Suspended Solids | mg/L | NURP | 174 | 113 | 2000 |
|  |  | CDM | 78.4 | 54.5 | 3047 |
|  |  | NSQD | 79.1 | 49.8 | 3404 |
| Biochemical Oxygen Demand | mg/L | NURP | 10.4 | 8.4 | 474 |
|  |  | $\mathrm{CDM}^{\text {a }}$ | 14.1 | 11.5 | 1035 |
|  |  | NSQD | 10.9 | 8.6 | 2973 |
| Chemical Oxygen Demand | mg/L | NURP | 66.1 | 55.0 | 1538 |
|  |  | CDM | 52.8 | 44.7 | 2639 |
|  |  | NSQD | 71.2 | 55.6 | 2699 |
| Total Phosphorus | mg/L | NURP | 0.337 | 0.266 | 1902 |
|  |  | CDM | 0.315 | 0.259 | 3094 |
|  |  | NSQD | 0.373 | 0.289 | 3162 |
| Dissolved Phosphorus | mg/L | NURP | 0.100 | 0.078 | 767 |
|  |  | $\mathrm{CDM}^{\text {b }}$ | 0.129 | 0.103 | 1091 |
|  |  | NSQD | 0.107 | 0.078 | 2093 |
| Total Kjeldahl Nitrogen | mg/L | NURP | 1.67 | 1.41 | 1601 |
|  |  | CDM | 1.73 | 1.47 | 2693 |
|  |  | NSQD | 1.74 | 1.37 | 3034 |
| Nitrite and Nitrate | mg/L | NURP | 0.837 | 0.666 | 1234 |
|  |  | CDM | 0.658 | 0.533 | 2016 |
|  |  | NSQD | 0.767 | 0.606 | 2983 |
| Copper | $\mu \mathrm{g} / \mathrm{L}$ | NURP | 66.6 | 54.8 | 849 |
|  |  | CDM | 13.5 | 11.1 | 1657 |
|  |  | NSQD | 17.8 | 14.2 | 2356 |
| Lead | $\mu \mathrm{g} / \mathrm{L}$ | NURP | 175 | 131 | 1579 |
|  |  | CDM | 67.5 | 50.7 | 2713 |
|  |  | NSQD | 24.4 | 16.5 | 2250 |
| Zinc | $\mu \mathrm{g} / \mathrm{L}$ | NURP | 176 | 140 | 1281 |
|  |  | CDM | 162 | 129 | 2234 |
|  |  | NSQD | 110 | 88 | 2888 |

Note: a. No $\mathrm{BOD}_{5}$ for USGS dataset. b. No DP for CDM portion of NPDES dataset

In an effort to recognize why differences were observed between the NURP and NSQD databases, further examinations of two communities that monitored stormwater during both NURP and the Phase I NPDES program were made. As part of their MS4 Phase I application, Denver and Milwaukee both returned to some of their earlier sampled monitoring stations used during the local NURP projects (EPA 1983). In the time between the early 1980s (NURP) and the early 1990s (MS4 permit applications), they did not detect any significant differences, except for large decreases in lead concentrations. Figure 51 compares suspended solids, copper, lead, and zinc concentrations at the Wood Center NURP monitoring site in Milwaukee. The average site concentrations remained the same, except for lead, which decreased from about 450 to about $110 \mu \mathrm{~g} / \mathrm{L}$, as expected due to the decrease in leaded gasoline during this period.


Figure 51. Comparison of pollutant concentrations collected during NURP (1981) to MS4 application data (1990) at the same location (personal communication, Roger Bannerman, WI DNR)

Urban Drainage and Flood Control District performed similar comparisons in the Denver Metropolitan area. Table 43 compares stormwater quality for commercial and residential areas for 1980/81 (NURP) and 1992/93 (MS4 application). Although there was an apparent difference in the averages of the event concentrations between the sampling dates, they concluded that the differences were all within the normal range of stormwater quality variations, except for lead, which decreased by about a factor of four.

Trends of stormwater concentrations with time can also be examined using the NSQD data. A classical example would be for lead, which is expected to decrease over time with the increased use of unleaded gasoline. Older stormwater samples from the 1970s typically have had lead concentrations of about 100 to $500 \mu \mathrm{~g} / \mathrm{L}$, or higher (as indicated above for Milwaukee and Denver), while most current data indicate concentrations as low as 1 to $10 \mu \mathrm{~g} / \mathrm{L}$.

Table 43. Comparison of Commercial and Residential Stormwater Runoff Quality from 1980/81 to 1992/93 (Doerfer, 1993)

| Constituent | Commercial |  | Residential |  |
| :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{1 9 8 0} \mathbf{- 1 9 8 1}$ | $\mathbf{1 9 9 2 - 1 9 9 3}$ | $\mathbf{1 9 8 0} \mathbf{- 1 9 8 1}$ | $\mathbf{1 9 9 2 - 1 9 9 3}$ |
| Total suspended solids $(\mathrm{mg} / \mathrm{L})$ | 251 | 165 | 226 | 325 |
| Total nitrogen $(\mathrm{mg} / \mathrm{L})$ | 3.0 | 3.9 | 3.2 | 4.7 |
| Nitrate plus nitrite $(\mathrm{mg} / \mathrm{L})$ | 0.80 | 1.4 | 0.61 | 0.92 |
| Total phosphorus $(\mathrm{mg} / \mathrm{L})$ | 0.46 | 0.34 | 0.61 | 0.87 |
| Dissolved phosphorus $(\mathrm{mg} / \mathrm{L})$ | 0.15 | 0.15 | 0.22 | 0.24 |
| Copper, total recoverable $(\mathrm{gg} / \mathrm{L})$ | 27 | 81 | 28 | 31 |
| Lead, total recoverable $(\boldsymbol{\mu g} / \mathrm{L})$ | 200 | 59 | 190 | 53 |
| Zinc, total recoverable $(\mathrm{mg} / \mathrm{L})$ | 220 | 290 | 180 | 180 |

The differences found in both the NURP and the NSQD databases are therefore most likely due to differences in geographical areas emphasized by each database. Figure 10 is a national map showing the percentage of events collected in each state as contained in the NSQD database, while Figure 11 shows the percentage of events contained in the NURP database. Half of the events included in the NSQD database were collected in EPA Rain Zone 2 (Maryland, Virginia, North Carolina, Kentucky and Tennessee), while half of the events contained in the

NURP database were collected in EPA Rain Zone 1 (Minnesota, Wisconsin, Michigan, Illinois, New York, Massachusetts and New Hampshire). Only 3\% of the events in the NSQD are located in EPA Rain Zone 1, while $50 \%$ of the NURP data is from this area. Twenty four percent of the NURP data is located in the Mid-Atlantic and southeast states, while $60 \%$ of the NSQD data is from this area (the area that was emphasized for this EPA-funded project). The NSQD is slightly better representative of other parts of the country compared to NURP. As an example, the percentage of the total event data from the west coast is similar for both databases, but the NSQD represents 10 communities with almost 60 different sites, while NURP has only 3 communities and only 7 sites. The total number of sites, communities and events collected in the NURP study are shown in Table 10.

Table 10. Total Events Monitored During NURP by EPA Rain Zones

| Rain Zone | Total Events | Percentage of <br> Events | Number of <br> Communities | Number of <br> Sites |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 804 | 51 | 12 | 42 |
| 2 | 324 | 20 | 3 | 10 |
| 3 | 65 | 4.1 | 1 | 5 |
| 4 | 0 | 0 | 0 | 0 |
| 5 | 24 | 1.5 | 1 | 2 |
| 6 | 45 | 2.8 | 2 | 5 |
| 7 | 136 | 8.6 | 1 | 2 |
| 8 | 0 | 0 | 0 | 0 |
| 9 | 188 | 12 | 3 | 12 |



Figure 10. Distribution of collected events using the NSQD database.


Figure 11. Distribution of collected events using the NURP database.

Figure 12 presents example plots for selected residential area data for different EPA Rain Zones for the country as contained in the NSQD. Rain Zones 3 and 7 (the wettest areas of the country) had the lowest concentrations for most of the constituents, while Rain Zone 1 has some of the highest concentrations.


Figure 12. Example of constituents collected in residential land use by EPA Rain Zone

It is likely that the few data from EPA Rain Zone 1 (having relatively high concentrations) in the NSQD and the few data in EPA Rain Zones 2 and 3 (having relatively low concentrations) in NURP are the main reason for the differences in the database summary values.

## Land use effects

Another factor that may affect the difference in reported concentrations between the NURP and NSQD databases is the percentage of samples collected for each different land use category. Although each database summarized observed concentrations by land use, having few data from few sites in a land use category reduces the reliability of the estimate. Almost $45 \%$ of the NURP database represents residential sites, while residential sites comprise about $30 \%$ of the NSQD. The percentage of industrial sites in the NSQD is $15 \%$, while industrial sites in the NURP database represent only $6 \%$ of the total. The NSQD contains samples for freeways sites, which are not included in the NURP database. The percentages of mixed land uses and commercial areas are similar for both databases. However, a better representation of open space land uses was observed in the NURP database ( $10 \%$ of the total) compared with the NSQD ( $3 \%$ of the total).

## Other Factors

Other factors may influence the differences in reported EMCs in the different databases. Figure 13 shows the probability plot for drainage areas for sites included in the NSQD and NURP databases.

This plot shows that the NURP watersheds are larger than those observed in the NSQD. The median NSQD drainage area was about 50 acres, while it was about twice as large during NURP. The NSQD also has about $10 \%$ of the
watersheds smaller than 10 acres, representing freeways sites. No literature was found that indicates that there is a relationship between the drainage area and the concentration of stormwater constituents.


Figure 13. Distribution of collected events using the NURP database

## Chapter Summary

This chapter describes the National Stormwater Quality Database. The information collected from the NPDES Phase I stormwater monitoring program was stored in a spreadsheet containing more than 3,700 rows and 250 columns. Each row represents a single monitored event. The main structure of the database is divided into six sections: site descriptions, hydrologic information, conventional constituents, nutrients, metals, and additional constituents. The collected data is grouped into 11 land use categories: residential, commercial, industrial, open space, freeways, mixtures of these land uses, and institutional. Support documents were also created for each community. These documents include aerial photos of the watershed and outfall area (when available), narrative descriptions about the main activities and land uses in the watersheds, sampling and quality control procedures, analytical methods, and equipment used during the collection and analysis of the samples. The last part of the support documents describe the problems that occurred during the collection and analyses of the samples, and meeting discharge permit requirements that specified sampling requirements. This information is useful for interpreting the reported monitoring data and as guidance for future stormwater programs in other communities around the country.

The data from the NSQD was compared with information from the most commonly used stormwater database, the EPA's Nationwide Urban Runoff Program (NURP) conducted more than 20 years ago. It was observed the concentrations in the NSQD were in general lower than those found during the NURP program. The analysis indicates that the main reason of these differences is the geographical differences represented by the monitoring locations represented in the databases. Most of the samples during the NURP program were collected in the upper Midwest and northeast coast areas of the country, while most of the samples represented in the NSQD were collected in the mid-east coast and southeast areas of the country. The preliminary regional analyses shown in this chapter indicate that southeast areas have lower stormwater concentrations than northeast areas.

## Chapter 3: QA/QC Procedures

## Introduction

This chapter presents the quality assurance and quality control procedures followed during the creation of the database. These tasks relied on two basic activities: identification of unusual observations and monitoring locations, and the examination of alternative methods to address non-detected pollutant concentration observations (leftcensored data).

## Quality Control/ Quality Assurance

More than 70 communities were contacted to request information concerning their NPDES Phase I monitoring activities. Communities submitted their reports in either electronic media or on paper. In cases where the data were in electronic form, the data were manipulated with macros and stored in the main Excel spreadsheet. For those communities with data only on paper, the information was typed directly into the spreadsheet.

Once the database was completed, the main table was first reviewed by rows (corresponding to individual runoff events) and then by columns (corresponding to measured constituents). Each row and column in the database was reviewed at least once and compared to information contained in the original reports (when available). For each constituent, probability plots, box and whisker plots, and time series plots were used to identify possible errors (likely associated with the transcription of the information, or as typographical errors in the original reports). Most of the identified errors were associated with the transcription process and, in some cases, errors associated with incorrect units (such as some metal results reported as $\mathrm{mg} / \mathrm{L}$ when they were really as $\mu \mathrm{g} / \mathrm{L}$ ).

Additional "logical" plots were used to identify possible errors in the database. A plot of the dissolved (filtered) concentrations against the total concentrations for metals should indicate that the dissolved concentrations are lower than the total forms, for example (Figure 14). Other plots included TKN versus $\mathrm{NH}_{3}, \mathrm{COD}$ versus $\mathrm{BOD}_{5}$, SS versus turbidity and TDS versus conductivity.

In all cases, suspect values were carefully reviewed and many were found to be associated with simple transcription errors, or obviously improper units, which could be corrected. However, about 300 suspect values were removed from the database as they could not be verified. None of the data were deleted without sufficient evidence of a highly probable error. For example, if a set of samples from the same community had extremely high concentrations (in one case, 20 times larger than the typical concentrations reported for other events for the same community) at different sites, but for the same event, this will indicate a very likely error during the collection or analysis of the sample. If just a single site had high concentrations (especially if other related constituents were also high), it would not normally be targeted for deletion, but certainly subject to further scrutiny. If a value was deleted from the database, or otherwise modified, a question mark notation was assigned to the respective constituent in the qualifier column. Appendix B includes all the modifications performed in the database.

In order to calculate the standard deviations for the site quality control tests, each location must have at least two observations. Nine sites were not included in that analysis because they had only one observation. These sites were: ALHUDRAV, KYLXEHL4, KYLXEHL5, KYLXNEL1, MABOA007, ORCCA001, ORODA001, ORODA002, and ORODA004

Many specific statistical methods were used as part of the QA/QC review, in addition to simple data comparisons on multiple generations of data sheets, and logical patterns. The following is one example that was used to identify unusual monitoring locations and to verify the associated data observations with site characteristics.


Figure 14. Example scatter plots of stormwater data (line of equivalent concentration shown)

## Unusual Monitoring Locations

Box and whisker plots can be used as a preliminary examination of the principal factors and interactions between EPA Rain Zones and land use for any constituent. These plots can also be used to identify sites that do not fit within an established pattern shown by other land use locations from other regions of the country. Figure 23 shows box and whisker plots for residential, commercial, and industrial land uses for EPA Rain Zones 1 through 9 . These plots indicate that there are significant differences between EPA Rain Zones and between land uses. Statistical tests also found that the interaction of these two factors was also significant. The median observations by land use have patterns similar to those found during NURP (EPA 1983), and other studies. Residential and open space areas have lower concentrations than commercial and industrial land use areas.



Figure 23. Box and whisker plots for TSS, total phosphorus and total copper by EPA Rain Zone and land use (Two-way ANOVA analyses indicate that both land use and geographical main factors and the interaction are significant at the $\mathbf{0 . 1 \%}$ level)

Residential, commercial and industrial areas are the single land uses having the most observations in the database. These three land uses were analyzed separately to identify those sites with different characteristics than the remaining sites in the same land use and EPA Rain Zone. The following is an example using TSS at residential land use sites to demonstrate the method used to detect unusual monitoring sites in the database. Summaries of additional constituents in residential, commercial and industrial land uses are given in Appendix D.

## Example Using Single Residential Land Use

The following example explains the steps used to identify unusual locations in the database. This analysis was performed in three steps. First, box and whisker plots we used to identify any site with concentrations unusually high or low compared with the other residential locations. The plot was used to identify preliminary differences between and within EPA Rain Zones. Figure 24 shows that there are some sites in EPA Rain Zone 2 having lower TSS concentrations than the remaining residential sites included in the database. On the other hand, it seems that sites located in EPA Rain Zone 4 have higher concentrations than other groups. The second step was to identify those single residential sites that failed the Xbar and S chart tests for all the observations and by EPA Rain Zone.

A total of 10 Xbar and S charts were created for each EPA Rain Zone and for all the zones combined. An indication of geographical differences is if the Xbar chart using all observations shows clusters close or outside the control limits. The effect will be confirmed if none of the sites failed the Xbar test within EPA Rain Zones. The S chart identifies those sites that have a larger or smaller variation than the overall sites in the set.

Figure 25 shows the Xbar and S chart for the residential land use sites. Six sites have mean TSS values different from the remaining sites in the same group. One important characteristic of this plot is that the control limits change with the number of samples collected at each site. The S chart identifies those sites with standard deviations different than the pooled deviation of the data set. In this case, two sites are outside the control limits. Table 15 shows the sites that failed the Xbar and S chart for all residential sites and for each EPA Rain Zone. Table 15 shows that most of the sites located below the lower control limit were located in North Carolina, Virginia (EPA Rain Zone 2) or Oregon (EPA Rain Zone 7). Sites above the upper control limit were located in Arizona (EPA Rain Zone 6), Kansas (EPA Rain Zone 4), and Colorado (EPA Rain Zone 9).

Xbar plots by EPA Rain Zones also indicate differences within groups. EPA Rain Zones 2, 3, and 4 showed nine sites failing the Xbar test. Six sites out of 54 failed the Xbar chart test in residential land use EPA Rain Zone 2. Each of these sites will be described individually.

The first site was located in Kentland Village (Flagstaff Street), in Prince George County, Maryland (Location_ID = MDPGCOS2, median TSS = $132 \mathrm{mg} / \mathrm{L}$ ). This site with 63 events has the largest number of observations in the database. An industrial park and a commercial area surrounded this high-density residential site. A special characteristic of this site is the construction of a stadium close of the watershed during the monitoring period.


Figure 24. TSS box and whisker plots in residential land use by EPA Rain Zone and location


Figure 25. Xbar and S chart for residential land use in EPA Rain Zone 2

Table 15. Sites Failing Xbar and R Chart in Residential Land Uses

| Rain Zone | Sites Failing Xbar chart | Sites Failing S Chart |
| :---: | :---: | :---: |
| ALL | AZTUA001(H) CODEA005(H) GAATAT02(L) GACLCOTR(H) KATOATWO(H) KATOBROO(H) KYLXEHL7(L) MDPGCOS2(H) MNMISD01(H) NCCHSIMS(L) NCFVCLEA(L) NCFVTRYO(L) NCGRWILL(L) ORCCA004(L) TXHCA006(H) TXHOA003(L) VAARLCV2(L) VAARLLP1(L) VACHCOF3(L) VACHCOF5(L) VAHATYH5(L) VAPMTYP5(L) | VAVBTYV2(L) |
| 1 | None | None |
| 2 | $\begin{aligned} & \text { MDPGCOS2(H) MDSHDTPS(H) NCCHSIMS(L) VACHCOF3(L) } \\ & \text { VACHCOF5(L) VAVBTYV1(H) } \end{aligned}$ | MDSHDTPS(H) <br> VAVBTYV2(L) |
| 3 | GACLCOTR(H) | None |
| 4 | TXHCA006(H) TXHOA003(L) | None |
| 5 | None | None |
| 6 | None | None |
| 7 | None | None |
| 8 | None | None |
| 9 | None | None |

The second site has 13 observations and was operated by the Maryland State Highway Department (MDSHDTPS, median TSS $=135 \mathrm{mg} / \mathrm{L}$ ). This 51 -acre site is considered $96 \%$ single family residential, with $4 \%$ agricultural land use. The site is located close to the intersection of two highways. Observed concentrations ranged from $10 \mathrm{mg} / \mathrm{L}$ up to $750 \mathrm{mg} / \mathrm{L}$. The highest concentrations were observed in summer and the lowest in spring. Another site in EPA Rain Zone 2 with elevated values has 26 observations and is located close to Bow Creek in Virginia Beach, VA (VAVBTYV1, median TSS $=69$ ). This site is located close to a golf course and is drained by a natural channel.

The site with a standard deviation below the lower control limit (VAVBTYV2) is located next to VAVBTYV1. It has also a high TSS concentration but inside the control limits. A total of 30 samples were collected at VAVBTYV2. The aerial photograph did not indicate any unusual conditions at this site.

In EPA Rain Zone 4, only one site had high concentrations compared with the remaining residential sites. This site (TXHCA006) is located in Harris County, TX. Six samples were collected, having a median TSS of $550 \mathrm{mg} / \mathrm{L}$. This site is also analyzed in Chapter 5 and seems to be affected by flooding or erosion activity. In EPA Rain Zone 3, site GACLCOTR is a new development in Tara Road, Clayton County, and Georgia. Twenty-two samples were collected at this location. The median TSS was $200 \mathrm{mg} / \mathrm{L}$. No unusual conditions were identified when examining the aerial photographs.

Site mean concentrations below the lower control limit in the Xbar chart were located in Virginia, North Carolina and Texas. The two sites located in Virginia are located in Chesterfield County. The first site is located in King Mills Road (VACHCOF3, 10 observations, median TSS $=4 \mathrm{mg} / \mathrm{L}$ ) and is located in a forested area with less than $20 \%$ impervious. The second site (VACHCOF5, 14 observations, median TSS $=15 \mathrm{mg} / \mathrm{L}$ ) is $50 \%$ impervious, but surrounded by a forested area. Only four events were collected at the site between March and August 1993, in Silo Lane, Charlotte, North Carolina (NCCHSIMS, median TSS = $10 \mathrm{mg} / \mathrm{L}$ ), no unusual characteristics were observed from the aerial photographs. The unusual low concentration site in Houston, Texas is located on Lazybrook Street (TXHOA003, median TSS $=21 \mathrm{mg} / \mathrm{L}$ ). Freeways ( $\mathrm{I}-610$ ) are located in the north and west part of the watershed. Tall trees surrounding the houses were also observed inside the watershed.

The final step was using ANOVA to evaluate if any EPA Rain Zone was different than the others. The ANOVA table indicated a p-value close to zero, indicating that there are significant differences in the TSS concentration among at least two of the different EPA Rain Zones. The Dunnett's comparison test with a family error of 5\% indicate that concentrations in EPA Rain Zones 4 (median TSS $=91 \mathrm{mg} / \mathrm{L}$ ), 5 (median TSS $=83 \mathrm{mg} / \mathrm{L}$ ), 6 (median $\mathrm{TSS}=118 \mathrm{mg} / \mathrm{L}$ ), 7 (median TSS $=69 \mathrm{mg} / \mathrm{L}$ ), and 9 (median TSS $=166 \mathrm{mg} / \mathrm{L}$ ) are significantly higher than the concentrations observed in EPA Rain Zone 2 (median TSS $=49 \mathrm{mg} / \mathrm{L}$ ).

This same procedure was performed for the following 13 additional constituents in residential, commercial and industrial land use areas: hardness, TSS, TDS, oil and grease, BOD, COD, NO2 + NO3, ammonia, TKN, dissolved phosphorus, total phosphorus, copper, lead and zinc.

## Identification of Unusual Sites

The Xbar charts were created for residential, commercial and industrial land uses. In residential areas, 54 sites were identified with at least one constituent out of control. These sites failed when compared with sites in the same EPA Rain Zone. Table 16 shows the sites with more than 4 constituents outside the control limits.

These eight sites were located in EPA Rain Zone 2. Three sites show elevated concentrations, one in all constituents, and another in metals and the third in nutrients. The site located near a golf course in Virginia Beach (VAVBTYV1) shows elevated concentrations in TSS, phosphorus and COD. The site located in Prince George County close to an industrial park (MDPGCOS2), indicated elevated concentrations of total phosphorus, lead and zinc.

The site with the highest number of constituents outside the control limits ( 10 out of 14 constituents evaluated) was located in Mt. Vernon, Lexington, Kentucky (KYLXTBL1). This site was monitored between 1992 an 1997; it is located close to two high schools and the University of Kentucky. It is interesting that one of the sites having elevated concentrations is located next to one of the sites with a large number of constituents below the lower control limit (VAVBTYV2 is located close to VAVBTYV1). VAVBTYV1 has low concentrations for 6 out of 14 constituents. This indicates that not only can geographical differences be expected; there are also differences between locations in the same EPA Rain Zone. Lead was most frequently found with high concentrations within the same EPA Rain Zone. Eight sites had elevated lead concentrations, while 11 sites had lower concentrations in the same group. The least frequent out-of-bound constituent was oil and grease: none of the sites indicated elevated concentrations of oil and grease when compared with other locations in the same EPA Rain Zone.

Table 16. Sites Failing Xbar Chart in Residential Land Uses

| SITE | HA | TSS | TDS | OG | BOD | COD | NO2 | NH3 | TKN | DP | TP | Cu | Pb | Zn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1MABOA006 |  |  |  |  |  |  |  |  |  |  | L | H |  |  |
| 2KYLOTSR3 |  |  | H |  |  |  |  |  |  |  |  |  |  |  |
| 2KYLXTBL1 | H |  | H |  | H | H | H |  | L | H | H | H | H | H |
| 2MDAACORK |  |  |  |  |  |  |  |  |  |  |  |  | L |  |
| 2MDBACOSC |  |  |  |  |  |  | H |  |  |  |  |  | L |  |
| 2MDBCTYHR |  |  |  |  |  |  |  |  | H |  |  | H |  |  |
| 2MDCLCOCE |  |  |  |  |  |  |  |  |  |  |  |  |  | H |
| 2MDHACOBP |  |  |  |  | L |  |  |  | L |  | L | L | L | L |
| 2MDHOCOGM |  |  |  |  | H |  |  |  |  |  |  |  |  |  |
| 2MDPGCOS2 |  | H |  |  |  |  |  |  |  |  | H | L | H | H |
| 2MDSHDTPS |  | H |  |  |  |  |  |  |  |  |  | L |  |  |
| 2NCCHHIDD |  |  |  |  |  |  | H | H |  |  |  |  |  |  |
| 2NCCHNANC |  |  |  |  | L |  |  |  |  |  |  |  |  |  |
| 2NCCHSIMS |  | L |  |  |  |  |  | H |  |  |  |  |  |  |
| 2NCFVCLEA |  |  | L |  |  |  |  |  |  |  | L | L |  |  |
| 2NCFVTRYO |  |  |  |  |  |  |  |  |  |  |  | L | L |  |
| 2NCGRWILL |  |  |  |  |  |  |  | L |  | H |  |  |  |  |
| 2VAARLCV2 |  |  |  |  | L | L |  |  |  |  | L |  | L |  |
| 2VAARLLP1 |  |  | H |  |  |  |  |  |  |  |  |  | L | L |
| 2VACHCN2A |  |  |  |  |  |  |  |  |  |  |  | L |  | L |
| 2VACHCOF3 | L | L | L | L |  | L |  |  |  | L | L | L |  |  |
| 2VACHCOF5 |  | L | L |  |  | L |  |  |  |  | L |  |  |  |
| 2VACPTSF2 |  |  |  |  |  |  |  |  |  |  |  | L |  | L |
| 2VAFFCOF1 |  |  | L |  |  |  |  |  |  |  |  |  |  |  |
| 2VAHATYH3 |  |  |  |  |  |  |  |  | L | L |  |  |  |  |
| 2VAHATYH5 |  |  |  |  |  |  |  |  |  | H |  |  |  | L |
| 2VANFTYN2 |  |  |  |  |  |  |  |  |  | H | H |  |  |  |
| 2VANFTYN3 |  |  |  |  |  | H |  |  |  |  |  |  | H |  |
| 2VANFTYN5 |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 2VAPMTYP2 |  |  |  |  |  |  |  |  |  | L |  |  | H |  |
| 2VAPMTYP4 |  |  |  |  |  |  |  | L |  | L |  |  |  |  |
| 2VAPMTYP5 |  |  |  |  |  |  |  |  |  |  |  |  |  | L |
| 2VAVBTYV1 |  | H |  |  |  | H |  |  |  | H | H |  |  |  |
| 2VAVBTYV2 |  |  | H |  | L |  | L | L |  | L |  | L |  | L |
| 3GAATAT02 |  |  |  |  |  |  |  |  |  |  |  | H |  |  |
| 3GACOC1A3 |  |  |  |  | L |  |  |  |  |  |  |  |  |  |
| 3GACLCOTR |  | H |  |  |  |  |  |  |  |  |  |  |  |  |
| 4KATOATWO |  |  |  | L |  |  |  |  |  |  |  |  |  |  |
| 4KATOBROO |  |  |  | L |  |  |  |  |  |  |  |  |  |  |
| 4TXHOA003 |  | L | H |  |  |  |  |  |  |  |  | L |  |  |
| 5TXARA002 |  |  |  |  |  | L |  |  |  |  |  |  | L |  |
| 5TXARA003 |  |  |  |  |  |  |  |  |  |  | H |  |  |  |
| 5TXDAA005 |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 5TXIRA001 |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 5TXMEA002 | H |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5TXMEA003 |  |  |  |  |  |  |  |  |  |  |  |  | L |  |
| 6AZMCA006 |  |  |  |  |  |  |  |  |  |  |  | H |  |  |
| 6AZTUA001 |  |  | H |  |  | H |  |  |  |  |  |  |  |  |
| 6AZTUA002 |  |  |  |  |  |  |  |  |  |  |  |  | L |  |
| 7ORCCA004 |  |  |  |  |  |  |  |  |  |  |  |  | L |  |
| 7OREUA003 | H |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 7ORGRA003 |  |  |  | L |  |  |  |  |  |  |  |  |  |  |
| 70RPOA006 | L |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 7ORSAA004 |  |  |  |  |  |  |  |  | L |  |  |  | L | L |

Note: H: Site with mean concentrations larger than UCL. L: Site with mean concentrations lower than LCL

In commercial land use areas, six out of 25 locations indicated more than three constituents outside of the control limits (Table 20). Five sites have more than one constituent above the upper detection limit. The site with the largest number was located in Wilhite Drive behind a K-Mart large shopping center in Lexington, Kentucky
(KYLXWHL1). This site was monitored between 1992 and 1996. The site indicates elevated nutrients, BOD, hardness and TDS concentrations. The second site was also located in Kentucky. East Land is located in an old commercial area in Lexington (KYLXNEL3). This site has elevated total and dissolved phosphorus concentrations.

Table 17. Sites Failing Xbar Chart in Commercial Land Uses

| SITE | HA | TSS | TDS | OG | BOD | COD | NO2 | NH3 | TKN | DP | TP | Cu | Pb | Zn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2KYLXNEL3 |  |  |  |  |  |  |  | L |  | H | H |  |  |  |
| 2KYLXWHL1 | H |  | H |  | H |  |  |  | H | H | H |  |  |  |
| 2MDAACOPP |  |  |  |  |  |  | L |  |  |  |  |  | L |  |
| 2MDHOCODC |  |  |  |  | L |  |  |  |  |  | L |  |  |  |
| 2MDHOCODC |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2MDPGCOS1 |  | H |  |  |  |  |  |  |  |  |  | H | H | H |
| 2NCGRATHE |  |  |  |  |  |  |  |  |  |  | H |  |  |  |
| 2NCGRMERR |  |  |  |  |  |  |  |  |  |  |  |  |  | L |
| 2VAARLRS3 |  |  |  |  | L | L |  |  |  |  |  |  | L | L |
| 2VACHCCC4 | H | L |  | L | L | L |  |  |  |  |  |  |  |  |
| 2VAHATYH1 |  |  |  |  |  |  |  |  |  | L |  |  |  |  |
| 2VAHCCOC2 |  |  |  |  |  |  |  |  |  |  |  |  |  | H |
| 2VAPMTYP1 |  |  |  |  |  |  |  |  |  | L |  |  |  |  |
| 3ALHUMASM |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 3ALHUWERP |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 3ALMODAPH |  |  |  |  | H |  |  |  |  |  |  |  |  |  |
| 4KATOJACK |  | H |  | H |  |  |  |  |  |  |  | H |  | H |
| 4TXHOA004 |  |  |  | L |  |  |  |  |  |  |  |  |  |  |
| 6AZTUA003 |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 7OREUA001 | H |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 7ORPOA001 | L |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 9CODEA001 |  |  |  |  |  |  |  |  |  |  |  | H |  | H |
| 9CODEA002 |  |  |  | H |  |  |  |  |  |  |  |  |  |  |
| 9KAWITOWN |  |  |  | L |  |  |  |  |  |  |  |  |  |  |

Note: H: Site with mean concentrations larger than UCL. L: Site with mean concentrations lower than LCL
A third site having elevated stormwater concentrations was found in Brightseat Road adjacent to Landover Mall in Prince George County, Maryland (MDPGCOS1). This site was monitored between 1992 and 1996. It has elevated TSS, copper, lead and zinc concentrations. A fourth site with elevated stormwater concentrations is located in Topeka, Kansas (KATOJACK). This site is located close to a sand quarry. Median TSS concentrations at this location were close to $600 \mathrm{mg} / \mathrm{L}$. Elevated oil and grease, total lead and total zinc were also found at this location. The last elevated concentration site is located in Denver, Colorado. Cherry Creek at Colfax Avenue (CODEA001) has elevated copper and zinc concentrations. The site is $87 \%$ commercial and contains a convention center, hotels and restaurants on 16th Street Mall, the State Capital and other government buildings.

Four out of 25 industrial land use locations indicated more than three constituents with median concentrations outside the upper control limit (Table 18). One site is located in Boston, Massachusetts. The Brighton (MABOA004) watershed drains runoff from warehouses and manufacturing operations associated with mechanical, roofing and electrical activities. According to the site description, there is a large potential for storage of rainfall on rooftops and poorly maintained parking lots and roadways. Extremely high ammonia and TKN concentrations were observed at this location. Another industrial site having high concentrations is located in Greensboro, North Carolina. The site is located at Husband Street (NCGRHUST). Zinc and especially copper concentrations were elevated (median copper $=29 \mu \mathrm{~g} / \mathrm{L}$ ).

A site located at Santa Fe Shops in Topeka, Kansas (KATOSTFE) had elevated metal concentrations. Railroad activity was present in the watershed. Another industrial site of interest is located on 27th Avenue at the Salt River in Maricopa, Arizona (AZMCA003). It had a median TSS of $668 \mathrm{mg} / \mathrm{L}$. Copper, lead and zinc had extremely high concentrations at this location compared with many other single land uses sites in the database.

Table 18. Sites Failing Xbar Chart in Industrial Land Uses

| SITE | HA | TSS | TDS | OG | BOD | COD | NO2 | NH3 | TKN | DP | TP | Cu | Pb | Zn |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1MABOA004 |  |  |  |  |  |  |  | H | H |  |  | H |  |  |
| 1MNMISD03 |  |  |  |  |  |  |  |  |  |  |  | L |  |  |
| 2KYLOTSR2 |  |  |  |  | H |  |  |  |  |  |  |  |  |  |
| 2KYLXTBL2 |  |  |  |  |  |  |  | L |  | H | H |  |  | L |
| 2MDBACOTC |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 2MDCHCOIP |  |  |  |  |  |  |  |  |  |  |  | L |  |  |
| 2MDPGCOS6 |  | H |  |  |  |  |  |  |  |  |  |  | H |  |
| 2NCCHBREV |  |  |  |  |  | H | H |  |  |  |  |  |  |  |
| 2NCCHHOSK |  |  |  |  |  |  |  | H |  |  |  |  |  | H |
| 2NCFVWINS |  |  |  |  |  |  |  |  |  |  |  |  |  | H |
| 2NCGRHUST |  |  |  |  | H |  |  |  |  |  |  | H |  | H |
| 2VAARLTC4 |  |  | H |  |  |  |  |  |  |  |  | L | L | L |
| 2VACHCOF1 |  |  |  |  |  |  |  |  |  |  |  | L |  |  |
| 2VACPTYC5 | L | L |  | L | L | L | L | L | L | L | L |  |  |  |
| 2VAFFOF10 |  |  |  |  |  |  |  |  | H |  |  |  | L |  |
| 2VAFFOF11 |  |  |  |  | H |  |  |  | H |  |  |  |  |  |
| 2VAHATYH2 |  |  |  |  |  |  |  |  |  |  | H |  |  |  |
| 2VAVBTYV4 | L | L |  |  |  |  |  |  |  |  | L |  |  |  |
| 3ALHUCHIP |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 3GAATAT01 |  | L |  |  |  |  |  |  |  |  |  | L |  |  |
| 3GACLCOSI |  |  |  |  |  |  |  |  |  | H |  |  |  |  |
| 4KATOSTFE |  |  |  |  |  |  |  |  |  |  |  | H | H | H |
| 4TNMET211 |  |  |  |  |  |  |  | H |  |  |  |  |  |  |
| 4TXHCA004 |  |  |  |  |  |  |  | L |  |  |  |  |  |  |
| 5TXDAA001 |  |  |  |  |  |  |  |  |  | H |  | L |  |  |
| 5TXDAA002 |  | L |  |  |  | L |  |  |  |  |  |  |  | L |
| 5TXFWA004 |  |  |  |  |  |  |  |  |  |  |  |  | H |  |
| 6AZMCA001 |  |  |  |  |  |  | L |  |  |  |  |  | L |  |
| 6AZMCA003 |  | H |  |  |  |  | H |  |  |  |  | H | H |  |
| 6AZTUA004 |  |  |  |  |  |  |  |  |  |  | L |  |  |  |
| 6CAALAL09 |  |  | L |  |  |  |  |  | L |  |  |  |  |  |
| 7ORSAA003 |  |  |  | L |  |  |  |  |  | L |  |  | L |  |
| 9CODEA007 |  |  |  |  |  |  |  |  |  |  | H |  |  |  |

Note: H: Site with mean concentrations larger than UCL. L: Site with mean concentrations lower than LCL

## Non-Detected Analyses

Left-censored data refers to observations that are reported as below the limits of detection, while right-censored data refers to over-range observations. Unfortunately, many important stormwater measurements (such as for filtered heavy metals) have large fractions of undetected values. These missing data greatly hinder many statistical tests. A number of methods have been used over the years to substitute appropriate values for these missing data in order to perform statistical tests:

- ignore the non-detects and report only using the detected values (also report the detection limit and the frequency of missing data). This may be suitable for the most basic summaries of the data.
- replace the non-detects with zero. This is the method suggested by the EPA for reporting discharge quantities associated with discharge permits. This method results in a decreased discharge estimate by assuming that the non-detects are actually associated with no pollutants in the waste stream.
- replace the non-detects with the detection limit. This would result in an increased discharge amount when conducting mass balances.
- replace with half the detection limit. This is usually the most common method used, but still may result in biased results. The biggest problem with any of these set value replacement methods is that a single value is used for each missing data value. This can therefore have dramatic effects on the calculated variance of the data set and makes statistical comparison tests error prone.
- replace with a randomly generated value based on the measured variation of the available observations. This is usually the preferred method as the variation of the data set is preserved, allowing suitable non-paired comparison tests. Paired tests cannot be conducted as there is no knowledge of which values belong with which observation.
- report the actual instrument reading, even if below the "minimum quantification limit" or "method detection limit." This is the best method, from a statistical standpoint, but is rarely available. Most of the detection limits are extremely conservative, especially in comparison with the other errors associated with a monitoring program. The use of "substandard" detection limits enables the use of all statistical tests, however, care must be taken to describe the detection limit methodology and the actual instrumentation errors.

Berthouex and Brown (2002) has an extended discussion of some of these methods applied to environmental analyses. To estimate the problems associated with censored values, it is important to identify the probability distributions of the data in the dataset and the level of censoring. Most of the constituents in the NSQD followed a lognormal distribution (See Chapter 4). Appendix C shows several approaches to analyze censoring observations with single and multiple detection limits. Different comparisons substantiated the conclusion that the non-detected values in the NSQD can be best estimated using the Cohen's maximum likelihood method (a method that randomly generates the missing data based on the known probability distributions of the data), compared to other traditional methods.

The values of the detection limits and their frequencies varied among the different constituents and monitoring locations. This made handling the non-detectable values even more confusing, as each constituent had several detection limits. Therefore, the first step in evaluating the different methods to address censored data was to identify the probability distribution of the dataset. The second step was applying and evaluating the different estimation methods.

## Censored Data Distribution

The level of censoring for each constituent was calculated for each land use and site, for 18 selected constituents. These constituents contained low levels of censored values. The National Council of the Paper Industry for Air and Stream Improvement found that for levels of censoring (non-detectable observations) above $60 \%$, the use of any estimation method is not appropriate (NCASI 1995). Table 11 shows the maximum, minimum, and percentage of detected values by constituent for each main land use for the complete dataset. In general, freeway sites have the largest percentage of detected observations, while open space sites have the highest percentage of non-detected observations. This is expected as freeway areas have the highest concentrations and open space areas have the lowest concentrations of most reported constituents.

The constituents having greater than $95 \%$ detected observations (of these 18 ) are conductivity, pH , hardness, TSS, TDS, and COD (except for open space areas). Most of the non-detected observations of these 18 constituents were for oil and grease, dissolved phosphorus, lead, and nickel analyses. The percentage of detected observations for these constituents in open space areas varied between $18 \%$ and $75 \%$, while freeways recorded valid values for $89 \%$ to $100 \%$ of the analyses for the metals.

Residential, commercial and industrial land uses have similar percentages of detected observations for each constituent shown in Table 11. The most frequent detection limit for each constituent was also identified. Because of
the duration of the monitoring activities reported in the NSQD, the large number of municipalities involved, and the large number of analytical methods used, each constituent usually had several reported detection limits. The number and percentage of non-detected observations at each detection limit was calculated with respect to the total number of non-detected observations. For example, there are a total of 60 oil and grease observations at freeway sites: 43 detected and 17 non-detected. There were three separate detection limits reported for the non-detected oil and grease observations: $<0.5,<1$ and $<3 \mathrm{mg} / \mathrm{L}$ with 1,2 and 14 observations reported for each, respectively. The frequency distribution of non-detected oil and grease observations at freeways sites was therefore $5.8 \%, 11.8 \%$ and $82.3 \%$, respectively. The results for the remaining land uses and constituents are shown in Table 12. A discussion about the percentage of the detected values and their distributions for each constituent is presented in Appendix D.

Table 11. Percentages of Detected Values by Land Use Category and for the Complete Database

| Constituent | Land use* | Total Events | Minimum Detected Concentration | Maximum Detected Concentration | Percentage with detected values |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ) | RE | 106 | 27.3 | 2020 | 100 |
|  | CO | 66 | 17 | 894 | 100 |
|  | ID | 108 | 42 | 1958 | 100 |
|  | OP | 2 | 75 | 150 | 100 |
|  | FW | 86 | 20 | 870 | 100 |
|  | TOTAL | 685 | 16.8 | 5955 | 100 |
| Hardness (mg/L) | RE | 250 | 3 | 401 | 100 |
|  | CO | 139 | 1.9 | 356 | 100 |
|  | ID | 138 | 5.5 | 888 | 96.4 |
|  | OP | 8 | 11 | 270 | 100 |
|  | FW | 127 | 5 | 1000 | 100 |
|  | TOTAL | 1082 | 1.9 | 1100 | 98.7 |
| Oil and Grease (mg/L) | RE | 533 | 0.2 | 2980 | 57.8 |
|  | CO | 308 | 0.8 | 359 | 70.8 |
|  | ID | 327 | 0.5 | 11000 | 65.1 |
|  | OP | 19 | 0.5 | 4 | 36.8 |
|  | FW | 60 | 3 | 30 | 71.7 |
|  | TOTAL | 1834 | 0.2 | 11000 | 66.1 |
| Total Dissolved Solids (mg/L) | RE | 861 | 3 | 1700 | 99.2 |
|  | CO | 399 | 4 | 3860 | 99.5 |
|  | ID | 412 | 4.5 | 11200 | 99.5 |
|  | OP | 45 | 32 | 542 | 97.8 |
|  | FW | 97 | 12 | 470 | 90.0 |
|  | TOTAL | 2956 | 3 | 17900 | 99.3 |
| Total Suspended Solids (mg/L) | RE | 991 | 3 | 2426 | 98.6 |
|  | CO | 458 | 3 | 2385 | 98.3 |
|  | ID | 427 | 3 | 2490 | 99.1 |
|  | OP | 44 | 3 | 980 | 95.5 |
|  | FW | 134 | 3 | 4800 | 99.3 |
|  | TOTAL | 3389 | 3 | 4800 | 98.8 |
| BOD (mg/L) | RE | 941 | 1 | 350 | 97.6 |
|  | CO | 432 | 2 | 150 | 97.4 |
|  | ID | 406 | 1 | 6920 | 95.3 |
|  | OP | 44 | 1 | 20 | 86.4 |
|  | FW | 26 | 2 | 89 | 84.6 |
|  | TOTAL | 3105 | 1 | 6920 | 96.2 |
| COD (mg/L) | RE | 796 | 5 | 620 | 98.9 |
|  | CO | 373 | 4 | 635 | 98.4 |
|  | ID | 361 | 2 | 1,260 | 98.9 |
|  | OP | 43 | 8 | 476 | 76.7 |
|  | FW | 67 | 2.44 | 1,013 | 98.5 |
|  | TOTAL | 2,750 | 1 | 1,260 | 98.4 |

$\mathrm{RE}=$ residential; $\mathrm{CO}=$ commercial; $\mathrm{ID}=$ industrial; $\mathrm{OP}=$ open space; $\mathrm{FW}=$ freeways Total=total database, all land uses combined, including mixed land uses

Table 11. Percentages of Detected Values by Land Use Category and for the Complete Database Continuation

| Constituent | Land use | Total Events | Minimum Detected Concentration | Maximum Detected Concentration | Percentage of detected values |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fecal Coliform (Colonies/100mL) | RE | 446 | 1 | 5,230,000 | 88.3 |
|  | CO | 233 | 4 | 610,000 | 88.0 |
|  | ID | 297 | 2 | 2,500,000 | 87.9 |
|  | OP | 23 | 650 | 63,000 | 91.3 |
|  | FW | 49 | 50 | 70,000 | 100 |
|  | TOTAL | 1704 | 1 | 5,230,000 | 91.2 |
| Fecal Streptococcus (Colonies $/ 100 \mathrm{~mL}$ ) | RE | 305 | 20 | 840,000 | 89.59 |
|  | CO | 181 | 20 | 1,100,000 | 91.79 |
|  | ID | 195 | 22 | 6,000,000 | 93.9 |
|  | OP | 22 | 160 | 101,000 | 90.9 |
|  | FW | 25 | 560 | 130,000 | 100 |
|  | TOTAL | 1141 | 20 | 6,000,000 | 94.0 |
| Ammonia (mg/L) | RE | 595 | 0.01 | 6 | 81.5 |
|  | CO | 299 | 0.02 | 8 | 83.3 |
|  | ID | 253 | 0.03 | 10 | 83.4 |
|  | OP | 32 | 0.07 | 2 | 18.8 |
|  | FW | 79 | 0.08 | 12 | 87.3 |
|  | TOTAL | 1908 | 0.01 | 12 | 71.3 |
| $\mathrm{NO} 2+\mathrm{NO} 3$ (mg/L) | RE | 927 | 0.01 | 18 | 97.4 |
|  | CO | 425 | 0.03 | 8.21 | 98.1 |
|  | ID | 417 | 0.02 | 8.4 | 96.2 |
|  | OP | 44 | 0.09 | 3.33 | 84.1 |
|  | FW | 25 | 0.1 | 3 | 96.0 |
|  | TOTAL | 3075 | 0.01 | 18 | 97.3 |
| TKN (mg/L) | RE | 957 | 0.05 | 36 | 95.6 |
|  | CO | 449 | 0.05 | 15 | 96.8 |
|  | ID | 439 | 0.05 | 25 | 97.3 |
|  | OP | 45 | 0.2 | 5 | 95.9 |
|  | FW | 125 | 0.2 | 36 | 71.1 |
|  | TOTAL | 3191 | 0.05 | 66 | 96.8 |
| Dissolved Phosphorus (mg/L) | RE | 738 | 0.01 | 2 | 84.2 |
|  | CO | 323 | 0.01 | 2 | 81.1 |
|  | ID | 325 | 0.02 | 2 | 87.4 |
|  | OP | 44 | 0.01 | 1 | 79.6 |
|  | FW | 22 | 0.06 | 7 | 95.5 |
|  | TOTAL | 2477 | 0.01 | 7 | 85.1 |
| Total Phosphorus (mg/L) | RE | 963 | 0.01 | 7 | 96.9 |
|  | CO | 446 | 0.02 | 3 | 95.7 |
|  | ID | 434 | 0.02 | 8 | 95.9 |
|  | OP | 46 | 0.02 | 15 | 84.8 |
|  | FW | 128 | 0.06 | 7 | 99.2 |
|  | TOTAL | 3285 | 0.01 | 15 | 96.5 |

Table 11. Percentages of Detected Values by Land Use Category and for the Complete Database Continuation

| Constituent | Land use | Total Events | Minimum Detected Concentration | Maximum Detected Concentration | Percentage of detected values |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE | 799 | 1 | 590 | 83.6 |
|  | CO | 387 | 1.5 | 384 | 92.8 |
|  | ID | 415 | 1.97 | 1360 | 89.6 |
|  | OP | 39 | 2 | 210 | 74.4 |
|  | FW | 97 | 5 | 244 | 99.0 |
|  | TOTAL | 2723 | 0.6 | 1360 | 87.4 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE | 788 | 0.5 | 585 | 71.3 |
|  | CO | 377 | 1 | 689 | 85.4 |
|  | ID | 411 | 1 | 1200 | 76.4 |
|  | OP | 45 | 0.2 | 150 | 42.2 |
|  | FW | 107 | 1.6 | 450 | 100 |
|  | TOTAL | 2949 | 0.2 | 1200 | 77.7 |
| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE | 419 | 1 | 100 | 45.4 |
|  | CO | 232 | 2 | 110 | 59.5 |
|  | ID | 249 | 1 | 110 | 62.7 |
|  | OP | 38 | 12 | 120 | 18.4 |
|  | FW | 99 | 2.8 | 100 | 89.9 |
|  | TOTAL | 1430 | 1 | 120 | 59.8 |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE | 810 | 3 | 1580 | 96.4 |
|  | CO | 392 | 5 | 3050 | 99.0 |
|  | ID | 432 | 5.77 | 8100 | 98.6 |
|  | OP | 45 | 5 | 390 | 71.1 |
|  | FW | 93 | 6 | 1829 | 96.8 |
|  | TOTAL | 3007 | 2 | 22500 | 96.6 |

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database

| Constituent | Land use* | <** | <0.5 <0 | <0.6 | <1 | <1.2 | <1.4 | <1.9 | <2 | <2.24 | <2.47 | <2.5 | <2.9 | <3 | <3.97 | <5 | <5.2 | <6 | <6.5 | <7 | <8.3 | <10 | <14 | <17 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Oil and Grease (mg/L) | RE | 0.44 | 1.78 |  | 36.89 |  | 2.22 |  | 5.78 | 0.44 | 0.44 | 3.56 |  | 1.78 |  | 39.56 |  | 2.67 | 0.89 |  | 0.44 | 3.11 |  |  |
|  | CO |  | 2.221. | 1.11 | 34.44 |  | 2.22 |  | 5.56 |  |  | 3.33 | 1.11 | 4.44 |  | 40.00 |  | 4.44 |  | 1.11 |  |  |  |  |
|  | ID |  | 2.63 |  | 51.75 |  | 1.75 |  | 6.14 |  |  | 3.51 |  |  |  | 28.95 | 1.75 | 2.63 |  |  |  | 0.88 |  |  |
|  | OP |  |  |  | 100.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | FW |  | 5.88 |  | 11.76 |  |  |  |  |  |  |  |  | 82.35 |  |  |  |  |  |  |  |  |  |  |
|  | TOTAL | 0.16 | 1.930. | 0.16 | 44.69 | 690.16 | 2.09 | 0.16 | 5.14 | 0.16 | 0.16 | 2.57 | 0.16 | 4.02 | 0.16 | 32.64 | 0.32 | 2.09 | 0.32 | 0.16 | 0.16 | 2.25 | 0.16 | 0.16 |
| *see footnote for Table 11 for definitions of land use categories <br> ** the < sign without a value implies a non-detected value that was not identified |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Constituent | $\begin{gathered} \text { Land } \\ \text { use } \\ \hline \end{gathered}$ | <1 | <5 |  | <6 | <10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\underset{(\mathrm{mg} / \mathrm{L})}{\text { Total Dissolved Solids }}$ | RE | 14.29 | 71.43 |  |  | 14.29 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | CO |  |  |  |  | 100.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | ID |  | 50.00 |  | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | OP |  | 100.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | FW | 100.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | TOTAL | 15.00 | 55.00 |  | 5.00 | 25.00 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Constituent | Land <br> use | $<0.5$ | $<1$ | $<2$ | $<5$ | $<10$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Suspended <br> Solids (mg/L) | RE | 7.14 | 7.14 |  | 85.71 |  |
|  | CO | 12.50 |  | 12.50 | 62.50 | 12.50 |
|  | ID |  | 25.00 |  | 75.00 |  |
|  | OP |  | 50.00 |  | 50.00 |  |
|  | FW |  |  |  |  |  |
|  | TOTAL | 4.76 | 9.52 | 4.76 | 78.57 | 2.38 |


| Constituent | $\begin{gathered} \text { Land } \\ \text { use } \end{gathered}$ | <1 | <2 | <3 | <4 | <5 | <6 | <10 | <15 | <20 | <100 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $B O D$ (mg/L) | RE | 8.70 | 13.04 | 8.70 | 4.35 | 4.78 | 0 | 4.35 |  | 8.70 |  |
|  | CO | 9.09 | 18.18 | 9.09 |  | 18.18 | 9.09 | 9.09 |  |  |  |
|  | ID | 5.26 | 31.58 | 5.26 |  | 47.37 |  |  | 5.26 |  |  |
|  | OP | 16.67 | 50.00 | 16.67 |  |  |  |  |  | 16.67 |  |
|  | FW |  |  | 50.00 |  | 50.00 |  |  |  |  |  |
|  | TOTAL | 33.05 | 20.34 | 6.78 | 3.39 | 18.64 | 2.54 | 1.69 | 0.85 | 3.39 | 1.69 |

Table 12．Percentages of Non－detected Values for Different Reported Detection Limits by Land Use and for the Total Database－ Continued

| Constituent | Land <br> use | $<1$ | $<2$ | $<5$ | $<10$ | $<20$ | $<25$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| COD（mg／L） | RE | 33.33 |  | 55.56 | 11.11 |  |  |
|  | CO | 50.00 |  | 33.33 | 16.67 |  |  |
|  | ID |  |  |  | 50.00 | 25.00 | 25.00 |
|  | OP |  |  |  | 20.00 | 80.00 |  |
|  | FW |  |  |  | 100.00 |  |  |
|  | TOTAL | 15.91 | 2.27 | 15.91 | 40.91 | 22.73 | 2.27 |


| Constituent | Land use | ＜ | ＜1 | ＜2 | ＜3 | ＜10 | ＜20 | ＜30 | ＜100 | ＜200 | ＜2000 | ＞ | ＞1．6K | ＞2．4K | ＞3K | ＞6K | ＞12K | ＞16K | ＞24K | ＞30K | ＞35K | ＞40K | ＞60K | ＞80K | ＞160K |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fecal Coliform （Colonies $/ 100 \mathrm{~mL}$ ） | RE |  | 7.84 | 3.92 |  | 1.96 |  |  |  | 3.92 |  |  | 25.49 | 3.92 |  | 1.96 |  | 13.73 | 1.96 |  |  | 1.96 | 21.57 | 3.92 | 7.84 |
|  | CO |  | 17.86 |  | 3.57 | 3.57 |  |  | 3.57 | 3.57 |  |  | 10.71 | 3.57 | 3.57 |  |  | 10.71 |  |  | 3.57 | 3.57 | 28.57 |  | 3.57 |
|  | ID |  | 44.44 | 11.11 |  |  | 2.78 | 2.78 | 2.78 | 5.56 |  |  | 2.78 |  |  | 2.78 |  | 8.33 |  |  |  |  | 11.11 | 2.78 | 2.78 |
|  | OP |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 100.00 |
|  | FW |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | TOTAL | 0.67 | 22.00 | 4.00 | 0.67 | 2.67 | 0.67 | 0.67 | 1.33 | 3.33 | 2.67 | 0.67 | 11.33 | 2.00 | 2.00 | 2.67 | 2.67 | 13.33 | 0.67 | 0.67 | 0.67 | 1.33 | 15.33 | 2.67 | 5.33 |


| $\begin{array}{\|c} \hline \text { 는 } \\ \text { N } \\ \hline \end{array}$ | $\frac{m}{m}$ |  |  |  | $\stackrel{\text { F}}{\stackrel{-}{+}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \text { y } \\ & \frac{6}{1} \\ & \hline \end{aligned}$ | $\left\lvert\, \begin{gathered} \stackrel{N}{0} \\ \underset{0}{2} \\ \hline \end{gathered}\right.$ |  |  |  | $\stackrel{ \pm}{\text { i }}$ |
| $\begin{aligned} & \frac{1}{\circ} \\ & \stackrel{1}{1} \\ & \hline \end{aligned}$ | $\left\lvert\, \begin{gathered} \stackrel{\sim}{n} \\ \underset{0}{2} \\ \hline \end{gathered}\right.$ |  |  |  | $\stackrel{+}{\text { i }}$ |
| $\stackrel{\text { volu}}{\wedge}$ | $\left\|\begin{array}{c} \sim \\ \\ \hline \end{array}\right\|$ |  | $\begin{gathered} 8 \\ 0 \\ i \end{gathered}$ |  | $\stackrel{\infty}{\infty}$ |
| $\begin{aligned} & \text { Y. } \\ & \text { íl } \end{aligned}$ | $\left\|\begin{array}{l} \infty \\ \stackrel{\infty}{N} \end{array}\right\|$ | $\begin{aligned} & \hat{6} \\ & \stackrel{\circ}{\circ} \\ & \stackrel{1}{2} \end{aligned}$ | $\underset{\substack{\infty \\ \underset{\infty}{\infty} \\ \hline}}{ }$ |  | ¢ |
|  | $\underset{m}{m}$ |  |  |  | $\stackrel{\text { F}}{\stackrel{-}{+}}$ |
| $\frac{\stackrel{y}{6}}{\grave{\wedge}}$ | $\begin{array}{\|c\|} \hline 0 \\ \stackrel{n}{2} \\ \hline \end{array}$ | $\begin{array}{\|l\|} \hline \stackrel{\imath}{\circ} \\ \hline \end{array}$ | $\mathfrak{c}$ |  |  |
| $\stackrel{\mathrm{N}}{\mathrm{~V}}$ | $\left\|\begin{array}{c} \infty \\ \underset{\sim}{\infty} \end{array}\right\|$ |  | $\underset{\infty}{\infty}$ |  | $\stackrel{0}{\infty}$ |
| v |  | $\stackrel{\bigcirc}{6}$ |  |  | $\stackrel{\sim}{+}$ |
| v | $\left\|\begin{array}{c} \underset{\sim}{\sim} \\ \underset{\sim}{n} \end{array}\right\|$ | － | $\underset{\infty}{\infty}$ |  | $\stackrel{ \pm}{\text { N }}$ |
| $\overline{\mathrm{v}}$ | $\begin{array}{\|c\|} \infty \\ \stackrel{\infty}{N} \\ \hline \end{array}$ | － | $\begin{aligned} & 8 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ |  | $\stackrel{\square}{\text { N }}$ |
| 薳 | $\underset{\text { w }}{\text { 区 }}$ | O | $\bigcirc 0$ | 3 | 号 |
|  |  |  |  |  |  |


| $\overline{\mathrm{v}}$ |  |  | $\stackrel{\sim}{\sim}$ |  |  | N－ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\stackrel{n}{\dot{v}}$ | $\begin{array}{\|c} \hline 0 \\ \stackrel{\rightharpoonup}{c} \\ \hline \end{array}$ | $\dot{c}$ |  | $\begin{gathered} i \\ i \\ \vdots \\ 0 \\ \hline \end{gathered}$ |  | $\stackrel{-}{\infty}$ |
| $\stackrel{m}{v}$ | $\begin{array}{\|c} \infty \\ \stackrel{\infty}{\square} \\ \hline \end{array}$ |  |  |  |  | へ－0． |
| $\begin{aligned} & \text { N } \\ & \text { vin } \end{aligned}$ |  | $\left\lvert\, \begin{aligned} & \mathrm{O} \\ & \mathbf{N} \end{aligned}\right.$ |  |  |  | $\stackrel{\infty}{\circ}$ |
| $\stackrel{\sim}{\text { V }}$ | $\left\|\begin{array}{l} \stackrel{n}{n} \\ \dot{\sim} \end{array}\right\|$ | $3$ | $\left.\begin{array}{c} 3 \\ 0 \\ 0 \end{array}\right)$ |  |  | （1） |
| $\bar{i}$ | $\begin{array}{\|c} \infty \\ \stackrel{\infty}{\infty} \\ \underset{\sim}{2} \\ \hline \end{array}$ | $\underset{\substack{\mathrm{j}} \underset{\mathrm{~N}}{\mathrm{O}}}{ }$ | $\underset{\sim}{\mathrm{S}} \underset{\sim}{\underset{\sim}{\sim}}$ | － |  | $\stackrel{\wedge}{\infty}$ |
| $\begin{aligned} & \stackrel{\circ}{\circ} \\ & \stackrel{\rightharpoonup}{v} \end{aligned}$ | $\begin{aligned} & \mathrm{c} \\ & \stackrel{\dot{m}}{ } \\ & \hline \end{aligned}$ | $\dot{d}$ | $\begin{array}{c\|c\|c} \substack{\infty \\ \underset{\sim}{2} \\ \hline} \\ \hline \end{array}$ |  | ọ. | $\stackrel{+}{\text {＋}}$ |
| $\begin{aligned} & \text { İ } \\ & \text { i } \end{aligned}$ | $$ |  |  | $\begin{array}{\|c\|} \hline \infty \\ \infty \\ \infty \\ \hline \end{array}$ | $\stackrel{\sim}{\infty}$ | ¢ |
| $\begin{aligned} & \text { No } \\ & \underset{\sim}{\mathrm{V}} \end{aligned}$ | $\begin{aligned} & \dot{\sigma} \\ & \hline \mathbf{o} \\ & \hline \end{aligned}$ |  |  |  |  | $\stackrel{\infty}{\square}$ |
| $\begin{aligned} & \bar{\circ} \\ & \stackrel{\rightharpoonup}{v} \end{aligned}$ |  |  |  |  |  | － |
| v | $\begin{array}{\|c} \hline \mathrm{O} \\ \hline \mathbf{c} \\ \hline \end{array}$ |  |  |  |  | $\stackrel{セ}{+}$ |
|  | $\underset{\text { ¢ }}{\text { ¢ }}$ | O | $\bigcirc$ | 0 |  | ¢ |
|  |  |  |  |  |  |  |

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database Continued

| Constituent | Land use | < | <0.01 | <0.02 | $<0.03$ | <0.05 | <0.06 | <0.1 | <0.2 | <0.3 | <0.5 | <1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{NO} 2+\mathrm{NO} 3(\mathrm{mg} / \mathrm{L})$ | RE | 16.67 |  | 8.33 | 8.33 |  |  | 37.50 | 4.17 |  |  | 25.00 |
|  | CO | 25.00 |  | 12.50 |  |  |  | 50.00 |  | 12.50 |  |  |
|  | ID | 6.25 |  | 6.25 |  | 6.25 |  | 43.75 | 6.25 | 6.25 |  | 25.00 |
|  | OP |  |  |  |  |  |  | 100.00 |  |  |  |  |
|  | FW |  |  |  |  |  |  |  |  | 100.00 |  |  |
|  | TOTAL | 15.85 | 1.22 | 8.54 | 4.88 | 4.88 | 1.22 | 35.37 | 4.88 | 6.10 | 3.66 | 13.41 |


| Constituent | Land <br> use | $<$ | $<0.01$ | $<0.05$ | $<0.1$ | $<0.2$ | $<0.28$ | $<0.3$ | $<0.5$ | $<1$ | $<1.5$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TKN (mg/L) | RE | 25.81 | 3.23 |  | 3.23 | 19.35 |  |  | 25.81 | 19.35 | 3.23 |
|  | CO | 25.00 | 8.33 |  |  | 16.67 |  |  | 25.00 | 8.33 | 16.67 |
|  | ID | 16.67 | 5.56 |  |  | 11.11 |  |  | 22.22 | 38.89 | 5.56 |
|  | OP |  |  |  |  |  |  |  | 100.00 |  |  |
|  | FW |  |  |  | 50.00 |  |  |  |  | 50.00 |  |
|  | TOTAL | 15.71 | 2.86 | 0.71 | 2.86 | 30.71 | 0.71 | 7.86 | 22.14 | 13.57 | 2.86 |


| Constituent | $\begin{aligned} & \text { Land } \\ & \text { use } \end{aligned}$ | < | <0.001 | <0.01 | <0.016 | <0.02 | <0.03 | <0.04 | <0.05 | <0.06 | <0.1 | <0.12 | <0.15 | <0.2 | $<0.5$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Dissolved Phosphorus$(\mathrm{mg} / \mathrm{L})$ | RE | 41.03 |  | 0.85 | 1.71 | 11.97 |  | 1.71 | 19.66 | 0.85 | 12.82 | 0.85 |  |  | 8.55 |
|  | CO | 37.70 |  | 6.56 | 1.64 | 18.03 | 1.64 |  | 11.48 | 1.64 |  | 3.28 |  |  | 18.03 |
|  | ID | 36.59 |  |  |  | 4.88 |  |  | 29.27 |  | 12.20 |  |  | 2.44 | 14.63 |
|  | OP |  |  | 11.11 |  |  |  |  | 22.22 |  | 11.11 |  |  |  | 55.56 |
|  | FW |  |  | 100.00 |  |  |  |  |  |  |  |  |  |  |  |
|  | TOTAL | 34.24 | 0.27 | 3.80 | 2.17 | 13.04 | 1.90 | 0.82 | 14.40 | 0.82 | 14.95 | 0.82 | 1.09 | 1.09 | 10.60 |


| $\stackrel{\bullet}{\square}$ |  |  | Nơ Nion |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\stackrel{n}{\stackrel{n}{v}}$ |  |  |  |  | $\stackrel{\circ}{6}$ |
| $\begin{aligned} & \stackrel{N}{\dot{v}} \\ & \hline \end{aligned}$ |  | - |  |  | $\stackrel{+}{\text { ¢ }}$ |
| $\stackrel{\rightharpoonup}{v}$ | $\begin{aligned} & 8 \\ & \hline \mathbf{i} \\ & \hline \end{aligned}$ |  |  |  | - |
| $\begin{aligned} & \circ \\ & \stackrel{\circ}{\mathrm{v}} \end{aligned}$ |  |  | - |  | $\stackrel{\square}{\text { ¢ }}$ |
| $\begin{aligned} & \stackrel{\circ}{\circ} \\ & \stackrel{\rightharpoonup}{v} \end{aligned}$ | $\begin{array}{\|l\|l\|} \hline 0 \\ \hline 0 \\ \hline \end{array}$ |  | - |  | $\stackrel{\infty}{\sim}$ |
| $\begin{aligned} & \text { to } \\ & \stackrel{\rightharpoonup}{i} \\ & \hline \end{aligned}$ | m |  |  |  | $\stackrel{0}{0}$ |
| $\begin{aligned} & \hline \stackrel{0}{\mathrm{v}} \\ & \hline \end{aligned}$ |  |  |  |  | \% |
| $\begin{aligned} & \text { No } \\ & \text { iv } \end{aligned}$ |  | - | $?$ |  | へị |
| $\begin{aligned} & \bar{\sigma}_{\dot{v}} \\ & \hline \end{aligned}$ | $\begin{array}{\|l\|l\|l\|l\|l\|} \hline \stackrel{y}{0} \\ \hline \end{array}$ | $\dot{S}$ |  |  | $\stackrel{\sim}{\substack{++}}$ |
| $\checkmark$ |  | - | - |  | - |
| $\stackrel{\text { coid }}{\text { ¢ }}$ | ¢ | 0 | 00 | 3 | 容 |
|  |  |  |  |  |  |

Table 12. Percentages of Non-detected Values for Different Reported Detection Limits by Land Use and for the Total Database Continued

| Constituent | Land use |  | <0.45 |  | <1.5 | <2 | <4 | <5 | <7 | <8 | <10 | <15 | <20 | <25 | <40 | <41 | <60 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE |  | 0.76 | 1.53 | 1.53 | 3.82 | 0.76 | 19.85 | 0.76 | 0.76 | 41.98 | 2.29 | 25.95 |  |  |  |  |
|  | CO | 3.57 |  |  |  | 7.14 |  | 14.29 |  |  | 57.14 |  | 17.86 |  |  |  |  |
|  | ID |  |  |  |  | 9.30 |  | 9.30 |  | 4.65 | 18.60 |  | 48.84 | 2.33 | 2.33 | 2.33 | 2.33 |
|  | OP | 20.00 |  |  |  |  |  |  |  |  | 70.00 |  | 10.00 |  |  |  |  |
|  | FW |  |  |  |  |  |  |  |  |  | 100.00 |  |  |  |  |  |  |
|  | TOTAL | 2.62 | 0.29 | 0.87 | 0.58 | 4.94 | 0.87 | 14.24 | 0.29 | 1.16 | 39.53 | 0.87 | 32.56 | 0.29 | 0.29 | 0.29 | 0.29 |


| Constituent | Land use | < | <0.2 | <0.65 | <0.7 | <1 |  |  | <2 | <2.5 | <3 | <4 | <5 | <6 | <7 | <9 | <10 | <15 | <20 | <25 | <30 | <40 | <42 | <50 | <53 | <55 | <60 | <100 | <200 | <250 | <500 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE | 10.6 | 0.4 | 0.4 |  | 0.4 |  |  | 2.2 | 4.0 | 9.3 | 1.3 | 21.7 | 0.4 |  | 7.5 | 5.8 | 1.3 | 4.4 | 5.8 | 0.9 | 3.5 | 0.4 | 8.9 |  | 0.9 | 1.3 | 5.3 | 3.1 |  |  |
|  | CO | 38.2 |  |  | 1.8 | 1.8 |  |  | 3.6 | 1.8 |  | 1.8 | 9.1 |  |  | 1.8 | 1.8 |  |  |  | 3.6 | 14.6 | 3.6 | 1.8 |  |  |  | 14.6 |  |  |  |
|  | ID | 17.5 |  |  |  | 3.1 | 1.0 | 1.0 | 3.1 |  | 3.1 |  | 13.4 |  |  | 8.3 | 2.1 | 2.1 | 8.3 | 5.2 |  | 3.1 | 1.0 | 9.3 | 1.0 | 3.1 | 2.1 | 10.3 |  | 1.0 | 1.0 |
|  | OP | 11.5 | 3.9 |  |  |  |  |  | 3.9 |  | 7.7 |  | 11.5 |  |  |  |  |  |  |  |  | 19.2 |  | 19.2 |  |  |  | 23.1 |  |  |  |
|  | FW |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | TOTAL | 14.7 | 0.3 | 0.2 | 0.2 | 6.2 | 0.2 | 0.2 | 2.6 | 1.7 | 4.9 | 3.3 | 12.4 | 0.2 | 2.0 | 6.41 | 14.3 | 1.1 | 2.9 | 4.3 | 0.9 | 3.6 | 0.6 | 6.1 | 0.3 | 1.1 | 0.9 | 6.4 | 2.0 | 0.2 | 0.2 |



Total lead had the largest number of different detection limits ( 31 in total) with $<10 \mu \mathrm{~g} / \mathrm{L}$ as the most frequent censored observation at $14.3 \%$. The constituent with the lowest number of detection limits was TDS, with four levels: $<1,<5,<6 \mathrm{and}<10 \mathrm{mg} / \mathrm{L}$. Less than $5 \mathrm{mg} / \mathrm{L}$ was the most common reported censored TDS observation occurring $55 \%$ of the time.

## Expected Percentages of Observations at Different Levels of Detection

There are different approved methods to calculate the concentration of a specific constituent in a water sample. Standard Methods for the Examination of Water and Waste Water (APHA 1995 and more recent) lists several approved methods for the detection of many of these constituents. The choice of methods presents a problem as these methods have varying features and costs. The objective is usually to select a method with a detection limit that results in useable data for most samples.

The distribution of the data, including the non-detected values, can be used to estimate the percentage of observations that will be detected using different analytical methods. Table 13 shows the expected percentage of observations below a specific detection limit for each of these constituents using the cumulative density function for each constituent and land use. For example, if a stormwater sample is collected at a freeway site and the detection limit of the conductivity method is $100 \mu \mathrm{~S} / \mathrm{cm}$, about $51 \%$ of the observations will be not-detects.

Table 13. Percentages of Observations below Specific Concentrations

| Constituent | Land | Percentage of observations smaller than |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{2 0} \mathbf{\mu \mathbf { S } / \mathbf { c m }}$ | $\mathbf{1 0 0} \mathbf{\mu \mathbf { S } / \mathbf { c m }}$ | $\mathbf{2 0 0} \boldsymbol{\mu \mathbf { S } / \mathbf { c m }}$ | $\mathbf{2 0 0 0} \mathbf{\mu \mathbf { S } / \mathbf { c m }}$ |
| Conductivity <br> ( $\mu \mathrm{S} / \mathrm{cm})$ | RE | 0 | 54 | 84 | 99 |
|  | CO | 0 | 39 | 82 | 100 |
|  | ID | 0 | 26 | 72 | 100 |
|  | OP | - | - | - | - |
|  | FW | 0 | 51 | 85 | 100 |
|  | TOTAL | 0 | 39 | 73 | 99 |


| Constituent | Land use | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $1 \mathrm{mg} / \mathrm{L}$ | $4 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ | 160 mg/L | $2500 \mathrm{mg} / \mathrm{L}$ |
| Hardness (mg/L) | RE | 0 | 0 | 5 | 98 | 100 |
|  | CO | 0 | 4 | 7 | 91 | 100 |
|  | ID | 0 | 0 | 3 | 95 | 100 |
|  | OP | - | - | - | - | - |
|  | FW | 0 | 0 | 2 | 96 | 100 |
|  | TOTAL | 0 | 0.1 | 3 | 94 | 100 |


| Constituent | Land use | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $0.5 \mathrm{mg} / \mathrm{L}$ | $1 \mathrm{mg} / \mathrm{L}$ | $2 \mathrm{mg} / \mathrm{L}$ | $5 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| Oil and Grease (mg/L) | RE | 2 | 19 | 31 | 75 | 91 |
|  | CO | 1 | 11 | 23 | 64 | 87 |
|  | ID | 1 | 20 | 31 | 66 | 86 |
|  | OP | - | 74 | - | - | - |
|  | FW | 2 | 5 | 5 | 55 | 75 |
|  | TOTAL | 0.3 | 17 | 29 | 67 | 84 |

Table 13. Percentages of Observations below Specific Concentrations (continued)

| Constituent | Land use | Percentage of observations smaller than |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | $1 \mathrm{mg} / \mathrm{L}$ | $5 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| Total Dissolved Solids (mg/L) | RE | 0 | 0.8 | 1.5 |
|  | CO | 0 | 0 | 0.5 |
|  | ID | 0 | 0 | 2 |
|  | OP | 0 | 0 | 0 |
|  | FW | 0 | 0 | 0 |
|  | TOTAL | 0.1 | 0.7 | 1.5 |


| Constituent | Land use | Percentage of observations smaller than |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | $1 \mathrm{mg} / \mathrm{L}$ | $5 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| Total Suspended Solids (mg/L) | RE | 0.2 | 4 | 11 |
|  | CO | 0.2 | 3 | 9 |
|  | ID | 0.2 | 3 | 5 |
|  | OP | 0 | 11 | 23 |
|  | FW | 0 | 2 | 2 |
|  | TOTAL | 0.2 | 3 | 7 |


| Constituent | Land | Percentage of observations smaller than |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{1} \mathbf{~ m g} / \mathbf{L}$ | $\mathbf{2 ~ m g} / \mathbf{L}$ | $\mathbf{5} \mathbf{~ m g} / \mathbf{L}$ |
| $\mathrm{BOD}_{5}(\mathrm{mg} / \mathrm{L})$ | RE | 0.2 | 2 | 18 |
|  | CO | 0.2 | 1 | 16 |
|  | ID | 0.2 | 3 | 18 |
|  | OP | 2 | 11 | 55 |
|  | FW | 0 | 0 | 31 |
|  | TOTAL | 1 | 3 | 22 |


| Constituent | Land use | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $0.7 \mathrm{mg} / \mathrm{L}$ | $1 \mathrm{mg} / \mathrm{L}$ | $5 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ | $20 \mathrm{mg} / \mathrm{L}$ |
| COD (mg/L) | RE | 0 | 0.4 | 1 | 3 | 9 |
|  | CO | 0 | 1 | 2 | 3 | 7 |
|  | ID | 0 | 0 | 0.5 | 2 | 7 |
|  | OP | 0 | 0 | 0 | 7 | 37 |
|  | FW | 0 | 0 | 1 | 4 | 7 |
|  | TOTAL | 0 | 0.2 | 2 | 5 | 13 |


| Constituent | Land | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{0 . 0 1} \mathbf{~ m g / L}$ | $\mathbf{0 . 0 5} \mathbf{~ m g / L}$ | $\mathbf{0 . 1} \mathbf{~ m g / L}$ | $\mathbf{0 . 2} \mathbf{~ m g / L}$ | $\mathbf{0 . 5} \mathbf{~ m g / L}$ |
| Ammonia (mg/L) | RE | 0 | 3 | 12 | 36 | 71 |
|  | CO | 0 | 2 | 9 | 28 | 53 |
|  | ID | 0 | 1 | 7 | 21 | 57 |
|  | OP | 0 | 11 | 15 | 22 | 93 |
|  | FW | 0 | 0 | 5 | 20 | 27 |
|  | TOTAL | 0.1 | 2 | 10 | 37 | 65 |

Table 13. Percentages of Observations below Specific Concentrations (continued)

| Constituent | Land use | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $0.01 \mathrm{mg} / \mathrm{L}$ | $0.05 \mathrm{mg} / \mathrm{L}$ | 0.1 mg/L | $0.2 \mathrm{mg} / \mathrm{L}$ | $0.5 \mathrm{mg} / \mathrm{L}$ |
| $\mathrm{NO} 2+\mathrm{NO} 3$ (mg/L) | RE | 0 | 2 | 5 | 11 | 40 |
|  | CO | 0 | 1 | 4 | 11 | 40 |
|  | ID | 0 | 2 | 6 | 11 | 31 |
|  | OP | 0 | 0 | 18 | 21 | 50 |
|  | FW | 0 | 0 | 0 | 28 | 72 |
|  | TOTAL | 0 | 2 | 4 | 10 | 40 |


| Constituent | Land | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{0 . 0 1} \mathbf{~ m g} / \mathbf{L}$ | $\mathbf{0 . 0 5} \mathbf{~ m g / L}$ | $\mathbf{0 . 1} \mathbf{~ m g} / \mathbf{L}$ | $\mathbf{0 . 2 ~ \mathbf { ~ m g } / \mathbf { L }}$ | $\mathbf{0 . 5} \mathbf{~ m g / L}$ |
| TKN (mg/L) | RE | 0.1 | 0.1 | 0.5 | 2 | 6 |
|  | CO | 0.2 | 0.2 | 0.7 | 2 | 6 |
|  | ID | 0.2 | 0.5 | 0.7 | 2 | 8 |
|  | OP | 0 | 0 | 0 | 0 | 44 |
|  | FW | 0 | 0 | 2 | 2 | 6 |
|  | TOTAL | 0.1 | 0.2 | 0.6 | 2 | 10 |


| Constituent | Land | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{0 . 0 1 ~ \mathbf { m g } / \mathbf { L }}$ | $\mathbf{0 . 0 2} \mathbf{~ m g / L}$ | $\mathbf{0 . 0 5} \mathbf{~ m g / L}$ | $\mathbf{0 . 1} \mathbf{~ m g / L}$ | $\mathbf{0 . 5} \mathbf{~ m g / L}$ |
| Dissolved <br> Phosphorus (mg/L) | RE | 0.3 | 3.5 | 11 | 32 | 93 |
|  | CO | 1 | 6 | 21 | 48 | 91 |
|  | ID | 0.3 | 2.2 | 16 | 46 | 95 |
|  | OP | 2 | 7 | 23 | 45 | 93 |
|  | FW | 5 | 5 | 5 | 14 | 82 |
|  | TOTAL | 0.7 | 4.5 | 17.5 | 44.5 | 94 |


| Constituent | Land use | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $0.01 \mathrm{mg} / \mathrm{L}$ | $0.02 \mathrm{mg} / \mathrm{L}$ | $0.05 \mathrm{mg} / \mathrm{L}$ | 0.1 mg/L | $0.5 \mathrm{mg} / \mathrm{L}$ |
| Total Phosphorus (mg/L) | RE | 0.2 | 0.4 | 1.5 | 10 | 28 |
|  | CO | 0.2 | 0.6 | 3 | 16 | 82 |
|  | ID | 0 | 0.2 | 3 | 14 | 74 |
|  | OP | 2 | 2 | 11 | 24 | 80 |
|  | FW | 0 | 0 | 0 | 3 | 83 |
|  | TOTAL | 0.1 | 0.5 | 3 | 12 | 78 |


| Constituent | Land | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{2 \mu \mathbf { g } / \mathbf { L }}$ | $\mathbf{5} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{1 0} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{2 0} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{4 0} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ |
| Total Copper $(\mu \mathrm{g} / \mathrm{L})$ | RE | 2.3 | 14 | $\mathbf{4 4}$ | 76 | 92 |
|  | CO | 0.7 | 6 | 26 | 58 | 84 |
|  | ID | 1.2 | 6 | 16 | 46 | 75 |
|  | OP | 0 | 32 | 54 | 73 | 92 |
|  | FW | 0 | 0 | 8 | 26 | 58 |
|  | TOTAL | 1.4 | 9 | 31 | 63 | 85 |

Table 13. Percentages of Observations below Specific Concentrations (continued)

| Constituent | Land | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{1} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{3} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{5} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{1 0} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{5 0} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L})$ | RE | 2 | 14 | 28 | 47 | 88 |
|  | CO | 0.6 | 3 | 8 | 23 | 80 |
|  | ID | 0.7 | 7 | 12 | 24 | 72 |
|  | OP | 12 | 21 | 33 | 38 | 76 |
|  | FW | 0 | 3 | 9 | 22 | 72 |
|  | TOTAL | 2 | 9 | 17 | 36 | 82 |


| Constituent | Land | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | use | $\mathbf{1} \boldsymbol{\mu \mathbf { g } / \mathbf { L }}$ | $\mathbf{2} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{5} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{1 0} \boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ | $\mathbf{2 0} \mathbf{\mu \mathbf { g } / \mathbf { L }}$ |
| Total Nickel $(\mu \mathrm{g} / \mathrm{L})$ | RE | 1 | 6 | 3 | 55 | 91 |
|  | CO | 0.5 | 3 | 29 | 56 | 92 |
|  | ID | 0 | 2 | 12 | 33 | 64 |
|  | OP | 0 | 30 | 39 | 39 | 73 |
|  | FW | 0 | 1 | 19 | 55 | 84 |
|  | TOTAL | 0.6 | 5 | 26 | 52 | 84 |


| Constituent | Land use | Percentage of observations smaller than |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $5 \mu \mathrm{~g} / \mathrm{L}$ | $10 \mu \mathrm{~g} / \mathrm{L}$ | $20 \mu \mathrm{~g} / \mathrm{L}$ | $100 \mu \mathrm{~g} / \mathrm{L}$ | $200 \mu \mathrm{~g} / \mathrm{L}$ |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | RE | 1 | 3 | 7 | 65 | 87 |
|  | CO | 0 | 0.2 | 1 | 28 | 51 |
|  | ID | 0.2 | 0.7 | 1 | 24 | 48 |
|  | OP | 5 | 25 | 35 | 85 | 92 |
|  | FW | 1 | 2 | 3 | 20 | 51 |
|  | TOTAL | 0.6 | 2 | 4 | 44 | 73 |

Appendix D describes the methods used to analyze censored observations for each constituent. Based on the results presented in Table 13 and these methods, it is possible to estimate the percentage of non-detected observations that can be obtained by constituent and land use. For example, the most frequently reported non-detected ammonia detection limit was $0.2 \mathrm{mg} / \mathrm{L}$. About $37 \%$ of the detected and non-detected observations were located below this detection limit. One of the EPA approved methods to measure ammonia has a detection limit close to $0.02 \mathrm{mg} / \mathrm{L}$. If this method was commonly used, the number of non-detected ammonia observations would have been significantly reduced. This is especially evident for metals analyses. Many commercial laboratories use ICP (inductively coupled plasma) procedures for heavy metals, as it is an approved method and generally more efficient than older atomic absorption methods using a graphite furnace. Unfortunately, standard ICP units have greatly reduced sensitivities compared to graphite furnace methods. When filtered heavy metals are to be analyzed, graphite furnace (or ICPmass spec) methods should be used. It is important that the person conducting a stormwater monitoring program take care in specifying the analytical methods to be used to ensure that most of the data will be usable. Of course, other factors, besides detection limits, must also be considered when selecting analytical methods, including sample preparation, sample storage limits, sample volume needed, safety, cost, disposal problems associated with wastes, interferences, and comparisons with other methods, etc. Burton and Pitt (2002) present a review of many alternative analytical methods that are suitable for stormwater sample analyses.

## Effects of Non-detected Observations on Calculating Mean and Standard Deviation Values

The selection of the proper procedure to deal with non-detected values is not an easy task. One option is to ignore the non-detected values and make a statement indicating the percentage of non-detected values found in the dataset. The problem arrives when it is desired to calculate the mean and standard deviation values of a dataset. The presence of non-detected values can strongly bias these parameters, depending on their prevalence. Three methods for dealing with non-detected values were explored during this research: 1) Ignore them; 2) Estimate them with the

Cohen's multi level MLE method for left censored data (NCASI 1995); and 3) replace them with half of the detection limit. In cases were Cohen's method could not be used (i.e. when only two values were detected), half of the detection limit was used as the estimated value to replace the non-detected observations.

Appendix D shows the results for each constituent and land use using the three substitution methods. In general, it was observed that if the censored data were deleted, the mean of the constituent was increased compared to the case where the non-detected values were replaced by half of the detection limit. The same pattern was observed for the standard deviation calculations. The behavior for the coefficient of variation was opposite: the coefficient of variation was reduced when the censored observations were deleted.

When the frequencies of the censored observations were lower than $5 \%$, the means, standard deviations and coefficients of variation were almost identical when the censored observations were replaced by half of the detection limit, or estimated using Cohen's Method. As the percentage of non-detected values increases, replacing the censored observation by half of the detection limit instead of estimating them using the Cohen's maximum likelihood method produces lower means and larger standard deviations.

## Effects on Mean, Median and Coefficient of Variation Values at Different Percentages of Censored Observations

As noted above, when the percentage of detected values is high, there are minimal changes in the calculated means, standard deviations, and coefficients of variation for any of the replacement methods. In this discussion, the ratios of the calculated values using the different methods for different frequencies of detection are examined. This analysis identifies the sensitivity of the detection frequencies for each substitution method.

The first task was to evaluate the effect of the substitutions and detection frequencies on the calculated means. When the percentage of detected values is close to $100 \%$, all of the substitution methods produce the same mean, as expected. As the percentage of non-detected values increases, the Cohen's estimated values and half of detection limit methods produces smaller means than if ignored.

Figure 15 is a scatter plot of both ratios (Cohen estimated/ignore and half of the detection limit/ignore) of the calculated mean values. If the scatter plot values formed a line near the 1.0 ratio value, then the "ignore" and the other option would be accurate. If the scatter plot values formed the same line for both of the sets of ratios, then either substitution method would be accurate. The regression equation 3.1 for the Cohen estimated/ignore ratio of calculated mean values has a coefficient of determination of almost $93 \%$. The coefficients in the equation are significant, with a probability that the coefficients are equal to zero smaller than 0.0001 .

$$
\begin{equation*}
\text { Ratio Mean }(\text { Estimated/Ignore })=0.316+0.0068 * D \tag{3.1}
\end{equation*}
$$

Where D is the percentage of detected values ( 0 to 100 ).


Figure 15. Effects on the mean when using random estimated values versus ignoring the nondetected observations, at different percentage of detected values

For percentages of detected values smaller than $60 \%$, the ratios are located away from the line formed by the other observations. The residual plot of the regression indicates those observations that are most affecting the departure from the regression line. Six observations are considered influential in this plot: oil and grease in open space (most influential), residential and industrial land uses, plus ammonia and lead in open space land uses. The Cook's distance procedure was used to remove the overly influential points in the regression. After removing the influential observations the final regression is therefore:

$$
\begin{equation*}
\text { Ratio Mean }(\text { Estimated/Ignore })=0.248+0.0075 * D \tag{3.2}
\end{equation*}
$$

Equation 3.2 indicates that a stormwater dataset having $30 \%$ non-detectable observations would have an expected reduction in the calculated mean of $23 \%$ when the censored data is appropriately estimated instead of being ignored. The standard deviation of the residuals is 0.014 . The coefficient of determination in this case was higher than $96 \%$ with no potential or influential points. This equation can be used to estimate the mean of the distribution for data sets with percentages of detected values higher than $60 \%$. When the non-detected observations are replaced by half of the detection limit, the coefficient of determination was reduced to $92 \%$ of the actual value. Equation 3.3 describes the relationship between the ratio of the means and the percentage of detected observations.

$$
\begin{equation*}
\text { Ratio Mean (Half Detection/Ignore) }=0.250+0.0075 * D \tag{3.3}
\end{equation*}
$$

From the regression of the ratios "estimated/ignore" and "half detection/ignore," replacing by half of the detection limit, or estimating the censored observations using Cohen's method, will produce the same results when the percentage of detected observations is larger than $80 \%$.

The effects on the median are similar to those observed in the mean. When the non-detected values are estimated with Cohen's method instead of ignoring the non-detected values, the regression of the coefficient of determination reduces to $86 \%$.


Figure 16. Effect of ignoring the non-detected observations on the median
Equation 3.4 shows the estimated regression line for the median case.

$$
\begin{equation*}
\text { Ratio Median (Estimate /Ignore) }=-0.326+0.0134^{*} D \tag{3.4}
\end{equation*}
$$

This equation is valid for percentage of detected observations higher than $70 \%$. A reduction of $40 \%$ in the median value is expected in a $30 \%$ censored dataset when the non-detected observations are estimated using Cohen's method instead of being ignored. The standard deviation of the residuals for this equation is 0.05 .

When the censored observations are replaced by half of the detection limit, the coefficient of determination is about $73 \%$. The regression equation for the ratio of the median is therefore not as good in explaining the variability as it was for the mean.

Equation 3.5 shows the calculated regression line for the median when the non-detected values are replace by half of the detection limit.

Ratio Median (Half Detection/Ignore) $=-0.195+0.012 * D$

This equation is valid when the percentage of detected observations is higher than $70 \%$. Replacing the censored observations by half of the detection limit has the same effect on the median as estimating them using Cohen's method, except for dissolved and total phosphorus in open space and lead in residential land uses.

The effects on the calculated standard deviation values also indicate a good correlation between the level of detected observations and the ratio between the "estimate the non-detected or ignore them" values. Figure 17 shows the scatter plot of the median values as a function of the percentage of detected observations. Equation 3.6 presents the estimated regression line of these data.

Ratio Standard Deviation (Estimate/Ignore) $=0.68+0.003226^{*} D$


Figure 17. Effect of ignoring the non-detected observations in the median

The regression has a low coefficient of determination (56\%) compared to the prior regressions. Oil and grease at freeway sites was considered unusual according to its Cook's distance. The data for this case was examined and no reason was found to eliminate it from the analysis. It was also observed that $\mathrm{BOD}_{5}$ in commercial land use areas had 3 right-censored observations. Because the Cohen method must be used with left censored observations, these data were eliminated from this analysis. Observations where the percentage of detection was smaller than $70 \%$ were not included. Equation 3.7 shows the estimated regression line for those constituents with more than $60 \%$ detected observations.

This equation indicates that for a dataset with $30 \%$ censored observations, the standard deviation will be reduced by $9.5 \%$ when the non-detected observations are estimated instead of ignored. The standard deviation of the residuals is 0.023 . When the censored observations are replaced by half of the detection limits, the coefficient of determination and the equation coefficients were almost the same. Equation 3.8 presents the estimated regression equation for the standard deviation when the censored observations are replaced by half of the detection limits.

$$
\begin{equation*}
\text { Ratio Standard Deviation (Half Detection/Ignore) }=0.6778+0.00325 * D \tag{3.8}
\end{equation*}
$$

The last parameter examined was the coefficient of variation. The coefficient of determination (69\%) for the fitted regression equation was better than for the standard deviation regression, but not as high as for the median and mean regressions. The calculated regression equation is presented as equation 3.9

$$
\begin{equation*}
\text { Ratio Coefficient of Variation (Estimate/Ignore) }=1.53-0.0053 * D \tag{3.9}
\end{equation*}
$$

The standard deviation of the residuals is 0.033 . As the number of non-detected observations increases, the coefficient of variation also increases. The regression equation is valid for percentages of detected values higher than $70 \%$. For a data set with $30 \%$ censored observations, the expected coefficients of variation using Cohen's method will be $16 \%$ higher than if the non-detected values are ignored.

In the case that the censored observations are replaced by half of the detection limits, the coefficient of determination of the resulting equation (equation 3.10) is reduced to $58 \%$. Figure 18 shows the scatter plot for the ratios "estimated/ignore" and "half detection/ignore" for the coefficient of variation.

$$
\begin{equation*}
\text { Ratio Coefficient of Variation (Half Detection/Ignore) }=1.543-0.0054 * D \tag{3.10}
\end{equation*}
$$



Figure 18. Effect of ignoring the non-detected observations on the coefficient of variation

## Total Suspended Solids Analyses at Different Levels of Censoring

To evaluate the effect of the non-detected values in the mean and standard deviation observations at different levels of censoring, one of the constituents with low percentages of non-detected observation (TSS) was trimmed in the lower tail until reduced to $50 \%$ of the original distribution. All TSS observations were used during this analysis.

The results are similar to those observed during the analysis of the censoring observations within multiple constituents and land uses. Real mean, median and standard deviation are smaller than the calculated values when censored observations are ignored (Figure 19). The true coefficients of variation are larger than those calculated when the level of trimming is increased.

Table 14. Descriptive Statistics for TSS Truncated at Different Levels

|  |  |  | RATIO |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total number of samples | \% of original samples | Minimum concentration in set (mg/L) | Average | Median | Standard Deviation | Coefficient of Variation |
| 2025 | 100.00 | 3 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2015 | 99.51 | 3 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 98.52 | 4 | 0.99 | 0.98 | 0.99 | 1.01 |
| 1974 | 97.48 | 5 | 0.98 | 0.96 | 0.99 | 1.02 |
| 1954 | 96.49 | 6 | 0.97 | 0.95 | 0.99 | 1.02 |
| 1934 | 95.51 | 7 | 0.96 | 0.95 | 0.98 | 1.03 |
| 1914 | 94.52 | 8 | 0.95 | 0.93 | 0.98 | 1.03 |
| 1873 | 92.49 | 10 | 0.93 | 0.90 | 0.97 | 1.05 |
| 1833 | 90.52 | 11 | 0.91 | 0.87 | 0.96 | 1.06 |
| 1792 | 88.49 | 13 | 0.89 | 0.84 | 0.96 | 1.07 |
| 1752 | 86.52 | 15 | 0.87 | 0.81 | 0.95 | 1.08 |
| 1711 | 84.49 | 17 | 0.86 | 0.78 | 0.94 | 1.10 |
| 1671 | 82.52 | 18 | 0.84 | 0.74 | 0.93 | 1.11 |
| 1630 | 80.49 | 20 | 0.82 | 0.73 | 0.92 | 1.12 |
| 1589 | 78.47 | 22 | 0.80 | 0.71 | 0.91 | 1.14 |
| 1545 | 76.30 | 24 | 0.78 | 0.68 | 0.90 | 1.15 |
| 1496 | 73.88 | 26 | 0.76 | 0.65 | 0.89 | 1.17 |
| 1468 | 72.49 | 27 | 0.75 | 0.64 | 0.89 | 1.18 |
| 1428 | 70.52 | 29 | 0.73 | 0.63 | 0.88 | 1.20 |
| 1387 | 68.49 | 31 | 0.72 | 0.60 | 0.87 | 1.21 |
| 1347 | 66.52 | 33 | 0.70 | 0.58 | 0.86 | 1.23 |
| 1306 | 64.49 | 35 | 0.68 | 0.57 | 0.85 | 1.24 |
| 1266 | 62.52 | 37 | 0.67 | 0.55 | 0.84 | 1.26 |
| 1225 | 60.49 | 40 | 0.65 | 0.53 | 0.83 | 1.28 |
| 1185 | 58.52 | 42 | 0.63 | 0.52 | 0.82 | 1.29 |
| 1144 | 56.49 | 44 | 0.62 | 0.50 | 0.81 | 1.31 |
| 1104 | 54.52 | 47 | 0.60 | 0.47 | 0.80 | 1.33 |
| 1063 | 52.49 | 50 | 0.58 | 0.46 | 0.79 | 1.35 |
| 1023 | 50.52 | 52 | 0.57 | 0.44 | 0.78 | 1.37 |



Figure 19. Effect on the mean when TSS observations are truncated

The effect on the mean indicates that when only about $5 \%$ of the data is censored or is trimmed, the ratios "replace/ignore," "estimated/ignore," or "trimmed/total" observations produced the same results in the mean of the distribution. When the percentage of non-detected observations is increased, the ratios "estimate/ignore" and "half detection/ignore" are higher than the ratio "trimmed/complete" in the TSS distribution. This means that trimming the data set has a larger effect than when the observations are censored. This is explained because for the trimmed/complete ratios, all the censored observations were at one value. In the other case, several detection limits were used during the analysis.

In the previous discussion, it was observed that censored levels less than $30 \%$ can be used for predicting simple statistics describing the distribution. The previous figure indicates that levels of censoring close to $45 \%$ followed the trend indicated by the ratio "trimmed/complete." This indicates that even if the regression analysis was recommended for levels of non-detected values smaller than $30 \%$, they can be used for levels of censoring up to $45 \%$.

The effects on the medians are stronger than on the means. When the level of censored observations is close to $30 \%$, the ratio "trimmed/complete" is close to 0.6 , compared with 0.75 in the case of the mean (Figure 20). Levels of censoring around $5 \%$ do not show the straight-line pattern that was observed with the mean. The trend for censoring levels between 5 and $45 \%$ is similar for the "estimated/ignore" ratio; however the dispersion around the trend line is higher.


Figure 20. Effect on the median when the TSS dataset is truncated

The effect on the standard deviation of the trimming the TSS is similar to the effect in the mean (Figure 21). When the level of censoring is close to $30 \%$, the ratio "trimmed/complete" is close to 0.85 . The dispersion around the trend line is lower than in the median case. When the percentage of non-detected values is lower than $5 \%$, the ratios "estimated/ignore," "half detection/ignore," and "trimmed/complete" are almost the same. For levels of censored observations larger than $15 \%$, the differences among the ratios increase.


Figure 21. Effect on the standard deviation when the TSS dataset is truncated

The ratio of the effects on the calculated coefficients of variation has a different slope than the previous statistics. As in the mean case when the level of censoring is smaller than $5 \%$, a linear trend between the percentage of detected and the ratio was observed (Figure 22). When the percentage of censored observations is larger than $15 \%$, the differences among the three ratios increase.


Figure 22. Effect on the coefficient of variation when the TSS dataset is truncated

## Summary

The level of censoring observations in a dataset affects the calculated mean, median, standard deviation and coefficient of variation values. As the level of non-detected observations increase, the mean, median and standard deviation are larger than if the censored observations are detected. The opposite behavior is expected for the coefficient of variation. Different laboratories report different detection limits for the same constituents. In many cases, the detection limits are calculated by each laboratory based on their measured repeatability (precision) for a specific laboratory test. Using methods with low precision increases the percentage of non-detected values and the uncertainty of the real mean and standard deviation values.

Open space has the largest number of non-detected observations among land uses. The largest percentages of detected observations were observed in freeways and industrial land uses.

Estimating or replacing by half of the detection limit for levels of censoring smaller than 5\% does not have a significant effect on the mean, standard deviation and coefficient of variation values.

Substituting the censored observations by half of the detection limit produces smaller values than when using Cohen's maximum likelihood method. Replacing the censored observations by half of the detection limit is not recommended for levels of censoring larger than $15 \%$.

The censored observations in the database were replaced using estimated values using Cohen's maximum likelihood method for each site before the statistical tests. Because this method uses the detected observations to estimate the non-detected values, it is not very accurate, and therefore not recommended, when the percentage of censored observations is larger than $40 \%$. Table 14 shows those constituents having percentages of non-detected observations smaller than $40 \%$ for the three main land uses.

All the methods used in this chapter are approximations to calculate the EMC when censored observations are present. These problems would not exist if appropriate analytical methods were used to analyze the samples. It is very important to select analytical methods capable of detecting the desired range of concentrations in the samples in order to reduce the numbers of censored observations to acceptable levels. Table 3 XX summarizes the recommended minimum detection limits for various stormwater constituents to obtain manageable non-detection frequencies ( $<5 \%$ ). Some of the open space stormwater measurements (oil and grease and lead, for example), would likely have greater than $5 \%$ non-detects, even with the detection limits shown. The detection limits for filtered heavy metals would be substantially less than shown on this table.

## Table 3XX. Suggested Analytical Detection Limits for Stormwater Monitoring Programs to Obtain <5\% Non-detects

|  | Residential, commercial, industrial, freeway | Open Space |
| :---: | :---: | :---: |
| Conductivity | $20 \mu \mathrm{~S} / \mathrm{cm}$ | $20 \mu \mathrm{~S} / \mathrm{cm}$ |
| Hardness | $10 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| Oil and grease | $0.5 \mathrm{mg} / \mathrm{L}$ | $0.5 \mathrm{mg} / \mathrm{L}$ |
| TDS | $10 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| TSS | $5 \mathrm{mg} / \mathrm{L}$ | $1 \mathrm{mg} / \mathrm{L}$ |
| $\mathrm{BOD}_{5}$ | $2 \mathrm{mg} / \mathrm{L}$ | $1 \mathrm{mg} / \mathrm{L}$ |
| COD | $10 \mathrm{mg} / \mathrm{L}$ | $5 \mathrm{mg} / \mathrm{L}$ |
| Ammonia | $0.05 \mathrm{mg} / \mathrm{L}$ | 0.01 mg/L |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3}$ | 0.1 mg/L | $0.05 \mathrm{mg} / \mathrm{L}$ |
| TKN | 0.2 mg/L | 0.2 mg/L |
| Dissolved P | $0.02 \mathrm{mg} / \mathrm{L}$ | 0.01 mg/L |
| Total P | $0.05 \mathrm{mg} / \mathrm{L}$ | $0.02 \mathrm{mg} / \mathrm{L}$ |
| Total Cu | $2 \mu \mathrm{~g} / \mathrm{L}$ | $2 \mu \mathrm{~g} / \mathrm{L}$ |
| Total Pb | $3 \mu \mathrm{~g} / \mathrm{L}$ (residential $1 \mu \mathrm{~g} / \mathrm{L}$ ) | $1 \mu \mathrm{~g} / \mathrm{L}$ |
| Total Ni | $2 \mu \mathrm{~g} / \mathrm{L}$ | $1 \mu \mathrm{~g} / \mathrm{L}$ |
| Total Zn | $20 \mu \mathrm{~g} / \mathrm{L}$ (residential $10 \mu \mathrm{~g} / \mathrm{L}$ ) | $5 \mu \mathrm{~g} / \mathrm{L}$ |

# Chapter 4: Stormwater Quality Descriptions Using the Three Parameter Lognormal Distribution 

## Introduction

Knowing the statistical distribution of observed stormwater data is a critical step in data analysis. The selection of the correct statistical analyses tools is dependent on the data distribution, and many QA/QC operations depend on examining the distribution behavior. However, much data is needed for accurate determinations of the statistical distributions of the data, especially when examining unusual behavior. The comparison of probability distributions between different data subsets is also a fundamental method to identify important factors affecting data observations. Statistical analyses basically are intended to explain data variability by identifying significantly different subsets of the data. The remaining variability that can not be explained must be described. In all cases, accurate descriptions of the data probability distributions are needed. This chapter explores these distributions for the NSQD data.

The Nationwide Urban Runoff Program (NURP) evaluated the characteristics of stormwater discharges at 81 outfalls in 28 communities throughout the U.S. (EPA 1983). One of the conclusions was that most of the stormwater constituent concentration probability plots could be described using lognormal distributions. More recently, Van Buren (1997) also found that stormwater concentrations were best described using a lognormal distribution for almost all constituents, with the exception of some dissolved constituents that were better described with a normal distribution. Beherra (2000) also found that some stormwater constituent concentrations were better described using a lognormal distribution, while others were better described with gamma or exponential distributions. The constituents that were best described with a gamma distribution included total solids, total Kjeldahl nitrogen (TKN), total phosphorous, chemical oxygen demand (COD), barium and copper. The constituents that were best described with an exponential distribution included suspended solids, nitrates and aluminum. In both of these recent studies, fewer than 50 samples (collected at the same site) were available for evaluation.

During the research reported in this chapter, statistical tests were used to evaluate the log-normality of a selection of the constituents in the NSQD database. Statistical descriptions were obtained of each set of data including box and whisker and probability plots for each land use category and for the pooled dataset. It was found in almost all cases that the log-transformed data followed a straight line between the 5th and 95th percentile, as illustrated in Figure 26 for total dissolved solids (TDS) in residential areas.

For many statistical tests focusing on the central tendency (such as for determining the average concentration that is used for mass balance calculations), this may be a suitable fit. As an example, WinSLAMM, the Source Loading and Management Model (Pitt and Voorhees 1995), uses a Monte Carlo component to describe the likely variability of stormwater source flow pollutant concentrations using either lognormal or normal probability distributions for each constituent. However, if the extreme values are of importance (such as when dealing with the influence of many non-detectable values on the predicted concentrations, or determining the frequency of observations exceeding a numerical standard), a better description of the extreme values may be important.

## Probability Plot for Different Land Uses Total Dissolved Solids (mg/L)



Figure 26. Log-probability plot of total dissolved solids in residential land use

The NSQD underwent an extensive data evaluation process, including multiple comparisons of the all data values in the database to original documents. In some cases, data was available from the local agency in electronic form. These spreadsheets were reformatted to be consistent to the NSQD format. However, it was found that all of the submitted electronic data needed to be verified against original data sheets and reports. When reviewing the NSQD, it was assumed that some of the events in the upper and lower tails of the distributions were caused by errors, most likely due to faulty transcription of the data (such as mislabeling the units for heavy metals or nutrients as $\mathrm{mg} / \mathrm{L}$ instead of $\mu \mathrm{g} / \mathrm{L}$, for example). Unusual values were verified with the original reports and datasets. While some values (less than $5 \%$ of the complete dataset) were found to be in error and were corrected, most of the suspected values were found to be correct stormwater observations. Besides the targeted extreme values, many constituents were also examined in relationship to other related constituents (COD vs. BOD; total metal concentrations vs. dissolved metal concentrations; TKN vs. NH3; TDS vs. specific conductivity; SS vs. turbidity; etc) and unusual behavior was further checked and corrected, as necessary. In some cases, unusual values could not be verified and were therefore eliminated from the dataset, although this was very unusual. After the extensive QA/QC activities
and corrections were made to the NSQD, the next step was to conduct a sensitivity analysis to determine the effects of the remaining unusual high and low values on the probability distribution parameters.

## The Effects of Unusual High and Low Values on Probability Distribution Parameters

For this evaluation, 10,000 sets of 200 samples each were randomly generated following a lognormal distribution (1, 1 ), but having differing amounts of extreme values in each data set. For each set, the mean, variance and coefficient of variation were calculated. Two main factors were analyzed using these data: the extreme value factor and percentage of extreme values in each sample. The following percentages of extreme values were selected for evaluation: $0.5,1,5,10,25$ and $50 \%$. For each percentage of extreme values, the following factors were analyzed: $0.001,0.01,0.1,10,100,1.000,10,000,100,000$ and $1,000,000$. For example $(5 \%, 100)$ indicates that in each set, $5 \%$ of the data were increased by a factor of 100 . The coefficient of variation was then calculated for each set of data. The medians of the coefficients of variation for the 10,000 runs are shown in Figure 27 for each level of extreme values.


Figure 27. Effect of unusual values on the coefficient of variation (based on LN(1,1))

For a lognormal distribution $(1,1)$ the coefficient of variation is equal to one. Figure 27 shows how this original value is changed for different amounts of extreme values in the data sets, and for different factors in these extreme values. The horizontal axis represents the factor used in the extreme values. As an example, many of the incorrect extreme values observed in the NSQD for heavy metals were because the units were originally incorrectly reported as $\mathrm{mg} / \mathrm{L}$ in the submitted information, while the correct units were actually $\mu \mathrm{g} / \mathrm{L}$. This would be an extreme value factor of 1,000 . Extreme value factors of 10 were also fairly common and were associated with simple misplacements of decimal points in the data.

Figure 27 also shows that for small error factors $(0.1,0.01$ and 0.001$)$ there is not a large effect in the coefficient of variation for percentages smaller than $10 \%$. For larger percentages, the effect in the coefficient of variation is important. When $50 \%$ of the data are affected by an error factor of 0.01 , the coefficient of variation was increased by almost three times.

High extreme value factors can have a much more important effect on the coefficient of variation. When $10 \%$ of the data were increased by a factor of 10 , the coefficient of variation was increased almost three times. Notice that affecting $10 \%$ of the data by a factor of ten have almost the same effect as affecting $50 \%$ of the data by a factor of a hundredth. This effect is reduced when the percentage of elevated values in the dataset is smaller than $10 \%$.

For factors larger than a hundred, the effect on the coefficient of variation is much greater. Very low percentages of elevated values can increase the coefficient of variation by up to 15 times. For example, when only $0.5 \%$ of the sample is affected by a factor of a thousand, the coefficient of variation increases almost 12 times more than the correct value. As noted earlier this is important because it is not unusual to find reported values affected by a factor larger than a hundred (See Figure 26). Some of these values can be due to incorrect reporting units, but in many cases they were considered as valid observations because they were supported by similarly high values of other closely related constituents. For factors greater than 10,000 the multiplying value of the coefficient of variation remains stable at the maximum value obtained.

The above analyses indicate that in lognormal distributions, the presence of just a few unusual elevated values is important and can dramatically affect the reported coefficient of variation for the distribution of concentration. This observation is critical in the relatively common case where one or a very few observations are affected by a factor larger than a hundred. In the other extreme, factors smaller than one do not have a large impact on the reported coefficient of variation, except when the percentage of errors is greater than $50 \%$.

The effect of extreme values on the mean and standard deviation was also analyzed. Figure 28 shows the effect of the extreme values on calculated standard deviation. For large extreme value factors (larger than one) the standard deviation increases as the percentage of extreme values increases.


Figure 28. Effect of unusual values on the standard deviation (based on LN(1,1))

Percentages smaller than $25 \%$ do not have an important effect on the standard deviation for small extreme value factors. For a specific extreme value factor, changing the extreme value percentages from $0.5 \%$ to $50 \%$ increases the standard deviation close to 10 times.

The effect of the presence of extreme values on the distribution mean is shown in Figure 29. For small extreme value factors, the mean is reduced almost $80 \%$ when the extreme value percentage is close to $50 \%$. This is expected because in a lognormal distribution $(1,1)$ most of the values are located in the lower tail of the distribution. For extreme value occurrences less than $25 \%$, the mean value is reduced by less than $20 \%$.

Large extreme value factors have much larger effects on the distribution means. As the extreme value percentage increases, the calculated means also increase. If $0.5 \%$ of the values are affected by a factor of a hundred, the mean value is doubled. If $50 \%$ of the values are affected by the same factor, the mean values are increased by almost 50 times. For factors larger than a thousand, increasing the percentages of extreme values from $0.5 \%$ to $50 \%$ increases the mean values by up to two orders of magnitude.

These evaluations are important because it points out that for a lognormal distribution, the effects of few elevated values in the upper tail have a much greater effect on common statistics than unusual values in the lower tail. Many
stormwater researchers have focused on the lower tail, especially when determining how to handle the detection limits and unreported data. Stormwater constituents usually have unusual values in both tails of the probability distribution. It is common to delete elevated values from the observations assuming they are expendable "outliers". This practice is not recommended unless there is sufficient evidence that the observed values are a mistake. Actual elevated values can have a large effect on the calculated distribution parameters. If these are arbitrarily removed, the data analyses will likely be flawed.


Figure 29. Effect of unusual values on the mean (based on LN(1,1))

## Analysis of Lognormality of Stormwater Constituents Parameters

The goodness of fit of twenty nine stormwater constituent probability distributions was evaluated using the Kolmogorov-Smirnov test. Figure 30 shows how the test accepts or rejects the null hypothesis that the empirical and the estimated distributions are the same. If the null hypothesis is valid, then the constituent can be adequately represented by the lognormal distribution. The observations are sorted and a probability is assigned by its rank. The distribution generated by this ranking is known as the empirical distribution. The estimated distribution function is also compared on the same plot. The estimated distribution function is calculated with the mean and standard deviation of the original data. If the distance between the empirical and the estimated distributions is higher than a
critical value $\mathrm{d}_{\alpha}$ or $\mathrm{D}_{\text {max }}$, the hypothesis of lognormality is rejected. Notice in Figure 30 that the horizontal axis has a logarithmic scale.


Figure 30. Cumulative and empirical probability distributions of total copper for residential land use data (Goodness of fit test, Kolmogorov-Smirnov)

There are many options to assign probability to a data observation based on ranks. Most methods assign the probability as a percentage of the total range. The probability of the observation is calculated as its rank divided by the number of observations. Kottegoda (1998) suggested that for extreme event analysis, the plotting position can be calculated as:

$$
\begin{equation*}
p=\frac{i-0.5}{n} \tag{4.1}
\end{equation*}
$$

Where $p$ is the cumulative probability of the observation, $i$ is the rank of the observation and $n$ is the total number of observations. This plotting position was used for the analyses during this research because it does not set the probability of the largest observation as one.

In the Kolmogorov-Smirnov test, the null hypothesis is that the observed data follow a lognormal distribution. If the sample size is small, and the distance between the empirical and the observed distributions is smaller than the critical value $D_{\text {max }}$, the test is interpreted as "there is not enough evidence to reject the hypothesis that the distribution is lognormal." In most cases, the NSQD contains enough samples to be able to accept or reject the null hypothesis with acceptable levels of confidence and power.

The NSQD contains many factors for each sampled event that likely affect the observed concentrations. These include such factors as seasons, geographical zones, rain intensities, etc. These factors may affect the shape of the probability distribution. As more data become available, the critical value $D_{\text {max }}$ is reduced in the test. There will always be a specific number of samples that will lead to rejection of the null hypothesis because the maximum distance between the empirical and estimated probability distributions became larger than the critical value $D_{\max }$. The only way to evaluate the required number of samples in each category is using the power of the test. Power is the probability that the test statistic will lead to a rejection of the null hypothesis when it is false (Gibbons and Chakraborti 2003). Masey (1950) states that the power of the Kolmogorov-Smirnov test can be written as:

$$
\begin{equation*}
\text { power }=1-\operatorname{Pr}\left(\frac{-d_{\alpha} \pm \Delta \sqrt{n}}{\sqrt{F_{1}\left(x_{0}\right)\left(1-F_{1}\left(x_{0}\right)\right)}}<\frac{\left\{S_{n}\left(x_{0}\right)-F_{1}\left(x_{0}\right)\right\} \sqrt{n}}{\sqrt{F_{1}\left(x_{0}\right)\left(1-F_{1}\left(x_{0}\right)\right)}}<\frac{d_{\alpha} \pm \Delta \sqrt{n}}{\sqrt{F_{1}\left(x_{0}\right)\left(1-F_{1}\left(x_{0}\right)\right)}}\right) \tag{4.2}
\end{equation*}
$$

where:

| $\mathrm{d}_{\alpha}$ | $=$ | Dmax: critical distance at the level of significance $\alpha$ (confidence of the test), |
| :--- | :--- | :--- |
| $\mathrm{S}_{\mathrm{n}}$ | $=$ | Cumulative empirical probability distribution, |
| $\mathrm{F}_{1}$ | $=$ | Cumulative alternative probability distribution, |
| $\Delta$ | $=$ | Maximum absolute difference between the cumulative estimated probability <br> distribution and the alternative cumulative probability distribution. |

Massey (1951) also found that for large sample sizes, the power can be never be smaller than

$$
\begin{equation*}
\text { power }>1-\int_{2\left(\Delta \sqrt{n}-d_{\alpha} \sqrt{n}\right)}^{2\left(\Delta \sqrt{n}+d_{\alpha} \sqrt{n}\right)} \frac{1}{\sqrt{2 \pi}} e^{-\frac{t^{2}}{2}} d t \tag{4.3}
\end{equation*}
$$

This reduced expression can be used to calculate the number of samples required to reject the null hypothesis with a desired power. Figure 31 shows the power of the D test for $1 \%, 5 \%$, and $10 \%$ levels of confidence of the test for samples size larger than 35 (Massey 1951). For example, assume that the maximum distance between the alternative cumulative and the estimated cumulative probability distributions is 0.2 , and we want an $80 \%$ power ( 0.8 ) against the alternative at a $5 \%$ level of confidence. To calculate the number of required samples, we read that $\Delta(N)^{0.5}$ is 1.8 for a power of 0.8 and $5 \%$ level of confidence. Solving for $\mathrm{N}=(1.8 / 0.2)^{2}=81$ samples. If we want to calculate the number of samples when the difference between the alternative cumulative and the estimated cumulative probability function is 0.05 , with the same power and level of confidence, then 1,296 samples would be required. When the lines are very close together, it is obviously very difficult to statistically show that they are different, and many samples are needed.

Lower Bounds for the Power of the D test


Figure 31. Lower bounds for the power of the $D$ test for $\alpha=1 \%, 5 \%$ and $10 \%(N>35)$

The Kolmogorov-Smirnov test was used to indicate if the cumulative empirical probability distribution of the NSQD residential stormwater constituents can be adequately represented with a lognormal distribution. Table 19 shows the resulting power of the test for $\mathrm{D}=0.05$ and $\mathrm{D}=0.1$, when applied to selected constituents that had high levels of detection in residential land uses.

Table 19. Power of the Test When Applied to Selected Constituents in Residential Land Uses

| Constituent | N | Percentage Detected | $\begin{gathered} \Delta N^{0.5} \\ (\alpha=0.05) \end{gathered}$ | Power $\begin{gathered} (D=0.05, \\ \beta=5 \%) \end{gathered}$ | $\begin{gathered} \Delta N^{0.5} \\ (\alpha=0.1) \end{gathered}$ | Power $\begin{gathered} (D=0.1, \\ \beta=10 \%) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TDS (mg/L) | 861 | 99.2 | 1.46 | 0.60 | 2.92 | 1 |
| TSS (mg/L) | 991 | 98.6 | 1.56 | 0.65 | 3.12 | 1 |
| BOD (mg/L) | 941 | 97.6 | 1.52 | 0.65 | 3.04 | 1 |
| COD (mg/L) | 796 | 98.9 | 1.40 | 0.55 | 2.80 | 1 |
| NO2+NO3 (mg/L) | 927 | 97.4 | 1.50 | 0.60 | 3.00 | 1 |
| TKN (mg/L) | 957 | 96.8 | 1.52 | 0.65 | 3.04 | 1 |
| TP (mg/L) | 963 | 96.9 | 1.53 | 0.65 | 3.06 | 1 |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | 799 | 83.6 | 1.29 | 0.50 | 2.58 | 1 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | 788 | 71.3 | 1.19 | 0.40 | 2.38 | 1 |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | 810 | 96.4 | 1.40 | 0.55 | 2.80 | 1 |

Table 19 shows that the number of collected samples is sufficient to detect if the empirical distribution is located inside an interval of width 0.1 above and below the estimated cumulative probability distribution. If the interval is reduced to 0.05 , the power varies between 40 and $65 \%$. To estimate the interval width, 10 cumulative distributions of 1,000 random data points, having a lognormal $(1,1)$ distribution, were compared with the estimated cumulative distribution for normal, gamma and exponential distributions. The maximum distance between the cumulative lognormal and the cumulative normal distributions was 0.25 . The maximum distance with cumulative gamma (the same for exponential in this case) was 0.28 . An interval width of 0.1 was considered appropriate for the analysis.

Another factor that must be considered is the importance of relatively small errors in the selected distribution and the problems of a false negative determination. It may not be practical to collect as many data observations as needed when the distributions are close (such as when the width interval is 0.05 ). Therefore, it is important to understand what types of further statistical and analysis problems may be caused by having fewer samples than optimal. For example, Figure 32 (total phosphorus in residential area) shows that most of the data fall along the straight line (indicating a lognormal fit), with fewer than 10 observations (out of 933) in the tails being outside of the obvious path of the line.


Figure 32. Normality test for total phosphorus in residential land uses using the NSQD

The calculated p-value for the Kolmogorov-Smirnov test is 0.022 , indicating that the null hypothesis could be rejected and that there is not enough evidence that the empirical distribution is adequately represented by a lognormal distribution. Notice that errors in the tails are smaller than 0.049 . However, the tails are not responsible for the rejection of the null hypothesis (see Figure 33).


Figure 33. $\mathrm{D}_{\text {max }}$ was located in the middle of the distribution

In this case, $D_{\text {max }}$ is located close to a total phosphorus concentration of $0.2 \mathrm{mg} / \mathrm{L}(-0.7 \mathrm{in} \log$ scale $)$. As in this case, the hypothesized distributions are usually rejected because of the departures in the middle of the distribution, not in the tails. However, as previously pointed out, a small number of observations in the upper tail can change the shape of the estimated cumulative probability distribution by affecting the mean and standard deviation of the data. The methods used previously by Van Buren and Beherra evaluated the probability distributions only using two parameters, the median and the standard deviation. They suggested the gamma and exponential distributions as alternatives to the lognormal for some stormwater constituents. Table 20 shows the comparison for the goodness of fit using the 2-parameter gamma, exponential and lognormal distributions using the method of moments.
Table 20. Comparison of Goodness of Fit for Gamma, Exponential and Lognormal Distributions Using the NSQD v.1.1

| CONSTITUENT | PDF | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  | OPEN SPACE |  |  | FREEWAYS |  |  |
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|  |  | $\mathbf{N}_{\text {Det }}$ | Dmax | P-value | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value |
| $\begin{aligned} & \text { Conductivity } \\ & (\mathrm{mS} / \mathrm{cm}) \end{aligned}$ | Gamma | $\begin{gathered} 106 \\ 100 \% \end{gathered}$ | 0.381 | 0 | $\begin{aligned} & 66 \\ & 100 \end{aligned}$ | 0.230 | 0.002 | $\begin{aligned} & 108 \\ & 100 \end{aligned}$ | 0.348 | 0 | $\underset{100}{2}$ | - | - | $\begin{gathered} 86 \\ 100 \end{gathered}$ | 0.238 | 0 |
|  | Exponential |  | 0.195 | 0.001 |  | 0.237 | 0.001 |  | 0.228 | 0 |  | - | - |  | 0.232 | 0 |
|  | Lognormal |  | 0.081 | 0.493 |  | 0.100 | 0.530 |  | 0.074 | 0.619 |  | - | - |  | 0.129 | 0.113 |
| Hardness ( $\mathrm{mg} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 250 \\ & 100 \% \end{aligned}$ | 0.217 | 0 | $\begin{aligned} & 139 \\ & 100 \end{aligned}$ | 0.141 | 0.008 | $\begin{aligned} & 138 \\ & 96.4 \end{aligned}$ | 0.323 | 0 | $\begin{gathered} 8 \\ 100 \end{gathered}$ | 0.304 | 0.458 | $\begin{aligned} & 127 \\ & 100 \end{aligned}$ | 0.451 | 0 |
|  | Exponential |  | 0.203 | 0 |  | 0.115 | 0.067 |  | 0.133 | 0.018 |  | 0.369 | 0.228 |  | 0.161 | 0.003 |
|  | Lognormal |  | 0.071 | 0.166 |  | 0.090 | 0.206 |  | 0.080 | 0.369 |  | 0.354 | 0.268 |  | 0.077 | 0.447 |
| $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Oil} \text { and Gease }}$ | Gamma | $\begin{gathered} 533 \\ 57.8 \% \end{gathered}$ | 0.876 | 0 | $\begin{aligned} & 308 \\ & 70.8 \end{aligned}$ | 0.629 | 0 | $\begin{aligned} & 327 \\ & 65.1 \end{aligned}$ | 0.939 | 0 | $\begin{gathered} 19 \\ 36.84 \end{gathered}$ | 0.210 | 1.080 | $\begin{gathered} 60 \\ 71.7 \end{gathered}$ | 0.103 | 0.810 |
|  | Exponential |  | 0.514 | 0 |  | 0.304 | 0 |  | 0.697 | 0 |  | 0.265 | 0.750 |  | 0.286 | 0.002 |
|  | Lognormal |  | 0.112 | 0.001 |  | 0.103 | 0.019 |  | 0.098 | 0.032 |  | 0.202 | 1.127 |  | 0.101 | 0.827 |
| Total Dissolved Solids (mg/L) | Gamma | $\begin{gathered} 861 \\ 99.3 \% \end{gathered}$ | 0.234 | 0 | $\begin{aligned} & 399 \\ & 99.5 \end{aligned}$ | 0.457 | 0 | $\begin{aligned} & 413 \\ & 99.5 \end{aligned}$ | 0.645 | 0 | $\begin{gathered} 45 \\ 97.8 \end{gathered}$ | 0.109 | 0.698 | $\begin{aligned} & 97 \\ & 99 \end{aligned}$ | 0.082 | 0.553 |
|  | Exponential |  | 0.207 | 0 |  | 0.150 | 0 |  | 0.172 | 0 |  | 0.195 | 0.070 |  | 0.171 | 0.007 |
|  | Lognormal |  | 0.050 | 0.029 |  | 0.049 | 0.303 |  | 0.066 | 0.053 |  | 0.120 | 0.561 |  | 0.054 | 1.136 |
| Total Suspended Solids (mg/L) | Gamma | $\begin{gathered} 991 \\ 98.6 \% \end{gathered}$ | 0.288 | 0 | $\begin{aligned} & 458 \\ & 98.3 \end{aligned}$ | 0.363 | 0 | $\begin{aligned} & 428 \\ & 99.1 \end{aligned}$ | 0.206 | 0 | $\begin{gathered} 44 \\ 95.5 \end{gathered}$ | 0.132 | 0.464 | $\begin{aligned} & 134 \\ & 99.3 \end{aligned}$ | 0.534 | 0 |
|  | Exponential |  | 0.141 | 0 |  | 0.214 | 0 |  | 0.108 | 0 |  | 0.289 | 0.002 |  | 0.168 | 0.011 |
|  | Lognormal |  | 0.032 | 0.280 |  | 0.064 | 0.053 |  | 0.029 | 0.995 |  | 0.113 | 0.683 |  | 0.066 | 0.627 |
| BOD5 (mg/L) | Gamma | $\begin{gathered} 941 \\ 97.6 \% \end{gathered}$ | 0.321 | 0 | $\begin{aligned} & 432 \\ & 97.5 \end{aligned}$ | 0.191 | 0 | $\begin{aligned} & 40 \\ & 95.3 \end{aligned}$ | 0.921 | 0 | $\begin{gathered} 44 \\ 86.4 \end{gathered}$ | 0.112 | 0.770 | $\begin{gathered} 26 \\ 84.6 \end{gathered}$ | 0.272 | 0.076 |
|  | Exponential |  | 0.140 | 0 |  | 0.142 | 0 |  | 0.355 | 0 |  | 0.261 | 0.011 |  | 0.168 | 0.580 |
|  | Lognormal |  | 0.058 | 0.004 |  | 0.054 | 0.166 |  | 0.105 | 0 |  | 0.114 | 0.746 |  | 0.103 | 1.252 |
| COD (mg/L) | Gamma | $\begin{gathered} 796 \\ 98.9 \% \end{gathered}$ | 0.129 | 0 | $\begin{aligned} & 373 \\ & 98.4 \end{aligned}$ | 0.137 | 0 | $\begin{aligned} & 362 \\ & 98.9 \end{aligned}$ | 0.216 | 0 | $\begin{gathered} 43 \\ 76.7 \end{gathered}$ | 0.373 | 0 | $\begin{gathered} 67 \\ 98.5 \end{gathered}$ | 0.163 | 0.061 |
|  | Exponential |  | 0.161 | 0 |  | 0.136 | 0 |  | 0.119 | 0 |  | 0.168 | 0.312 |  | 0.139 | 0.157 |
|  | Lognormal |  | 0.036 | 0.250 |  | 0.038 | 0.695 |  | 0.074 | 0.040 |  | 0.128 | 0.684 |  | 0.107 | 0.445 |
| Fecal Coliform (Colonies/ 100 mL ) | Gamma | $\begin{gathered} 446 \\ 88.3 \% \end{gathered}$ | 0.655 | 0 | $\begin{gathered} 233 \\ 88 \end{gathered}$ | 0.333 | 0 | $\begin{aligned} & 297 \\ & 87.9 \end{aligned}$ | - | - | $\begin{gathered} 23 \\ 91.3 \end{gathered}$ | 0.179 | 0.520 | $\begin{aligned} & 49 \\ & 100 \end{aligned}$ | 0.239 | 0.007 |
|  | Exponential |  | 0.374 | 0 |  | 0.396 | 0 |  | 0.504 | 0 |  | 0.208 | 0.324 |  | 0.355 | 0 |
|  | Lognormal |  | 0.080 | 0.013 |  | 0.076 | 0.192 |  | 0.051 | 0.510 |  | 0.181 | 0.503 |  | 0.105 | 0.677 |
| Fecal Streptococcus (Colonies/100 mL) | Gamma | $\begin{gathered} 305 \\ 89.5 \% \end{gathered}$ | 0.158 | 0 | $\begin{aligned} & 181 \\ & 91.7 \end{aligned}$ | 0.354 | 0 | $\begin{aligned} & 195 \\ & 93.8 \end{aligned}$ | - | - | $\begin{gathered} 22 \\ 90.9 \end{gathered}$ | 0.144 | 0.869 | $\begin{aligned} & 25 \\ & 100 \end{aligned}$ | 0.096 | 1.262 |
|  | Exponential |  | 0.202 | 0 |  | 0.278 | 0 |  | 0.399 | 0 |  | 0.142 | 0.892 |  | 0.164 | 0.518 |
|  | Lognormal |  | 0.077 | 0.081 |  | 0.097 | 0.091 |  | 0.083 | 0.161 |  | 0.181 | 0.538 |  | 0.119 | 0.990 |
| Ammonia (mg/L) | Gamma | $\begin{gathered} 595 \\ 81.5 \% \end{gathered}$ | 0.132 | 0 | $\begin{aligned} & 299 \\ & 83.3 \end{aligned}$ | 0.131 | 0 | $\begin{aligned} & 254 \\ & 85.8 \end{aligned}$ | 0.154 | 0 | $\begin{gathered} 32 \\ 18.7 \end{gathered}$ | - | - | 7987.3 | 0.216 | 0.003 |
|  | Exponential |  | 0.101 | 0 |  | 0.066 | 0.228 |  | 0.071 | 0.221 |  | - | - |  | 0.105 | 0.440 |
|  | Lognormal |  | 0.044 | 0.305 |  | 0.050 | 0.589 |  | 0.047 | 0.758 |  | - | - |  | 0.133 | 0.173 |
| NO2+NO3 (mg/L) | Gamma | $\begin{gathered} 927 \\ 97.4 \% \end{gathered}$ | 0.197 | 0 | $\begin{aligned} & 425 \\ & 98.1 \end{aligned}$ | 0.147 | 0 | $\begin{aligned} & 418 \\ & 96.2 \end{aligned}$ | 0.080 | 0.011 | $\begin{gathered} 44 \\ 84.1 \end{gathered}$ | 0.123 | 0.654 | $\begin{aligned} & 25 \\ & 96 \end{aligned}$ | 0.274 | 0.055 |
|  | Exponential |  | 0.141 | 0 |  | 0.120 | 0 |  | 0.132 | 0 |  | 0.120 | 0.686 |  | 0.177 | 0.443 |
|  | Lognormal |  | 0.070 | 0 |  | 0.040 | 0.531 |  | 0.080 | 0.011 |  | 0.141 | 0.463 |  | 0.139 | 0.789 |
| TKN (mg/L) | Gamma | $\begin{gathered} 957 \\ 96.8 \% \end{gathered}$ | 0.203 | 0 | $\begin{aligned} & 449 \\ & 97.3 \end{aligned}$ | 0.127 | 0 | $\begin{aligned} & 440 \\ & 95.9 \end{aligned}$ | 0.195 | 0 | $\begin{gathered} 45 \\ 71.1 \end{gathered}$ | 0.169 | 0.323 | $\begin{aligned} & 125 \\ & 96.8 \end{aligned}$ | 0.280 | 0 |
|  | Exponential |  | 0.182 | 0 |  | 0.156 | 0 |  | 0.134 | , |  | 0.141 | 0.556 |  | 0.138 | 0.020 |
|  | Lognormal |  | 0.035 | 0.218 |  | 0.042 | 0.423 |  | 0.048 | 0.292 |  | 0.147 | 0.500 |  | 0.074 | 0.539 |

Table 20. Comparison of Goodness of Fit for Gamma, Exponential and Lognormal Distributions Using the NSQD - Continued

| CONSTITUENT | PDF | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  | OPEN SPACE |  |  | FREEWAYS |  |  |
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|  |  | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value | $\mathbf{N}_{\text {Det }}$ | Dmax | P-value | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value | $\mathrm{N}_{\text {Det }}$ | Dmax | P-value | N ${ }_{\text {Det }}$ 22 95.5 | Dmax | P-value |
| Dissolved Phosphorus (mg/L) | Gamma | $\begin{gathered} 738 \\ 84.1 \end{gathered}$ | 0.117 | 0 | $\begin{gathered} 323 \\ 81.1 \end{gathered}$ | 0.177 | 0 | $\begin{gathered} 325 \\ 87.1 \end{gathered}$ | 0.200 | 0 | $\begin{gathered} 44 \\ 79.5 \end{gathered}$ | 0.154 | 0.127 | $\begin{gathered} 22 \\ 95.5 \end{gathered}$ | $0.449$ | 0 |
|  | Exponential |  | 0.144 | 0 |  | 0.129 | 0 |  | 0.135 | 0 |  | 0.384 | 0.657 |  | 0.350 | 0.012 |
|  | Lognormal |  | 0.043 | 0.199 |  | 0.075 | 0.104 |  | 0.124 | 0.682 |  | 0.124 | 0.682 |  | 0.170 | 0.593 |
| Total Phosphorus$(\mathrm{mg} / \mathrm{L})$ | Gamma | $\begin{aligned} & 9639 \\ & 969 \end{aligned}$ | 0.184 | 0 | $\begin{aligned} & 444 \\ & 95.7 \end{aligned}$ | 0.179 | 0 | $\begin{aligned} & 434 \\ & 96.3 \end{aligned}$ | 0.227 | 0 | $\begin{gathered} 46 \\ 84.8 \end{gathered}$ | 0.666 | 0 | $\begin{gathered} 128 \\ 99.2 \end{gathered}$ | 0.456 | 0 |
|  | Exponential |  | 0.129 | 0 |  | 0.114 | 0 |  | 0.107 | 0 |  | 0.320 | 0.001 |  | 0.187 | 0 |
|  | Lognormal |  | 0.049 | 0.022 |  | 0.038 | 0.582 |  | 0.049 | 0.273 |  | 0.116 | 0.696 |  | 0.085 | 0.325 |
| Total Antimony$(\mu \mathrm{g} / \mathrm{L})$ | Gamma | $\begin{gathered} 288 \\ 2.8 \end{gathered}$ | 0.268 | 0.636 | $\begin{aligned} & 142 \\ & 2.1 \end{aligned}$ | - | - | $\begin{array}{r} 164 \\ 14.6 \end{array}$ | 0.282 | 0.045 | $\begin{gathered} 17 \\ 0 \end{gathered}$ | - | - | $\begin{aligned} & 14 \\ & 50 \end{aligned}$ | 0.423 | 0.164 |
|  | Exponential |  | 0.417 | 0.213 |  | - | - |  | 0.173 | 0.473 |  | - | - |  | 0.465 | 0.096 |
|  | Lognormal |  | 0.233 | 0.841 |  | - | - |  | 0.096 | 1.279 |  | - | - |  | 0.419 | 0.171 |
| Total Arsenic ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{gathered} 426 \\ 42 \end{gathered}$ | 0.531 | 0 | $\begin{aligned} & 213 \\ & 32.9 \end{aligned}$ | 0.643 | 0 | $\begin{gathered} 267 \\ 54.3 \end{gathered}$ | 0.291 | 0 | $\begin{gathered} 19 \\ 31.6 \end{gathered}$ | 0.271 | 0.828 | $\begin{gathered} 61 \\ 55.7 \end{gathered}$ | 0.125 | 0.694 |
|  | Exponential |  | 0.224 | 0 |  | 0.249 | 0 |  | 0.141 | 0.006 |  | 0.462 | 0.154 |  | 0.266 | 0.016 |
|  | Lognormal |  | 0.154 | 0 |  | 0.164 | 0.046 |  | 0.129 | 0.016 |  | 0.273 | 0.819 |  | 0.149 | 0.441 |
| Total Beryllium$(\mu \mathrm{g} / \mathrm{L})$ | Gamma | $\begin{gathered} 301 \\ 7.3 \end{gathered}$ | 0.464 | 0 | $\begin{aligned} & 163 \\ & 4.29 \end{aligned}$ | 0.305 | 0.542 | $\begin{aligned} & 209 \\ & 10.5 \end{aligned}$ | 0.390 | 0.002 | $\begin{gathered} 19 \\ 0 \end{gathered}$ | - | - | $\begin{gathered} 12 \\ 16.7 \end{gathered}$ | - | - |
|  | Exponential |  | 0.471 | 0 |  | 0.530 | 0.039 |  | 0.539 | 0 |  | - | - |  | - | - |
|  | Lognormal |  | 0.200 | 0.342 |  | 0.205 | 1.108 |  | 0.163 | 0.620 |  | - | - |  | - | - |
| Total Cadmium$(\mu \mathrm{g} / \mathrm{L})$ | Gamma | $\begin{gathered} 723 \\ 30.3 \end{gathered}$ | 0.643 | 0 | $\begin{gathered} 358 \\ 43 \end{gathered}$ | 0.511 | 0 | $\begin{gathered} 395 \\ 49.4 \end{gathered}$ | 0.445 | 0 | $\begin{gathered} 38 \\ 55.3 \end{gathered}$ | 0.295 | 0.051 | $\begin{gathered} 95 \\ 71.6 \end{gathered}$ | 0.110 | 0.388 |
|  | Exponential |  | 0.358 | 0 |  | 0.311 | 0 |  | 0.237 | , |  | 0.560 | 0 |  | 0.153 | 0.083 |
|  | Lognormal |  | 0.120 | 0.004 |  | 0.113 | 0.039 |  | 0.083 | 0.136 |  | 0.206 | 0.338 |  | 0.052 | 1.380 |
| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 435 \\ & 55.4 \end{aligned}$ | 0.292 | 0 | $\begin{gathered} 235 \\ 58.7 \end{gathered}$ | 0.151 | 0.004 | $\begin{gathered} 256 \\ 72.7 \end{gathered}$ | 0.122 | 0.008 | $\begin{gathered} 36 \\ 36.1 \end{gathered}$ | 0.252 | 0.386 | $\begin{gathered} 76 \\ 98.7 \end{gathered}$ | 0.058 | 1.208 |
|  | Exponential |  | 0.132 | 0 |  | 0.201 | 0 |  | 0.067 | 0.381 |  | 0.272 | 0.290 |  | 0.176 | 0.019 |
|  | Lognormal |  | 0.069 | 0.206 |  | 0.086 | 0.262 |  | 0.062 | 0.480 |  | 0.180 | 0.861 |  | 0.084 | 0.685 |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 799 \\ & 83.6 \end{aligned}$ | 0.394 | 0 | $\begin{aligned} & 387 \\ & 92.8 \end{aligned}$ | 0.296 | 0 | $\begin{aligned} & 416 \\ & 89.9 \end{aligned}$ | 0.408 | 0 | $\begin{gathered} 39 \\ 74.4 \end{gathered}$ | 0.107 | 0.226 | $\begin{aligned} & 97 \\ & 99 \end{aligned}$ | 0.451 | 0 |
|  | Exponential |  | 0.149 | 0 |  | 0.137 | 0 |  | 0.177 | 0 |  | 0.127 | 0.092 |  | 0.231 | 0.090 |
|  | Lognormal |  | 0.067 | 0.005 |  | 0.070 | 0.060 |  | 0.080 | 0.017 |  | 0.131 | 0.742 |  | 0.038 | 1.507 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 788 \\ & 71.3 \end{aligned}$ | 0.300 | 0 | $\begin{gathered} 377 \\ 85.4 \end{gathered}$ | 0.297 | 0 | $\begin{gathered} 412 \\ 76.5 \end{gathered}$ | 0.276 | 0 | $\begin{gathered} 45 \\ 42.2 \end{gathered}$ | 0.177 | 0.608 | $\begin{aligned} & 107 \\ & 100 \end{aligned}$ | 0.203 | 0 |
|  | Exponential |  | 0.173 | 0 |  | 0.136 | 0 |  | 0.225 | 0 |  | 0.389 | 0.006 |  | 0.125 | 0.072 |
|  | Lognormal |  | 0.044 | 0.218 |  | 0.057 | 0.250 |  | 0.059 | 0.223 |  | 0.132 | 1.034 |  | 0.039 | 1.451 |
| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 419 \\ & 45.3 \end{aligned}$ | 0.292 | 0 | $\begin{gathered} 232 \\ 59.5 \end{gathered}$ | 0.260 | 0 | $\begin{aligned} & 250 \\ & 62.4 \end{aligned}$ | 0.090 | 0.159 | $\begin{gathered} 38 \\ 18.4 \end{gathered}$ | 0.164 | 1.373 | $\begin{gathered} 99 \\ 89.9 \end{gathered}$ | 0.188 | 0.004 |
|  | Exponential |  | 0.203 | 0 |  | 0.176 | 0 |  | 0.111 | 0.044 |  | 0.261 | 0.772 |  | 0.227 | 0 |
|  | Lognormal |  | 0.081 | 0.160 |  | 0.056 | 0.831 |  | 0.065 | 0.525 |  | 0.166 | 1.360 |  | 0.091 | 0.460 |
| Total Selenium$(\mu \mathrm{g} / \mathrm{L})$ | Gamma | $\begin{gathered} 318 \\ 6.9 \end{gathered}$ | 0.263 | 0.095 | $\begin{aligned} & 169 \\ & 7.7 \end{aligned}$ | 0.169 | 0.952 | $\begin{gathered} 203 \\ 5.9 \end{gathered}$ | 0.434 | 0.022 | $\begin{gathered} 19 \\ 21.1 \end{gathered}$ | - | - | $\begin{aligned} & 16 \\ & 6.3 \end{aligned}$ | - | - |
|  | Exponential |  | 0.254 | 0.117 |  | 0.174 | 0.907 |  | 0.256 | 0.416 |  | - | - |  | - | - |
|  | Lognormal |  | 0.253 | 0.119 |  | 0.196 | 0.735 |  | 0.190 | 0.841 |  | - | - |  | - | - |
| Total Silver ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 406 \\ & 12.6 \end{aligned}$ | 0.421 | 0 | $\begin{aligned} & 222 \\ & 11.3 \end{aligned}$ | 0.143 | 0.718 | $\begin{aligned} & 287 \\ & 17.4 \end{aligned}$ | 0.263 | 0.002 | $\begin{gathered} 19 \\ 5.3 \end{gathered}$ | - | - | $\begin{aligned} & 21 \\ & 19 \end{aligned}$ | - | - |
|  | Exponential |  | 0.333 | 0 |  | 0.159 | 0.563 |  | 0.340 | 0 |  | - | - |  | - | - |
|  | Lognormal |  | 0.271 | 0.001 |  | 0.184 | 0.370 |  | 0.146 | 0.236 |  | - | - |  | - | - |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | Gamma | $\begin{aligned} & 810 \\ & 96.4 \end{aligned}$ | 0.244 | 0 | $\begin{gathered} 392 \\ 99 \end{gathered}$ | 0.234 | 0 | $\begin{aligned} & 433 \\ & 98.6 \end{aligned}$ | 0.273 | 0 | 71.1 | 0.180 | 0.253 | 9396.8 | 0.158 | 0.023 |
|  | Exponential |  | 0.122 | 0 |  | 0.141 | 0 |  | 0.083 | 0.005 |  | 0.167 | 0.336 |  | 0.155 | 0.027 |
|  | Lognormal |  | 0.054 | 0.020 |  | 0.040 | 0.585 |  | 0.044 | 0.389 |  | 0.105 | 0.981 |  | 0.063 | 0.985 | * P-values greater than one are used only for comparison. NDet: Number of collected samples and percentage detected

Table 20 shows that for residential, commercial and industrial land uses, the lognormal distribution better fits the empirical data, except for selenium and silver in commercial land uses. In open space land uses, about $50 \%$ of the constituents were adequately fitted by the lognormal distribution, $30 \%$ by the gamma distribution and the remaining by the exponential distribution. In freeway areas, lognormal distributions better fit most of the constituents, except that fecal streptococcus, total arsenic and total chromium were better fitted by the gamma distribution and ammonia was better fitted by the exponential distribution. Also note in Table 20 that residential, commercial and industrial land uses had larger sample sizes than the other two land uses. It seems that for small sample sizes, gamma and exponential distributions better represent actual stormwater constituent distributions, but once the number of samples increases, the lognormal distribution is best. The few cases were the gamma distribution was a better fit was for $\mathrm{NO}_{2}+\mathrm{NO}_{3}$ in industrial land uses, and chromium in freeway areas. The exponential distribution better represents total ammonia in freeway areas (with around 70 detected samples) than the other two distribution types.

Other transformations were also tested, such as the square root, and other power functions, but the results were not improved. It was therefore decided to investigate if a three-parameter lognormal distribution function can be used to improve the overall goodness of fit for stormwater constituent probability distributions. As shown in the following section, this third parameter, in some cases, allows a much better fit of the cumulative empirical and estimated probability distributions.

## Three Parameter Lognormal Calculations

Goodness of fit was evaluated using 3-parameter lognormal probability distribution. The probability distributions were created for residential, commercial, industrial, open space and freeways land uses. The distribution parameters were calculated using the maximum likelihood and the L-moments methods. The maximum likelihood method requires that it be solved iteratively using three equations (see Appendix C). The results were compared with the 2 parameter standard model and the actual data. The model with the smaller maximum distance between the empirical and the estimated function was selected as the best model. All the calculations were made using only the detected values. In general, the L-moments method provided a better fit for the upper tail of the distribution whereas the maximum likelihood method provided a better fit for the lower tail. Figure 34 shows the three estimated models for TSS in commercial land use areas.

## PROBABILITY PLOT



Figure 34. Estimated models for TSS in commercial land uses

In this graph, it is observed that the empirical distribution has higher values in the upper tail compared with any of the three models. In the lower tail, the maximum likelihood method using the 3-parameters better fit the observed values. In this case, the maximum likelihood method was better than the other two models, although none of the methods adequately represented the extreme high values. The L-moments method generally betters fits the upper tail distribution, but typically trims or overestimates the lower tail. Figure 35 shows the results for TDS in industrial land uses. The L-moments better fits the empirical distribution in the upper tail, but it trims any observation smaller than $35 \mathrm{mg} / \mathrm{L}$ (almost 20\% of the total dataset) in the lower tail. The 2-parameter lognormal and the maximum likelihood method provide better results although both were worse than the L-moments in the upper tail region.

## PROBABILITY PLOT



Figure 35. Estimated models for TDS in industrial land use

Table 21 presents the results for 15 constituents in five land uses. For each of the three methods, the p-value was calculated. The higher the p-value, the better is the fit between the empirical and the estimated function. Some of the p -values in the table are larger than one. When the number of samples is large, the p -value is calculated as a chi square distribution with two degrees of freedom. This probability is calculated only with one tail of the chi square distribution. The $p$-value is two times this probability. The maximum $p$-value is one, but for effects of comparison this presents two times the probability calculated from a one tail chi square distribution.

The maximum likelihood method with 3-parameters, or the lognormal 2-parameter distribution produced the best descriptions for most of the constituents. For almost all constituents the function estimated by the L-moments method failed the lognormal assumption. Low p-values were obtained because the function was truncated and does not estimate the lower tail of the distribution.

It seems that when the numbers of samples increase, the L-moments method tends to truncate the function. The maximum likelihood method seems to improve the fit of the distribution, but when the number of samples is large, the cumulative estimated probability distribution is far from the cumulative empirical probability distribution, or no convergence is possible during the iteration process.

In commercial, industrial and freeways land uses, the numbers of samples available were between 100 and 500 samples. According to the prior discussion, this number of samples will result in an analysis having a power close or above 0.5 . In these cases, most of the better fits were obtained using the L-moments method. In commercial and industrial land uses, more than half of the constituents also had the highest p-values when the L-moments method was used.

In open space areas, there were not many samples available. The small number of samples results in a low power. In this case, the higher $p$-values results were observed when the 2 -parameter lognormal distribution was used. The use of the third parameter in constituents having small numbers of sample observations did not improve the fit of the estimated cumulative probability distribution.

## Summary

Most of the stormwater constituents can be assumed to follow a lognormal distribution with little error. The use of the third parameter does not show a significant improvement in estimating the empirical distribution compared with the 2-parameter lognormal distribution. When the number of samples is very large per category (approximately more than 400 samples) the maximum likelihood and the 2-parameter lognormal distribution better fit the empirical distribution. For large sample sizes, the L-moments method usually unacceptably truncates the distribution in the lower tail. When the sample size is small ( $<100$ samples), the use of the third parameter does not improve the fit with the empirical distribution and the 2-parameter lognormal distribution produces a better fit than the other two methods.

The lognormal distribution is a skewed distribution when plotted in real space coordinates. When the sample size is small, the calculated skewness is smaller than the skewness of the real distribution. Insufficient sample sizes are not likely to accurately represent the extreme observations in the actual distribution of the data.
Table 21. Goodness of Fit for Different Land Uses ( p -values Larger than One are Used Only for Comparison)

| RESIDENTIAL |  |  | Critical | Obs | rved | 2 Par | neter Log |  | 3 para | er maximu | likeliho |  |  | 3 para | neter L-mome |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONSTITUENT | N | \% Detected | $\mathrm{D}_{0.05}$ | $\mu$ | $\sigma$ | D2 | p-value | $\mu$ | $\sigma$ | c | D3m | p-value | к | $\alpha$ | $\xi$ | D31 | p-value |
| Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ) | 106 | 100 | 0.132 | 4.638 | 0.710 | 0.081 | 0.493 | 4.327 | 0.919 | 20.767 | 0.052 | 1.133 | -1.240 | 61.461 | 88.366 | 0.087 | 0.398 |
| Hardness (mg/L) | 250 | 100 | 0.086 | 3.497 | 0.706 | 0.071 | 0.166 | 3.539 | 0.675 | -1.114 | 0.066 | 0.231 |  |  |  | 0.635 | 0 |
| Oil \& Grease (mg/L) | 533 | 57.8 | 0.077 | 1.428 | 1.204 | 0.112 | 0.001 | 1.356 | 1.267 | 0.164 | 0.102 | 0.003 | -2.559 | 2.065 | 2.482 | 0.208 | 0 |
| Fecal Coliform ( $\mathrm{Col} / 100 \mathrm{ml}$ ) | 446 | 88.3 | 0.069 | 8.7205 | 2.40448 | 0.080 | 0.013 | 8.734 | 2.367 | -2.992 | 0.078 | 0.017 | -1.929 | 17283.541 | 8236.423 | 0.096 | 0.001 |
| Fecal Streptococcus (C/100 ml) | 305 | 89.5 | 0.082 | 9.8344 | 1.88029 | 0.077 | 0.081 | 9.907 | 1.725 | -210.924 | 0.066 | 0.190 | -1.309 | 38433.290 | 24834.760 | 0.077 | 0.078 |
| Ammonia (mg/L) | 595 | 81.5 | 0.062 | -1.1672 | 0.9166 | 0.044 | 0.305 | 0.015 | 0.220 | -0.684 | 0.139 | 0.000 |  | - |  |  |  |
| DP ( $\mathrm{mg} / \mathrm{L}$ ) | 738 | 84.1 | 0.055 | -1.8303 | 0.85689 | 0.043 | 0.199 | 0.003 | 0.056 | -0.877 | 0.327 | 0.000 | - | - | - | - |  |
| Antimony ( $\mu \mathrm{g} / \mathrm{L}$ ) | 288 | 2.8 | 0.454 | 1.3554 | 1.71904 | 0.233 | 0.841 | 6.569 | 0.023 | -700.000 | 0.368 | 0.228 | -1.117 | 9.631 | 5.281 | 0.270 | 0.626 |
| Arsenic ( $\mu \mathrm{g} / \mathrm{L}$ ) | 426 | 42 | 0.102 | 1.2098 | 0.85043 | 0.154 | 0.000 | 1.047 | 0.971 | 0.356 | 0.166 | 0.000 | -1.588 | 1.981 | 2.632 | 0.208 | 0 |
| Beryllium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 301 | 7.3 | 0.281 | 0.0283 | 1.5958 | 0.200 | 0.342 | -0.423 | 2.040 | 0.136 | 0.186 | 0.436 | -2.063 | 1.154 | 0.675 | 0.227 | 0.207 |
| Total Cadmium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 723 | 30.3 | 0.092 | -0.3532 | 1.21891 | 0.120 | 0.004 | -0.444 | 1.301 | 0.033 | 0.110 | 0.010 | -1.880 | 0.626 | 0.511 | 0.093 | 0.045 |
| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 435 | 55.4 | 0.088 | 1.5794 | 0.89199 | 0.069 | 0.206 | 1.473 | 0.980 | 0.328 | 0.067 | 0.236 | -1.157 | 4.003 | 4.382 | 0.066 | 0.242 |
| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | 419 | 45.3 | 0.099 | 1.7909 | 0.75669 | 0.081 | 0.160 | 1.601 | 0.890 | 0.768 | 0.083 | 0.147 | -1.078 | 4.163 | 5.384 | 0.084 | 0.137 |
| Total Selenium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 318 | 6.9 | 0.281 | 1.0969 | 0.83323 | 0.253 | 0.119 | 0.479 | 1.348 | 0.876 | 0.259 | 0.061 | -1.178 | 2.103 | 2.577 | 0.257 | 0.11 |
| Total Silver ( $\mu \mathrm{g} / \mathrm{L}$ ) | 406 | 12.6 | 0.19 | 1.0686 | 1.3707 | 0.271 | 0.001 | 0.984 | 1.469 | 0.089 | 0.278 | 0.001 | -1.522 | 3.390 | 2.767 | 0.294 | 0 |


| COMMERCIAL |  |  | Observed |  |  | 2 Parameter Log |  | 3 parameter maximum likelihood |  |  |  |  | 3 parameter L-moments |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONSTITUENT | N | \% Detected | $\mu$ | $\sigma$ | $\mathrm{D}_{0.05}$ |  | p-value | $\mu$ | $\sigma$ | c | D3m | p-value | $\kappa$ | $\alpha$ | $\xi$ | D31 | p-value |
| Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ) | 66 | 100 | 4.779 | 0.721 | 0.167 | 0.100 | 0.530 | 4.736 | 0.746 | 3.865 | 0.097 | 0.581 | -1.011 | 76.386 | 108.055 | 0.093 | 0.633 |
| Hardness (mg/L) | 139 | 100 | 3.689 | 0.988 | 0.115 | 0.090 | 0.206 | 3.828 | 0.844 | -3.808 | 0.063 | 0.653 | -0.935 | 36.911 | 40.394 | 0.072 | 0.474 |
| Oil \& Grease (mg/L) | 308 | 70.8 | 1.609 | 1.070 | 0.092 | 0.103 | 0.019 | 1.300 | 1.358 | 0.737 | 0.092 | 0.062 | -1.853 | 3.648 | 3.638 | 0.126 | 0.002 |
| TDS (mg/L) | 399 | 99.5 | 4.332 | 0.791 | 0.068 | 0.049 | 0.303 | 4.393 | 0.741 | -3.495 | 0.049 | 0.289 | -1.066 | 55.106 | 70.423 | 0.071 | 0.035 |
| TSS (mg/L) | 458 | 98.3 | 3.883 | 1.180 | 0.064 | 0.064 | 0.053 | 3.735 | 1.218 | 1.988 | 0.042 | 0.416 | -1.207 | 55.082 | 46.245 | 0.048 | 0.250 |
| BOD (mg/L) | 432 | 97.5 | 2.493 | 0.868 | 0.066 | 0.054 | 0.166 | 2.302 | 1.026 | 1.396 | 0.040 | 0.527 | -1.002 | 10.321 | 11.447 | 0.044 | 0.380 |
| COD (mg/L) | 373 | 98.4 | 4.167 | 0.865 | 0.071 | 0.038 | 0.695 | 4.163 | 0.868 | 0.194 | 0.037 | 0.719 | -0.911 | 54.903 | 63.127 | 0.034 | 0.860 |
| Fecal Coliform ( $\mathrm{Col} / 100 \mathrm{ml}$ ) | 233 | 88 | 8.202 | 2.380 | 0.095 | 0.076 | 0.192 | 8.191 | 2.398 | 1.870 | 0.077 | 0.175 | -1.768 | 11370.330 | 5408.770 | 0.150 | 0.000 |
| Fecal Streptococcus (C/100 ml) | 181 | 91.7 | 8.940 | 2.061 | 0.106 | 0.097 | 0.091 | 8.936 | 2.061 | 2.494 | 0.096 | 0.093 | -1.640 | 16795.600 | 9532.540 | 0.056 | 0.702 |
| Ammonia (mg/L) | 299 | 83.3 | -0.706 | 1.083 | 0.086 | 0.050 | 0.589 | -0.697 | 1.072 | -0.002 | 0.048 | 0.632 | -0.947 | 0.549 | 0.522 | 0.040 | 0.888 |
| NO2+NO3 (mg/L) | 425 | 98.1 | -0.523 | 0.882 | 0.067 | 0.040 | 0.531 | -0.432 | 0.800 | -0.039 | 0.034 | 0.837 | -0.849 | 0.510 | 0.600 | 0.030 | 0.954 |
| TKN (mg/L) | 449 | 97.3 | 0.471 | 0.828 | 0.065 | 0.042 | 0.423 | 0.575 | 0.734 | -0.126 | 0.050 | 0.228 | -0.866 | 1.246 | 1.571 | 0.032 | 0.816 |
| DP (mg/L) | 323 | 81.1 | -2.077 | 1.016 | 0.084 | 0.075 | 0.104 | -2.157 | 1.092 | 0.006 | 0.062 | 0.273 | -1.077 | 0.127 | 0.121 | 0.059 | 0.315 |
| TP (mg/L) | 446 | 95.7 | -1.473 | 0.881 | 0.066 | 0.038 | 0.582 | -1.537 | 0.935 | 0.010 | 0.041 | 0.466 | -0.991 | 0.196 | 0.220 | 0.049 | 0.264 |
| Arsenic ( $\mu \mathrm{g} / \mathrm{L}$ ) | 213 | 32.9 | 0.9336 | 0.92361 | 0.163 | 0.164 | 0.046 | 0.729 | 1.098 | 0.301 | 0.195 | 0.010 | -1.736 | 1.565 | 1.935 | 0.280 | 0 |
| Total Cadmium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 358 | 43 | 0.047 | 1.309 | 0.11 | 0.113 | 0.039 | 0.023 | 1.335 | 0.011 | 0.109 | 0.052 | -1.640 | 1.180 | 0.879 | 0.063 | 0.591 |
| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 235 | 58.7 | 1.8134 | 0.71608 | 0.116 | 0.086 | 0.262 | 1.711 | 0.787 | 0.453 | 0.089 | 0.225 | -0.734 | 4.462 | 6.100 | 0.083 | 0.295 |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | 387 | 92.8 | 2.8829 | 0.90117 | 0.072 | 0.070 | 0.060 | 2.807 | 0.965 | 0.874 | 0.063 | 0.117 | -1.251 | 14.103 | 15.641 | 0.069 | 0.067 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | 377 | 85.4 | 3.0328 | 1.03226 | 0.076 | 0.057 | 0.250 | 3.015 | 1.049 | 0.220 | 0.058 | 0.225 | -1.251 | 19.633 | 18.805 | 0.053 | 0.329 |
| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | 232 | 59.5 | 1.9782 | 0.8075 | 0.116 | 0.056 | 0.831 | 1.668 | 1.070 | 1.282 | 0.089 | 0.220 | -0.969 | 5.691 | 6.842 | 0.076 | 0.406 |
| Total Selenium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 169 | 7.7 | 1.366 | 0.953 | 0.361 | 0.196 | 0.735 | 0.829 | 1.492 | 0.863 | 0.167 | 0.670 | -0.803 | 3.940 | 3.621 | 0.210 | 0.638 |
| Total Silver ( $\mu \mathrm{g} / \mathrm{L}$ ) | 222 | 11.3 | 0.9637 | 1.35108 | 0.272 | 0.184 | 0.370 | 1.080 | 1.174 | -0.141 | 0.182 | 0.379 | -0.911 | 3.587 | 3.133 | 0.165 | 0.513 |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | 392 | 99 | 5.0388 | 0.84183 | 0.069 | 0.040 | 0.585 | 5.082 | 0.803 | -4.834 | 0.039 | 0.619 | -1.021 | 120.091 | 144.868 | 0.052 | 0.243 |

Table 21. Goodness of Fit for Different Land Uses (p-values Larger than One are Used Only for Comparison) - Continued

| INDUSTRIAL |  |  | Observed |  |  | 2 Parameter Log |  | 3 parameter maximum likelihood |  |  |  |  | 3 parameter L-moments |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONSTITUENT | N | \% Detected | $\mu$ | $\sigma$ | $\mathrm{D}_{0.05}$ | D2 | p-value | $\mu$ | $\sigma$ | c | D3m | p-value | $\kappa$ | $\alpha$ | $\xi$ | D31 | p-value |
| Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ) | 108 | 100 | 5.011 | 0.673 | 0.131 | 0.074 | 0.619 | 4.743 | 0.848 | 27.365 | 0.067 | 0.767 | -1.197 | 83.673 | 129.715 | 0.106 | 0.174 |
| Hardness (mg/L) | 138 | 96.4 | 3.794 | 0.842 | 0.118 | 0.080 | 0.369 | 3.701 | 0.914 | 2.758 | 0.081 | 0.346 | -1.272 | 31.221 | 38.285 | 0.119 | 0.047 |
| Oil \& Grease (mg/L) | 327 | 65.1 | 1.623 | 1.153 | 0.093 | 0.098 | 0.032 | 1.456 | 1.298 | 0.447 | 0.093 | 0.048 | -3.227 | 1.068 | 2.750 | 0.313 | 0.000 |
| TDS (mg/L) | 413 | 99.5 | 4.516 | 0.870 | 0.067 | 0.066 | 0.053 | 4.539 | 0.849 | -1.484 | 0.065 | 0.064 | -1.496 | 62.528 | 76.123 | 0.111 | 0.000 |
| TSS (mg/L) | 428 | 99.1 | 4.287 | 1.200 | 0.066 | 0.029 | 0.995 | 4.292 | 1.193 | -0.169 | 0.028 | 1.023 | -1.133 | 88.174 | 74.697 | 0.026 | 1.119 |
| BOD (mg/L) | 406 | 95.3 | 2.4121 | 0.992 | 0.069 | 0.105 | 0.000 | 2.303 | 1.085 | 0.729 | 0.095 | 0.002 | -2.246 | 5.718 | 7.369 | 0.184 | 0.000 |
| COD (mg/L) | 362 | 98.9 | 4.2217 | 0.91441 | 0.072 | 0.074 | 0.040 | 4.237 | 0.899 | -0.714 | 0.076 | 0.032 | -1.096 | 57.774 | 63.159 | 0.046 | 0.437 |
| Fecal Coliform ( $\mathrm{Col} / 100 \mathrm{ml}$ ) | 297 | 87.9 | 7.6064 | 2.65965 | 0.084 | 0.051 | 0.510 | 7.574 | 2.732 | 1.560 | 0.055 | 0.417 | -2.356 | 5638.993 | 2369.488 | 0.045 | 0.688 |
| Fecal Streptococcus (C/100 ml) | 195 | 93.8 | 9.1491 | 1.82906 | 0.101 | 0.083 | 0.161 | 9.190 | 1.741 | -64.211 | 0.073 | 0.280 | -2.253 | 11045.202 | 7378.455 | 0.199 | 0.000 |
| Ammonia (mg/L) | 254 | 85.8 | -0.7071 | 1.00903 | 0.092 | 0.047 | 0.758 | -0.685 | 0.985 | -0.007 | 0.049 | 0.715 | -0.864 | 0.518 | 0.524 | 0.046 | 0.789 |
| NO2+NO3 (mg/L) | 418 | 96.2 | -0.3857 | 0.93116 | 0.068 | 0.080 | 0.011 | -0.142 | 0.703 | -0.132 | 0.043 | 0.454 | -0.689 | 0.608 | 0.739 | 0.045 | 0.406 |
| TKN (mg/L) | 440 | 95.9 | 0.4238 | 0.88424 | 0.066 | 0.048 | 0.292 | 0.471 | 0.840 | -0.050 | 0.050 | 0.239 | -1.023 | 1.256 | 1.444 | 0.040 | 0.502 |
| DP (mg/L) | 325 | 87.1 | -2.1766 | 0.87102 | 0.081 | 0.124 | 0.682 | -2.141 | 0.837 | -0.003 | 0.051 | 0.450 | -1.002 | 0.093 | 0.108 | 0.063 | 0.211 |
| TP (mg/L) | 434 | 96.3 | -1.2683 | 0.98202 | 0.067 | 0.049 | 0.273 | -1.299 | 1.010 | 0.005 | 0.044 | 0.387 | -1.068 | 0.271 | 0.271 | 0.035 | 0.724 |
| Antimony ( $\mu \mathrm{g} / \mathrm{L}$ ) | 164 | 14.6 | 1.4793 | 1.01264 | 0.269 | 0.096 | 1.279 | 1.275 | 1.183 | 0.479 | 0.113 | 1.088 | -1.334 | 3.747 | 3.661 | 0.150 | 0.684 |
| Arsenic ( $\mu \mathrm{g} / \mathrm{L}$ ) | 267 | 54.3 | 1.5218 | 0.95205 | 0.113 | 0.129 | 0.016 | 1.19121.2 | 2628 | 0.752 | 0.128 | 0.018 | -1.069 | 4.359 | 4.359 | 0.116 | 0.039 |
| Beryllium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 209 | 10.5 | -0.3588 | 1.94765 | 0.281 | 0.163 | 0.620 | -0.892 | 2.658 | 0.060 | 0.197 | 0.362 | -2.074 | 1.346 | 0.568 | 0.231 | 0.191 |
| Total Cadmium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 395 | 49.4 | 0.7417 | 1.12552 | 0.097 | 0.083 | 0.136 | 0.588 | 1.276 | 0.161 | 0.095 | 0.060 | -1.611 | 1.898 | 1.686 | 0.115 | 0.012 |
| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 256 | 72.7 | 2.5512 | 1.06906 | 0.1 | 0.062 | 0.480 | 2.599 | 1.015 | -0.359 | 0.059 | 0.543 | -0.911 | 13.859 | 13.657 | 0.050 | 0.803 |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | 416 | 89.9 | 3.2275 | 1.02866 | 0.07 | 0.080 | 0.017 | 3.179 | 1.076 | 0.716 | 0.073 | 0.030 | -1.343 | 22.969 | 22.081 | 0.057 | 0.172 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | 412 | 76.5 | 3.3651 | 1.32824 | 0.077 | 0.059 | 0.223 | 3.333 | 1.367 | 0.379 | 0.057 | 0.263 | -1.374 | 38.025 | 27.991 | 0.054 | 0.316 |
| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | 250 | 62.4 | 2.8058 | 0.97034 | 0.109 | 0.065 | 0.525 | 2.802 | 0.971 | 0.042 | 0.066 | 0.512 | -0.772 | 17.094 | 17.834 | 0.088 | 0.182 |
| Total Selenium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 203 | 5.9 | 1.1472 | 1.27671 | 0.375 | 0.190 | 0.841 | 6.916 | 0.014 | -1000.000 | 0.364 | 0.083 | -1.851 | 2.381 | 2.025 | 0.202 | 0.753 |
| Total Silver ( $\mu \mathrm{g} / \mathrm{L}$ ) | 287 | 17.4 | 0.113 | 1.80819 | 0.192 | 0.146 | 0.236 | 0.158 | 1.708 | -0.010 | 0.153 | 0.194 | -1.358 | 2.181 | 1.394 | 0.157 | 0.170 |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | 433 | 98.6 | 5.305 | 0.970 | 0.066 | 0.044 | 0.389 | 5.359 | 0.915 | -7.026 | 0.034 | 0.743 | -0.899 | 198.198 | 208.408 | 0.030 | 0.951 |

Table 21. Goodness of Fit for Different Land Uses ( $p$-values Larger than One are Used Only for Comparison) - Continued

| OPEN SPACE |  |  | Observed |  |  | 2 Parameter Log |  | 3 parameter maximum likelihood |  |  |  |  | 3 parameter L-moments |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONSTITUENT | N | \% Detected | $\mu$ | $\sigma$ | $\mathrm{D}_{0.05}$ | D2 | p-value | $\mu$ | $\sigma$ | c | D3m | p-value | K | $\alpha$ | $\xi$ | D31 | p-value |
| TDS (mg/L) | 45 | 97.8 | 4.762 | 0.744 | 0.205 | 0.120 | 0.561 | 4.480 | 0.962 | 20.400 | 0.115 | 0.621 | - | - | - | 0.759 | 0 |
| TSS (mg/L) | 44 | 95.5 | 3.945 | 1.717 | 0.21 | 0.113 | 0.683 | 3.672 | 2.096 | 2.777 | 0.095 | 0.942 | -1.184 | 121.932 | 72.238 | 0.173 | 0.162 |
| BOD (mg/L) | 44 | 86.4 | 1.6211 | 0.66954 | 0.215 | 0.114 | 0.746 | 1.624 | 0.659 | -0.013 | 0.115 | 0.733 | -0.614 | 3.421 | 5.098 | 0.110 | 0.801 |
| COD (mg/L) | 43 | 76.7 | 3.548 | 0.785 | 0.231 | 0.128 | 0.684 | 3.323 | 0.946 | 5.000 | 0.151 | 0.441 | -1.221 | 23.447 | -16.159 | 0.184 | 0.215 |
| Fecal Coliform (Col/100 ml) | 23 | 91.3 | 8.9527 | 1.43209 | 0.287 | 0.181 | 0.503 | 8.725 | 1.692 | 534.506 | 0.217 | 0.278 | -0.791 | 13914.660 | 10684.240 | 0.187 | 0.458 |
| Fecal Streptococcus (C/100 ml) | 22 | 90.9 | 9.6472 | 1.62819 | 0.294 | 0.181 | 0.538 | 9.847 | 1.248 | -1070.380 | 0.165 | 0.676 | -0.599 | 27514.897 | 24175.705 | 0.139 | 0.921 |
| NO2+NO3 (mg/L) | 44 | 84.1 | -0.478 | 1.07398 | 0.234 | 0.141 | 0.463 | -0.378 | 1.173 | 0.135 | 0.141 | 0.464 | -0.579 | 0.746 | 0.759 | 0.122 | 0.664 |
| TKN (mg/L) | 45 | 71.1 | -0.097 | 0.92495 | 0.234 | 0.147 | 0.500 | -0.378 | 1.173 | 0.135 | 0.140 | 0.569 | -0.716 | 0.918 | 0.976 | 0.160 | 0.389 |
| DP (mg/L) | 44 | 79.5 | -2.1844 | 1.08407 | 0.224 | 0.124 | 0.682 | -2.137 | 1.017 | -0.003 | 0.131 | 0.606 | -0.595 | 0.135 | 0.137 | 0.142 | 0.484 |
| TP (mg/L) | 46 | 84.8 | -1.4264 | 1.12592 | 0.215 | 0.116 | 0.696 | -1.499 | 1.188 | 0.009 | 0.126 | 0.584 | -1.999 | 0.161 | 0.171 | 0.211 | 0.062 |
| Total Cadmium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 38 | 55.3 | 0.0291 | 2.62317 | 0.287 | 0.206 | 0.338 | 6.921 | 0.244 | -1000.000 | 0.327 | 0.022 | -1.561 | 7.207 | 2.856 | 0.319 | 0.028 |
| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 36 | 36.1 | -1.79 | 1.66154 | 0.325 | 0.799 | 0.000 | 6.926 | 0.030 | -1000.000 | 0.275 | 0.279 | -1.446 | 10.225 | 5.944 | 0.193 | 0.756 |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | 39 | 74.4 | 2.1184 | 1.20044 | 0.246 | 0.131 | 0.742 | 6.926 | 0.035 | -1000.000 | 0.334 | 0.003 | -1.539 | 8.380 | 6.792 | 0.187 | 0.262 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | 45 | 42.2 | 1.8882 | 1.95431 | 0.301 | 0.132 | 1.034 | 1.634 | 2.355 | 0.188 | 0.203 | 0.417 | -1.406 | 16.381 | 8.735 | 0.174 | 0.632 |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | 45 | 71.1 | 3.6048 | 1.19833 | 0.234 | 0.105 | 0.981 | 3.315 | 1.527 | 3.980 | 0.132 | 0.658 | -1.142 | 44.515 | 36.591 | 0.113 | 0.885 |


| FREEWAYS |  |  | Observed |  |  | 2 Parameter Log |  | 3 parameter maximum likelihood |  |  |  |  | 3 parameter L-moments |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CONSTITUENT | N | \% Detected | $\mu$ | $\sigma$ | $\mathrm{D}_{0.05}$ | D2 | p-value | $\mu$ | $\sigma$ | c | D3m | p-value | $\boldsymbol{\kappa}$ | $\alpha$ | $\xi$ | D31 | p-value |
| Conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ) | 86 | 100 | 4.586 | 0.681 | 0.147 | 0.129 | 0.113 | 4.404 | 0.795 | 12.765 | 0.113 | 0.226 | -1.060 | 58.592 | 87.231 | 0.096 | 0.411 |
| Hardness (mg/L) | 127 | 100 | 3.604 | 0.791 | 0.121 | 0.077 | 0.447 | 3.485 | 0.875 | 3.045 | 0.077 | 0.440 | -1.395 | 22.761 | 30.345 | 0.114 | 0.074 |
| Oil \& Grease (mg/L) | 60 | 71.7 | 1.974 | 0.575 | 0.207 | 0.101 | 0.827 | 1.395 | 0.976 | 2.433 | 0.161 | 0.214 | -0.543 | 4.299 | 7.228 | 0.098 | 0.881 |
| TDS (mg/L) | 97 | 99 | 4.279 | 0.771 | 0.139 | 0.054 | 1.136 | 4.342 | 0.720 | -3.500 | 0.051 | 1.210 | -0.660 | 56.217 | 74.574 | 0.044 | 1.368 |
| TSS (mg/L) | 134 | 99.3 | 4.464 | 1.068 | 0.118 | 0.066 | 0.627 | 4.513 | 1.010 | -2.484 | 0.059 | 0.797 | -1.447 | 76.753 | 75.337 | 0.120 | 0.043 |
| BOD (mg/L) | 26 | 84.6 | 2.262 | 0.894 | 0.281 | 0.103 | 1.252 | 2.013 | 1.084 | 1.385 | 0.099 | 1.304 | -1.212 | 7.478 | 8.170 | 0.117 | 1.097 |
| COD (mg/L) | 67 | 98.5 | 4.552 | 0.980 | 0.167 | 0.107 | 0.445 | 4.798 | 0.715 | -17.925 | 0.073 | 0.999 | -0.881 | 81.577 | 97.057 | 0.082 | 0.817 |
| Fecal Coliform ( $\mathrm{Col} / 100 \mathrm{ml}$ ) | 49 | 100 | 7.585 | 1.716 | 0.194 | 0.105 | 0.677 | 7.514 | 1.811 | 30.900 | 0.128 | 0.398 | -1.569 | 3543.87 | 2053.58 | 0.108 | 0.635 |
| Fecal Streptococcus (C/100 ml) | 25 | 100 | 9.263 | 1.662 | 0.264 | 0.119 | 0.990 | 8.854 | 2.363 | 540.500 | 0.211 | 0.215 | -0.896 | 21116.59 | 15451.91 | 0.132 | 0.833 |
| Ammonia (mg/L) | 79 | 87.3 | 0.009 | 1.094 | 0.164 | 0.133 | 0.173 | 0.141 | 0.942 | -0.083 | 0.106 | 0.423 | -1.065 | 1.00 | 1.01 | 0.123 | 0.252 |
| NO2+NO3 (mg/L) | 25 | 96 | -1.097 | 0.868 | 0.269 | 0.139 | 0.789 | -1.628 | 1.301 | 0.087 | 0.114 | 1.073 | -1.275 | 0.241 | 0.275 | 0.117 | 1.038 |
| TKN (mg/L) | 125 | 96.8 | 0.750 | 0.887 | 0.124 | 0.074 | 0.539 | 0.723 | 0.907 | 0.038 | 0.075 | 0.512 | -1.159 | 1.669 | 1.907 | 0.071 | 0.598 |
| DP ( $\mathrm{mg} / \mathrm{L}$ ) | 22 | 95.5 | -1.226 | 1.188 | 0.287 | 0.170 | 0.593 | -1.670 | 1.595 | 0.053 | 0.207 | 0.333 | -1.992 | 0.181 | 0.190 | 0.255 | 0.131 |
| TP ( $\mathrm{mg} / \mathrm{L}$ ) | 128 | 99.2 | -1.266 | 0.772 | 0.121 | 0.085 | 0.325 | -1.545 | 0.961 | 0.051 | 0.065 | 0.686 | -1.394 | 0.171 | 0.231 | 0.089 | 0.27 |
| Antimony ( $\mu \mathrm{g} / \mathrm{L}$ ) | 14 | 50 | 0.967 | 0.284 | 0.483 | 0.419 | 0.171 | 4.415 | 0.008 | -80.000 | 0.397 | 0.220 | 0.660 | 0.563 | 2.922 | 0.270 | 0.719 |
| Arsenic ( $\mu \mathrm{g} / \mathrm{L}$ ) | 61 | 55.7 | 0.952 | 0.666 | 0.227 | 0.149 | 0.441 | 0.290 | 1.180 | 0.886 | 0.106 | 0.937 | -0.679 | 1.787 | 2.547 | 0.096 | 1.07 |
| Total Cadmium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 95 | 71.6 | 0.016 | 0.838 | 0.165 | 0.052 | 1.380 | -0.108 | 0.937 | 0.081 | 0.067 | 1.087 | -0.781 | 0.028 | 1.574 | 0.051 | 1.409 |
| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | 76 | 98.7 | 2.096 | 0.734 | 0.157 | 0.084 | 0.685 | 2.165 | 0.680 | -0.450 | 0.075 | 0.854 | -0.555 | 6.127 | 8.575 | 0.055 | 1.272 |
| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | 97 | 99.0 | 3.525 | 0.842 | 0.139 | 0.038 | 1.507 | 3.433 | 0.915 | 2.047 | 0.048 | 1.295 | -0.857 | 28.576 | 33.493 | 0.041 | 1.443 |
| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | 107 | 100 | 3.226 | 1.164 | 0.131 | 0.039 | 1.451 | 3.178 | 1.212 | 0.587 | 0.045 | 1.285 | -1.155 | 28.993 | 24.965 | 0.040 | 1.424 |
| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | 99 | 89.9 | 2.325 | 0.673 | 0.144 | 0.091 | 0.460 | 1.989 | 0.896 | 2.228 | 0.062 | 1.013 | -0.960 | 6.390 | 9.308 | 0.073 | 0.769 |
| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | 93 | 96.8 | 5.273 | 0.877 | 0.143 | 0.063 | 0.985 | 5.392 | 0.757 | -17.226 | 0.054 | 1.000 | -0.920 | 156.885 | 189.601 | 0.064 | 0.959 |

Experimental design procedures enable the required sample size to be estimated, according to desired confidence and power of the experimental results. It may be possible, without being able to identify the real skewness, that the best distribution fit could be the gamma or exponential distribution.

The utility of the third parameter has been questioned, especially because one of the objectives in modeling is to be parsimonious. Only in cases where it is important to include the effect of unusual elevated values in the model, is the third parameter recommended. In all the other cases, the use of the 2-parameter distribution is adequate to explain the distribution of most of the contaminants.

When the mean and the standard deviation values are not known, Lilieford's test is recommended to evaluate the goodness of fit to a specific distribution. During this research, the Kolmogorov-Smirnov test was used based on the assumption that the large sample sizes minimized errors associated with small sample sizes and uncertainty in the mean and standard deviation values.

Some constituents (such as TKN, TP, COD and Cu ) show an increase in the p-value when the number of samples is acceptable and the 3-parameter lognormal probability distribution is used. The use of the lognormal distribution also has an advantage over the other distributions because it can be easily transformed to a normal distribution.

The few cases where the gamma distribution seems to be a better model was for cases with low counts (constituents in open space or arsenic, chromium and fecal streptococcus in freeways areas; for example). The exponential distribution better fit total ammonia in freeway areas. The remaining constituents were well represented by the lognormal distribution.

The 2-parameter lognormal distribution is considered the most appropriate distribution to represent stormwater constituents. Its use facilitates statistical analyses of the data, because procedures such as ANOVA or regression require the errors to be normally distributed. If the number of observations is small, the use of nonparametric methods will be required, as the distributions cannot be accurately determined. Some nonparametric methods require symmetry in the data distribution. The log transformed constituent concentrations usually satisfy these assumptions.

# Chapter 5: Identification of Significant Factors Affecting Stormwater Quality Using the NSQD 

## Introduction

The normal approach to classify urban sites for estimating stormwater characteristics is based on land use. This approach is generally accepted because it is related to the activity in the watershed, plus many site features are generally consistent within each land use. Two drainage areas with the same size, percentage of imperviousness, ground slope, sampling methods, and stormwater controls will produce different stormwater concentrations if the main activity in one watershed is an automobile manufacturing facility (industrial land use) while the other is a shopping center (commercial land use) for example. There will likely be higher concentrations of metals at the industrial site due to the manufacturing processes, while the commercial site may have higher concentrations of PAHs (polycyclic aromatic hydrocarbons) due to the frequency and numbers of customer automobiles entering and leaving the parking lots.

The results from the previous chapter indicated that there are significant differences in stormwater constituents for different land use categories. This is supported for other databases like NURP (EPA 1983) and USGS (Driver, et al. 1985). The main question to be addressed in this chapter is if there is a different classification method that better describes stormwater quality, possibly by also considering such factors as geographical area (EPA Rain Zone), season, percentage of imperviousness cover, type of conveyance, controls in the watershed, sampling method, and type of sample compositing, and possible interactions between these factors.

This chapter presents several approaches to explain the variability of stormwater quality by considering these additional factors. As shown in Chapter 3, ignoring the non-detected observations can adversely affect the mean, median and standard deviations of the dataset, and the resulting statistical test results. Therefore, the calculations presented in this chapter were preceded by substituting the censored observations using the Cohen's maximum likelihood method.

## Main Factors Affecting Stormwater Quality

The EPA Rain Zone, percentage of imperviousness, watershed size, land use, type of conveyance, controls in the watershed, sample analysis method, and type of sampling procedures were selected as potential influencing factors affecting stormwater quality for the preliminary analyses in this chapter. Data from sites having a single land use will be used in the basic analyses, while data from the mixed land use sites could be used for verification. The first step was to inventory the total number of events in each of the possible combinations of these factors. The EPA Rain Zone, land use, type of conveyance, type of controls present in the watershed, sampling methods and type of compositing procedures are discrete variables, while percentage of imperviousness and watershed area are continuous variables. The total counts and percentage for each discrete variable option is shown in Table 22.

Table 22. Numbers and percentage of samples by discrete site variable category

| Land use | Events | \% |
| :--- | :---: | :---: |
| Residential | 1042 | 28 |
| Mixed Residential | 611 | 16 |
| Commercial | 527 | 14 |
| Mixed Commercial | 324 | 8.6 |
| Industrial | 566 | 15 |
| Mixed Industrial | 249 | 6.6 |
| Institutional | 18 | 0.48 |
| Open Space | 49 | 1.3 |
| Mixed Open Space | 168 | 4.5 |
| Freeways | 185 | 4.9 |
| Mixed Freeways | 26 | 0.69 |


| EPA Rain Zone | Events | $\%$ |
| :---: | :---: | :---: |
| 1 | 69 | 1.8 |
| 2 | 2000 | 53 |
| 3 | 266 | 7.1 |
| 4 | 212 | 5.6 |
| 5 | 485 | 13 |
| 6 | 356 | 9.5 |
| 7 | 229 | 6.1 |
| 8 | 24 | 0.64 |
| 9 | 124 | 3.3 |


| Controls | Events | \% |
| :--- | :---: | :---: |
| Channel Weirs (CW) | 30 | 0.80 |
| Dry Pond (DP) | 50 | 1.3 |
| Detention Storage (enlarged | 17 | 0.45 |
| pipe) (DS) | 113 | 3.0 |
| Wet Pond at Outfall (WP) | 182 | 4.8 |
| Wet Pond in Watershed |  |  |
| (WP_W) | 42 | 1.1 |
| Wet Pond in Series at Outfall | 4331 | 88 |


| Sample Analysis | Events | \% |
| :--- | :---: | :---: |
| Composite, type not specified | 718 | 19 |
| Flow Composite | 2752 | 73 |
| Time Composite | 295 | 7.8 |


| Type of Conveyance | Events | $\%$ |
| :--- | :---: | :---: |
| Curb and gutter | 2454 | 65 |
| Grass swale | 344 | 9.1 |
| Not specified | 967 | 26 |


| Sampler | Events | \% |
| :--- | :---: | :---: |
| Automatic | 3055 | 81 |
| Manual | 393 | 10 |
| Not specified | 317 | 8.4 |

About $80 \%$ of the samples were collected using automatic samplers. It was observed that manual sampling can result in lower TSS concentrations compared to automatic sampling procedures. This may occur, for example, if the manual sampling team arrives after the start of runoff and therefore misses the first flush (if it exists for the site), resulting in reduced event mean concentrations. For those sites using automatic samplers, about $73 \%$ of the events
were collected using flow-composite samplers, $8 \%$ were collected using time-composite samplers, and about 19\% did not have any designation available. Flow-composite samples are considered more accurate than time-composite samples when obtaining data for event mean concentrations, unless very large numbers of subsamples are obtained (Roa-Espinosa and Bannerman 1995).

Almost $66 \%$ of the events were collected at sites drained with conventional curbs and gutters, $9 \%$ were collected at sites having roadside grass swales, and it was not possible to determine the drainage system for about $25 \%$ of the samples. Grass swales can reduce the concentrations of suspended solids and metals, especially during low flows. They can also infiltrate large quantities of the stormwater, reducing pollutant mass discharges, runoff volume, and peak flows.

## Effects of Stormwater Controls on Stormwater Quality

It is hoped that stormwater controls located in a watershed, or at an outfall, would result in significant reductions in stormwater pollutant concentrations. Figure 36 shows the effects on effluent TSS concentrations when using various controls in residential area watersheds in EPA Rain Zone 2 (Maryland, Virginia, North Carolina, Tennessee and Kentucky), an area having enough samples for an effective statistical analysis. The controls noted for these locations included:

- Channel weir: a flow measurement weir in an open channel that forms a small pool (a very small wet pond). - Dry pond (DP): a dry detention pond that drains completely between each storm event.
- Wet pond (WP): a wet detention pond that retains water between events, forming a small lake or pond. If the pond is in the watershed but not at the outfall, this will be considered a wet pond inside of the watershed (WPW), which would only treat a fraction of the total stormwater from the site
- Detention storage (DS): Oversize pipes with small outlet orifices, usually under parking lots.

The stormwater monitoring was conducted at the outfalls of the drainage areas, after the stormwater controls. Wet ponds are seen to reduce the TSS concentration in the stormwater more than the other controls (about 78\%) compared to the "no control" median value. Detention storage units and dry ponds also reduced the TSS concentrations, but to a smaller extent (about $60 \%$ and $37 \%$ respectively). Only one site (located in Virginia Beach) had a channel weir control, but that site did not reduce the observed TSS concentrations compared to the "no control" category. The effectiveness of the stormwater controls were evaluated for each constituent separately. The effects of sample analysis method, sampler instrument, and type of conveyance were also examined.

The first step was to identify the suitable subsets that could be examined, based on suitable numbers of samples in each category. The following four land uses and EPA Rain Zones had suitable numbers of sites having controls that could be examined: residential, commercial and industrial areas in EPA Rain Zone 2 and industrial areas in EPA Rain Zone 3. For each group, one-way ANOVA analyses were used to identify if there were any differences in the concentrations of 13 constituents (after log-transformations and substitutions for non-detectable values) for those sites that included different controls. Dunnet's method was also used to compare sites with each specific stormwater control type with sites without stormwater controls, using a family error rate of $5 \%$. Table 23 shows the results for these analyses for each of these groups.

Tables 23 through 26 show that there are no significant differences between sites with or without wet ponds for all constituents having observations in industrial land uses in EPA Rain Zone 3. Nitrite-nitrate, total phosphorus, total copper and total zinc were significantly lower in concentrations at sites located in EPA Rain Zone 2, having wet ponds before the outfall, compared to sites without stormwater controls. Wet ponds did not reduce the TKN concentrations in any of the four groups. Significant reductions in TSS concentrations were also observed for sites having wet ponds in residential and commercial land uses, but not in industrial land uses.


Figure 36. TSS distribution by controls in residential areas and EPA Rain Zone 2 (the cross circles indicate the average concentrations, while the median concentrations are written next to the median bar in the box diagrams)
Table 23. One-Way ANOVA Results by Control Type in Residential Land Use, Rain Zone 2



|  | Total Lead $\mu \mathrm{g} / \mathrm{L}$ |  |  | Total Zinc $\mu \mathrm{g} / \mathrm{L}$ |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  |  |  |  |  |  |  |  | $\boldsymbol{0}$ |  |  |  |  |
|  | n | median | Dunnet | n | median | Dunnet |  |  |  |  |  |  |  |
| Weir | 3 | 6.41 | $<$ | 3 | 4.11 | $<$ |  |  |  |  |  |  |  |
| DP | 21 | 1.50 | $<$ | 21 | 29.63 | $<$ |  |  |  |  |  |  |  |
| DS | 9 | 1.16 | $<$ | 9 | 103.25 | $>$ |  |  |  |  |  |  |  |
| No Control | 364 | 7.73 |  | 405 | 67.56 |  |  |  |  |  |  |  |  |
| WP | 4 | 1.00 | $<$ | 4 | 10.44 | $<$ |  |  |  |  |  |  |  |

Table 24. One-Way ANOVA Results by Control Type in Commercial Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0.717 |  |  | 0.082 |  |  | 0.477 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| DS | 8 | 58.17 = |  | 8 | 1.84 | $=$ | 8 | 100.69 | $=$ | 819.54 |  | < |
| No Control | 35 | 58.97 |  | 100 | 4.20 |  | 174 | 74.89 |  | 244 | 48.13 |  |
| WP | 11 | 71.80 |  | 17 | 2.84 |  | 26 | 89.99 |  | 26 | 19.47 | $<$ |
| WPW | 9 | 47.11 |  | 13 | 3.36 |  | 13 | 71.12 |  | 13 16.85 < |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ |  |  | COD mg/L |  |  | Ammonia mg/L |  | $\mathrm{NO}_{2}+\mathrm{NO}_{3} \mathrm{mg} / \mathrm{L}$ |  |  |
| p -value |  | 0 |  |  | 0 |  |  | 0 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| DS | 8 | 4.44 < |  | 8 | 27.18 | < | 8 | 0.30 | = | 8 | 1.18 | $=$ |
| No Control | 241 | 14.66 |  | 174 | 73.62 |  | 174 | 0.39 | $<$ | 242 | 0.60 | $=$ |
| WP | 26 | 7.06 | $<$ | 26 | 35.99 | < | 26 | 0.13 |  | 26 | 0.48 |  |
| WPW | 12 | 5.41 | < | 13 | 23.88 | < | 13 | 0.16 | < | 13 | 0.22 | $<$ |


|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p -value |  | 0.057 |  |  | 0 |  |  | 0 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| DS | 8 | 1.04 | = | 7 | 0.09 | $=$ | 8 | 0.16 | = | 8 | 14.14 | = |
| No Control | 241 | 1.59 |  | 161 | 0.11 |  | 238 | 0.25 |  | 194 | 17.53 |  |
| WP | 26 | 1.19 | = | 25 | 0.05 | = | 26 | 0.13 | < | 6 | 5.57 | $<$ |
| WPW | 13 | 1.03 | = | 13 | 0.03 | = | 13 | 0.08 | < | 4 | 6.00 | < |


|  | Total Lead $\mu \mathbf{g} / \mathrm{L}$ |  |  | Total Zinc $\boldsymbol{\mu g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | $\boldsymbol{0}$ |  |  | $\boldsymbol{0}$ |  |
|  | n | median | Dunnet | n | median | Dunnet |
| DS | 8 | 1.61 | $<$ | 8 | 82.57 | $<$ |
| No Control | 194 | 16.41 |  | 197 | 188.02 |  |
| WP | 7 | 4.90 | $<$ | 7 | 44.26 | $<$ |
| WPW | 4 | 2.49 | $<$ | 4 | 39.68 | $<$ |

Table 25. One-Way ANOVA Results by Control Type in Industrial Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | none |  |  | 0 |  |  | none |  |  | 0.693 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| No Control WP |  |  |  | $\begin{aligned} & 81 \\ & 37 \end{aligned}$ | $\begin{aligned} & 3.85 \\ & 1.43 \end{aligned}$ | < |  |  |  | $\begin{gathered} 205 \\ 29 \end{gathered}$ | $\begin{aligned} & \hline 51.96 \\ & 48.05 \end{aligned}$ | = |



|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0.166 |  |  | none |  |  | 0 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| No Control WP | $\begin{aligned} & 198 \\ & 29 \\ & \hline \end{aligned}$ | $\begin{aligned} & \hline 1.22 \\ & 0.98 \end{aligned}$ | = |  |  |  | $\begin{gathered} 200 \\ 29 \end{gathered}$ | $\begin{aligned} & 0.23 \\ & 0.09 \end{aligned}$ | < | 150 29 | $\begin{aligned} & \hline 16.00 \\ & 7.38 \end{aligned}$ | < |


|  | Total Lead $\mu \mathrm{g} / \mathrm{L}$ |  |  | Total Zinc $\boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0.353 |  |  | $\boldsymbol{0}$ |  |
|  | n | median | Dunnet | n | median | Dunnet |
| No Control | 142 | 11.16 |  | 157 | 180.01 |  |
| WP | 29 | 8.66 | $=$ | 29 | 60.44 | $<$ |

Table 26. One-Way ANOVA Results by Control Type in Industrial Land Use, Rain Zone 3

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | None |  |  | None |  |  | 0.112 |  |  | 0.281 |  |
|  | n | median | Dunnet | n | median | Dunnet | N | median | Dunnet | n | median | Dunnet |
| No Control WP |  |  |  |  |  |  | $\begin{aligned} & 44 \\ & 25 \end{aligned}$ | $\begin{aligned} & 69.53 \\ & 49.84 \end{aligned}$ | = | $\begin{aligned} & 44 \\ & 25 \end{aligned}$ | $\begin{aligned} & 48.35 \\ & 70.40 \\ & \hline \end{aligned}$ | = |


|  | $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ |  |  | COD mg/L |  |  | Ammonia mg/L |  |  | $\mathrm{NO}_{2}+\mathrm{NO}_{3} \mathrm{mg} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0.221 |  |  | 0.395 |  |  | 0.165 |  |  | 0.193 |  |
|  | n | median | Dunnet | n | median | Dunnet | N | median | Dunnet | n | median | Dunnet |
| No Control WP | 44 23 | 6.41 5.14 | $=$ | 44 25 | $\begin{aligned} & 37.00 \\ & 43.06 \end{aligned}$ | $=$ | $\begin{gathered} 3 \\ 25 \end{gathered}$ | $\begin{aligned} & 0.12 \\ & 0.03 \end{aligned}$ | $=$ | 30 25 | $\begin{aligned} & 0.57 \\ & 0.40 \\ & \hline \end{aligned}$ | $=$ |


|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0.807 |  |  | 0.191 |  |  | 0.438 |  |  | 0.106 |  |
|  | n | median | Dunnet | n | median | Dunnet | N | median | Dunnet | n | median | Dunnet |
| No Control WP | 43 25 | 1.18 1.12 | $=$ | 39 25 | $\begin{aligned} & 0.07 \\ & 0.06 \end{aligned}$ | = | 43 25 | $\begin{aligned} & \hline 0.16 \\ & 0.19 \end{aligned}$ | = | 38 25 | $\begin{aligned} & 16.66 \\ & 12.58 \end{aligned}$ | = |

> | WP | 25 | 6.73 | $=$ | 25 |
| :---: | :---: | :---: | :---: | :---: |
| Note. The bold, italicized probability values indicate "statistically significant" findings at the 0.05 level, or better. |  |  |  |  |

Dunnet test compared if sites with control produces larger concentrations " $>$ ", smaller concentrations " $<$ " or not statistically difference " $=$ " than sites without control at a family error of $5 \%$.
"None" indicates no samples were collected for this constituent in the group.

Dry ponds were only available for evaluation in the residential land use category in EPA Rain Zone 2. No significant differences were found for TSS or nitrite-nitrate for sites having dry ponds. However, significant reductions of $\mathrm{BOD}_{5}$, TKN, total phosphorus, total copper, total lead and total zinc were noted.

Some communities have installed detention-storage facilities (enlarged pipes) under parking lots to reduce runoff flow rates. More than 400 of these underground pipes are located in Arlington, Virginia, for example. A significant reduction in the $\mathrm{TSS}, \mathrm{BOD}_{5}, \mathrm{COD}$, total lead, and total zinc concentrations were observed at sites with these underground devices. On the other hand, these controls did not indicate a significant difference in the concentrations of nutrients (ammonia, nitrite-nitrate, TKN, dissolved phosphorus and total phosphorus), compared to comparable sites not having stormwater controls. A conflicting situation was observed in EPA Rain Zone 2 for total zinc for sites having underground enlarged pipes. Zinc concentrations at residential land uses were significantly higher, while zinc concentrations at commercial areas were significantly lower, compared to sites with no stormwater controls. It is possible that the sites having elevated zinc concentrations used galvanized metal enlarged pipe systems.

## Sampling Method Effects on Stormwater Concentrations

The use of manual or automatic sampling is a factor that is sometimes mentioned as having a possible effect on the quality of the collected samples. Manual sampling is usually preferred when the number of samples is small and when there are not available resources for the purchase, installation, operation, and maintenance of automatic samplers. Manual sampling may also be required when the constituents being sampled require specific handling (such as for bacteria, oil and grease, and volatile organic compounds) (ASCE/EPA 2002). Automatic samplers are recommended for larger sampling programs, when better representations of the flows are needed, and especially when site access is difficult or unsafe. In most cases, where a substantial number of samples are to be collected and when composite sampling is desired, automatic sampling can be much less expensive. Automatic samples also improve repeatability by reducing additional variability induced by the personnel from sample to sample (Bailey 1993). Most importantly, automatic samplers can be much more reliable compared to manual sampling, especially when the goal of a monitoring project is to obtain data for as many of the events that occur as possible, and sampling must start near the beginning of the rainfall (Burton and Pitt 2002).

Residential, commercial and industrial sites located in EPA Rain Zone 2 were used to evaluate any significant differences between the two sampling methods. One-way ANOVA analyses were used to identify any statistical differences between the two groups. Dunnet's test was used to compare manual sampling against automatic sampling. Tables 27 through 29 show the results from these ANOVA analyses.
Table 27. One-Way ANOVA Results by Type of Sampler in Residential Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0 |  |  |  |  |  | 0.004 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Automatic | 23 | 51.9 |  | All manual |  |  | 318 | $\begin{aligned} & 65.4 \\ & 50.0 \end{aligned}$ | < | 420 | $\begin{aligned} & 45.5 \\ & 19.2 \\ & \hline \end{aligned}$ |  |
| Manual | 28 | 22.4 | $<$ |  |  |  | 66 |  |  | 78 |  | < |


Table 28. One-Way ANOVA Results by Type of Sampler in Commercial Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0 |  |  | 0.009 |  |  | 0.25 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Automatic | 23 | 97.86 |  | 70 | 4.75 |  | 123 | 76.36 |  | 179 | 52.29 |  |
| Manual | 12 | 22.34 | < | 19 | 2.30 | < | 18 | 60.80 | = | 24 | 20.55 | < |

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|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0.117 |  |  | 0.554 |  |  | 0.003 |  |  | 0.001 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Automatic | 177 | 1.63 |  | 113 | 0.097 | = | 176 | 0.261 | < | 127 | $\begin{aligned} & 20.27 \\ & 11.80 \end{aligned}$ | < |
| Manual | 24 | 1.21 | = | 17 | 0.115 |  | 23 | 0.157 |  | 23 |  |  |


Table 29. One-Way ANOVA Results by Type of Sampler in Industrial Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | none |  |  | 0.723 |  |  | 0.362 |  |  | 0.402 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Automatic Manual | $62 \quad 3.68$ |  |  |  |  |  | $\begin{gathered} 128 \\ 10 \end{gathered}$ | $\begin{gathered} 73.2 \\ 100.0 \end{gathered}$ | $=$ | $\begin{gathered} 171 \\ 19 \end{gathered}$ | $\begin{aligned} & 51.45 \\ & 62.82 \end{aligned}$ | $=$ |
|  | $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ |  |  | COD mg/L |  |  | Ammonia mg/L |  |  | $\mathrm{NO}_{2}+\mathrm{NO}_{3} \mathrm{mg} / \mathrm{L}$ |  |  |
| p-value |  | 0.112 |  |  | 0.371 |  |  | 0 |  |  | 0.021 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Automatic Manual | $\begin{gathered} 166 \\ 19 \end{gathered}$ | 9.65 |  | $\begin{gathered} 127 \\ 10 \end{gathered}$ | 55.02 |  | $\begin{gathered} 122 \\ 10 \end{gathered}$ | 0.243 |  | 163 19 | 0.558 | > |


|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0.008 |  |  | 0.870 |  |  | 0.056 |  |  | 0.797 |  |
|  | n | median | Dunnet | n | median | Dunnet | N | median | Dunnet | N | median | Dunnet |
| Automatic | 164 | 1.135 |  | 109 | 0.091 |  | 166 | 0.214 |  | 108 | 15.66 |  |
| Manual | 19 | 1.944 | > | 10 | 0.086 | $=$ | 19 | 0.315 | $=$ | 22 | 14.97 | $=$ |


|  | Total Lead $\mu \mathbf{g} / \mathbf{L}$ |  |  | Total Zinc $\boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0.908 |  |  | $\mathbf{0 . 0 2 8}$ |  |  |
|  | n | median | Dunnet | n | median | Dunnet |  |
| Automatic | 109 | 11.27 |  | 115 | 156 |  |  |
| Manual | 16 | 10.83 | $=$ | 22 | 233 | $>$ |  |

Note. The bold, italicized probability values indicate "statistically significant" findings at about the 0.05 level, or better.
Dunnet's test compared if sites with controls produce larger concentrations " $>$ ", smaller concentrations " $<$ " or not statistically difference " $=$ " than sites without
controls at a family error of $5 \%$.
"None" indicates no samples were collected for this constituent in the group.

Tables 27 through 29 indicated that $\mathrm{BOD}_{5}$ and dissolved phosphorus measurements are not affected by differences in sampling methods used in residential, commercial or industrial areas in EPA Rain Zone 2. In residential and commercial land uses, TSS and COD concentrations obtained using automatic samplers were almost twice the concentrations obtained when using manual sampling methods. Median total phosphorus concentrations were about $50 \%$ higher using automatic samplers, while no effects were noted for other nutrients. Figure 37 contains box and whisker plots comparing automatic versus manual sampling methods in residential land uses in EPA Rain Zone 2. TSS, total copper and total zinc have lower concentrations using manual sampling compared with automatic sampling ( p -values of $0,0.025$ and 0.02 respectively). The opposite pattern was observed for nitrate-nitrate; manual sampling shows higher concentrations than samples collected with automatic samples ( p -value of 0.005 ).

In industrial land uses, the pattern was found to be opposite. Ammonia, nitrate-nitrite, TKN and total zinc indicated higher concentrations when using manual sampling methods compared to using automatic samplers. Concentrations for these constituents were almost twice as high when using manual sampling, except for ammonia that was almost six times higher when manual sampling was used compared to automatic sampling methods. These elevated concentrations were observed in industrial sites located in Fairfax County Virginia, Howard County Maryland and the city of Charlotte in North Carolina. Sites with controls were not included in this analysis of the effects of sampling method.


Figure 37. Comparison of reported concentrations in residential land use and EPA Rain Zone 2 for automatic vs. manual sampling methods

## Sample Compositing Procedures

Time and flow-weighted composite options were also evaluated in residential, commercial, and industrial land uses in EPA Rain Zone 2 and in industrial land uses in EPA Rain Zone 3. With time-compositing, individual subsamples are combined for even time increments. As an example, automatic samplers can be programmed to collect a subsample every 15 minutes for deposit into a large composite bottle. An automatic sampler can also collect discrete subsamples at even time increments, keeping each sample in a separate smaller sample bottle. After the sampled event, these samples can be manually combined as a composite. With flow-weighted sampling, an automatic sampler can be programmed to deposit a subsample into a large composite bottle for each set increment of flow.

The Wisconsin Department of Natural Resources conducted a through evaluation of alternative sampling modes for stormwater sampling to determine the average pollutant concentrations for individual events (Roa-Espinosa and Bannerman 1995). Four sampling modes were compared at outfalls at five industrial sites, including: flow-weighted composite sampling, time-discrete sampling, time-composite sampling, and first flush sampling during the first 30 minutes of runoff. Based on many attributes, they concluded that time-composite sampling at outfalls is the best method due to simplicity, low cost, and good comparisons to flow-weighted composite sampling (assumed to be the most accurate). The time-composite sampling cost was about $25 \%$ of the cost of the time- discrete and flowweighted sampling schemes, but was about three times the cost of the first flush sampling only. The accuracy and reproducibility of the composite samples were all good, while these attributes for the first flush samples were poor. Burton and Pitt (2001) stress that it is important to ensure that acceptable time-weighted composite sampling include many subsamples. Any sampling scheme is very inaccurate if too few samples are collected. Samples need to be collected to represent the extreme conditions during the event, and the total storm duration. Experimental design methods can be used to determine the minimum number of subsamples needed considering likely variations. It is more common to now include the use of "continuous" water quality probes at sampling locations, with in-situ observations obtained every few minutes. Unfortunately, these details were not available for the NSQD sampling sites; some sites may have had too few subsamples to represent the storm conditions, while others may have had sufficient numbers of subsamples. Also, most of the NSQD samples only represented the first 3 hours of runoff events. If events were longer, the later storm periods were likely not represented. These issues are discussed more in the next subsection.

One-way ANOVA tests were used to evaluate the presence of significant differences between these two composite sampling schemes. Dunnet's comparison test was used to evaluate if concentrations associated with timecompositing were larger or lower than concentrations associated with flow- compositing. Tables 30 through 33 show the results of these tests.
Table 30. One-Way ANOVA Results by Sample Compositing Scheme in Residential Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | none |  |  |  |  |  | 0.229 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | Median | Dunnet | n | median | Dunnet |
| Flow composite |  |  |  | No composite |  |  | 351 | 64.02 |  | 398 | 36.08 |  |
| Time composite |  |  |  |  |  |  | 14 | 76.90 | = | 80 | 90.30 | > |


Table 31. One-Way ANOVA Results by Sample Compositing Scheme in Commercial Land Use, Rain Zone 2



|  | Total Lead $\boldsymbol{\mu g} / \mathbf{L}$ |  |  | Total Zinc $\boldsymbol{\mu} \mathbf{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | $\mathbf{0}$ |  | $\mathbf{0}$ |  |  |
|  | n | median | Dunnet | n | median | Dunnet |
| Flow composite | 115 | 11.96 |  | 115 | 156 |  |
| Time composite | 30 | 52.23 | $>$ | 30 | 408 | $>$ |

Table 32. One-Way ANOVA Results by Sample Compositing Scheme in Industrial Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | none |  |  | Few samples |  |  | 0.076 |  |  | 0 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Flow composite Time composite |  |  |  |  |  |  | $\begin{gathered} 101 \\ 9 \end{gathered}$ | $\begin{gathered} \hline 68.5 \\ 132.9 \end{gathered}$ | = | $\begin{gathered} 116 \\ 40 \end{gathered}$ | $\begin{aligned} & 44.2 \\ & 84.6 \end{aligned}$ | $>$ |




|  | Total Lead $\boldsymbol{\mu g} / \mathbf{L}$ |  |  | Total Zinc $\mu \mathbf{g} / \mathbf{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0.279 |  |  | 0.163 |  |
|  | n | median | Dunnet | n | median | Dunnet |
| Flow composite | 10 | 9.75 |  | 15 | 108.9 |  |
| Time composite | 19 | 6.08 | $=$ | 20 | 161.4 |  |

Note. Dunnet's test compared if sites with control produces larger concentrations " $>$ ", smaller concentrations " $<$ " or not statistically difference " $=$ " than sites
without control at a amily error or $5 \%$.
"None" indicates no samples were collected for this constituent in the group.

Tables 30 through 33 show that no significant differences were observed for $\mathrm{BOD}_{5}$ concentrations using either of the compositing schemes for any of the four categories. A similar result was observed for COD except for commercial land uses in EPA Rain Zone 2, where not enough samples were collected to detect a significant difference. TSS and total lead median concentrations in EPA Rain Zone 2 were two to five times higher in concentration when timecompositing was used instead of flow-compositing.

Nutrients in EPA Rain Zone 2 collected in residential, commercial and industrial areas showed no significant differences using either compositing method. The only exceptions were for ammonia in residential and commercial land use areas and total phosphorus in residential areas where time-composite samples had higher concentrations. Metals were higher when time-compositing was used in residential and commercial land use areas. No differences were observed in industrial land use areas, except for lead. Figure 38 shows box and whiskers plots for TSS using both methods.


Figure 38. Comparisons between time- and flow-composite options for TSS

## Sampling Period during Runoff Event and Selection of Events to Sample

Another potential factor that may affect stormwater quality is the sampling period during the runoff event. Automatic samplers can initiate sampling very close to the beginning of flow, while manual sampling usually requires travel time and other delays before sampling can be started. It is also possible for automatic samplers to represent the complete storm, especially if the storm is of long duration, as long as proper sampler setup programming is performed (Burton and Pitt 2001). However, automatic samplers are not capable of sampling bed load material, and are less effective in sampling larger particles ( $>500 \mu \mathrm{~m}$ ). Manual sampling, if able to collect a sample from a cascading flow, can collect from the complete particle size distribution. Bed load samples and special floatable capture nets may be needed to supplement automatic samplers to obtain information for the complete range of solids.

The NPDES stormwater sampling protocols only required collecting composite samples over the first three hours of the event instead of during the whole event. Truncating the sampling before the runoff event ended may have adversely affected the measured stormwater quality.

Selecting a small subset of the annual events can also bias the monitoring results. In most stormwater research projects, the goal is to sample and analyze as many events as possible during the monitoring period. As a minimum, about 30 samples are usually desired in order to adequately determine the stormwater characteristics with an error level of about $25 \%$ (assuming $95 \%$ confidence and $80 \%$ power) (Burton and Pitt 2001). With only three events per year required per land use for the NPDES stormwater permits, the accuracy of the calculated EMC is questionable until many years have passed. Also, the three storms need to be randomly selected from the complete set of rains in order to be most statistically representative, not just for a narrow range of rain depths as specified in the NPDES sampling protocol.

Flagstaff Street, in Prince George MD, had the most events collected for any site in the NSQD. They collected 28 events during two years of sampling (1998 and 1999). A statistical test was made choosing 6 events (three for each year) from this set, creating 5,600 different possibilities. Figure 39 shows the histogram of these possibilities. The median TSS of the 28 events was $170 \mathrm{mg} / \mathrm{L}$, with a $95 \%$ confidence interval between 119 and $232 \mathrm{mg} / \mathrm{L}$. Only $60 \%$ of the 5,600 possibilities were inside this confidence interval. Almost half ( $40 \%$ ) of the possibilities for the observed EMC would therefore be outside the $95 \%$ confidence interval for the true median concentration if only three events were available for two years. As the number of samples increase, there will be a reduction in the bias of the EMC estimates. In Southern California, Leecaster (2002) determined that ten years of collecting three samples per year was required in order to reduce the error to $10 \%$.


Figure 39. Histogram of possible TSS concentrations in Flagstaff Street based on collecting three samples per year for two years (the measured median TSS concentration was $170 \mathrm{mg} / \mathrm{L}$ )

## Type of Conveyance

Almost all of the samples in the NSQD were collected using automatic samplers and flow compositing. Statistical tests investigating the effects of the type of conveyance only used information from flow-weighted composite samples to reduce potential errors associated with other sampling schemes, as discussed above. Grass swales are considered to be effective stormwater controls compared to conventional curb and gutter stormwater collection systems. Grass swales are commonly found in residential areas with low levels of imperviousness, especially in low density residential areas. NSQD data from residential and mixed residential sites in Virginia, Georgia, and Texas were used to compare stormwater concentrations in areas drained by grass swales and by concrete curbs and gutters.

Historical swale performance tests usually focused on pollutant mass discharges and not concentrations. Swales normally infiltrate significant amounts of the flowing water, resulting in large mass discharge decreases. Most swales operate with relatively deep water, and any "filtering" benefits of the grass (and hence concentration reductions) are usually minimal. Very shallow flows in swales do have particulate pollutant concentration reductions, but these are rarely observed during moderate to large flows (Nara and Pitt 2005).

One-way ANOVA analyses were used to identify any significant differences in stormwater pollutant concentrations between watersheds drained with grass swales or with curbs and gutters. Dunnett's test was used to determine if grass swales produced different concentrations than curbs and gutters. The results are shown in Tables 34 through 37.
Table 35. One-Way ANOVA Results by Type of Conveyance in Industrial Land Use, Rain Zone 2

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | none |  |  | none |  |  | 0 |  |  | 0.023 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter Grass Swale |  |  |  |  |  |  | $\begin{aligned} & 67 \\ & 77 \end{aligned}$ | $\begin{gathered} 45.5 \\ 184 \end{gathered}$ | > | $\begin{gathered} 69 \\ 7 \end{gathered}$ | $\begin{aligned} & 37.52 \\ & 97.70 \end{aligned}$ | > |


|  | $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ |  |  | COD mg/L |  |  | Ammonia mg/L |  |  | $\mathrm{NO}_{\mathbf{2}}+\mathrm{NO}_{3} \mathbf{m g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0 |  |  | 0.035 |  |  | 0.492 |  |  | none |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter Grass Swale | 67 5 | 6.84 39.98 | > | 6650.16 |  |  | 610.223 |  |  |  |  |  |


|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | none |  |  | 0.012 |  |  | 0.468 |  |  | 0.905 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter Grass Swale |  |  |  | $\begin{gathered} 50 \\ 4 \end{gathered}$ | $\begin{aligned} & 0.07 \\ & 0.23 \end{aligned}$ | > | $\begin{gathered} 64 \\ 7 \end{gathered}$ | $\begin{aligned} & 0.174 \\ & 0.232 \end{aligned}$ | $=$ | $\begin{gathered} 20 \\ 7 \end{gathered}$ | $\begin{gathered} 13.0 \\ 12.36 \end{gathered}$ | = |


|  | Total Lead $\boldsymbol{\mu g} / \mathbf{L}$ |  |  | Total Zinc $\boldsymbol{\mu g} / \mathbf{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | none |  |  | 0.447 |  |
|  | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter |  |  |  |  | 20 | 225.7 |
| Grass Swale |  |  |  |  |  |  |

Table 36. One-Way ANOVA Results by Type of Conveyance in Residential Land Uses, Rain Zone 3

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | none |  |  | none |  |  | 0.049 |  |  | 0.425 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter Grass Swale |  |  |  |  |  |  | 11 6 | $\begin{aligned} & 94.06 \\ & 47.84 \end{aligned}$ | < | 12 6 | $\begin{aligned} & \hline 19.2 \\ & 29.6 \end{aligned}$ | = |

\[

\]



|  | Total Lead $\boldsymbol{\mu g} / \mathbf{L}$ |  |  | Total Zinc $\boldsymbol{\mu} \mathbf{g} / \mathbf{L}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p -value |  | 0.154 |  |  | 0.781 |  |  |
|  | n | median | Dunnet | n | median | Dunnet |  |
| Curb and Gutter | 9 | 12.9 |  | 11 | 49.5 |  |  |
| Grass Swale | 6 | 4.20 | $=$ | 6 | 43.0 | $=$ |  |

Table 37. One-Way ANOVA Results by Type of Conveyance in Industrial Land Use, Rain Zone 3

|  | Hardness mg/L |  |  | Oil and Grease mg/L |  |  | TDS mg/L |  |  | TSS mg/L |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p -value |  | none |  |  | none |  |  | 0.134 |  |  | 0.014 |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter Grass Swale |  |  |  |  |  |  | $\begin{gathered} 10 \\ 6 \end{gathered}$ | $\begin{aligned} & \hline 76.74 \\ & 131.6 \end{aligned}$ | $=$ | $\begin{gathered} 10 \\ 6 \end{gathered}$ | 9.68 91.2 | > |


|  | $\mathbf{B O D}_{\mathbf{5}} \mathbf{m g} / \mathbf{L}$ |  |  | COD mg/L |  |  |  | Ammonia mg/L |  |  |  | $\mathbf{N O}_{\mathbf{2}} \mathbf{+} \mathbf{N O}_{\mathbf{3}} \mathbf{m g} / \mathbf{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0.461 |  |  | 0.446 |  |  | none |  |  | none |  |  |  |
|  | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet | n | median | Dunnet |  |  |
| Curb and Gutter | 10 | 4.68 |  | 10 | 29.40 |  |  |  |  |  |  |  |  |  |
| Grass Swale | 6 | 6.61 | $=$ | 6 | 41.26 | $=$ |  |  |  |  |  |  |  |  |


|  | TKN mg/L |  |  | Dissolved Phosphorus mg/L |  |  | Total Phosphorus mg/L |  |  | Total Copper $\mu \mathrm{g} / \mathrm{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $p$-value |  | 0.299 |  |  | 0.077 |  |  | 0.460 |  |  | 0.098 |  |
|  | n | median | Dunnet | n | median | Dunnet | N | median | Dunnet | N | median | Dunnet |
| Curb and Gutter Grass Swale | 6 | $\begin{aligned} & 0.515 \\ & 0.885 \\ & \hline \end{aligned}$ | = | 5 | $\begin{aligned} & \hline 0.046 \\ & 0.027 \\ & \hline \end{aligned}$ | = | 9 6 | 0.138 0.202 | = | 9 6 | 8.57 22.32 | = |


|  | Total Lead $\boldsymbol{\mu g} / \mathbf{L}$ |  |  | Total Zinc $\boldsymbol{\mu g} / \mathbf{L}$ |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| p-value |  | 0.157 |  |  | $\mathbf{0 . 0 0 7}$ |  |
|  | n | median | Dunnet | n | median | Dunnet |
| Curb and Gutter | 4 | 4.86 |  | 9 | 72.86 |  |
| Grass Swale | 6 | 15.5 | $=$ | 6 | 198.9 | $>$ |
| Note. Dunnet's test determined if sites with controls produce larger concentrations ">", smaller |  |  |  |  |  |  | without control at a family error of $5 \%$.

"None" indicates no samples were collected for this constituent in the group.

Total lead and total phosphorus did not have any significant differences in concentrations when comparing the two conveyance systems in both land use areas. Total copper concentrations from residential land uses in EPA Rain Zones 2 and 3 were lower when grass swales were used instead of curbs and gutters. No copper concentration differences were observed at industrial land uses having different conveyance systems.

Figure 40 shows box and whiskers plots for TSS in industrial land uses, EPA Rain Zones 2 and 3 and residential areas in EPA Rain Zone 2. The median concentrations in industrial land uses were smaller in locations where curbs and gutters were used compared to sites having grass swales. The statistical tests did not identify a significant difference between the median concentrations in residential areas in EPA Rain Zone 3 (the residential boxes have much more overlap than for the industrial sites).

## Concentration Effects Associated with Varying Amounts of Impervious Cover

The reported values for imperviousness do not reflect the amount of pavement and roofs that are not directly connected to the drainage system. Directly connected impervious areas (DCIA) are also referred to as effective impervious areas (EIA). For example, imagine a park with a single paved basketball court surrounded by turf; the area of the court will be counted as part of the total impervious area, but would not be considered as part of the effective impervious area. The runoff from the paved court would likely be totally infiltrated by the grass and will not be discharged to the drainage system. In this case, even if we have a value for "total imperviousness," the "effective percentage of imperviousness" is zero.


Figure 40. TSS concentration by type of conveyance (Significant differences were observed in industrial land uses)

It is therefore difficult to compare database concentrations with the imperviousness values due to these potential uncertainties in the actual effective imperviousness. Figure 41 is an example plot of the percent imperviousness values of different land uses for COD. Each vertical set of observations represent a single monitoring location (all of the events at a single location have the same percent imperviousness). The variation of COD at any one monitoring location is seen to vary greatly, typically by about an order of magnitude. These large variations will make trends difficult to identify. All of the lowest percentage imperviousness sites are open space land uses, while all of the highest percentage imperiousness sites are freeway and commercial land uses. This plot shows no apparent trend in concentration that can be explained by imperviousness. However, it is very likely that a significant and important trend does exist between percent effective imperviousness and pollutant mass that is discharged. While the relationship between imperviousness and concentration is not clear, the relationship between effective imperviousness and total runoff volume is much stronger and more obvious as the non-paved areas can infiltrate much water.


Figure 41. Plot of COD concentrations against watershed area percent imperviousness values for different land uses (CO: commercial; FW: freeway; ID: industrial; OP: open space; and RE: residential)

One important feature in the percentage of imperviousness is that most of the residential sites have low levels of imperviousness, while commercial and industrial sites usually have high percentages of imperviousness. Figure 42 shows the mean TSS concentration for residential, commercial and industrial land uses in the database. Only four of the monitored residential watersheds have percentage imperviousness values larger than $60 \%$. Two commercial sites have less than $60 \%$ imperviousness, with the remaining commercial sites above this value. Analyses concerning the effects of impervious cover on stormwater concentrations for each land use separately are difficult as there are limited ranges of impervious cover within each land use category.


Figure 42. TSS concentrations by impervious cover and single land use

Regression analyses were used to identify possible relationships between constituent concentrations and the percentage of imperviousness for residential land use data. Table 38 shows the results from these regression analyses. Residential land uses in EPA Rain Zone 2 were examined during these analyses. Median concentrations from sites using automatic, flow-weighted samplers, and not having any controls and with curb and gutter conveyance systems were selected for analyses. Data from the site KYLOTSR3 were not used during these analyses because sewage disposal facilities were located in the test watershed. Solids and heavy metal median concentrations were higher at this location than for the remaining residential sites in the same Rain Zone.

Only nitrate-nitrite indicated a significant regression relationship between percentage of imperviousness and constituent concentration for these sites, as shown in Figure 43. In this case, the slope was negative, indicating a reduction in the concentration as the level of imperviousness increased. One possible explanation is that the nutrients are associated with landscaped areas and the use of fertilizers which all decrease with increasing impervious areas. This does not indicate that the total mass of nitrate-nitrite will be reduced. The load of this constituent depends on the total runoff volume that is discharged during the event. As the percentage of imperviousness increases, the runoff volume also increases due to lack of infiltration. Even if the concentration is shown to decrease, the total mass discharged may still increase with increasing amounts of pavement or roofs. There was not enough evidence to indicate a relationship between concentration and percentage of imperviousness for the other 11 constituents examined.

Table 38. Regression of Median Concentrations by Percentage of Impervious in Residential land Use, EPA Rain Zone 2

|  |  | Constant |  | Impervious |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Constituent | n | Coefficient | p-value | Coefficient | $p$-value | $R^{2}$ adjusted | Significant at 0.05 level? |
| TDS mg/L | 10 | 71.94 | 0.002 | -0.386 | 0.446 | 0 | Not significant |
| TSS mg/L | 10 | 74.44 | 0.002 | -0.715 | 0.172 | 0.121 | Not significant |
| $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ | 10 | 8.74 | 0.117 | 0.076 | 0.619 | 0 | Not significant |
| COD mg/L | 10 | 53.94 | 0.027 | 0.332 | 0.578 | 0 | Not significant |
| Ammonia mg/L | 10 | 0.319 | 0.052 | -0.002 | 0.639 | 0 | Not significant |
| $\mathrm{NO}_{3}-\mathrm{NO}_{2} \mathrm{mg} / \mathrm{L}$ | 9 | 0.756 | 0 | -0.009 | 0.013 | 0.556 | Significant |
| TKN mg/L | 9 | 1.817 | 0.003 | -0.016 | 0.247 | 0.069 | Not significant |
| DP mg/L | 10 | 0.237 | 0.033 | -0.003 | 0.349 | 0 | Not significant |
| TP mg/L | 10 | 0.561 | 0.002 | -0.006 | 0.13 | 0.171 | Not significant |
| $\mathrm{Cu} \square \mathrm{g} / \mathrm{L}$ | 11 | 16.51 | 0.005 | -0.140 | 0.225 | 0.065 | Not significant |
| $\mathrm{Pb} \square \mathrm{g} / \mathrm{L}$ | 11 | 46.64 | 0.336 | -0.337 | 0.767 | 0 | Not significant |
| $\mathrm{Zn} \mu \mathrm{g} / \mathrm{L}$ | 11 | 98.13 | 0.027 | -0.572 | 0.542 | 0 | Not significant |



Figure 43. Total nitrates regression at different percentages of impervious

The same regression analysis was performed for commercial and industrial land uses in EPA Rain Zone 2. The results of the regression analyses are shown in Table 39.

Table 39. Regression of Median Concentrations by Percentage of Impervious in Commercial and Industrial land use, EPA Rain Zone 2

|  |  | Constant |  | Impervious |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Constituent | n | Coefficient | $\begin{gathered} \mathrm{p}- \\ \text { value } \end{gathered}$ | Coefficient | pvalue | $\begin{gathered} R^{2} \\ \text { adjusted } \\ \hline \end{gathered}$ | Significant at 0.05 level? |
| TDS mg/L | 5 | -4.80 | 0.854 | 0.821 | 0.103 | 0.523 | Not significant |
| TSS mg/L | 5 | -22.01 | 0.406 | 0.805 | 0.097 | 0.541 | Not significant |
| $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ | 5 | -1.80 | 0.879 | 0.153 | 0.410 | 0 | Not significant |
| COD mg/L | 5 | 1.41 | 0.968 | 0.748 | 0.215 | 0.268 | Not significant |
| Ammonia mg/L | 5 | -0.05 | 0.906 | 0.005 | 0.439 | 0 | Not significant |
| $\mathrm{NO}_{3}-\mathrm{NO}_{2} \mathrm{mg} / \mathrm{L}$ | 5 | 0.01 | 0.985 | 0.007 | 0.438 | 0 | Not significant |
| TKN mg/L | 5 | -0.84 | 0.467 | 0.030 | 0.140 | 0.426 | Not significant |
| DP mg/L | 5 | -0.02 | 0.858 | 0.001 | 0.516 | 0 | Not significant |
| TP mg/L | 5 | -0.10 | 0.649 | 0.004 | 0.271 | 0.168 | Not significant |
| Cu $\mu \mathrm{g} / \mathrm{L}$ | 5 | 4.26 | 0.759 | 0.089 | 0.679 | 0 | Not significant |
| $\mathrm{Pb} \mu \mathrm{g} / \mathrm{L}$ | 6 | 15.69 | 0585 | -0.021 | 0.961 | 0 | Not significant |
| Zn $\mu \mathrm{g} / \mathrm{L}$ | 6 | 247.9 | 0.269 | -0.949 | 0.765 | 0 | Not significant |

None of the median stormwater constituents in commercial and industrial areas seem to be affected by changes in impervious cover. There is not enough evidence to indicate a significant relationship between constituent concentration and percentage of imperviousness. More samples will be required to identify those regression relationships.

## Seasonal Effects on Stormwater Quality

Another factor that may affect stormwater quality is the season when the sample was obtained. If the few samples collected for a single site were all collected in the same season, the results may not be representative of the whole year. The NPDES sampling protocols were designed to minimize this effect by requiring the three samples per year to be separated by at least 1 month. The few samples still could be collected within a single season, but at least not within the same week. Seasonal variations for residential stormwater data are shown in Figure 44. These variations are not as obvious as the land use or geographical variations, except for bacteria which appear to be lowest during the winter season and highest during the summer and fall (a similar conclusion was obtained during the NURP, EPA 1983, data evaluations). The database does not contain any snowmelt data, so all of the data corresponds to rainrelated runoff only.


Figure 44. Example residential area stormwater pollutant concentrations sorted by season

## Precipitation Depth Effects on Stormwater Quality

A common assumption is that higher runoff concentrations are associated with smaller rain events. While this has been shown to be true during controlled washoff studies (Pitt 1987), or for sheetflows taken from relatively small paved areas during rains (see Chapter 6 discussion about first flush observations), this has not been frequently detected for samples collected at outfalls for areas having a mixture of surfaces and for typical random periods of high rain intensities. Figure 45 contains several scatter plots showing concentrations plotted against rain depth. There are no obvious trends of concentration associated with rain depth for the NSQD data.


Figure 45. Examples of scatter plots by precipitation depth

Figure 46 shows scatter plots of rainfall and runoff depth for each land use. These should follow a 45 degree line for areas having very large amounts of directly connected impervious areas.

# Scatterplot of Runoff vs Precipitation Depth 



Panel variable: Landuse

Figure 46. Precipitation depth and runoff depth plotted by land use

These plots show much greater scatter than expected. The freeway plot even indicated larger amounts of runoff than precipitation. This may have occurred due to several reasons: (1) the rainfall was not representative of the drainage area being monitored (especially possible for those sites that relied on off-site rain data); (2) the runoff monitoring was inaccurate (possible when the runoff monitoring relied on stage recording devices and the Manning's equation was applied without local calibration); (3) the drainage area was inaccurately delineated; or (4) when base flows contributed significant amounts of runoff during the event. When reviewing the runoff plots provided in some of the annual reports, significant base flows were observed. It was also apparent that these base flows were not subtracted from the total flows recorded during the rain event. The magnitude of the error would be greater for smaller rain events when the base flows could be much larger than the direct runoff quantity. Base flows commonly occur when a local spring or high groundwater levels enter the storm drainage system. In addition, runoff may still be occurring from a prior large event that ended soon before the current event started (the 3 day antecedent dry period requirement for monitored events was intended to minimize this last cause of base flows).

## Antecedent Period without Rain before Monitored Event

The EPA Rain Zones with the longest reported dry interevent periods having data in the NSQD are EPA Rain Zones 6 (southern California) and 7 (Oregon). In these EPA Rain Zones, some antecedent dry periods were reported to be longer than 100 days. Monitored events with the shortest interevent periods of no rains were monitored along the east and south east coasts of the country (EPA Rain Zones 2 and 3). The mean interevent dry period in the western states was about 18 days, while eastern states had mean interevent dry periods of about 5 days. Figure 47 shows box and whisker plots of the number of days having no rain before the monitored event by each EPA Rain Zone.

Samples collected using automatic flow-weighted samplers from watersheds having curbs and gutters and without stormwater controls were used during the following analyses. Only EPA Rain Zone 2 has enough observations to evaluate possible effects of the antecedent dry period on the concentration of stormwater pollutants. Table 40 shows
the results from the regression analyses. In residential land uses, 7 out of 12 constituents indicated that antecedent dry period had a significant effect on the median concentrations. All the regression slope coefficients were positive, indicating that as the number of days having no rain increased, the concentrations also increased.


Figure 47. Box and whisker plot of dry days preceding rain event by EPA Rain Zone

Table 40. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period for Residential Land Use, EPA Rain Zone 2

|  |  | Constant |  | Days since last event |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Constituent | $\mathbf{n}$ | Coefficient | $\mathbf{p -}$ <br> value | Coefficient | p-value | $\mathbf{R}^{\mathbf{2}}$ <br> adjusted | Significant <br> at <br> level? |
| Oil - Grease mg/L | 35 | 0.737 | 0 | -0.364 | 0.062 | 0.074 | No |
| $\mathrm{TDS} \mathrm{mg/L}$ | 208 | 1.761 | 0 | 0.094 | 0.120 | 0.007 | No |
| $\mathrm{TSS} \mathrm{mg/L}$ | 214 | 1.524 | 0 | 0.116 | 0.254 | 0.001 | No |
| $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ | 211 | 0.887 | 0 | 0.211 | $\mathbf{0 . 0 0 4}$ | 0.035 | Yes |
| $\mathrm{COD} \mathrm{mg/L}^{\mathrm{Ammonia} \mathrm{mg} / \mathrm{L}}$ | 206 | 1.682 | 0 | 0.151 | $\mathbf{0 . 0 3 2}$ | 0.018 | Yes |
| $\mathrm{NO}_{3}-\mathrm{NO}_{2} \mathrm{mg} / \mathrm{L}$ | 204 | -0.826 | 0 | 0.300 | $\mathbf{0 . 0 0 3}$ | 0.039 | Yes |
| $\mathrm{TKN} \mathrm{mg/L}$ | 208 | -0.428 | -0.066 | 0.193 | 0.232 | $\mathbf{0 . 0 0 1}$ | 0.049 |
| $\mathrm{DP} \mathrm{mg/L}$ | 203 | -1.061 | 0 | 0.282 | $\mathbf{0 . 0 0 2}$ | 0.043 | Yes |
| $\mathrm{TP} \mathrm{mg/L}$ | 214 | -0.629 | 0 | 0.183 | $\mathbf{0 . 0 0 5}$ | 0.031 | Yes |
| $\mathrm{Cu} \mu \mathrm{g} / \mathrm{L}$ | 58 | 1.082 | 0 | 0.025 | 0.830 | 0 | Yes |
| $\mathrm{Pb} \mu \mathrm{g} / \mathrm{L}$ | 53 | 1.305 | 0 | -0.311 | 0.277 | 0.004 | No |
| $\mathrm{Zn} \mu \mathrm{g} / \mathrm{L}$ | 58 | 1.872 | 0 | -0.058 | 0.764 | 0 | No |

All nutrients (plus organic matter) in residential land uses showed a positive correlation between days since last event and constituent concentration. In all cases, the coefficients of determination $\left(\mathrm{R}^{2}\right)$ were smaller than 0.05 , indicating that relatively little of the total variation was explained by antecedent dry period. Solids and metals were not affected by the antecedent dry period. Figure 48 shows the regression lines and $95 \%$ confidence intervals for four nutrients in residential land uses.


Figure 48. Nutrient concentrations affected by dry periods since last rain in residential land use

Table 41 shows the results from the regression analyses in commercial land uses. Except for nitrates, all the nutrients have positive regressions inside the $95 \%$ confidence interval. In commercial land uses, the effects of antecedent dry periods on the median concentrations were less important. Only total phosphorus and total lead had significant regression results. As in the residential case, phosphorus has a positive coefficient with a small coefficient of determination. However, lead decreases with the number of dry days before the storm.

Table 41. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period for Commercial Land Use, EPA Rain Zone 2

|  |  | Constant |  | Impervious |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Constituent | n | Coefficient | $p$-value | Coefficient | $p$-value | $\begin{gathered} \mathbf{R}^{2} \\ \text { adjusted } \end{gathered}$ | $\begin{aligned} & \hline \text { Significant } \\ & \text { at } 0.05 \\ & \text { level? } \\ & \hline \end{aligned}$ |
| Oil - Grease mg/L | 25 | 0.783 | 0.001 | -0.202 | 0.402 | 0 | No |
| TDS mg/L | 64 | 1.715 | 0 | 0.215 | 0.169 | 0.015 | No |
| TSS mg/L | 82 | 1.506 | 0 | 0.018 | 0.872 | 0 | No |
| $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ | 83 | 0.971 | 0 | 0.149 | 0.176 | 0.01 | No |
| COD mg/L | 64 | 1.670 | 0 | 0.221 | 0.093 | 0.029 | No |
| Ammonia mg/L | 64 | -0.591 | 0 | 0.258 | 0.175 | 0.014 | No |
| NO2 mg/L | 83 | -0.235 | 0 | -0.208 | 0.176 | 0.01 | No |
| TKN mg/L | 83 | -0.006 | 0.949 | 0.196 | 0.109 | 0.019 | No |
| DP mg/L | 61 | -1.329 | 0 | 0.241 | 0.160 | 0.017 | No |
| TP mg/L | 83 | -0.784 | 0 | 0.198 | 0.028 | 0.047 | Yes |
| $\mathrm{Cu} \mu \mathrm{g} / \mathrm{L}$ | 33 | 1.081 | 0 | 0.959 | 0.501 | 0 | No |
| $\mathrm{Pb} \mu \mathrm{g} / \mathrm{L}$ | 33 | 1.498 | 0 | -1.02 | 0.001 | 0.261 | Yes |
| $\mathrm{Zn} \mu \mathrm{g} / \mathrm{L}$ | 32 | 2.21 | 0 | -0.082 | 0.527 | 0 | No |

Figure 49 shows the regression equations for total phosphorus and total lead for data from commercial land uses. The $95 \%$ confidence interval of the regression line for total phosphorus can include zero slope lines. This indicates that there is not a strong correlation between antecedent dry period and total phosphorus concentrations. For total lead, the reduction in concentrations with increasing dry periods is more obvious, but not very explicable.


Figure 49. Total phosphorus and total lead concentrations as a function of antecedent dry period in commercial land use areas

The effect of the antecedent dry period on stormwater concentrations at industrial land uses was not significant, except for TSS, as shown on Table 42. Figure 50 is a plot of the TSS concentrations increasing with increasing dry periods.

Table 42. Regression of Logarithm of Constituent Concentrations by Logarithm of Antecedent Dry Period in Industrial Land Use, EPA Rain Zone 2

|  |  | Constant |  | Impervious |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Constituent | $\mathbf{n}$ | Coefficient | $\mathbf{p -}$ <br> value | Coefficient | p-value | $\mathbf{R}^{\mathbf{2}}$ <br> adjusted | Significant <br> at <br> level? |
| Oil - Grease mg/L | 3 | 0.271 | 0.773 | -0.451 | 0.700 | 0 | No |
| TDS mg/L | 30 | 1.651 | 0 | -0.009 | 0.958 | 0 | No |
| TSS mg/L | 31 | 1.190 | 0 | 0.656 | $\mathbf{0 . 0 2 5}$ | 0.134 | Yes |
| $\mathrm{BOD}_{5} \mathrm{mg} / \mathrm{L}$ | 32 | 0.780 | 0 | 0.201 | 0.202 | 0.022 | No |
| $\mathrm{COD} \mathrm{mg/L}$ | 29 | 1.685 | 0 | 0.071 | 0.622 | 0 | No |
| Ammonia mg/L | 27 | -0.487 | 0.014 | -0.084 | 0.753 | 0 | No |
| $\mathrm{NO2} \mathrm{mg/L}$ | 32 | -0.154 | 0.233 | -0.124 | 0.493 | 0 | No |
| TKN mg/L | 32 | -0.151 | 0.215 | 0.218 | 0.207 | 0.021 | No |
| $\mathrm{DP} \mathrm{mg/L}$ | 28 | -1.176 | 0 | 0.190 | 0.406 | 0 | No |
| $\mathrm{TP} \mathrm{mg/L}$ | 32 | -0.966 | 0 | 0.373 | 0.11 | 0.053 | No |
| $\mathrm{Cu} \mu \mathrm{g} / \mathrm{L}$ | 3 | 1.109 | 0.124 | 0.216 | 0.565 | 0 | No |
| $\mathrm{Pb} \mu \mathrm{g} / \mathrm{L}$ | 3 | 0.882 | 0.197 | 0.119 | 0.787 | 0 | No |
| $\mathrm{Zn} \mu \mathrm{g} / \mathrm{L}$ | 3 | 2.072 | 0.056 | 0.186 | 0.555 | 0 | No |



Figure 50. TSS concentrations vs. dry days since rain event in industrial land use areas

## Trends in Stormwater Quality with Time

Figure 52 shows a plot of lead concentrations for residential areas only (in EPA Rain Zone 2), for the time period from 1991 to 2002. This plot shows likely decreasing lead concentrations with time. Statistically however, the trend line is not significant due to the large variation in observed concentrations ( $p=0.41$; there is insufficient data to show that the slope term is significantly different from zero). Likewise the COD concentrations have an apparent downward trend with time, but again, the slope term is not significant $(\mathrm{p}=0.12)$.


Figure 52. Residential lead and COD concentrations with time (EPA Rain Zone 2 data only)

Except for lead, it is not likely that time between the data collection efforts is the reason why the NURP and NSQD databases have different values.

## Summary

Several factors were evaluated using data from the NSQD. Only residential, commercial and industrial land uses in EPA Rain Zone 2 and industrial areas in EPA Rain Zone 3 have enough samples to evaluate factors affecting stormwater concentrations. The effect of each factor cannot be extrapolated to the rest of the country. However they can be used as guidance for communities in other EPA Rain Zones. Additional data from communities that were not included in this first phase of the NSQD database would enable more complete and sensitive analyses. Also, this chapter examined most of these factors in isolation, more as sensitivity analyses and to help identify significant factors. These analyses did not consider factors together and possible interactions.

There is a significant reduction in TSS, nitrite-nitrate, total phosphorus, total copper, and total zinc concentration at sites having wet ponds, the control practice having the largest concentration reductions. No reductions in TKN concentrations were found using wet ponds, but TKN seems to be reduced by dry ponds. Locations with detention storage facilities had smaller reductions of $\mathrm{TSS}, \mathrm{BOD}_{5}, \mathrm{COD}$, total lead and total zinc concentrations compared to wet pond sites. Unfortunately, there were few sites in the database having grass swales that could be compared with data from sites having curbs and gutters.

The decision to use automatic or manual sampling methods is not always clear. There were statistical differences found between both methods in residential areas for several constituents. Most communities calculate their EMC values using flow-composited sample analyses. If first flush effects are present, manual sampling may likely miss these more concentrated flows due to delays in arriving at the site to initiate sampling. If the first flush is for a very short duration, time-composited samples may overly emphasize these higher flows. Flow compositing produces more accurate EMC values than time composite analyses. An automatic sampler with flow-weighted samples, in conjunction with a bed load sampler, is likely the most accurate sampling alternative.

There is a certain amount of redundancy (self-correlation) between land use and the percentage of impervious areas, as each land use category generally has a defined narrow range of paved and roof areas. Therefore, it is not possible to test the hypothesis that different levels of impervious (surface coverage) are more important than differences in land use (activities within the area). Residential land uses cover only the lower range of imperviousness, while
commercial sites have imperviousness amounts larger than $50 \%$. In order to perform a valid comparison test, the range of imperviousness needs to be similar for both test cases.

Antecedent dry periods before sampling was found to have a significant effect for $\mathrm{BOD}_{5}, \mathrm{COD}$, ammonia, nitrates, TKN, dissolved, and total phosphorus concentrations at residential land use sites. As the number of days increased, there was an increase in the concentrations of the stormwater constituents. This relationship was not observed for freeway sites. This may be associated with the very small drainage areas associated with the freeway sites (drainage areas close to 1 acre), while the drainage areas for residential, commercial and industrial areas ranged between 50 and 100 acres (Figure 2).

No seasonal effects on concentrations were observed, except for bacteria levels that appear to be lower in winter and higher in summer. No effects on concentration were observed according to precipitation depth. Rainfall energy determines erosion and washoff of particulates, but sufficient runoff volume is needed to carry the particulate pollutants to the outfalls. Different travel times from different locations in the drainage areas results in these materials arriving at different times, plus periods of high rainfall intensity (that increase pollutant washoff and movement) occur randomly throughout the storm. The resulting outfall stormwater concentration patterns for a large area having various surfaces is therefore complex and rain depth is just one of the factors involved. The next chapter examines time delivery of pollutants in more detail.

# Chapter 6: Comparisons of First 30-minute Samples to 3-hour Composite Samples 

## Introduction

Sample collection conducted for some of the NPDES MS4 Phase I permits required both a grab and a composite sample for each event. A grab sample was to be taken during the first 30 minutes of discharge, and a flow-weighted composite sample for the entire time of discharge (up to three hours). The initial grab sample was used for the analysis of the "first flush effect," which assumes that more of the pollutants are discharged during the first period of runoff than during later periods. The composite sample was obtained with aliquots collected about every 15 to 20 minutes for at least 3 hours, or until the event ended.

## First Flush

First flush refers to an assumed elevated load of pollutants discharged during the beginning of a runoff event. The first flush effect has been observed more often in small catchments than in large catchments (Thompson, et al, 1995, cited by WEF and ASCE 1998). In another study, large catchments ( $>162 \mathrm{Ha}, 400 \mathrm{acres}$ ) had the highest concentrations observed at the times of flow peak (Soeur, et al. 1994; Brown, et al. 1995). The presence of a first flush also has been reported to be associated with runoff duration by the City of Austin, TX (Swietlik, et al. 1995). Peak pollutant concentrations can occur after the peak discharge, thus some pollutant discharges can be significant for events longer than the time of concentration (Ellis 1986). Adams and Papa (2000), and Deletic (1998) both concluded that the presence of a first flush depends on numerous site and rainfall characteristics.

In this chapter, pollutant characteristics are evaluated using the NSQD database for events that included separate samples collected during both the first 30 minutes and for the entire event (the composite sample), using nonparametric statistical methods. A better analysis of first flush conditions could be performed by using mass discharge curves that relate the total mass discharge as a function of the total runoff volume; however, this procedure requires high resolution flow and concentration information. The NSQD database only contains concentration data from composite samples (and selected first flush samples) and few flow data.

## Methodology

A total of 417 storm events having paired first flush and composite samples were available from the NPDES MS4 database. The majority of the events were located in North Carolina ( $76.2 \%$ ), but some events were also from Alabama (3.1\%), Kentucky (13.9\%) and Kansas (6.7\%). Table 44 shows the events that were used for this analysis, separated by land use and community. All the events correspond to end-of-pipe samples in separate storm drainage systems.

## Table 44. Preliminary Number of Storm Events Selected

| State | Community | CO | FW | ID | IS | OP | RE | Total <br> Events | $\%$ |
| :---: | :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AL | Jefferson County | 5 | 2 | 0 | 0 | 0 | 6 | 13 | 3.1 |
| NC | City of Charlotte | 8 | 0 | 8 | 0 | 3 | 16 | 35 | 8.4 |
| NC | City of Fayetteville | 18 | 0 | 18 | 18 | 6 | 46 | 106 | 25.4 |
| NC | City of Greensboro | 33 | 0 | 33 | 0 | 15 | 33 | 114 | 27.3 |
| KY | City of Lexington | 12 | 3 | 2 | 0 | 2 | 18 | 37 | 8.9 |
| KY | City of Louisville | 0 | 0 | 7 | 0 | 0 | 14 | 21 | 5.0 |
| NC | City of Raleigh | 18 | 0 | 18 | 0 | 9 | 18 | 63 | 15.1 |
| KA | City of Wichita | 7 | 0 | 7 | 0 | 0 | 14 | 28 | 6.7 |
| Total |  | 101 | 5 | 93 | 18 | 35 | 165 | 417 |  |
| Events |  | 24.2 | 1.2 | 22.3 | 4.3 | 8.4 | 39.6 |  | 100 |
| $\%$ |  |  |  |  |  |  |  |  |  |

Note: CO (commercial), FW (freeway), ID (industrial), IS (institutional), OP (Open Space) and RE (residential) land uses

The initial task was to select the constituents and land uses that meet the requirements of the statistical comparison tests. Probability plots, box and whiskers plots, concentration vs. precipitation and standard descriptive statistic calculations were performed for 22 constituents for each land use and all areas combined. Nonparametric statistical analyses were performed after these initial analyses. Mann-Whitney and Fligner-Policello tests were most commonly used. Minitab and Systat statistical programs, along with Word and Excel macros, were used during the analysis.

## Initial Analyses

One of the conclusions of the NURP program was that most of the constituents in stormwater generally follow a lognormal distribution, especially between the 5th and 95 th percentiles (EPA 1983). This characteristic was validated using probability plots during the initial analyses. Results from first flush and composite samples were logtransformed, for different pollutant types, in each land use category.

Figure 53 shows initial statistical results for both phosphorus and COD. Elevated first flush concentrations were evident for COD compared to phosphorus. Probability plots provide useful information about the characteristics of the sample population.


Figure 53. Cumulative probability and box and whiskers plots

Figure 53 is an example for total phosphorus observations from the open space land use. Both sample sets follow a lognormal distribution because most of the points lie on a straight line. The slopes of the lines are different, indicating unequal variances. In this case, about $40 \%$ of the first flush samples did not have detected concentrations for phosphorus, while about $20 \%$ of the composite samples had non-detected phosphorous concentrations. This plot also indicated that the median concentration of the composite samples is almost twice the median value for the first flush samples.

The next initial analysis used box plots. These plots also represent the distribution of the data, but only show the detectable concentrations. The middle line inside the box represents the median of the data. The top of the box represents the third quartile, and the bottom the first quartile. The whiskers are extended from the 5th to the 95th percentile limits. Values outside these limits are represented with asterisks. The exclusion of the non-detected values
changes the median of the data compared to the probability plots. In this example, both of the medians are similar, in contrast with the results of the probability plot. In this example, the variability of the first flush observations is also seen to be larger than the composite data set.

Descriptive statistics for each constituent and land use were calculated to determine if the distributions were symmetrical and if they had the same variance (see Appendix E). This evaluation is needed to select the most appropriate statistical tests. In some conditions, the number of sample pairs was not large enough to allow further analyses. Table 45 shows the results of the initial analysis. Samples having lognormal probability distributions and sufficient data sets were selected for further analyses.

Figure 54 shows the steps that were followed during the nonparametric analysis. The most useful test was the Fligner-Policello test. This test requires independent random samples symmetric about the medians for each data set. The advantage of this test is that it does not require normality or the same variance in each data set (Fligner and Policello 1981). The U statistic and the p-value are shown in the Appendix E for some constituents. Chakraborti (2003) presents a definition and explanation of the Mann-Whitney U test. P-values smaller than $5 \%(<0.05)$ indicate that the first flush and composite sample sets have different median concentrations at the $95 \%$, or greater, confidence level.

Table 45. Initial Analyses to Select Data Sets for First Flush Analyses

| Constituent | CO | ID | IS | OP | RE | ALL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Turbidity, NTU | Selected | No data | No data | Ned | Selected | Selected |
| pH, S.U. | Selected | Selected | No data | Ned | Selected | Selected |
| BOD5, mg/L | Selected | Selected | Box plot FF > Com | Selected | Selected | Selected |
| COD, mg/L | Selected | Selected | Selected | Selected | Selected | Selected |
| TSS, mg/L | Selected | Selected | Selected | Selected | Selected | Selected |
| TDS, mg/L | Selected | Selected | Selected | Selected | Selected | Selected |
| O\&G, mg/L | Selected | Ned | Ned | Ned | Selected | Selected |
| Fecal Coliform, col/100mL | Selected | Ned | Ned | Ned | Selected | Selected |
| Fecal Streptococcus, col/100 mL | Selected | Ned | Ned | Ned | Selected | Selected |
| Ammonia, mg/L | Selected | Selected | Box plot FF > com. | Ned | Selected | Selected |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3,} \mathrm{mg} / \mathrm{L}$ | Selected | Selected | Selected | Selected | Selected | Selected |
| N Total, mg/L | Selected | Selected | Ned | Selected | Selected | Selected |
| TKN, mg/L | Selected | Selected | Box plot FF > com. | Selected | Selected | Selected |
| P Total, mg/L | Selected | Selected | Selected | Selected | Selected | Selected |
| P Dissolved, mg/L | Selected | Selected | Selected | Selected | Selected | Selected |
| Ortho-P, mg/L | Ned | Selected | Ned | Ned | Selected | Selected |
| Cadmium Total, $\mu \mathrm{g} / \mathrm{L}$ | Selected | Selected | Ned | Selected | Selected | Selected |
| Chromium Total, $\mu \mathrm{g} / \mathrm{L}$ | Selected | Selected | Ned | Selected | Selected | Selected |
| Copper Total, $\mu \mathrm{g} / \mathrm{L}$ | Selected | Selected | Selected | Selected | Selected | Selected |
| Lead Total, $\mu \mathrm{g} / \mathrm{L}$ | Selected | Selected | Selected | Selected | Selected | Selected |
| Mercury, $\mu \mathrm{g} / \mathrm{L}$ | Ned | Ned | Ned | Ned | Ned | Ned |
| Nickel, $\mu \mathrm{g} / \mathrm{L}$ | Selected | Selected | Ned | Ned | Selected | Selected |
| Zinc, $\mu \mathrm{g} / \mathrm{L}$ | Selected | Selected | Selected | Selected | Selected | Selected |

* Ned: Not enough data. CO (commercial), FW (freeway), ID (industrial), IS (institutional), OP (Open Space) and RE (residential)


## Nonparametric Analyses

If the number of samples is large, and the distributions are normal and have the same variance, a paired Student's ttest is usually a better test to evaluate the hypothesis and support the results of the Fligner-Policello test. To verify that the data distributions are normal, the Anderson-Darling normality test was used (Kottegoda and Rosso 1997). This method uses an empirical cumulative distribution function to check normality. In Appendix E, the p-values of the paired differences are shown. P -values larger than $5 \%(>0.05)$ indicate that the normality requirement was met at the $95 \%$ or greater confidence level.

Finally, if the first flush and composite sample distributions are symmetrical (but not necessarily normal), and if they have the same variance, the Mann-Whitney test can be used. If the p-value is larger than $5 \%(>0.05)$, the medians of the sample distribution are assumed to be the same, at the $95 \%$ or greater confidence level. The preferred test would be the Student's t-test, if the sample characteristics warrant, followed by the Mann-Whitney test and finally the Fligner-Policello test. The selected cases are only for pairs with concentration values above the detection limits. The ratios between the first flush and composite sample median concentrations are also shown. Commercial and residential areas have the highest ratios for most constituents. The smallest ratios were found for open space sites.

Null Hypothesis: median first flush and composite concentrations are the same


Figure 54. Analysis flow chart

Alternative Hypothesis: median first flush and composite concentrations are different

## Results

About $83 \%$ of the possible paired cases were successfully evaluated. The remaining cases could not be evaluated because the data set did not have enough paired data or they were not symmetrical. Table 46 shows the results of the analysis.

Table 46. Significant First Flushes Ratios (first flush to composite median concentration)

| Parameter | Commercial |  |  |  | Industrial |  |  |  | Institutional |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | SC | R | ratio | n | SC | R | ratio | n | SC | R | ratio |
| Turbidity, NTU | 11 | 11 | = | 1.32 |  |  | X |  |  |  | X |  |
| pH, S.U. | 17 | 17 | = | 1.03 | 16 | 16 | = | 1.00 |  |  | X |  |
| COD, mg/L | 91 | 91 | $\neq$ | 2.29 | 84 | 84 | \# | 1.43 | 18 | 18 | \# | 2.73 |
| TSS, mg/L | 90 | 90 | $\neq$ | 1.85 | 83 | 83 | = | 0.97 | 18 | 18 | $\neq$ | 2.12 |
| $\mathrm{BOD}_{5}, \mathrm{mg} / \mathrm{L}$ | 83 | 83 | $\neq$ | 1.77 | 80 | 80 | $\neq$ | 1.58 | 18 | 18 | \# | 1.67 |
| TDS, mg/L | 82 | 82 | $\neq$ | 1.83 | 82 | 81 | \# | 1.32 | 18 | 18 | \# | 2.66 |
| O\&G, mg/L | 10 | 10 | $\neq$ | 1.54 |  |  | X |  |  |  | X |  |
| Fecal Coliform, col/100mL | 12 | 12 | = | 0.87 |  |  | X |  |  |  | X |  |
| Fecal Streptococcus, col/100 mL | 12 | 11 | = | 1.05 |  |  | X |  |  |  | X |  |
| Ammonia, mg/L | 70 | 52 | \# | 2.11 | 40 | 33 | $=$ | 1.08 | 18 | 16 | \# | 1.66 |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3,} \mathrm{mg} / \mathrm{L}$ | 84 | 82 | \# | 1.73 | 72 | 71 | \# | 1.31 | 18 | 18 | \# | 1.70 |
| N Total, mg/L | 19 | 19 | = | 1.35 | 19 | 16 | = | 1.79 |  |  | X |  |
| TKN, mg/L | 93 | 86 | $\neq$ | 1.71 | 77 | 76 | $\neq$ | 1.35 |  |  | X |  |
| P Total, mg/L | 89 | 77 | \# | 1.44 | 84 | 71 | = | 1.42 | 17 | 17 | = | 1.24 |
| P Dissolved, mg/L | 91 | 69 | = | 1.23 | 77 | 50 | = | 1.04 | 18 | 14 | = | 1.05 |
| Ortho-P, mg/L |  |  | X |  | 6 | 6 | = | 1.55 |  |  | X |  |
| Cadmium Total, $\mu \mathrm{g} / \mathrm{L}$ | 74 | 48 | \# | 2.15 | 80 | 41 | $=$ | 1.00 |  |  | X |  |
| Chromium Total, $\mu \mathrm{g} / \mathrm{L}$ | 47 | 22 | $\neq$ | 1.67 | 54 | 25 | = | 1.36 |  |  | X |  |
| Copper Total, $\mu \mathrm{g} / \mathrm{L}$ | 92 | 82 | $\neq$ | 1.62 | 84 | 76 | \# | 1.24 | 18 | 7 | $=$ | 0.94 |
| Lead Total, $\mu \mathrm{g} / \mathrm{L}$ | 89 | 83 | \# | 1.65 | 84 | 71 | \# | 1.41 | 18 | 13 | \# | 2.28 |
| Nickel, $\mu \mathrm{g} / \mathrm{L}$ | 47 | 23 | $\neq$ | 2.40 | 51 | 22 | = | 1.00 |  |  | X |  |
| Zinc, $\mu \mathrm{g} / \mathrm{L}$ | 90 | 90 | $\neq$ | 1.93 | 83 | 83 | $\neq$ | 1.54 | 18 | 18 | $\neq$ | 2.48 |
| Turbidity, NTU |  |  | X |  | 12 | 12 | = | 1.24 | 26 | 26 | = | 1.26 |
| pH, S.U. |  |  | X |  | 26 | 26 | = | 1.01 | 63 | 63 | = | 1.01 |
| COD, mg/L | 28 | 28 | = | 0.67 | 140 | 140 | \# | 1.63 | 363 | 363 | \# | 1.71 |
| TSS, mg/L | 32 | 32 | = | 0.95 | 144 | 144 | $\neq$ | 1.84 | 372 | 372 | \# | 1.60 |
| BOD5, mg/L | 28 | 28 | = | 1.07 | 133 | 133 | $\neq$ | 1.67 | 344 | 344 | $\neq$ | 1.67 |
| TDS, mg/L | 31 | 30 | = | 1.07 | 137 | 133 | \# | 1.52 | 354 | 342 | $\neq$ | 1.55 |
| O\&G, mg/L |  |  | X |  |  |  | X |  | 18 | 14 | \# | 1.60 |
| Fecal Coliform, col/100mL |  |  | X |  | 10 | 9 | = | 0.98 | 22 | 21 | = | 1.21 |
| Fecal Streptococcus, col/100 mL |  |  | X |  | 11 | 8 | = | 1.30 | 26 | 22 | = | 1.11 |
| Ammonia, mg/L |  |  | X |  | 119 | 86 | \# | 1.36 | 269 | 190 | \# | 1.54 |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3,} \mathrm{mg} / \mathrm{L}$ | 30 | 21 | = | 0.96 | 121 | 118 | \# | 1.66 | 324 | 310 | \# | 1.50 |
| N Total, mg/L | 6 | 6 | = | 1.53 | 31 | 30 | = | 0.88 | 77 | 73 | = | 1.22 |
| TKN, mg/L | 32 | 14 | = | 1.28 | 131 | 123 | $\neq$ | 1.65 | 335 | 301 | $\neq$ | 1.60 |
| P Total, mg/L | 32 | 20 | = | 1.05 | 140 | 128 | $\neq$ | 1.46 | 363 | 313 | \# | 1.45 |
| P Dissolved, mg/L | 32 | 14 | = | 0.69 | 130 | 105 | \# | 1.24 | 350 | 254 | = | 1.07 |
| Ortho-P, mg/L |  |  | X |  | 14 | 14 | = | 0.95 | 22 | 22 | = | 1.30 |
| Cadmium Total, $\mu \mathrm{g} / \mathrm{L}$ | 30 | 15 | = | 1.30 | 123 | 33 | \# | 2.00 | 325 | 139 | \# | 1.62 |
| Chromium Total, $\mu \mathrm{g} / \mathrm{L}$ | 16 | 4 | = | 1.70 | 86 | 31 | = | 1.24 | 218 | 82 | \# | 1.47 |
| Copper Total, $\mu \mathrm{g} / \mathrm{L}$ | 30 | 22 | = | 0.78 | 144 | 108 | \# | 1.33 | 368 | 295 | \# | 1.33 |
| Lead Total, $\mu \mathrm{g} / \mathrm{L}$ | 31 | 16 | = | 0.90 | 140 | 93 | \# | 1.48 | 364 | 278 | \# | 1.50 |
| Nickel, $\mu \mathrm{g} / \mathrm{L}$ |  |  | X |  | 83 | 18 | = | 1.20 | 213 | 64 | \# | 1.50 |
| Zinc, $\mu \mathrm{g} / \mathrm{L}$ | 21 | 21 | = | 1.25 | 136 | 136 | $\neq$ | 1.58 | 350 | 350 | \# | 1.59 |

Note: $\mathrm{n}=$ number of total possible events. $\mathrm{sc}=$ number of selected events with detected values. $\mathrm{R}=$ result. Not enough data (X); not enough evidence to conclude that median values are different $(=)$; median values are different $(\neq)$.

The " $\neq$ " sign indicates that the medians of the first flush and the composite data set are different. The " $=$ " sign indicates that there is not enough information to reject the null hypothesis at the desired level of confidence (at least at the $95 \%$ level). Events without enough data are represented with an " X ".

Also, shown on this table are the ratios of the medians of the first flush to the composite data for each constituent and land use combination. Generally, a statistically significant first flush is associated with a median concentration ratio of about 1.4 , or greater (the exceptions are where the number of samples in a specific category is much smaller). The largest ratios are about 2.5 , indicating that for these conditions, the first flush sample concentrations are about 2.5 times greater than the composite sample concentrations. More of the larger ratios are found for the commercial and institutional land use categories, areas where larger paved areas are likely to be found. The smallest ratios are associated with the residential, industrial, and open spaces land uses, locations where there may be larger areas of unpaved surfaces.

Results indicate that for $55 \%$ of the evaluated cases, the median of the first flush data set were different than the composite sample set. In the remaining $45 \%$ of the cases, both medians were likely the same, or the concentrations were possibly greater later in the events.

Approximately $70 \%$ of the constituents in the commercial land use category had elevated first flush concentrations, about $60 \%$ of the constituents in the residential, institutional and the mixed (mostly commercial and residential) land use categories had elevated first flushes, and only $45 \%$ of the constituents in the industrial land use category had elevated first flushes. In contrast, no constituents were found to have elevated first flushes in the open space category.
$\mathrm{COD}, \mathrm{BOD}_{5}, \mathrm{TDS}, \mathrm{TKN}$ and Zn all had first flushes in all areas (except for the open space category). In contrast, turbidity, pH , fecal coliform, fecal streptococcus, total N, dissolved and ortho-P never showed a statistically significant first flush in any category. The different findings for TKN and total nitrogen imply that there may be other factors involved in the identification of first flushes besides land use.

## Summary

It is expected that peak concentrations generally occur during periods of peak flows (and highest rain energy). On relatively small paved areas, however, it is likely that there will always be a short initial period of relatively high concentrations associated with washing off of the most available material (Pitt 1987). This peak period of high concentrations may be overwhelmed by periods of high rain intensity that may occur later in the event. In addition, in more complex drainage areas, the routing of these short periods of peak concentrations may blend with larger flows and may not be noticeable. A first flush in a separate storm drainage system is therefore most likely to be seen if a rain occurs at relatively constant intensities on a paved area having a simple drainage system.

If the peak flow (and highest rain energy) occurs later in the event, then there likely will not be a noticeable first flush. However, if the rain intensity peak occurs at the beginning of the event, then the effect is exaggerated. Figure 55 shows an example storm in Lexington, KY. Note that in this event there are two periods of elevated peaks, the first occurs one hour after the rain started, the second two hours later. If the concentration remains the same during the entire event, the maximum load will occur during the later periods having the maximum flows (the two peaks), and not during the initial period of the storm. Another factor that needs to be considered is the source of the contaminants and how fast they travel through the watershed. Streets and other impervious areas will contribute flows to the outfall monitoring location before the pervious areas in the drainage area.

Beaumont Center (SE-L2)
Sample 1- December 10, 1999

Rainfall ( 0.50 in )
Flow Rate ( 470752.3 gal)


10 Fri Dec 99
12/10/1999 12:00:00 AM - 12/10/1999 6:00:00 AM
Figure 55. Hydrograph for a storm event (Source: NPDES permit Lexington-KY 2000) ( 1 in $=25.4 \mathrm{~mm}, 1 \mathrm{~m}^{3}=264 \mathrm{gal}$ )


Figure 56. Contributing areas in urban watersheds (Pitt, 1999) ( $\mathbf{~ m}^{3} / \mathrm{s}=35 \mathrm{cfs}$ )
Figure 56 (Pitt 1999) shows that for an example constant rainfall, the source area flow contribution changes for different rain conditions in an area. If the percentage of impervious surfaces is high, many of the constituents will be discharged faster. This observation agrees with the results observed from the statistical analysis. Commercial areas have a larger frequency of high concentrations at the beginning of the event in contrast to open space areas.

Figure 57 shows that for events ( $<12 \mathrm{~mm}$, or 0.5 in ) in this example medium density residential area, most of the runoff is generated by impervious areas. The average percentage of imperviousness for the monitoring sites was examined. Commercial areas had an average of $83 \%$ imperviousness, followed by industrial areas at $70 \%$ imperviousness. Institutional and residential land uses were very similar, with $45 \%$ and $42 \%$ imperviousness respectively. The open space land use category had the smallest imperviousness area, at about $4 \%$. As indicated in Figure 57, larger events can generate more runoff from previous areas than impervious areas. However, it is likely that most of the runoff during the MS4 monitoring activities was associated with the more common small events, and hence, impervious areas were more important.


Figure 57. Contributing areas in urban watersheds (Pitt and Voorhees, 1995) ( 1 in = $\mathbf{2 5 . 4} \mathbf{~ m m}$ )

Probability plots of the precipitation associated with each monitored event for each land use category were prepared to see if there were any significant differences in the ranges of rains observed within each land use category that could have influenced the results. Figure 58 shows that precipitation has the same distribution for almost all the different land uses. The institutional land use category shows a slightly smaller median rain, but this is likely because of the smaller number of events observed in that land use category (18 events). The median precipitation observed during the monitoring at all land uses was about $8 \mathrm{~mm}(0.3 \mathrm{in})$, indicating the importance of runoff from the impervious areas.


Finally, another factor that must be considered is the effect of the sampling duration. The guidance provided for monitoring during the Phase I NPDES activities was to collect a sample during the first 30 minutes of the event, and a composite sample only during the first three hours of the event (or the complete event, if shorter than three hours). Figure 59 shows an example case when these conditions can lead to inappropriate conclusions for longer duration events.

Beaumont Center (SE-L2)
Sample 2 - January 3, 2000


Figure 59. Example of an event with peaks after the sampling period (Source: NPDES permit Lexington-KY, 2000)

The 12 aliquots sampled during the first three hours are shown on the left side of Figure 59. The peak discharge occurred four hours after the event started, as shown on the right side of the figure, and was not represented in the sampling effort. Missing these later storm periods can lead to inappropriate conclusions. It is suggested that for stormwater monitoring, samples should be collected during the complete event and composited before laboratory analyses.

Another sampling example was presented by Roa-Espinosa and Bannerman (1995) who collected samples from five industrial sites using five different monitoring methods. Table 47 shows the ranking of the best methods of sampling based in six criteria. In this table a value between one and five points is assigned to each criterion. Five points indicates that the method is excellent in the specific criterion. Rao-Espinosa and Bannerman concluded that many time-composite subsamples combined for a single composite analysis can provide improved accuracy compared to fewer samples associated with flow-weighted sampling. They also found that time composite subsamples provide better results than samples collected during the first 30 minutes of the event alone.

Table 47. Ranking by Methods of Sampling (Roa-Espinosa, Bannerman, 1995)

| Criteria | Flow <br> Composite | Time <br> Discrete | Time <br> Composite | Old Source <br> Sample | New Source <br> Sample | First 30 <br> Minutes |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Site Selection | 1 | 1 | 1 | 5 | 5 | 3 |
| Cost | 1 | 1 | 5 | 5 | 5 | 5 |
| Technical difficulty | 1 | 1 | 3 | 1 | 5 | 5 |
| Accuracy | 5 | 5 | 4 | 1 | 5 | 1 |
| Reproducibility | 5 | 5 | 5 | 5 | 1 |  |
| Representativeness | 1 | 1 | 3 | 22 | 3 | 1 |
| TOTAL POINTS | 14 | 14 | 19 | 30 | 16 |  |

## Conclusion

A major goal of the present study is to provide guidance to stormwater managers and regulators. Especially important will be the use of this data as an updated benchmark for comparison with locally collected data. In addition, this data may be useful for preliminary calculations when using the "simple method" for predicting mass discharges for unmonitored areas. These data can also be used as guidance when designing local stormwater monitoring programs (Burton and Pitt 2002), especially when determining the needed sampling effort based on expected water quality variations. Additional analyses reported in other chapters expand on these preliminary examples and also investigate other stormwater data and sampling issues.

This investigation of first flush conditions indicated that a first flush effect was not present for all the land use categories, and certainly not for all constituents. Commercial and residential areas were more likely to show this phenomenon, especially if the peak rainfall occurred near the beginning of the event. It is expected that this effect will be more likely to occur in a watershed with a high level of imperviousness, but even so, the data indicated first flushes for less than $50 \%$ of the samples for the most impervious areas. This reduced frequency of observed first flushes in these areas most likely to have first flushes is likely associated with the varying rain conditions during the different events, including composite samples that did not represent the complete runoff durations.

Groups of constituents showed different behaviors for different land uses. All the heavy metals evaluated showed higher concentrations at the beginning of the event in the commercial land use category. Similarly, all the nutrients showed higher initial concentrations in residential land use areas, except for total nitrogen and ortho-phosphorus. This phenomenon was not found in the bacteria analyses. None of the land uses showed a higher population of bacteria during the beginning of the event. Conventional constituents showed elevated concentrations in commercial, residential and institutional land uses.

## Chapter 7: Effects of Land Use and Geographical Location on Stormwater Quality

## Model Building using the NSQD

This chapter describes the methods used to analyze stormwater characterization data in the NSQD in order to determine the best simple method that can be used to calculate the EMC for a site, given the land use, geographical location, and season. These analyses only used those events obtained at single land use sites. This chapter stresses suspended solids analysis as the prototype evaluation procedure that can be used for the other constituents. The later section of this chapter presents results of detailed analyses for other pollutants.

## ANOVA Evaluation of Suspended Solids Data

Total suspended solids is one of the most important constituents in stormwater and is commonly used to measure the effectiveness of controls. Unfortunately, there is much controversy concerning TSS monitoring and laboratory analyses. Automatic samplers cannot include bed load and floatable fractions of the solids, and the samplers have reduced efficiency for larger particles (usually larger than about $300 \mu \mathrm{~m}$ ). In addition, some laboratories improperly allow the samples to settle before analyses in order to obtain only the suspended portion of the sample, and not the non-filterable fraction as defined by Standard Methods. The TSS data in the NSQD were all obtained from outfall monitoring locations, where the amount of particles larger than $300 \mu \mathrm{~m}$ are quite rare, and the laboratories followed proper TSS analytical methods. Analysis of variance (ANVOA) statistical tests were used on natural-log transformed TSS values to identify significant groupings of data, considering both main factors and interactions. The factors examined included land use (residential, commercial, industrial, open space, and freeways), season (spring, summer, fall, and winter) and EPA Rain Zone (the nine EPA rain zones, as shown on Figure 1).


Figure 1. EPA rain zones for the continental US.

## Descriptive TSS Statistics

The first step was to calculate simple descriptive statistics for TSS for each of the main factor categories. The TSS concentrations were log transformed (natural log) in order to preserve the normality assumption in the ANOVA analysis. The number of samples, mean, median, maximum, minimum, among other statistics, were calculated in each level of the main factors. Table 1 shows the descriptive statistics for these factors.

Table 1. Descriptive Statistics of the Natural Logarithm (Ln) of TSS mg/L for Single Landuse Categories

| Descriptive Statistics: LNTSS by Landuse |  |  |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: |
| Variable | Landuse | N | Mean | Median | TrMean | StDev |
| LTSS | CO | 450 | 3.8831 | 3.7377 | 3.8469 | 1.1801 |
|  | FW | 133 | 4.4644 | 4.5951 | 4.4636 | 1.0680 |
|  | ID | 423 | 4.2777 | 4.3567 | 4.2842 | 1.1913 |
|  | OP | 42 | 3.945 | 3.877 | 3.945 | 1.717 |
|  | RE | 977 | 3.8744 | 3.8918 | 3.8650 | 1.1804 |
|  |  |  |  |  |  | Q1 |

Descriptive Statistics: LNTSS by EPA Rain Zone

| Variable | EPA_Rain | N | Mean | Median | TrMean | StDev |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LTSS | 1 | 42 | 3.862 | 3.761 | 3.877 | 1.268 |
|  | 2 | 1161 | 3.7446 | 3.7612 | 3.7376 | 1.1086 |
|  | 3 | 120 | 3.906 | 3.880 | 3.898 | 1.389 |
|  | 4 | 218 | 4.5466 | 4.4426 | 4.5320 | 1.4053 |
|  | 5 | 152 | 4.3056 | 4.3437 | 4.3124 | 1.0898 |
|  | 6 | 159 | 4.6129 | 4.7005 | 4.6011 | 1.0135 |
|  | 7 | 141 | 4.1096 | 4.2047 | 4.1142 | 1.0561 |
|  | 8 | 7 | 4.221 | 3.970 | 4.221 | 0.794 |
|  | 9 | 25 | 5.412 | 5.587 | 5.414 | 0.882 |
| Variable | EPA_Rain | SE Mean | Minimum | Maximum | Q1 | Q3 |
| LTSS | 1 | 0.196 | 1.099 | 6.447 | 2.996 | 4.825 |
|  | 2 | 0.0325 | 1.0986 | 7.0867 | 2.9957 | 4.4543 |
|  | 3 | 0.127 | 1.099 | 7.030 | 2.773 | 4.940 |
|  | 4 | 0.0952 | 1.6094 | 7.8087 | 3.4657 | 5.6204 |
|  | 5 | 0.0884 | 1.0986 | 7.8200 | 3.6636 | 5.1044 |
|  | 6 | 0.0804 | 1.3863 | 8.4764 | 4.0431 | 5.1330 |
|  | 7 | 0.0889 | 1.0986 | 6.9847 | 3.4340 | 4.7664 |
|  | 8 | 0.300 | 3.367 | 5.858 | 3.829 | 4.477 |
|  | 9 | 0.176 | 3.714 | 7.056 | 4.684 | 6.019 |

There are enough samples to identify if there are any significant differences among the levels and factors, although EOA Rain Zones 1, 8, and 9 and open space have fewer than 50 samples. The range between the minimum and
maximum values are similar for all the groups，indicating that there are not any unusual extreme high or low concentration values in the data set．The mean and median values are also close（after the natural－log transformations）indicating data symmetry for each factor level．

During the ANOVA analyses，each factor was identified as a discrete variable．The partial sum of squares was used to identify the effects of the interactions．The results of the ANOVA（using DataDesk 6.1 from MBAWare）， including all the interactions are：

```
* 国 Dependent variables
\begin{tabular}{ll} 
Name & Code \\
LTSS & \(L T S\)
\end{tabular}
Tupe of aralusis：OLS ANOVA
\(\forall\) 国 Factors
\begin{tabular}{|c|c|c|c|c|c|}
\hline Name & Code & Nested in & F／R & Kind & \\
\hline Landuse & Lnd & （） & Fix & Dise & \\
\hline Season & SEn & C） & Fix & Dise & \\
\hline RZone & F Zn & ¢ & Fix & Dise & \\
\hline Fartiol & ype 3 ） & Sums of Squ & & & Deesign Help \\
\hline
\end{tabular}
＊国 Interactions up to All－way
（Add Interaction）（Remove Seleoted Terms）（Up）Down）
\begin{tabular}{|c|c|c|c|c|}
\hline Source & F／R & max df & EMS & F－Denom \\
\hline Const & － & 1 & Const & Error \\
\hline Lrid & F & 4 & Lrid & Error \\
\hline Sm & F & 3 & Sen & Error \\
\hline R Zn & F & 8 & r Zn & Error \\
\hline Lnd＊SEn & F & 12 & Lnd＊Sen & Error \\
\hline Lrid＊R： & F & 32 & Lrid＊REn & Error \\
\hline Sm＊RZn & F & 24 & Sm＊RZn & Error \\
\hline Lnd＊SEni＊RZn & F & 96 & Lnd＊SEn＊RZn & Error \\
\hline Error & R & 1846 & & \\
\hline Total & & 2025 & & \\
\hline
\end{tabular}
```


## $\geqslant$ 国 No Modifications <br> $\forall$ 國 General Results

```
2026 total coses
\(\forall\) 国 ANOVA
Analysis of Varianee For LTSS
No Seleotor
\begin{tabular}{|c|c|c|c|c|c|}
\hline Source & df & Surns of Squares & Mean Square & F－ratio & Prob \\
\hline Const & 1 & 32463.3 & 32463.3 & 25689 & \(\leq 0.0001\) \\
\hline Lnd & 4 & 16.0568 & 4.41271 & 3． 1754 & 0.6130 \\
\hline \(\operatorname{Sm}\) & 3 & 5.00983 & 1.66994 & 1.3215 & 0.2657 \\
\hline R：Zn & 8 & 29.529 & 3.69113 & 2.9269 & 0.61630 \\
\hline Lnd＊ \(\mathrm{Sm}_{\text {m }}\) & 12 & 5． 16695 & 0.430579 & 0.34073 & 0.9817 \\
\hline Lnd＊RZn & 18 & 32.4181 & 1.801 & 1.4252 & Q． 1099 \\
\hline SEn＊RZn & 20 & 31.7578 & 1.58789 & 1.2565 & 0． 1982 \\
\hline Lnd＊Ssn＊RZn & 39 & 48.9396 & 1.25486 & 0.99301 & 0.4830 \\
\hline Error & 1921 & 2427.55 & 1.26369 & & \\
\hline Total & 2625 & 2951.29 & & & \\
\hline
\end{tabular}
```


## $\rangle$ 罒 Results for fractor Lnd

The probability value for the 3－way interaction term（ 0.4830 ）shows that this interaction is not significant in the model．After deleting this three－way interaction，the new ANOVA table is：

| Analysis of Yariance For No Selector |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Source | df | Surns of Squares | Mean Square | F-ratio | Prob |
| Gonst | 1 | 32463.3 | 32463.3 | 25693 | $\leq 0.0601$ |
| Lhud | 4 | 19.8304 | 4.9576 | 3.9236 | 6. 016.35 |
| Sen | 3 | 8.5914 | 2.8638 | 2.266 .5 | 6.0789 |
| R Z ${ }^{\text {a }}$ | 8 | 55.6248 | 6.87899 | 5.4436 | $\leq$ 9.00] 1 |
| Lnd*SEn | 12 | 40.5535 | 3.37946 | 2.6746 | Q. 01614 |
| Lnd*RZn | 18 | 74.5744 | 4. 14362 | 3.279 | ¢ 日.00] |
| Sen* C Zr | 20 | 46.6732 | 2.60366 | 1.5858 | Q. 0477 |
| Error | 1960 | 2476.49 | 1.26352 |  |  |
| Total | 2025 | 2951.29 |  |  |  |

In this case, season and season-rain zone interaction seems not to be significant (probability $>0.05$ ), while all of the remaining factors seem to be important. The mean square error (MSE) is an estimator of the variance in the model. The lower the MSE, the better the model. It was observed that deleting any other source would increase the MSE. The assumption of normality and independence of the residuals for this result was also evaluated as shown in Figure 3.


Figure 3. ANOVA results for LN of TSS mg/L using single land uses.

There are not any unusual patterns in the predicted vs. studentized residuals plot. The residuals seem to be independent and normally distributed. The next step is to check if there are any values having large influences or residuals. The potential-residual plot (potential=influence) indicates the data points that have a high influence or residual in the model. A point with an elevated influence indicates that if the point is removed from the dataset, the slope and intercept of the regression line will be affected significantly. DataDesk uses Hadi's influence measure method in preparing the residual-potential plot. The plot identifies unusual observations if they are outside an area described by a hyperbolic trend. Another useful measure is the Cook's distance that considers the influence of each case in all the values. Figure 4 shows the potential residual plot for the natural logarithm of the TSS data.


Figure 4. Potential-Residual plot of the natural logarithm of TSS (single land uses).

In this case, the potential residual plot does not indicate any unusual observations in the dataset. All the observations followed a hyperbolic trend; there are not any points outside the area described by this hyperbola. The box plot of the Hadi's influence parameter also does not show any single observation that will influence the whole dataset. The box plot of the Cook's distance indicates a potentially unusual observation. This observation corresponds to a concentration of $46 \mathrm{mg} / \mathrm{L}$ in a residential area in EPA Rain Zone 8 during the summer. The unusual characteristic of this observation is that it is the only observation in the database with these characteristics. If this observation is not included in the data plot, the results do not change. The largest influence point is an observation having a concentration of $825 \mathrm{mg} / \mathrm{L}$ in an open space area in EPA Rain Zone 2 during the spring. This concentration is not common but it can occur. These data were not deleted from the dataset. Figure 5 shows the box plot of both influence methods.


Figure 5. Influence box plot (Cook's distance and Hadi's measure methods).
Because there are no unusual observations, it is possible to evaluate the coefficients for each factor with all the data. A complete examination (all single and multiple interactions) of the coefficients is shown in Table 2.

Table 2. Significant Coefficients for the Complete Factorial Model

| LN(TDS mg/L) COEFFICIENTS FOR DIFFERENT LAND USES, SEASONS AND RAIN ZONES |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Level | Coefficient | std. err. | t Ratio | prob |
| Constant |  | 4.642 | 0.1742 | 26.65 | < 0.0001 |
| Landuse | RE | -0.4871 | 0.1495 | -3.258 | 0.0011 |
| Season | None |  |  |  |  |
| Rain Zone | 2 | -0.8947 | 0.2155 | -4.152 | < 0.0001 |
|  | 4 | 0.2949 | 0.1463 | 2.016 | 0.0439 |
|  | 9 | 0.976 | 0.4261 | 2.291 | 0.0221 |
|  |  |  |  |  |  |
| Landuse*Season | SP,OP | -0.6637 | 0.3246 | -2.045 | 0.0410 |
|  | SU,CO | -0.4617 | 0.1825 | -2.53 | 0.0115 |
|  | SU,OP | 0.6597 | 0.3021 | 2.183 | 0.0291 |
|  |  |  |  |  |  |
| Landuse*Rain Zone | 1,ID | -2.492 | 0.7161 | -3.479 | 0.0005 |
|  | 1,RE | -0.4554 | 0.2031 | -2.242 | 0.0251 |
|  | 2,RE | 0.4554 | 0.2031 | 2.242 | 0.0251 |
|  | 5,CO | -1.385 | 0.6483 | -2.136 | 0.0328 |
|  |  |  |  |  |  |
| Season*Rain Zone | 2,FA | -0.3648 | 0.1386 | -2.632 | 0.0086 |
|  | 4,SP | 0.5739 | 0.1666 | 3.444 | 0.0006 |
|  | 4,SU | -0.5285 | 0.1961 | -2.695 | 0.0071 |
|  | 7,SU | 0.6614 | 0.3321 | 1.991 | 0.0466 |
|  |  |  |  |  |  |
| Landuse*Season*Rain Zone | None |  |  |  |  |

There are 180 possible combinations between the land uses (5), seasons (4) and rain zones (9). The estimated value for any combination is the sum of the coefficients under the conditions of the observation. If the term for a condition being examined is not shown, it was not significant and a zero value is used. Otherwise, the coefficient corresponding to the site condition is used. For example, the following is used to estimate the log value of the TSS
for an observation in EPA Rain Zone 4 during spring in a commercial land use. According to Table 2, the expected value is:

Concentration $=$ constant + landuse + season + rain zone + landuse*season + landuse*rain zone + season* rain zone
$\mathrm{Y}=4.642+0+0+0.2949+0+0+0.5739=5.511$
This corresponds to an expected mean concentration of $247 \mathrm{mg} / \mathrm{L}$. The TSS data in the database for this same group has a mean value of $299 \mathrm{mg} / \mathrm{L}$. This difference is well within the expected error.

After calculating the expected means for each of the 180 possible combinations, a dot plot was created to determine if some groups overlap. For example, it is expected that many of the observations in EPA Rain Zones 1, 3, 4, 6, 7 and 8 will have the same expected TSS concentration values because there were no variations by season for any land uses, except for the residential area. The dot plot of the 180 combinations is shown in Figure 6.


Figure 6. Dot plot of estimated concentrations of Ln TSS.
Figure 6 shows that there are about 17 different groups, at the most. The ANOVA model was reviewed to determine which of the main factors or interactions were the most important. The interaction of Land use*Rain Zone produces by itself the smallest MSE. The new ANOVA table using this interaction is shown in Table 3. The new MSE is 1.29 and is not much larger from the previous MSE using all the significant factors in the model (1.26). Table 4 shows the relevant coefficients using only the reduced model.

Table 3. ANOVA Table using Land Use - Rain Zone Interaction

## $\geqslant$ 䃈 General Results

## 2026 total gases

$\forall$ 國 ANOVA
Arolysis of Vorionee For LTSS
No Selector

| Source | df | Sums of | Squares | Mean Square | F-rotio | Prob |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Const | 1 | 32463.3 |  | 32463.3 | 25678 | $\leq 0.0601$ |
| Lnd*RZn | 29 | 367.524 |  | 12.6732 | 9.7908 | $\leq 6.061$ |
| Error | 1996 | 2583.76 |  | 1.29447 |  |  |
| Total | 2025 | 2951.29 |  |  |  |  |

$\Rightarrow$ Results for factor Lnd*RZn

Table 4. Reduced Model
LN (TDS mg/L) COEFFICIENTS FOR DIFFERENT LAND USES, SEASONS AND RAIN ZONES

|  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Level | Coefficient | std. err. | t Ratio | prob |
| Constant |  | 4.098 | 0.07133 | 57.45 | <0.0001 |
|  |  |  |  |  |  |
| Landuse | None |  |  |  |  |
|  |  |  |  |  |  |
| Season | None |  |  |  |  |
|  |  |  |  |  |  |
| Rain Zone | None |  |  |  |  |
|  |  |  |  |  |  |
| Landuse*Season | None |  |  |  |  |
|  |  |  |  |  |  |
| Landuse*Rain Zone | 1,ID | -0.8756 | 0.3235 | -2.706 | 0.0069 |
|  | 2,CO | -0.3258 | 0.0954 | -3.415 | 0.0007 |
|  | 2,FW | 1.273 | 0.331 | 3.845 | 0.0001 |
|  | 2,OP | -0.4027 | 0.2081 | -1.935 | 0.0531 |
|  | 2,RE | -0.3944 | 0.08116 | -4.859 | <0.0001 |
|  | 3,CO | -0.5965 | 0.2429 | -2.455 | 0.0142 |
|  | 4,CO | 0.4018 | 0.1705 | 2.357 | 0.0185 |
|  | 4,ID | 0.5327 | 0.1611 | 3.306 | 0.001 |
|  | 4,OP | 0.8354 | 0.3235 | 2.582 | 0.0099 |
|  | 4,RE | 0.3287 | 0.1309 | 2.512 | 0.0121 |
|  | 5,CO | -0.7134 | 0.2528 | -2.821 | 0.0048 |
|  | 5,ID | 0.8913 | 0.1876 | 4.751 | <0.0001 |
|  | 6,CO | 0.6862 | 0.3504 | 1.958 | 0.0503 |
|  | 6,FW | 0.3606 | 0.1324 | 2.723 | 0.0065 |
|  | 6,ID | 1.229 | 0.2706 | 4.54 | <0.0001 |
|  | 6,RE | 0.4203 | 0.217 | 1.937 | 0.0529 |
|  | 7,FW | 0.4948 | 0.2343 | 2.112 | 0.0348 |
|  | 7,ID | 0.6974 | 0.2429 | 2.87 | 0.0041 |
|  | 7,RE | -0.5442 | 0.1612 | -3.376 | 0.0007 |
|  | 9,CO | 0.9278 | 0.3859 | 2.404 | 0.0163 |
|  | 9,ID | 1.916 | 0.3859 | 4.966 | <0.0001 |
|  |  |  |  |  |  |
| Season*Rain Zone | None | 0 |  |  |  |
|  |  |  |  |  |  |
| Landuse*Season*Rain Zone | None | 0 |  |  |  |

All land uses in EPA Rain Zone two (except for freeways) have reduced TSS values when compared with the group average. On the other hand, conditions in EPA Rain Zones 4, 6 and 9 have higher TSS values for the land uses noted. Notice also that industrial and freeway land uses increase the TSS concentrations compared with the other land uses, as expected from the one-way ANOVA tests. Of the 45 possible EPA Rain Zone and land use
interactions, 21 have significantly different coefficients and resultant TSS concentrations. All of these possible TSS concentrations, based on this model, are shown in Table 4 b .

Table 4b. TSS Concentrations (mg/L) for Different Land Uses and Rain Zones (if values not shown, use $60 \mathrm{mg} / \mathrm{L}$ )

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Open space |  | 40 |  | 139 |  |  |  |  |  |
| Residential |  | 40 |  | 84 |  | 92 | 35 |  |  |
| Commercial |  | 43 | 33 | 90 | 30 | 120 |  |  | 152 |
| Industrial | 25 |  |  | 103 | 147 | 206 | 121 |  |  |
| Freeways |  | 215 |  |  |  | 86 | 99 |  | 409 |

Figure 7 shows the groups using the land use*rain zone model. A further reduction in the number of similar groups is not likely possible with this model.

Dotplot for LN(TSS mg/L) using Land Use - Rain Zone interaction


Each dot represents up to 2 observations
Figure 7. Dot plot using the reduced model

Out of the 45 total land use-rain zone groups, 24 (or $53 \%$ ) are not affected by significant land use - EPA Rain Zone interaction terms. Seven of the 21 significant groups are smaller than the overall average condition ( $60 \mathrm{mg} / \mathrm{L}$ ), while 14 are larger. Only 2 percent of the observations have very large concentrations, they were located in industrial land uses in EPA Rain Zone 9. Figure 8 shows the 5 groups identified with the ANOVA analysis. The variation within the groups is the same as the variation for the whole dataset. The two separate groups located in the upper tail are important. It is not recommended to merge these groups because their concentration differences are very large.


Figure 8. Probability plot using the reduced model (average of the tied points).


Each dot represents up to 2 observations.
Figure 9. TSS data groups in real space.

There are 2,025 TSS observations for single land uses sites in the NSQD. These observations were classified according to the five groups identified by the above ANOVA model. Figure 11 is a box-whisker plot showing the medians, and $25^{\text {th }}, 75^{\text {th }}, 5^{\text {th }}$, and $95^{\text {th }}$ percentiles for each of these groups.

## TSS mg/L for single landuses using NSQD



Figure 11. Box plots of five groups also showing $5^{\text {th }}$ and $95^{\text {th }}$ percentiles.

Figure 11 indicates that about half of the TSS single land use data in the NSQD database were in the first group ( $52 \%$ ). Most of this data are from residential areas and EPA Rain Zone 2. Twenty-four percent of the observations were not affected by the land use - EPA Rain Zone interaction. Only $1.5 \%$ of the data are present in groups 4 and 5 . These groups are significantly different than groups 1 and 2. Overall, there are three main levels of TSS concentrations in stormwater: Low (1), Medium (2) and High (3). Other minor categories correspond to groups 4 and 5 and contain the unusually high values.

Table 4c. Five TSS Concentration Categories in NSQD

|  | Land use*rain zone interactions (Rain Zone: land uses) | Concentrat ions (mean $\pm$ st. dev. In mg/L) | Range <br> (mean; mean - st. dev. and mean + st. dev., mg/L) | Number of single land use TSS observations in category in NSQD |
| :---: | :---: | :---: | :---: | :---: |
| Low | 1: residential <br> 2: open space; residential; commercial <br> 3: commercial <br> 5: commercial <br> 7: residential | $3.69 \pm 1.12$ | $40(13-123)$ | 1056 |
| Medium | All others not noted elsewhere | $4.02 \pm 1.11$ | 56 (18-169) | 478 |
| High | 4: residential; commercial; industrial; open space <br> 5: industrial <br> 6: freeways; residential; commercial <br> 7: freeways; industrial <br> 9: commercial | $4.60 \pm 1.20$ | 99 (30-330) | 460 |
| Unusually high 1 | 2: freeways <br> 6: industrial |  |  | 22 |
| Unusually high 2 | 9: industrial |  |  | 9 |

To evaluate if groups 1 (low) and 2 (medium) are from the same population, a two-sample $t$ test was calculated. The results are as follows:

Two-sample $T$ for LTSSG1 vs LTSSG2

|  | N | Mean | StDev | SE Mean |
| :--- | ---: | ---: | ---: | ---: |
| LTSSG1 | 1056 | 3.69 | 1.12 | 0.034 |
| LTSSG2 | 478 | 4.02 | 1.11 | 0.051 |

Difference $=$ mu LTSSG1 - mu LTSSG2
Estimate for difference: -0.3370
95\% CI for difference: (-0.4577, -0.2162)
T-Test of difference $=0$ (vs not $=$ ) : T-Value $=-5.47 \quad$ P-Value $=0.000 \quad$ DF $=1532$
Both use Pooled StDev $=1.12$

This test indicates that both groups are from different populations with a $p$ value close to zero. The assumption of equal variances is also valid. The same procedure can be used to evaluate if group 2 (medium) and group 3 (high) are from the same population. The results are as follows:

```
Two-sample T for LTSSG2 vs LTSSG3
\begin{tabular}{lrrrr} 
& \(N\) & Mean & StDev & SE Mean \\
LTSSG2 & 478 & 4.02 & 1.11 & 0.051 \\
LTSSG3 & 460 & 4.60 & 1.20 & 0.056
\end{tabular}
Difference = mu LTSSG2 - mu LTSSG3
Estimate for difference: -0.5789
95% CI for difference: (-0.7273, -0.4305)
T-Test of difference = 0 (vs not =) : T-Value = -7.66 P-Value = 0.000 DF = 936
Both use Pooled StDev = 1.16
```

The variance of both samples are within $10 \%$. The T statistic and p value corroborates that both distributions are from different populations. A grouped probability plot of the five groups is shown in Figure 12.

## Lognormal base 10 Probability Plot for TSS (mg/l) By Group <br> ML Estimates



Figure 12. Probability plot of the TSS data in the NSQD using the 5 main sample groups.

The three main groups are clearly defined. Groups 4 and 5 do not have the same numbers of observations as the other groups, so the parameters are not as well described. The upper $5 \%$ of the tails in groups 1 and 2 overlap. Group 3 has a slightly larger variance ( 1.2 vs 1.1 ) compared with groups 1 and 2 . The tails fit the lognormal distribution almost perfectly. The normality test using the Anderson Darling test statistic resulted in a p-value of close to zero for group 1 , while for group 2 , the p -value was 0.78 and for group 3 , the p -value was 0.53 . Group 1 fails the normality assumption because of distortion in the upper tail.

## Land Use and Geographical Area Effects for All Constituents

This chapter section summarizes the analyses that were conducted to identify significant land use and geographical interactions affecting stormwater concentrations contained in the NSQD. The first step was to select the data for analysis. Only samples collected using flow-weighted automatic samplers were used, in areas not having detention ponds. Also, no sites having only a single monitored event were used.

The second step was to select the following single land uses from the NSQD. The following cross-tabulation summarizes the data counts for samples meeting the above selection criteria in the main three land uses being investigated ( CO is for commercial areas, ID is for industrial areas, and RE is for residential areas). The other land uses had many instances of few observations in the EPA Rain Zones.

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | All |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CO | 3 | 123 | 6 | 16 | 42 | 34 | 41 | 0 | 9 | 274 |
| ID | 3 | 109 | 16 | 17 | 47 | 70 | 33 | 0 | 3 | 298 |
| RE | 6 | 331 | 18 | 31 | 71 | 38 | 40 | 7 | 7 | 549 |
| All | 12 | 563 | 40 | 64 | 160 | 142 | 114 | 7 | 19 | 1121 |

EPA Rain Zones 1, 8 and 9 do have not enough data observations in each land use group. Therefore, only EPA Rain Zones 2 thru 7 were included in these analyses. A single land use site corresponds to a watershed with a predominant land uses, where other land uses present in the watershed represent $10 \%$ or less of the total area. Therefore, these analyses represent stormwater observations from about the southern half of the country, plus the Pacific Northwest. Data in the NSQD are much sparser in the northeastern states, the upper mid west, the northern Great Plains and western mountain states. The initial NSQD data collection efforts focused on the mid-Atlantic and southern states, while additional data also became available in the southwest and west coast states, allowing at least this partial geographical analysis.

The third step was to estimate the non-detected observations using the Cohen's Maximum likelihood method. The estimation was performed site by site; only samples collected at the same location were used to estimate the censored observations.

For these calculations, the General Linear Model (GLM) was used to identify significant two-way interactions between these land uses and the EPA Rain Zones. The associated Minitab file used is: RECOID NOSINGLE NOPOND AU FLOW.MPJ. In all cases, an $\alpha=5 \%$ criterion was high-lighted, although all $p$ values are tabulated for comparison.

## Significant Land Use and Geographical Interactions affecting MS4 Stormwater Quality

The following tables summarize the most common stormwater constituents and how they are affected by the interaction of land use and geographical area (residential, commercial, and industrial areas only, and for EPA Rain Zones 2 through 7). The small tables summarize the overall statistics for the constituent. The larger tables summarize a similar summary for each land use/geographical area subset of data. Overall land use summaries are also shown. Only data collected with flow-weighted automatic samplers, with no ponds, are used for these summaries, as described above. In addition, left-censored (non-detected) values were substituted using Cohen's Maximum likelihood method. Calculated $p$ values are located at the top of each cell on this matrix describing the probability that the data in the subset is different from the overall set of data. The grayed-out cells represent conditions where the p-value is greater than 0.05 , the usually selected critical value for identifying significant differences. The other cells are therefore usually interpreted as being significantly different from the overall conditions. Some of the cells have no observations and are therefore left blank, except for the zero sample size. Also, some cells are highlighted because they have few sites represented ( 0,1 , or 2 sites). The data in these grayedout and highlighted cells should therefore be used with caution. Overall land use summary statistics are also shown. These could be used for those cells indicated in gray, and for those cells that have very few observations, depending on the test statistics comparing the different land uses for each pollutant (see Table XX below). These matrices display the interaction terms for geographical area (represented by EPA Rain Zone) and land use, plus the test statistics for the land uses separately. The detailed tests for statistical significance for the individual factors for each constituent are presented in Appendix F and were calculated using the General Linear Model (GLM) available in Minitab.

Table XX shows the calculated p values using the Tukey simultaneous tests and the General Linear Model for the land use effects alone. This can be used to help select the most appropriate data summary statistics to use for a specific situation, if the land use/geographical interaction data is not appropriate (with not significantly different, or too few data). If the individual cell values are not available, this table indicates that:

- Constituents that should clearly be separated by land use: copper, lead, and zinc
- Constituents that clearly did not have any significant differences for different land use categories, therefore use overall values: pH , temperature (obvious seasonal effects), TDS, and TKN
- Constituents where residential data should be separated from commercial plus industrial area data: TSS (possible) and nitrates plus nitrites
- Constituents where it is not clear; conflicts in p values when comparing different combinations of land uses: hardness, oil and grease, $\mathrm{BOD}_{5}, \mathrm{COD}$, ammonia, total P , and dissolved P

Table XX. Probability of Concentration Differences Between Land Use Categories (General Linear Model and Tukey Simultaneous Tests)*

| Constituent | Overall Land Use p | p for Resid. vs. Commercial | p for Resid. vs. Industrial | p for Commercial vs. Industrial | Comment |
| :---: | :---: | :---: | :---: | :---: | :---: |
| pH | 0.20 | n/a | 0.20 | n/a | use overall values |
| temperature | 0.99 | 1.00 | 1.00 | 1.00 | use overall values (obvious seasonal effect) |
| hardness | 0.008 | 0.18 | 0.24 | 0.005 | not clear |
| oil and grease | 0.010 | 0.01 | 0.89 | 0.06 | not clear |
| TDS | 0.065 | 0.15 | 0.81 | 0.06 | use overall values |
| TSS | <0.0001 | 0.08 | <0.0001 | 0.36 | not clear, or resid. vs. commercial plus industrial if willing to accept slightly higher p |
| $\mathrm{BOD}_{5}$ | 0.002 | 0.005 | 1.00 | 0.004 | not clear |
| COD | 0.036 | 0.03 | 0.62 | 0.45 | not clear |
| ammonia | 0.001 | 0.0005 | 0.28 | 0.09 | not clear |
| nitrates plus nitrites | <0.0001 | 0.0007 | 0.0006 | 1.00 | resid. vs. commercial plus industrial |
| TKN | 0.30 | 0.99 | 0.35 | 0.42 | use overall values |
| total phosphorus | 0.003 | 0.008 | 1.00 | 0.005 | not clear |
| dissolved P | 0.021 | 0.020 | 0.37 | 1.00 | not clear |
| copper | <0.0001 | <0.0001 | <0.0001 | <0.0001 | use individual land use values |
| lead | <0.0001 | 0.0015 | <0.0001 | 0.021 | use individual land use values |
| zinc | <0.0001 | <0.0001 | <0.0001 | 0.0007 | use individual land use values |

* the high-lighted $p$ values are $<0.05$, the usual critical value to identify differences between data categories

When examining the detailed land use and seasonal interactions in the following tables, it is clear that some of the constituents do not have many significant interactions in these factors, or that there are too few observations (or sites) represented in the NSQD. In these cases, the above Table XX can be used to help select either the significant land use value, or the overall value. The constituents that have few, if any clear geographical area/land use interactions include: pH (I6), temperature, hardness, oil and grease (I5 and I7), TDS (C2), ammonia (C7), and dissolved P (R2 and R5). The values in the parentheses are the significant interaction terms (the land use and the EPA Rain Zone).

|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.7469 | - | - | 0.1546 | 0.4828 | - | Residential |
|  |  | 14 | 8 | 0 | 71 | 22 | 25 | 140 |
|  |  | 3 | 1 |  | 6 | 3 | 3 | 16 |
|  |  | 7.36 | 6.79 |  | 7.52 | 7.34 | 7.32 | 7.40 |
|  |  | 7.43 | 6.50 |  | 7.50 | 7.40 | 7.20 | 7.40 |
|  |  | 0.39 | 0.42 |  | 0.58 | 0.56 | 0.71 | 0.60 |
|  |  | 0.05 | 0.06 |  | 0.08 | 0.08 | 0.10 | 0.08 |
|  |  | 6.70 | 6.50 |  | 6.20 | 6.30 | 6.20 | 6.20 |
|  |  | 7.87 | 7.40 |  | 9.90 | 8.30 | 9.10 | 9.90 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | - | - | - | 0.8281 | 0.058 | 0.6767 | Commercial |
|  |  | 0 | 0 | 0 | 41 | 27 | 27 | 95 |
|  |  |  |  |  | 2 | 2 | 3 | 7 |
|  |  |  |  |  | 7.70 | 6.74 | 7.36 | 7.33 |
|  |  |  |  |  | 7.70 | 6.70 | 7.40 | 7.40 |
|  |  |  |  |  | 0.44 | 0.63 | 0.66 | 0.69 |
|  |  |  |  |  | 0.06 | 0.09 | 0.09 | 0.09 |
|  |  |  |  |  | 6.90 | 5.70 | 5.30 | 5.30 |
|  |  |  |  |  | 9.10 | 8.60 | 8.38 | 9.10 |
| $\begin{aligned} & \frac{1}{\boxed{a}} \\ & \frac{\mathbf{x}}{6} \\ & \underset{2}{2} \end{aligned}$ | p observ. sites mean median st. dev. COV min. max. | 0.8245 | 0.2146 | - | 0.2121 | 0.0056 | 0.6767 | Industrial |
|  |  | 18 | 9 | 0 | 46 | 53 | 5 | 131 |
|  |  | 3 | 1 |  | 3 | 3 | 1 | 11 |
|  |  | 7.32 | 6.61 |  | 7.98 | 7.68 | 7.34 | 7.65 |
|  |  | 7.31 | 6.50 |  | 8.00 | 7.70 | 7.40 | 7.70 |
|  |  | 0.53 | 0.22 |  | 0.50 | 0.72 | 0.36 | 0.69 |
|  |  | 0.07 | 0.03 |  | 0.06 | 0.09 | 0.05 | 0.09 |
|  |  | 6.40 | 6.50 |  | 6.60 | 5.90 | 6.80 | 5.90 |
|  |  | 8.65 | 7.00 |  | 9.00 | 9.30 | 7.70 | 9.30 |

Temperature（ ${ }^{\circ} \mathrm{C}$ ）


| $\left.\begin{array}{\|c} \overline{\mathrm{W}} \\ \stackrel{\rightharpoonup}{0} \\ \partial \end{array} \right\rvert\,$ |  |  |  |
| :---: | :---: | :---: | :---: |
|  |  |  |  |
|  | $\mid \underset{N}{\sim}$ |  | Br |
|  |  |  |  |
| $\checkmark$ | $0$ | $\bigcirc$ | $\bigcirc$ |
| $\infty$ | $0$ | $\bigcirc$ | $\bigcirc$ |
| N |  |  | 0 |
|  |  |  |  |
|  | 7VIINヨaisヨy | าชเจษョwพา | 7vicisnani |

Hardness (mg/L)
245
20
35.0
29.0
37.4
1.07
1.90
443

|  |  | ©の |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\wedge$ |  |  |  | $\stackrel{\sim}{\sim} \sim \sim \sim$ |  |  |
| - |  | $\bigcirc$ |  |  |  | $\mid \infty-\bar{m} \stackrel{0}{\sim} \stackrel{\circ}{\sim}$ |
| م | N | $\left\lvert\, \begin{array}{\|ccc} 6 \\ \hline \end{array} \stackrel{m}{\sim}\right.$ $\dot{m} \dot{m} \underset{\sim}{\circ}$ |  |  <br>  | ? |  <br>  |
|  |  | - |  | $\bigcirc$ |  | - |
| $\infty$ |  | - |  | $\bigcirc$ |  | $\bigcirc$ |
|  |  |  |  |  | - |  |
|  |  |  |  |  |  |  |
|  |  | רVIINaaisay |  | าชเวบョwพ๐ |  | 7VİıISnaNI |

Oil and Grease (mg/L)


Total Dissolved Solids (mg/L)
891
78
115
68.0
416
3.63
3.00
11200

|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | p <br> observ. <br> sites <br> mean median st. dev. COV min. max. | 0.6452 | 0.8115 | 0.0608 | 0.5094 | 0.2516 | 0.1005 | Residential |
|  |  | 268 | 17 | 31 | 64 | 34 | 37 | 451 |
|  |  | 19 | 4 | 2 | 6 | 3 | 4 | 38 |
|  |  | 87.5 | 92.2 | 228.9 | 77.4 | 108.9 | 56.4 | 95.0 |
|  |  | 61.5 | 63.0 | 173.0 | 63.0 | 85.0 | 58.0 | 67.0 |
|  |  | 124.1 | 62.3 | 195.9 | 35.0 | 64.6 | 35.5 | 117.8 |
|  |  | 1.42 | 0.68 | 0.86 | 0.45 | 0.59 | 0.63 | 1.76 |
|  |  | 11.00 | 27.00 | 59.00 | 33.00 | 42.00 | 3.00 | 3.00 |
|  |  | 1700 | 211 | 1096 | 191 | 280 | 175 | 1700 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cOV <br> min. <br> max. <br> p | 0.0529 | 0.1439 | 0.0668 | 0.1313 | 0.689 | 0.8431 | Commercial |
|  |  | 82 | 6 | 15 | 39 | 20 | 37 | 199 |
|  |  | 6 | 2 | 1 | 2 | 2 | 4 | 17 |
|  |  | 124.1 | 53.3 | 204.0 | 54.4 | 117.6 | 56.8 | 101.2 |
|  |  | 61.0 | 42.0 | 140.0 | 47.0 | 81.5 | 53.0 | 62.0 |
|  |  | 421.9 | 33.5 | 200.3 | 22.2 | 90.0 | 36.2 | 280.5 |
|  |  | 3.40 | 0.63 | 0.98 | 0.41 | 0.77 | 0.64 | 4.52 |
|  |  | 23.00 | 29.00 | 64.00 | 23.00 | 33.00 | 4.00 | 4.00 |
|  |  | 3860 | 120 | 880 | 120 | 383 | 124 | 3860 |
| $\frac{1}{4}$$\frac{\mathbf{x}}{\mathbf{c}}$0$\underline{0}$$\underline{2}$ | p <br> porv. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.0938 | 0.1045 | 0.0004 | 0.3216 | 0.4684 | 0.0941 | Industrial |
|  |  | 86 | 16 | 16 | 43 | 56 | 24 | 241 |
|  |  | 9 | 3 | 1 | 3 | 4 | 3 | 23 |
|  |  | 259.2 | 112.3 | 126.3 | 85.6 | 134.4 | 76.7 | 162.5 |
|  |  | 56.5 | 101.6 | 88.0 | 71.0 | 127.0 | 70.5 | 77.0 |
|  |  | 1232.4 | 61.5 | 115.9 | 46.5 | 67.0 | 44.1 | 739.0 |
|  |  | 4.76 | 0.55 | 0.92 | 0.54 | 0.50 | 0.58 | 9.60 |
|  |  | 4.50 | 14.00 | 30.00 | 27.00 | 16.00 | 7.00 | 4.50 |
|  |  | 11200 | 224 | 524 | 238 | 373 | 154 | 11200 |

Total Suspended Solids (mg/L)
979
103
133
54.0
258
1.94
0.43
2490

|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \hline \mathbf{p} \\ \text { observ. } \\ \text { sites } \\ \text { mean } \\ \text { median } \\ \text { st. dev. } \\ \text { COV } \\ \text { min. } \\ \text { max. } \\ \hline \end{gathered}$ | 0.0046 | 0.5206 | 0.503 | 0.0198 | 0.3101 | 0.3709 | Residential |
|  |  | 309 | 18 | 30 | 64 | 33 | 37 | 491 |
|  |  | 32 | 4 | 2 | 6 | 3 | 4 | 51 |
|  |  | 64.6 | 37.3 | 376.5 | 102.1 | 112.4 | 75.1 | 91.5 |
|  |  | 43.0 | 20.5 | 188.5 | 72.0 | 95.0 | 49.0 | 51.0 |
|  |  | 79.8 | 43.3 | 559.0 | 105.4 | 80.2 | 120.5 | 176.3 |
|  |  | 1.24 | 1.16 | 1.48 | 1.03 | 0.71 | 1.60 | 3.46 |
|  |  | 2.55 | 4.20 | 11.00 | 4.00 | 3.00 | 7.00 | 2.55 |
|  |  | 823 | 180 | 2462 | 608 | 350 | 757 | 2462 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cov <br> min. <br> max. <br> p | 0.4554 | 0.0058 | <0.0001 | <0.0001 | 0.0011 | 0.1118 | Commercial |
|  |  | 113 | 6 | 15 | 40 | 20 | 37 | 231 |
|  |  | 12 | 2 | 1 | 2 | 2 | 4 | 23 |
|  |  | 59.0 | 76.7 | 829.2 | 67.3 | 111.8 | 86.4 | 119.8 |
|  |  | 35.0 | 84.5 | 609.0 | 29.5 | 100.5 | 50.0 | 47.0 |
|  |  | 79.2 | 38.1 | 579.1 | 109.9 | 109.1 | 92.3 | 251.4 |
|  |  | 1.34 | 0.50 | 0.70 | 1.63 | 0.98 | 1.07 | 5.35 |
|  |  | 6.70 | 25.00 | 176.00 | 2.00 | 8.00 | 8.00 | 2.00 |
|  |  | 629 | 110 | 2385 | 640 | 510 | 380 | 2385 |
|  | p <br> porv. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.0929 | 0.0039 | <0.0001 | <0.0001 | <0.0001 | 0.0206 | Industrial |
|  |  | 101 | 16 | 16 | 43 | 57 | 24 | 257 |
|  |  | 15 | 3 | 1 | 3 | 4 | 3 | 29 |
|  |  | 68.9 | 69.1 | 335.9 | 244.9 | 503.9 | 182.7 | 222.1 |
|  |  | 42.0 | 27.5 | 176.0 | 167.0 | 297.0 | 114.5 | 92.0 |
|  |  | 66.9 | 90.2 | 395.9 | 387.2 | 504.6 | 222.4 | 354.4 |
|  |  | 0.97 | 1.31 | 1.18 | 1.58 | 1.00 | 1.22 | 3.85 |
|  |  | 3.00 | 0.43 | 21.00 | 24.00 | 16.00 | 16.00 | 0.43 |
|  |  | 330 | 320 | 1183 | 2490 | 2325 | 1080 | 2490 |

Biochemical Oxygen Demand, 5-day ( $\mathrm{BOD}_{5}$ ) (mg/L)

|  |  | 2 | 3 | 4 | 5 | , | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | pobserv.sitesmeanmedianst. dev.COVmin.max.me | 0.0265 | 0.3419 | 0.0007 | 0.0014 | 0.7423 | <0.0001 | Residential |
|  |  | 290 | 16 | 31 | 67 | 15 | 37 | 456 |
|  |  | 25 | 4 | 2 | 6 | 3 | 4 | 44 |
|  |  | 14.7 | 9.2 | 19.6 | 9.7 | 32.7 | 7.0 | 14.0 |
|  |  | 11.0 | 8.5 | 14.0 | 7.6 | 22.0 | 5.0 | 9.4 |
|  |  | 17.0 | 6.5 | 15.2 | 8.6 | 30.4 | 7.7 | 16.3 |
|  |  | 1.16 | 0.71 | 0.77 | 0.89 | 0.93 | 1.10 | 1.75 |
|  |  | 2.00 | 2.00 | 2.00 | 2.40 | 10.00 | 1.34 | 1.34 |
|  |  | 226 | 28 | 69 | 50 | 130 | 41 | 226 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cov <br> min. <br> max. <br> m | 0.7121 | 0.0021 | 0.0403 | 0.0058 | 0.9511 | $<0.0001$ | Commercial |
|  |  | 114 | 6 | 15 | 40 | 13 | 37 | 225 |
|  |  | 9 | 2 |  | 2 | 2 | 4 | 20 |
|  |  | 15.8 | 29.2 | 22.3 | 6.9 | 46.0 | 10.1 | 15.8 |
|  |  | 13.0 | 21.5 | 17.0 | 7.3 | 43.0 | 6.4 | 10.1 |
|  |  | 13.5 | 28.8 | 15.4 | 3.1 | 30.3 | 9.8 | 16.5 |
|  |  | 0.85 | 0.99 | 0.69 | 0.44 | 0.66 | 0.97 | 1.63 |
|  |  | 2.93 | 7.30 | 4.00 | 2.50 | 9.00 | 0.75 | 0.75 |
|  |  | 96 | 83 | 57 | 17 | 100 | 42 | 100 |
| $\begin{aligned} & \frac{1}{4} \\ & \frac{\mathbf{x}}{6} \\ & \mathbf{N} \\ & \underline{0} \end{aligned}$ | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.1136 | 0.0031 | <0.0001 | 0.9638 | 0.6756 | <0.0001 | Industrial |
|  |  | 97 | 16 | 16 | 46 | 21 | 24 | 220 |
|  |  | 13 | 3 | 1 | 3 | 4 | 3 | 27 |
|  |  | 12.5 | 7.1 | 6.0 | 6.4 | 56.6 | 43.7 | 18.0 |
|  |  | 8.0 | 7.2 | 5.5 | 6.4 | 21.0 | 29.0 | 8.0 |
|  |  | 17.7 | 4.9 | 3.3 | 2.3 | 72.2 | 38.8 | 32.4 |
|  |  | 1.41 | 0.69 | 0.55 | 0.36 | 1.27 | 0.89 | 4.05 |
|  |  | 1.24 | 1.00 | 2.00 | 2.00 | 1.40 | 4.00 | 1.00 |
|  |  | 147 | 19 | 14 | 13 | 270 | 160 | 270 |

Chemical Oxygen Demand (COD) (mg/L)


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | p <br> observ. <br> sites <br> mean median st. dev. cov min. max. | 0.0451 | 0.6485 | - | <0.0001 | 0.0813 | 0.0007 | Residential |
|  |  | 265 | 16 | 0 | 69 | 37 | 37 | 424 |
|  |  | 19 | 4 |  | 6 | 3 | 4 | 36 |
|  |  | 77.4 | 47.5 |  | 80.7 | 157.5 | 47.4 | 81.2 |
|  |  | 60.0 | 37.0 |  | 60.0 | 130.0 | 31.0 | 59.5 |
|  |  | 69.9 | 34.2 |  | 69.2 | 95.4 | 54.5 | 74.5 |
|  |  | 0.90 | 0.72 |  | 0.86 | 0.61 | 1.15 | 1.25 |
|  |  | 5.00 | 6.00 |  | 10.00 | 32.00 | 9.00 | 5.00 |
|  |  | 620 | 140 |  | 480 | 370 | 300 | 620 |
|  | p <br> porv. <br> sites <br> mean <br> median <br> st. dev. <br> cov <br> min. <br> max. <br> p | 0.7521 | 0.0396 | - | 0.0238 | 0.7645 | 0.0274 | Commercial |
|  |  | 82 | 6 | 0 | 41 | 33 | 37 | 199 |
|  |  | 6 | 2 |  | 2 | 2 | 4 | 16 |
|  |  | 92.9 | 93.8 |  | 50.7 | 225.7 | 58.3 | 99.8 |
|  |  | 67.5 | 69.0 |  | 45.0 | 200.0 | 41.0 | 63.0 |
|  |  | 91.1 | 78.8 |  | 28.6 | 127.1 | 54.8 | 101.8 |
|  |  | 0.98 | 0.84 |  | 0.56 | 0.56 | 0.94 | 1.62 |
|  |  | 8.00 | 23.00 |  | 14.09 | 77.00 | 8.00 | 8.00 |
|  |  | 635 | 240 |  | 150 | 582 | 330 | 635 |
| $\begin{aligned} & \frac{1}{4} \\ & \frac{\mathbf{c}}{5} \\ & \mathfrak{b} \\ & \underline{\mathbf{0}} \end{aligned}$ | $\begin{gathered} \hline \text { p } \\ \text { observ. } \\ \text { sites } \\ \text { mean } \\ \text { median } \\ \text { st. dev. } \\ \text { cov } \\ \text { min. } \\ \text { max. } \\ \hline \end{gathered}$ | 0.0344 | 0.0311 | - | 0.0117 | 0.1328 | <0.0001 | Industrial |
|  |  | 85 | 16 | 0 | 45 | 56 | 24 | 226 |
|  |  | 8 | 3 |  | 3 | 3 | 3 | 20 |
|  |  | 66.3 | 42.9 |  | 48.9 | 242.6 | 104.2 | 108.9 |
|  |  | 54.0 | 35.0 |  | 35.0 | 225.0 | 84.5 | 60.5 |
|  |  | 51.5 | 28.8 |  | 39.7 | 165.7 | 75.8 | 121.7 |
|  |  | 0.78 | 0.67 |  | 0.81 | 0.68 | 0.73 | 2.01 |
|  |  | 2.00 | 4.00 |  | 9.55 | 18.00 | 18.00 | 2.00 |
|  |  | 340 | 116 |  | 250 | 906 | 284 | 906 |

Ammonia $\left(\mathbf{N H}_{3}\right)(\mathrm{mg} / \mathrm{L})$


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.3204 | - | - | - | 0.6461 | 0.238 | Residential |
|  |  | 262 | 0 | 0 | 0 | 20 | 26 | 308 |
|  |  | 19 |  |  |  | 1 | 3 | 23 |
|  |  | 0.33 |  |  |  | 1.10 | 0.48 | 0.40 |
|  |  | 0.25 |  |  |  | 0.96 | 0.11 | 0.26 |
|  |  | 0.28 |  |  |  | 0.72 | 1.14 | 0.49 |
|  |  | 0.84 |  |  |  | 0.66 | 2.40 | 1.89 |
|  |  | 0.02 |  |  |  | 0.19 | 0.00 | 0.00 |
|  |  | 1.49 |  |  |  | 3.40 | 5.60 | 5.60 |
|  | ppobserv.sitesmeanmedianst. dev.COVmin.max.mp | 0.793 | - | - | - | 0.0128 | 0.0079 | Commercial |
|  |  | 82 | 0 | 0 | 0 | 23 | 25 | 130 |
|  |  | 6 |  |  |  | 1 | 3 | 10 |
|  |  | 0.56 |  |  |  | 2.30 | 0.42 | 0.84 |
|  |  | 0.42 |  |  |  | 1.70 | 0.14 | 0.48 |
|  |  | 0.49 |  |  |  | 1.62 | 0.84 | 1.09 |
|  |  | 0.88 |  |  |  | 0.70 | 1.99 | 2.30 |
|  |  | 0.04 |  |  |  | 0.92 | 0.01 | 0.01 |
|  |  | 2.51 |  |  |  | 7.80 | 4.20 | 7.80 |
| $\begin{aligned} & \frac{1}{\mathbf{x}} \\ & \frac{\mathbf{x}}{\mathbf{k}} \\ & \mathbf{n} \\ & \underline{\mathbf{a}} \end{aligned}$ | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.2238 | - | - | - | 0.0011 | 0.0003 | Industrial |
|  |  | 81 | 0 | 0 | 0 | 51 | 18 | 150 |
|  |  | 8 |  |  |  | 2 | 2 | 12 |
|  |  | 0.34 |  |  |  | 1.05 | 0.43 | 0.59 |
|  |  | 0.28 |  |  |  | 0.83 | 0.30 | 0.35 |
|  |  | 0.28 |  |  |  | 0.88 | 0.47 | 0.66 |
|  |  | 0.85 |  |  |  | 0.83 | 1.10 | 1.89 |
|  |  | 0.02 |  |  |  | 0.03 | 0.03 | 0.02 |
|  |  | 1.60 |  |  |  | 5.20 | 1.70 | 5.20 |

Nitrites and Nitrates $\left(\mathrm{NO}_{2}+\mathrm{NO}_{3}\right)(\mathrm{mg} / \mathrm{L})$

|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | <0.0001 | <0.0001 | - | <0.0001 | <0.0001 | - | Residential |
|  |  | 302 | 7 | 0 | 69 | 38 | 36 | 452 |
|  |  | 23 | 1 |  | 6 | 3 | 4 | 37 |
|  |  | 0.62 | 1.61 |  | 0.82 | 1.24 | 0.81 | 0.73 |
|  |  | 0.49 | 1.29 |  | 0.66 | 1.15 | 0.75 | 0.60 |
|  |  | 0.44 | 1.31 |  | 0.85 | 0.54 | 0.66 | 0.60 |
|  |  | 0.71 | 0.82 |  | 1.03 | 0.44 | 0.82 | 1.01 |
|  |  | 0.04 | 0.08 |  | 0.11 | 0.50 | 0.08 | 0.04 |
|  |  | 2.90 | 3.16 |  | 7.17 | 2.96 | 3.50 | 7.17 |
|  | p$\qquad$ sites mean median st. dev. COV min. max. | <0.0001 | - | - | 0.001 | 0.0021 | <0.0001 | Commercial |
|  |  | 114 | 0 | 0 | 40 | 30 | 37 | 221 |
|  |  | 9 |  |  | 2 | 2 | 4 | 17 |
|  |  | 0.77 |  |  | 0.54 | 1.26 | 0.39 | 0.73 |
|  |  | 0.55 |  |  | 0.52 | 1.00 | 0.28 | 0.52 |
|  |  | 0.88 |  |  | 0.24 | 0.91 | 0.44 | 0.78 |
|  |  | 1.14 |  |  | 0.45 | 0.72 | 1.11 | 1.51 |
|  |  | 0.02 |  |  | 0.15 | 0.06 | 0.08 | 0.02 |
|  |  | 7.30 |  |  | 1.13 | 3.90 | 2.60 | 7.30 |
|  | p observ. <br> sites <br> mean median st. dev. cov min. max. | 0.0656 | 0.0093 | - | 0.1978 | 0.7055 | <0.0001 | Industrial |
|  |  | 94 | 9 | 0 | 46 | 58 | 24 | 231 |
|  |  | 11 | 1 |  | 3 | 3 | 3 | 21 |
|  |  | 0.67 | 0.44 |  | 0.68 | 1.86 | 0.21 | 0.92 |
|  |  | 0.59 | 0.24 |  | 0.67 | 1.70 | 0.17 | 0.68 |
|  |  | 0.49 | 0.53 |  | 0.30 | 1.01 | 0.18 | 0.84 |
|  |  | 0.73 | 1.20 |  | 0.44 | 0.54 | 0.84 | 1.23 |
|  |  | 0.03 | 0.05 |  | 0.07 | 0.46 | 0.02 | 0.02 |
|  |  | 2.36 | 1.66 |  | 1.69 | 4.70 | 0.70 | 4.70 |

Total Kjeldahl Nitrogen (TKN) (mg/L)


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.7294 | 0.3106 | - | <0.0001 | 0.3109 | <0.0001 | Residential |
|  |  | 302 | 17 | 0 | 69 | 37 | 37 | 462 |
|  |  | 23 | 4 |  | 6 | 3 | 4 | 40 |
|  |  | 1.76 | 1.19 |  | 2.07 | 3.90 | 1.34 | 1.92 |
|  |  | 1.32 | 1.00 |  | 1.40 | 3.50 | 0.70 | 1.33 |
|  |  | 2.39 | 0.48 |  | 1.80 | 2.37 | 2.10 | 2.32 |
|  |  | 1.36 | 0.40 |  | 0.87 | 0.61 | 1.56 | 1.75 |
|  |  | 0.10 | 0.56 |  | 0.20 | 0.50 | 0.00 | 0.00 |
|  |  | 36.00 | 2.39 |  | 10.00 | 11.00 | 11.60 | 36.00 |
|  | pobserv.sitesmeanmedianst. dev.covmin.max.ma | 0.0944 | 0.4593 | - | 0.0687 | 0.6993 | 0.2111 | Commercial |
|  |  | 114 | 6 | 0 | 40 | 34 | 37 | 231 |
|  |  | 9 | 2 |  | 2 | 2 | 4 | 19 |
|  |  | 1.94 | 1.27 |  | 1.14 | 4.41 | 1.46 | 2.07 |
|  |  | 1.44 | 0.97 |  | 0.87 | 3.65 | 0.80 | 1.39 |
|  |  | 1.51 | 0.86 |  | 0.83 | 2.84 | 1.59 | 1.96 |
|  |  | 0.78 | 0.67 |  | 0.72 | 0.64 | 1.09 | 1.41 |
|  |  | 0.05 | 0.72 |  | 0.26 | 1.04 | 0.02 | 0.02 |
|  |  | 8.68 | 3.00 |  | 4.00 | 12.00 | 8.30 | 12.00 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cOV <br> min. <br> max. | 0.0382 | 0.0556 | - | 0.0825 | 0.5193 | <0.0001 | Industrial |
|  |  | 94 | 15 | 0 | 46 | 64 | 24 | 243 |
|  |  | 11 | 3 |  | 3 | 4 | 3 | 24 |
|  |  | 1.31 | 0.84 |  | 1.05 | 4.66 | 2.09 | 2.19 |
|  |  | 0.98 | 0.80 |  | 0.83 | 3.85 | 1.70 | 1.20 |
|  |  | 1.28 | 0.48 |  | 0.80 | 3.54 | 1.30 | 2.54 |
|  |  | 0.97 | 0.57 |  | 0.77 | 0.76 | 0.62 | 2.12 |
|  |  | 0.16 | 0.05 |  | 0.10 | 0.25 | 0.01 | 0.01 |
|  |  | 10.00 | 1.77 |  | 5.00 | 16.00 | 5.90 | 16.00 |

Total Phosphorus (P) (mg/L)


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.003 | 0.8219 | - | <0.0001 | 0.0038 | $<0.0001$ | Residential |
|  |  | 309 | 18 | 0 | 69 | 38 | 37 | 471 |
|  |  | 25 | 4 |  | 6 | 3 | 4 | 42 |
|  |  | 0.39 | 0.35 |  | 0.49 | 0.66 | 0.29 | 0.41 |
|  |  | 0.31 | 0.16 |  | 0.43 | 0.49 | 0.20 | 0.33 |
|  |  | 0.28 | 0.62 |  | 0.22 | 0.77 | 0.38 | 0.37 |
|  |  | 0.72 | 1.81 |  | 0.45 | 1.16 | 1.32 | 1.12 |
|  |  | 0.05 | 0.07 |  | 0.19 | 0.14 | 0.04 | 0.04 |
|  |  | 1.89 | 2.78 |  | 1.12 | 4.96 | 2.20 | 4.96 |
|  | ppobserv.sitesmeanmedianst. dev.COVmin.max. | 0.022 | 0.1594 | - | 0.0002 | 0.2391 | 0.6938 | Commercial |
|  |  | 114 | 6 | 0 | 40 | 34 | 37 | 231 |
|  |  | 9 | 2 |  | 2 | 2 | 4 | 19 |
|  |  | 0.30 | 0.23 |  | 0.27 | 0.53 | 0.38 | 0.34 |
|  |  | 0.22 | 0.25 |  | 0.11 | 0.48 | 0.27 | 0.23 |
|  |  | 0.24 | 0.15 |  | 0.66 | 0.34 | 0.53 | 0.41 |
|  |  | 0.82 | 0.62 |  | 2.47 | 0.65 | 1.40 | 1.80 |
|  |  | 0.06 | 0.07 |  | 0.02 | 0.16 | 0.02 | 0.02 |
|  |  | 1.75 | 0.46 |  | 4.27 | 2.00 | 3.30 | 4.27 |
|  |  | <0.0001 | 0.1215 | - | 0.0013 | <0.0001 | <0.0001 | Industrial |
|  |  | 96 | 15 | 0 | 45 | 59 | 24 | 239 |
|  |  | 13 | 3 |  | 3 | 3 | 3 | 25 |
|  |  | 0.27 | 0.24 |  | 0.25 | 1.32 | 0.61 | 0.56 |
|  |  | 0.22 | 0.17 |  | 0.18 | 1.10 | 0.59 | 0.29 |
|  |  | 0.23 | 0.24 |  | 0.22 | 1.23 | 0.32 | 0.78 |
|  |  | 0.85 | 1.00 |  | 0.89 | 0.93 | 0.53 | 2.70 |
|  |  | 0.01 | 0.03 |  | 0.05 | 0.14 | 0.06 | 0.01 |
|  |  | 1.29 | 0.90 |  | 1.28 | 7.90 | 1.40 | 7.90 |

Dissolved Phosphorus (dissolved P) (mg/L)
696
56
0.18
0.12
0.18
1.02
0.01
1.60

|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | pobserv.sitesmeanmedianst. dev.COVmin.max. | 0.0103 | 0.5914 | - | <0.0001 | 0.0607 | 0.002 | Residential |
|  |  | 255 | 14 | 0 | 69 | 20 | 8 | 366 |
|  |  | 17 | 4 |  | 6 | 1 | 2 | 30 |
|  |  | 0.19 | 0.08 |  | 0.29 | 0.28 | 0.03 | 0.20 |
|  |  | 0.14 | 0.06 |  | 0.27 | 0.21 | 0.02 | 0.17 |
|  |  | 0.16 | 0.07 |  | 0.15 | 0.15 | 0.03 | 0.16 |
|  |  | 0.86 | 0.92 |  | 0.53 | 0.56 | 0.97 | 0.97 |
|  |  | 0.01 | 0.03 |  | 0.02 | 0.11 | 0.01 | 0.01 |
|  |  | 1.07 | 0.26 |  | 0.84 | 0.70 | 0.09 | 1.07 |
|  | p <br> observ. <br> sites mean median st. dev. COV min. max. | 0.1317 | 0.0324 | - | <0.0001 | 0.0056 | 0.8748 | Commercial |
|  |  | 73 | 6 | 0 | 40 | 26 | 7 | 152 |
|  |  | 4 | 2 |  | 2 | 1 | 2 | 11 |
|  |  | 0.10 | 0.10 |  | 0.05 | 0.42 | 0.04 | 0.14 |
|  |  | 0.07 | 0.06 |  | 0.05 | 0.39 | 0.03 | 0.07 |
|  |  | 0.09 | 0.10 |  | 0.03 | 0.33 | 0.03 | 0.20 |
|  |  | 0.91 | 0.95 |  | 0.67 | 0.79 | 0.75 | 2.84 |
|  |  | 0.01 | 0.03 |  | 0.01 | 0.05 | 0.01 | 0.01 |
|  |  | 0.36 | 0.25 |  | 0.17 | 1.60 | 0.09 | 1.60 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.4974 | 0.0618 | - | 0.0636 | 0.3767 | 0.0194 | Industrial |
|  |  | 67 | 11 | 0 | 46 | 52 | 2 | 178 |
|  |  | 6 | 3 |  | 3 | 2 | 1 | 15 |
|  |  | 0.10 | 0.04 |  | 0.09 | 0.29 | 0.09 | 0.15 |
|  |  | 0.08 | 0.03 |  | 0.06 | 0.21 | 0.09 | 0.09 |
|  |  | 0.08 | 0.03 |  | 0.08 | 0.26 | 0.07 | 0.18 |
|  |  | 0.75 | 0.64 |  | 0.89 | 0.90 | 0.83 | 1.99 |
|  |  | 0.01 | 0.03 |  | 0.02 | 0.05 | 0.04 | 0.01 |
|  |  | 0.45 | 0.10 |  | 0.39 | 1.50 | 0.14 | 1.50 |

Total Copper (Cu) ( $\mu \mathrm{g} / \mathrm{L}$ )


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} \hline \mathbf{p} \\ \text { observ. } \\ \text { sites } \\ \text { mean } \\ \text { median } \\ \text { st. dev. } \\ \text { COV } \\ \text { min. } \\ \text { max. } \\ \hline \end{gathered}$ | 0.0017 | 0.0006 | <0.0001 | 0.1063 | 0.0238 | 0.0401 | Residential |
|  |  | 140 | 17 | 31 | 68 | 38 | 37 | 331 |
|  |  | 31 | 4 | 2 | 6 | 3 | 4 | 50 |
|  |  | 16.1 | 21.4 | 33.0 | 9.4 | 18.2 | 12.1 | 16.4 |
|  |  | 10.0 | 5.0 | 20.0 | 8.0 | 9.5 | 8.0 | 10.0 |
|  |  | 25.6 | 26.2 | 27.0 | 8.1 | 29.8 | 14.2 | 23.5 |
|  |  | 1.59 | 1.22 | 0.82 | 0.86 | 1.64 | 1.17 | 2.35 |
|  |  | 0.71 | 5.00 | 7.00 | 1.24 | 1.40 | 1.63 | 0.71 |
|  |  | 240 | 102 | 103 | 63 | 180 | 81 | 240 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cOV <br> min. <br> max. <br> p | 0.0199 | 0.3758 | 0.0022 | 0.1283 | 0.0007 | 0.6792 | Commercial |
|  |  | 66 | 6 | 15 | 38 | 34 | 37 | 196 |
|  |  | 12 | 2 | 1 | 2 | 2 | 4 | 23 |
|  |  | 18.5 | 9.7 | 142.8 | 13.3 | 17.7 | 25.7 | 28.0 |
|  |  | 17.0 | 9.5 | 130.0 | 8.5 | 14.0 | 18.0 | 14.4 |
|  |  | 9.7 | 4.5 | 99.1 | 16.9 | 14.6 | 25.2 | 45.4 |
|  |  | 0.52 | 0.47 | 0.69 | 1.27 | 0.83 | 0.98 | 3.15 |
|  |  | 1.70 | 5.00 | 5.00 | 2.00 | 1.50 | 3.00 | 1.50 |
|  |  | 49 | 17 | 384 | 82 | 63 | 130 | 384 |
| $\begin{aligned} & \frac{1}{\mathbb{4}} \\ & \frac{\mathbf{k}}{5} \\ & \mathbf{N} \\ & \underline{0} \end{aligned}$ | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | <0.0001 | 0.0242 | 0.7863 | 0.8903 | <0.0001 | 0.1397 | Industrial |
|  |  | 54 | 15 | 15 | 44 | 66 | 25 | 219 |
|  |  | 15 | 3 | 1 | 3 | 4 | 3 | 29 |
|  |  | 19.5 | 20.0 | 257.4 | 21.2 | 90.6 | 46.6 | 60.7 |
|  |  | 15.0 | 14.0 | 118.0 | 17.5 | 89.5 | 31.0 | 25.0 |
|  |  | 17.0 | 18.3 | 358.0 | 13.3 | 73.4 | 34.1 | 117.9 |
|  |  | 0.87 | 0.91 | 1.39 | 0.63 | 0.81 | 0.73 | 4.72 |
|  |  | 2.20 | 2.00 | 10.00 | 4.00 | 2.00 | 11.00 | 2.00 |
|  |  | 87 | 59 | 1360 | 87 | 340 | 120 | 1360 |

Total Lead (Pb) ( $\mu \mathrm{g} / \mathrm{L}$ )


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | pobserv.sitesmeanmedianst. dev.COVmin.max.man | 0.0233 | 0.1786 | 0.0013 | 0.4 | 0.4616 | 0.1456 | Residential |
|  |  | 135 | 15 | 31 | 68 | 30 | 37 | 316 |
|  |  | 30 | 4 | 2 | 6 | 3 | 4 | 49 |
|  |  | 17.1 | 39.2 | 26.0 | 16.7 | 32.9 | 19.9 | 20.8 |
|  |  | 7.5 | 5.7 | 14.0 | 11.5 | 24.0 | 11.0 | 10.5 |
|  |  | 37.4 | 114.1 | 39.5 | 15.8 | 39.5 | 35.5 | 41.3 |
|  |  | 2.18 | 2.91 | 1.52 | 0.95 | 1.20 | 1.78 | 3.93 |
|  |  | 0.46 | 1.00 | 1.00 | 1.68 | 0.13 | 0.60 | 0.13 |
|  |  | 368 | 450 | 219 | 89 | 190 | 210 | 450 |
|  | p <br> p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. <br> p | 0.2046 | 0.811 | 0.1062 | 0.9323 | <0.0001 | 0.2779 | Commercial |
|  |  | 66 | 6 | 15 | 39 | 34 | 37 | 197 |
|  |  | 12 | 2 | 1 | 2 | 2 | 4 | 23 |
|  |  | 20.8 | 11.1 | 96.5 | 30.4 | 15.8 | 42.2 | 31.3 |
|  |  | 13.0 | 10.9 | 80.0 | 16.0 | 10.0 | 22.0 | 16.0 |
|  |  | 24.4 | 4.1 | 69.1 | 49.6 | 14.8 | 58.8 | 46.1 |
|  |  | 1.17 | 0.37 | 0.72 | 1.63 | 0.93 | 1.39 | 2.88 |
|  |  | 0.21 | 5.00 | 1.00 | 1.57 | 3.00 | 3.00 | 0.21 |
|  |  | 138 | 16 | 219 | 300 | 77 | 290 | 300 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cOV <br> min. <br> max. | 0.0017 | 0.132 | 0.2569 | 0.3919 | <0.0001 | 0.7662 | Industrial |
|  |  | 48 | 10 | 15 | 45 | 69 | 25 | 212 |
|  |  | 14 | 3 | 1 | 3 | 4 | 3 | 28 |
|  |  | 22.5 | 16.2 | 214.1 | 37.0 | 142.7 | 47.0 | 80.9 |
|  |  | 7.5 | 15.0 | 72.0 | 25.0 | 87.0 | 30.0 | 33.0 |
|  |  | 30.8 | 13.9 | 309.6 | 51.5 | 148.0 | 41.4 | 135.7 |
|  |  | 1.37 | 0.85 | 1.45 | 1.39 | 1.04 | 0.88 | 4.11 |
|  |  | 0.56 | 1.00 | 2.00 | 2.23 | 2.00 | 4.00 | 0.56 |
|  |  | 130 | 40 | 1200 | 270 | 620 | 170 | 1200 |

Total Zinc (Zn) ( $\mu \mathrm{g} / \mathrm{L}$ )


|  |  | 2 | 3 | 4 | 5 | 6 | 7 | Overall |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | p$\qquad$ sites mean median st. dev. cov min. max. | 0.0378 | 0.8403 | 0.2289 | 0.0036 | 0.0487 | 0.0879 | Residential |
|  |  | 140 | 17 | 31 | 68 | 38 | 37 | 331 |
|  |  | 31 | 4 | 2 | 6 | 3 | 4 | 50 |
|  |  | 73 | 69 | 199 | 72 | 220 | 83 | 102 |
|  |  | 56 | 54 | 130 | 60 | 155 | 58 | 70 |
|  |  | 67 | 66 | 286 | 40 | 249 | 109 | 146 |
|  |  | 0.93 | 0.96 | 1.44 | 0.55 | 1.13 | 1.31 | 2.08 |
|  |  | 1.11 | 8.17 | 10.00 | 16.00 | 35.00 | 6.92 | 1.11 |
|  |  | 532 | 290 | 1580 | 230 | 1500 | 650 | 1580 |
|  | p <br> observ. <br> sites <br> mean <br> median <br> st. dev. <br> cOV <br> min. <br> max. <br> p | 0.0005 | 0.3658 | 0.1004 | 0.0091 | 0.0131 | 0.0655 | Commercial |
|  |  | 66 | 6 | 15 | 39 | 34 | 37 | 197 |
|  |  | 12 | 2 | 1 | 2 | 2 | 4 | 23 |
|  |  | 199 | 121 | 463 | 98 | 235 | 160 | 196 |
|  |  | 150 | 106 | 410 | 63 | 205 | 92 | 140 |
|  |  | 156 | 55 | 297 | 97 | 139 | 173 | 182 |
|  |  | 0.78 | 0.45 | 0.64 | 0.99 | 0.59 | 1.08 | 1.30 |
|  |  | 52.00 | 71.00 | 46.00 | 28.00 | 64.00 | 39.00 | 28.00 |
|  |  | 892 | 220 | 930 | 560 | 660 | 920 | 930 |
| $\begin{aligned} & \frac{1}{4} \\ & \frac{\mathbf{c}}{6} \\ & \mathbf{6} \\ & \underline{0} \end{aligned}$ | p <br> porv. <br> sites <br> mean <br> median <br> st. dev. <br> COV <br> min. <br> max. | 0.084 | 0.1843 | 0.5258 | 0.9216 | 0.4551 | 0.001 | Industrial |
|  |  | 54 | 15 | 15 | 45 | 70 | 25 | 224 |
|  |  | 15 | 3 | 1 | 3 | 4 | 3 | 29 |
|  |  | 204 | 144 | 521 | 225 | 505 | 660 | 370 |
|  |  | 173 | 88 | 320 | 128 | 425 | 332 | 255 |
|  |  | 141 | 118 | 488 | 253 | 313 | 1566 | 599 |
|  |  | 0.69 | 0.82 | 0.94 | 1.13 | 0.62 | 2.37 | 2.35 |
|  |  | 10.00 | 25.00 | 120.00 | 28.00 | 70.81 | 47.00 | 10.00 |
|  |  | 550 | 420 | 1590 | 1400 | 1400 | 8100 | 8100 |

# Chapter 8: Example Application of the National Stormwater Quality Database (TSS and Nutrient Export Calculations for Chesapeake Bay Watersheds) 

## Overview

This chapter is a demonstration of how the data contained in the NSQD can be used, especially in conjunction with additional urban area flow data, and rural runoff data to estimate the relative contributions of pollutants in a region. This chapter first summarizes the data used, the statistical tests performed, and the results obtained, as part of our effort to identify the most appropriate nonpoint source runoff characteristics for the Chesapeake Bay watershed, the area having most of the collected data in the NSQD.

## Data Availability

Two sources of data were used to estimate nonpoint sources of pollution. The first data source corresponding to discharges from urban areas was obtained from the NSQD for the area. The second data source corresponding to discharges from agricultural land uses and forested land cover was obtained from regional data summaries provided by the EPA's Chesapeake Bay Program, "Smart Growth" project group (Office of Policy, Economics, and Innovation).

## Urban Data

Data from within the Chesapeake Bay watershed, as contained in the National Stormwater Quality Database (NSQD version 1.1), were used to determine the most appropriate concentrations for urban stormwater nutrients and suspended solids. The NSQD contains information of stormwater discharge concentrations for 19 counties in Virginia and Maryland (Table 48). More than 1,300 events were monitored in these areas representing residential, commercial and industrial land uses. There were no data reported for open space or freeway land uses. The watersheds monitored in Maryland and Virginia ranged from 3.5 and 882 acres and were between 7 and $90 \%$ impervious. Reported events used in these analyses were monitored from October 1990, through December 2000.

Table 48. Urban Monitoring Locations in the Chesapeake Bay Watershed Represented in the National Stormwater Quality Database (NSQD, version 1.1)

| Virginia | Maryland |
| :---: | :---: |
| Arlington County | Hartford County |
| Norfolk County | Baltimore County |
| Virginia Beach County | Baltimore City |
| Chesapeake County | Carroll County |
| Portsmouth County | Howard County |
| Hampton County | Anne Arundel County |
| Newport News County | Price George's County |
| Henrico County | Charles County |
| Chesterfield County | Montgomery County |
| Fairfax County |  |

Data for total nitrogen (the sum of total Kjeldahl nitrogen, TKN , and nitrite plus nitrate, $\mathrm{NO}_{2}+\mathrm{NO}_{3}$ ), total phosphorus, and total suspended solids (TSS) were evaluated for use in the Chesapeake Bay watershed.

## Rural Data

Chesapeake Bay rural water quality information was reported by the USGS in: Synthesis of Nutrient and Sediment Data for Watersheds within the Chesapeake Bay Drainage Basin (Langland, et al. 1995), prepared in corporation with the EPA. This report describes the comprehensive database of nutrient and sediment data collected from 1972 through 1992 from 1,058 non-tidal monitoring stations in the Chesapeake Bay watershed. Annual discharge loads were calculated at 48 locations for total nitrogen, at 99 locations for total phosphorus, and at 33 locations for suspended sediment. Many of the stations did not have sufficient samples, or flow data to enable load calculations. The fewer locations available for suspended sediment reflect those stations that evaluated suspended sediment, and not suspended solids. Gray, et al. (2000) concluded that suspended solids data are not reliable indicators of suspended sediment due to the laboratory processing associated with TSS analyses. The typical pipetting, or pouring, of a subsample for gravimetric analyses typically under predicts the mass associated with sand-sized particles ( $>63$ micrometers). If cone or churn splitters were used, then the TSS analyses were found to be reasonable. They also found that the results using the two methods are comparable if the mass of these larger particles comprise less than about $25 \%$ of the total sample mass. Since no particle size data was available for the TSS samples, they only used information for locations that had total sediment concentrations. Outfall urban runoff samples typically have less than $20 \%$ sand, although some early season samples in northern areas where sand is used for traction control may periodically have close to $50 \%$ sand, and some source area samples can also have large sand fractions. The TSS values used in the urban component of the analyses, described previously, are expected to be acceptable, as Chesapeake Bay region samples should not be influenced by appreciable winter sand applications, and these are all outfall samples.

Langland, et al. (1995) calculated annual nutrient and sediment loads for the selected locations using an unbiased log-linear regression model. This model enabled them to extrapolate the results to annual conditions, and to recognize both base flow conditions (groundwater recharge to the rivers is a major nitrate source, for example) and higher flows associated with surface runoff during storm periods. This analysis also enabled them to consider the potential septic tank and atmospheric deposition contributions to the annual soluble nitrogen loads. Numerous correlation analyses of annual yields of sediment and nutrients with respect to land use, physiographic province, and rock type. They found that river basins having larger percentages of agricultural land had larger nutrient and sediment yields, and that basins that were urbanized had substantially less yields. Table 49 shows the amount of each major land use category in the watershed, and in the portions of the major states within the watershed. In all cases, the land is dominated by forest and agricultural lands, with all urban lands making up about $12 \%$ for Maryland and $9 \%$ for Virginia portions of the watershed.

## Table 49. Land Uses in the Chesapeake Bay Watershed (Landland, et al. 1995)

|  | Percent of Bay <br> Basin | Pennsylvania | Maryland | Virginia |
| :---: | :---: | :---: | :---: | :---: |
| Woody (forest) | 53.9 | 62.5 | 32.6 | 52.4 |
| Herbaceous (agriculture) | 30.6 | 31.1 | 31.3 | 28.3 |
| High intensity urban | 0.6 | 0.3 | 1.2 | 0.6 |
| Low intensity urban | 4.0 | 2.9 | 6.4 | 4.5 |
| Woody urban | 1.1 | 0.6 | 1.9 | 1.4 |
| Herbaceous urban | 1.6 | 0.8 | 2.7 | 2.3 |
| Water | 7.2 | 11.1 | 20.8 | 9.6 |
| Exposed | 0.3 | 0.7 | 0.2 | 0.1 |
| Herbaceous wetland | 0.8 | 0 | 2.9 | 0.8 |
| Total | $\mathbf{1 0 0}$ | $\mathbf{1 0 0}$ | $\mathbf{1 0 0}$ | $\mathbf{1 0 0}$ |

Langland, et al. (1995) used Kendall's tau test to examine simple linear correlations between annual nutrient and sediment yields and land use, physiographic province, and rock type in the river basins above each station where the annual loads were calculate. They found that land use was the most important variable for predicting nutrient and sediment yield from a river basin. The strongest, most significant and most consistent correlations were between nutrient and sediment yields and agricultural land use. Table 50 shows selected annual yield and land use data for ten of the "load" stations evaluated by the USGS in their Chesapeake Bay report (Langland, et al. 1995).

Unfortunately, they did not determine the unit area yields corresponding to separate land uses. They presented these stations as representing the range of land uses for separate locations.

Table 50. Reported Mean Annual Yields and Land Use (Landland, et al. 1995)

|  | Percentage Land Use |  |  |  | Mean Annual Yields (Ib/acre/year) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Basin | Area ( $\mathrm{mi}^{2}$ ) | Urban | Agriculture | Forest | Total Nitrogen (TKN plus nitrates) | Total Phosphorus |
| All 127 "load" basins |  |  |  |  | 6.8 | 0.70 |
| Predominantly Urban Basins |  |  |  |  |  |  |
| 01571000 | 11.2 | 46.0 | 26.8 | 28 | 8.2 | 0.80 |
| 01589300 | 32.5 | 54.4 | 16.8 | 27 | 8.1 | na |
| 01593500 | 38.0 | 54.5 | 16.8 | 21 | na | 0.67 |
| 01646000 | 57.9 | 50.9 | 11.4 | 28 | 5.9 | 0.61 |
| 01657655 | 4.0 | 48.6 | 22.5 | 27 | na | 0.28 |
| Agricultural and Urban |  |  |  |  |  |  |
| Basins |  |  |  |  |  |  |
| 01586000 | 56.6 | 42.4 | 51.0 | 3.4 | na | 0.41 |
| 01616000 | 16.5 | 43.8 | 41.9 | 13 | 29.7 | 4.0 |
| Predominantly Agricultural |  |  |  |  |  |  |
|  |  |  |  |  |  |  |
| 01573810 | 0.38 | 1.4 | 91.0 | 6.7 | 42.1 | 6.3 |
| 0157608335 | 1.42 | 1.1 | 63.4 | 26 | 26.4 | 4.5 |
| 01639500 | 102 | 1.1 | 69.9 | 29 | 14.6 | na |

## Summary of Data and Load Calculations

The "simple" model (Schueler 1987) was used to calculate the nonpoint discharges of TSS, total phosphorus, and total nitrogen for Anne Arundel County, Maryland, for the EPA's Chesapeake Bay Program, "Smart Growth" project (Office of Policy, Economics, and Innovation). The simple model was developed by Schueler to enable rapid calculations of pollutant discharges by multiplying the event mean concentration values for a specific land use, the volumetric runoff coefficient for that land use, and the annual rainfall. With appropriate unit conversions, the result can be expressed as the unit area annual discharge for a specific pollutant. When multiplied by the area corresponding to each of the land uses in the area of concern, the total area pollutant discharges can be calculated, and the relative sources of the discharges can be identified. When working with large watersheds, these calculated values are usually much greater than the monitored in-stream values observed at the watershed outlet, because hindered pollutant transport in the stream or river is not considered. However, it is a suitable method to identify the relative pollutant contributions of different land uses in a county, as in this example.

The volumetric runoff coefficients for each land use category were based on analyses of typical land use surface configurations (mostly the impervious area characteristics) and the rain depth was determined from 50 years of rain records from the Baltimore (BWI) airport. The urban area concentration values were obtained by statistical evaluations of the Maryland and Virginia urban area data contained in the National Stormwater Quality Database, as described in the following subsections of this chapter. The urban runoff and concentration factors are assumed to have excellent reliability. However, some of the urban categories were not represented with regional Chesapeake Bay region data, so these factors were obtained from the national averaged values in the database and are labeled with a moderate reliability. The non-urban values are labeled as having poor to very good reliability, depending on the availability of local data. The agriculture values are from regional information summarized by Staver (1995) and Hartigan (1983) and are assumed to be of very good reliability. The forestlands data are from regional Chesapeake Bay regional data collected by the EPA, Office of Policy, Economics, and Innovation (Richards, personal communication) and are assumed to be of moderate reliability. The other land categories and the extraction lands data are unknown and are of poor reliability. Fortunately, as shown on the following summary tables, the best information is associated with the agricultural, forest, and urban categories which are responsible for almost the entire calculated discharges for the county.

The runoff factors are also indicated with varying reliabilities. The urban lands data all have excellent reliability due to the use of calibrated urban data for varying conditions. The agricultural runoff data is of the poorest reliability due to the uncertainties associated with the many agricultural operations that can have dramatic effects on these values. The natural land runoff values are expected to have moderate reliabilities. The USGS (Langland, et al. 1995) reported values are not comparable to these discharge values due to a number of reasons, most specifically because they are in-stream values and are affected by sediment and pollutant transport. The USGS report also did not report unit area loadings for specific land uses and the preliminary calculations resulted in unrealistic results that were highly variable. Tables 51, 52 and 53 list the nutrient and suspended solids data applicable for Chesapeake Bay watershed analyses, based on the analyses performed and outlined later in this chapter.

Table 51. Commercial TSS (mg/L), Mean and COV, (a function of season and rain depth)

|  | $<\mathbf{0 . 1}$ inches | $\mathbf{0 . 1}$ to $\mathbf{0 . 3 5}$ inches | $\mathbf{0 . 3 5}$ to $\mathbf{1}$ inch | $\boldsymbol{> 1}$ inch |
| :---: | :---: | :---: | :---: | :---: |
| Spring | $18(0.72)$ | $31(0.67)$ | $75(1.5)$ | no data |
| Summer | $75(1.5)$ | $18(0.72)$ | $75(1.5)$ | $75(1.5)$ |
| Fall | $18(0.72)$ | $75(1.5)$ | $18(0.72)$ | $18(0.72)$ |
| Winter | $18(0.72)$ | $75(1.5)$ | $75(1.5)$ | $75(1.5)$ |

Table 51 shows that storm events with precipitation depths larger than 0.35 inches are more likely to discharge higher TSS concentrations in spring, summer and winter than fall. Table 52 shows the expected total nitrogen concentrations in commercial land uses. There is a clear variation among the seasons and precipitation depth. Storm events smaller than 0.1 inch are expected to have higher total nitrogen discharges during the fall and winter than during the summer and spring seasons. For rain events between 0.35 and 1 inch , the highest concentrations were observed during the summer and fall. Table 53 shows the average concentrations and coefficients of variation for TSS, total phosphorus and total nitrogen for residential, commercial and industrial urban land use areas. This table also includes the expected concentrations in agricultural and forested areas.

Table 52. Commercial Total Nitrogen (mg/L), Mean and COV, (a function of season and rain depth)

|  | $\boldsymbol{< 0 . 1}$ inch | $\mathbf{0 . 1}$ to $\mathbf{0 . 3 5}$ inches | $\mathbf{0 . 3 5}$ to $\mathbf{1}$ inch | $\boldsymbol{> 1}$ inch |
| :---: | :---: | :---: | :---: | :---: |
| Spring | $2.0(0.49)$ | $3.2(0.50)$ | $2.0(0.49)$ | no data |
| Summer | $2.0(0.49)$ | $2.0(0.49)$ | $3.2(0.50)$ | $3.2(0.50)$ |
| Fall | $3.2(0.50)$ | $3.2(0.50)$ | $3.2(0.50)$ | $2.0(0.49)$ |
| Winter | $3.2(0.50)$ | $2.0(0.49)$ | $2.0(0.49)$ | $3.2(0.50)$ |

The total runoff discharges for the county can be determined based on the calculated total mass discharges for each land use, and the areas for each land use area. Table 54 shows the percentage of total annual runoff volume produced for each land use by season and rain depth range. About $61 \%$ of the total annual runoff volume was produced by events having more than 1 inch of rain, followed by rain events in the range 0.36 to 1 inches ( $31 \%$ of the annual runoff volume), rain events in the range of 0.1 to 0.35 inches ( $7 \%$ of the annual runoff volume), and rain events less than 0.1 inch in depth (with $1 \%$ of the annual runoff volume).

Table 53. Average Concentrations by Land Use

| Land Use | Constituent | Conditions | Average (COV) |
| :---: | :---: | :---: | :---: |
| Urban - Residential | TSS | Summer rains (between 0.1 and 0.35 inches in depth) | 143 (0.71) |
|  |  | All other rains | 58 (0.70) |
|  | TP | Sites having <27\% impervious cover: |  |
|  |  | Winter rains | 0.28 (0.59) |
|  |  | All other rains | 0.41 (0.65) |
|  |  | Sites having >27\% impervious cover: |  |
|  |  | Winter rains (less than 0.1 inches in depth) | 0.16 (0.86) |
|  |  | All other rains | 0.30 (0.63) |
|  | TN | Fall rains (less than 0.1 and greater than 1 inch in depth) | 1.4 (0.57) |
|  |  | Winter rains (0.35 and 1 inch in depth) | 1.5 (0.30) |
|  |  | Fall rains ( 0.35 and 1 inch) and Winter rains (between 0.1 and 0.35 inches in depth) | 1.9 (0.51) |
|  |  | All other rains | 2.4 (0.62) |
|  |  | Spring and summer rains (between 0.1 and 0.35 inches in depth) | 2.6 (0.38) |
| Urban - Commercial | TSS and TN | See tables 7.4 and 7.5 |  |
|  | TP | Summer rains >1 inch and fall rains between 0.1 and 0.35 inch | 0.46 (0.36) |
|  |  | All other rains | 0.23 (0.71) |
| Urban - Industrial | TSS | Fall, spring, and summer | 77 (1.48) |
|  |  | Winter | 81 (0.93) |
|  | TP | Rains less than 0.35 inches | 0.29 (0.81) |
|  |  | Rains greater than 0.35 inches | 0.22 (1.05) |
|  | TN | All conditions | 2.1 (0.79) |
| Rural - Agricultural | Sediment |  | $\begin{aligned} & \hline 1115 \mathrm{lb} / \mathrm{ac} / \mathrm{yr} \\ & \text { (unreliable estimate) } \\ & \hline \end{aligned}$ |
|  | TN |  | $40 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |
|  | TP |  | $5.4 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |
| Rural - Forest | Sediment |  | $\begin{aligned} & 4500 \mathrm{lb} / \mathrm{ac} / \mathrm{yr} \\ & \text { (unreliable estimate) } \end{aligned}$ |
|  | TN |  | $0 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |
|  | TP |  | $0 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |

Table 54. Fraction of Annual Flow Associated with Season and Rain Depth Categories (based on 50 years of rain records at Baltimore, BWI)

|  | rain range | Fraction of rain depth in category | Ultra low density residential | Low density residential | Medium density residential | High density residential | Commercial | Industrial | Institutional | Open urban land | Freeways |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Winter | <0.1 inch | 0.0078 | 0.0008 | 0.0011 | 0.0013 | 0.0016 | 0.0017 | 0.0014 | 0.0039 | 0.0004 | 0.0022 |
|  | 0.1 to 0.35 | 0.0275 | 0.0109 | 0.0137 | 0.0171 | 0.0193 | 0.0184 | 0.0173 | 0.0190 | 0.0086 | 0.0180 |
|  | 0.36 to 1 | 0.0917 | 0.0727 | 0.0786 | 0.0855 | 0.0860 | 0.0870 | 0.0874 | 0.0857 | 0.0727 | 0.0846 |
|  | >1 | 0.1025 | 0.1394 | 0.1319 | 0.1230 | 0.1202 | 0.1202 | 0.1212 | 0.1188 | 0.1420 | 0.1220 |
| Spring | <0.1 inch | 0.0091 | 0.0010 | 0.0013 | 0.0016 | 0.0018 | 0.0020 | 0.0017 | 0.0046 | 0.0005 | 0.0026 |
|  | 0.1 to 0.35 | 0.0387 | 0.0153 | 0.0193 | 0.0240 | 0.0272 | 0.0259 | 0.0243 | 0.0267 | 0.0121 | 0.0253 |
|  | 0.36 to 1 | 0.0999 | 0.0793 | 0.0857 | 0.0932 | 0.0937 | 0.0948 | 0.0953 | 0.0934 | 0.0792 | 0.0922 |
|  | >1 | 0.1075 | 0.1463 | 0.1383 | 0.1290 | 0.1261 | 0.1261 | 0.1271 | 0.1246 | 0.1489 | 0.1280 |
| Summer | <0.1 inch | 0.0089 | 0.0010 | 0.0010 | 0.0015 | 0.0018 | 0.0020 | 0.0017 | 0.0045 | 0.0005 | 0.0026 |
|  | 0.1 to 0.35 | 0.0366 | 0.0145 | 0.0183 | 0.0228 | 0.0258 | 0.0245 | 0.0230 | 0.0253 | 0.0115 | 0.0240 |
|  | 0.36 to 1 | 0.0857 | 0.0680 | 0.0735 | 0.0799 | 0.0804 | 0.0813 | 0.0817 | 0.0801 | 0.0679 | 0.0791 |
|  | >1 | 0.1438 | 0.1955 | 0.1849 | 0.1724 | 0.1686 | 0.1685 | 0.1699 | 0.1665 | 0.1991 | 0.1711 |
| Fall | <0.1 inch | 0.0064 | 0.0007 | 0.0009 | 0.0011 | 0.0013 | 0.0014 | 0.0012 | 0.0032 | 0.0004 | 0.0018 |
|  | 0.1 to 0.35 | 0.0248 | 0.0098 | 0.0124 | 0.0154 | 0.0174 | 0.0166 | 0.0156 | 0.0171 | 0.0078 | 0.0163 |
|  | 0.36 to 1 | 0.0691 | 0.0548 | 0.0593 | 0.0645 | 0.0648 | 0.0656 | 0.0659 | 0.0646 | 0.0548 | 0.0638 |
|  | >1 | 0.1391 | 0.1892 | 0.1789 | 0.1668 | 0.1631 | 0.1631 | 0.1644 | 0.1611 | 0.1926 | 0.1655 |

The flow weighting factors in Table 54 were used with the statistical analyses of the concentration data to obtain calculated long term averaged concentrations for mass loading calculations. Table 55 shows the urban area concentrations developed for Anne Arundel County using the Chesapeake Bay regional data contained in the National Stormwater Quality Database, along with concentrations and runoff quantities for other county land uses.

Table 55. Total Suspended Solids Concentrations and Volumetric Runoff Coefficients for Land Use Categories in Anne Arundel County, Maryland

| Land Use Description | $\begin{gathered} \text { \# of } \\ \text { acres in } \\ 2000 \end{gathered}$ | $\begin{gathered} \text { TSS } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Concentration reliability? | $\mathbf{R}_{\mathbf{v}}$ | $R_{V}$ reliability? |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Large lot subdivision (1 unit/ 5-10 ac) | 0 | 60 | excellent | 0.09 | excellent |
| Low-density residential (1 unit/ 5 acres to 2 units/acre) | 33,337 | 60 | excellent | 0.14 | excellent |
| Medium-density residential (2 to 8 units/acre) | 33,791 | 60 | excellent | 0.23 | excellent |
| High-density residential (8+ units/acre) | 6,274 | 60 | excellent | 0.34 | excellent |
| Commercial | 11,670 | 58 | excellent | 0.72 | excellent |
| Industrial | 3,249 | 80 | excellent | 0.52 | excellent |
| Institutional (schools, churches, military institutions, etc.) | 9,813 | 58 | moderate | 0.49 | excellent |
| Open urban land | 4,139 | 50 | moderate | 0.08 | excellent |
| Transportation | 1,557 | 99 | moderate | 0.41 | excellent |
| Extractive | 1,686 | 350 | poor | 0.3 | moderate |
| Deciduous forest | 43,901 | 90 | moderate | 0.08 | moderate |
| Evergreen forest | 4,891 | 90 | moderate | 0.08 | moderate |
| Mixed forest | 56,621 | 90 | moderate | 0.08 | moderate |
| Brush | 2,565 | 90 | poor | 0.08 | moderate |
| Wetlands | 1,643 | 0 | poor | 0.65 | moderate |
| Beaches | 29 | 0 | poor | 0.1 | moderate |
| Bare ground | 224 | 1000 | poor | 0.3 | moderate |
| Row and garden crops | 300 | 357 | very good | 0.2 | poor |
| Cropland | 42,368 | 357 | very good | 0.2 | poor |
| Orchards / vineyards / horticulture | 63 | 357 | very good | 0.15 | poor |
| Pasture | 4,690 | 145 | very good | 0.08 | moderate |
| Feeding operations | 49 | 145 | very good | 0.2 | poor |
| Agricultural building, breeding and training facilities | 163 | 145 | very good | 0.5 | poor |

Urban land uses produced slightly lower TSS concentrations compared with those observed in forest areas. However the volumetric runoff coefficients for forests are smaller than any other use, except for open urban land, resulting in the likely lowest annual yields. Bare ground, cropland, vineyards, horticulture, row and garden crops and extractive activities have the highest estimated concentrations amongst the land uses examined. Total phosphorus concentrations are presented in Table 56.

In this case, the highest concentrations were assumed for croplands, row and garden crops, orchards, vineyards and horticulture. The lowest concentrations were assumed for forested areas. One order of magnitude separates the minimum and maximum concentrations. This difference can be associated with the use of fertilizers and associated nutrient discharges. For urban areas, industrial and commercial land use areas had lower phosphorus concentrations than residential land use areas. Table 57 shows the average urban area concentrations for long term analyses, based on statistical analyses examining site factors for this regional data. Only phosphorus had different concentrations associated with different site categories that were tested.

Table 56. Total Phosphorus Concentrations for Land Use Categories in Anne Arundel County, Maryland

| Land Use Description | \# of <br> acres in <br> $\mathbf{2 0 0 0}$ | TP <br> (mg/L) | Concentration <br> reliability? |
| :--- | :---: | :---: | :---: |
| Large lot subdivision (1 unit/5- 10 acres) | 0 | 0.38 | excellent |
| Low-density residential (1 unit/ 5 acres to 2 <br> units/acre) | 33,337 | 0.38 | excellent |
| Medium-density residential (2 to 8 units/acre) | 33,791 | 0.3 | excellent |
| High-density residential (8+ units/acre) | 6,274 | 0.3 | excellent |
| Commercial | 11,670 | 0.25 | excellent |
| Industrial | 3,249 | 0.23 | excellent |
| Institutional (schools, churches, military <br> institutions, etc.) | 9,813 | 0.27 | moderate |
| Open urban land | 4,139 | 0.25 | moderate |
| Transportation | 1,557 | 0.25 | moderate |
| Extractive | 1,686 | 0.5 | poor |
| Deciduous forest | 43,901 | 0.1 | moderate |
| Evergreen forest | 4,891 | 0.1 | moderate |
| Mixed forest | 56,621 | 0.1 | moderate |
| Brush | 2,565 | 0.38 | poor |
| Wetlands | 1,643 | 0.38 | poor |
| Beaches | 29 | 0.1 | poor |
| Bare ground | 224 | 0.38 | poor |
| Row and garden crops | 300 | 1.00 | very good |
| Cropland | 42,368 | 1.00 | very good |
| Orchards / vineyards / horticulture | 63 | 1.00 | very good |
| Pasture | 4,690 | 0.38 | very good |
| Feeding operations | 49 | 0.38 | very good |
| Agricultural building, breeding and training <br> facilities | 163 | 0.38 | very good |
|  |  |  |  |

Table 57. Urban Area Stormwater Concentrations

| Land Use | Constituent | Conditions | Average value <br> for long-term <br> analyses (mg/L) |
| :---: | :---: | :---: | :---: |
|  | TSS |  | 60 |
|  | TP | Sites having <27\% impervious cover (ultra low <br> and low density areas) | 0.38 |
|  |  | Sites having <br> and high density areas) |  |
| Urban - Commercial |  | 0.30 |  |
|  | TN |  | 2.1 |
|  | TSS |  | 58 |
|  | TP |  | 0.25 |
|  | TN |  | 2.6 |
|  | TSS |  | 0.23 |
|  | TP |  | 2.1 |

Table 58 shows the summary for total nitrogen. Similar to the total phosphorus case, the largest nitrogen concentrations were predicted for croplands, vineyards, row and garden crops orchards and horticulture activities. The lowest concentrations were observed in open urban land and forested areas. The ratio between largest and smallest concentrations was approximately 2 to 1 .

Table 58. Total Nitrogen Calculated Concentrations for Land Use Categories in Anne Arundel County, Maryland

| Land Use Description | \# of <br> acres in <br> $\mathbf{2 0 0 0}$ | TN <br> ( $\mathbf{m g} / \mathrm{L}$ ) | Concentration reliability? |
| :--- | :---: | :---: | :---: |
| Large lot subdivision (1 unit/5- 10 acres) | 0 | 2.1 | excellent |
| Low-density residential (1 unit/ 5 acres to 2 units/acre) | 33,337 | 2.1 | excellent |
| Medium-density residential (2 to 8 units/acre) | 33,791 | 2.1 | excellent |
| High-density residential (8+ units/acre) | 6,274 | 2.1 | excellent |
| Commercial | 11,670 | 2.6 | excellent |
| Industrial | 3,249 | 2.1 | excellent |
| Institutional (schools, churches, military institutions, | 9,813 | 2 | moderate |
| etc.) | 4,139 | 1.3 | moderate |
| Open urban land | 1,557 | 2.3 | moderate |
| Transportation | 1,686 | 1.5 | poor |
| Extractive | 43,901 | 1.5 | moderate |
| Deciduous forest | 4,891 | 1.5 | moderate |
| Evergreen forest | 56,621 | 1.5 | moderate |
| Mixed forest | 2,565 | 1.5 | poor |
| Brush | 1,643 | 1.5 | poor |
| Wetlands | 29 | 1.5 | poor |
| Beaches | 224 | 1.5 | poor |
| Bare ground | 300 | 2.92 | very good |
| Row and garden crops | 42,368 | 2.92 | very good |
| Cropland | 63 | 2.92 | very good |
| Orchards / vineyards / horticulture | 4,690 | 2.2 | very good |
| Pasture | 49 | 2.2 | very good |
| Feeding operations | 163 | 2.2 | very good |
| Agricultural building, breeding and training facilities |  |  | 2 |

Using the simple model, it is possible to calculate the total annual discharges from these different non point sources. Table 59 shows the total estimated runoff discharged by year, and the total discharges of suspended solids, total nitrogen and total phosphorus for each of the major land use categories. Urban sites produced most of the runoff and total nitrogen, followed by agricultural and forested areas. Half of the total suspended solids were produced by agricultural activities, followed by urban areas (30\%), forested areas (12\%), and other lands (10\%). Urban and agricultural sites combined (in about equal fractions) produced almost $90 \%$ of the phosphorus loads. Forested areas only produced about $4 \%$ of the total phosphorus annual loads. The remaining phosphorus discharges were produced by other land uses.

## Table 59. Discharges by Major Land Use Categories in Anne Arundel County, Maryland

|  | Runoff Yield |  | Total Suspended Solids |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Total Discharge ( $\mathrm{ft}^{3} /$ year) | Percent of Total | Total Discharge (kg/year) | Percent of Total |
| Urban | $4.2 \times 10^{9}$ | 60.6 | $7.4 \times 10^{6}$ | 28.7 |
| Agricultural | $1.3 \times 10^{9}$ | 18.7 | $1.3 \times 10^{7}$ | 50.0 |
| Forest | $1.2 \times 10^{9}$ | 17.1 | $3.0 \times 10^{6}$ | 11.9 |
| Other lands | $2.6 \times 10^{8}$ | 3.7 | $2.4 \times 10^{6}$ | 9.4 |
| Total County | $7.0 \times 10^{9}$ |  | $2.6 \times 10^{7}$ |  |
|  | Total Phosphorus |  | Total Nitrogen |  |
|  | Total Discharge (kg/year) | Percent of Total | Total Discharge (kg/year) | Percent of Total |
| Urban | $3.6 \times 10^{4}$ | 46.4 | $2.6 \times 10^{5}$ | 60.5 |
| Agricultural | $3.6 \times 10^{4}$ | 45.8 | $1.1 \times 10^{5}$ | 25.0 |
| Forest | $3.4 \times 10^{3}$ | 4.3 | $5.1 \times 10^{4}$ | 11.9 |
| Other lands | $2.8 \times 10^{3}$ | 3.6 | $1.1 \times 10^{4}$ | 2.6 |
| Total County | $7.8 \times 10^{4}$ |  | $4.3 \times 10^{5}$ |  |

The final values used during for the calculations are summarized in Tables 60 and Table 61. For each of the main land uses, the percentage of impervious areas (indicating the percentage connected and disconnected), the volumetric runoff coefficient and the TSS, total phosphorus, and total nitrogen concentrations are shown. The volumetric runoff coefficients, and curve numbers, were calculated using 50 years of precipitation data from the BWI airport in Baltimore.

Table 60. Urban Land Use Categories Used in Anne Arundel County, Maryland

| Description | Note | Average percentage of Impervious areas | $\begin{aligned} & \text { TSS } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | Total P (mg/L) | Total N (mg/L) | Comments |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| High density residential | Rv = 0.34, 47\% pervious, 39.9\% dir con imp, and 13.1\% dis con impervious | 53 | 60 | 0.3 | 2.1 | Rv and CN calculated using 50 yrs of BWI rains and concentration factors from MD and VA MS4 data |
| Medium density residential | $\begin{gathered} \text { Rv }=0.23,62.3 \% \\ \text { pervious, } 24.2 \% \\ \text { directly con imp, and } \\ \text { 13.5\% disconnected } \\ \text { impervious } \\ \hline \end{gathered}$ | 37.8 | 60 | 0.3 | 2.1 | Rv and CN using 50 yr BWI rain and concentration factors from MD and VA MS4 data |
| Low density residential | $\mathrm{Rv}=0.14,79.6 \%$ pervious, $14.9 \%$ dir con impervious, and 5.5\% disconnected imp. | 20.4 | 60 | 0.38 | 2.1 | 1 unit/5 ac to 2 units/ac. Calc Rv and CN using 50 years BWI rains and concentration factors from MD and VA MS4 |
| Ultra low den residential | ```Rv=0.09, 90.4% pervious, 5.6%directly con imp and 4% discon impervious``` | 9.6 | 60 | 0.38 | 2.1 | 1 unit/5 to 10 ac, calc 50 yr , concentration factors from MD and VA MS4 data |
| Freeways and other main roads with paved drainage | Rv = 0.41, 49.5\% pervious, $50.5 \%$ dir con impervious. | 50.5 | 99 | 0.25 | 2.3 | Calc using 50 yrs of BWI rains and concentration factors from national MS4 data |
| Commercial (shopping centers) | $\mathrm{Rv}=0.72,8.3 \%$ pervious, and 91.7\% dir con imp. | 91.7 | 58 | 0.25 | 2.6 | 50 yr of BWI rains and concentration factors from MD and VA MS4 data |
| Institutional (schools, churches, military, etc.) | Rv=0.49, 36.4\% pervious, 61.3\%dir con imp, and 2.3\%discon imp. | 63.6 | 57.9 | 0.35 | 1.57 | Calculated from 50 yr BWI rains and concentration factors from national average institutional MS4 data. |
| Industrial (medium) | Rv=0.52, 16.7\% pervious, $62.8 \%$ dir con imp, and 20.5\% discon con imp. | 83.3 | 80 | 0.23 | 2.1 | CN calc using 50 yr BWI rain and concentration factors from MD and VA MS4 data. |
| Open urban area | $\begin{aligned} & \mathrm{Rv}=0.08,95.1 \% \\ & \text { pervious and } 4.9 \% \\ & \text { dir con impervious. } \end{aligned}$ | 4.9 | 70 | 0.12 | 1.5 | CN calc from 50 yr BWI rains and concentration factors from national average urban open area MS4 data |

The land uses having the largest amounts of directly connected impervious surfaces were the commercial, institutional, and industrial land use areas. Urban TSS concentrations ranged between 57 and $99 \mathrm{mg} / \mathrm{L}$, total phosphorus concentrations ranged between 0.12 and $0.40 \mathrm{mg} / \mathrm{L}$, and total nitrogen ranged between 1.5 and 2.6 $\mathrm{mg} / \mathrm{L}$. Table 61 shows the summaries for the other land uses.

Table 61. Other Land Use Categories Used in Anne Arundel County Calculations

| Description | Note | $\begin{aligned} & \text { TSS } \\ & \text { (mg/L) } \end{aligned}$ | Total Phosphorus (mg/L) | Total Nitrogen (mg/L) |
| :---: | :---: | :---: | :---: | :---: |
| Fallow | Straight Row. Concentration factors from prior regional data | 107 | 1.3 | 4.4 |
| Row Crops | Straight Row, small grain. Concentration factors from prior regional data | 357 | 1 | 2.92 |
| Row and garden Crops | Straight Row. <br> Concentrations factors from prior regional data | 357 | 1 | 2.92 |
| Orchards, vineyards, horticultural | Concentration factors from prior regional data | 357 | 1 | 2.92 |
| Pasture or Range | Concentration factors from prior regional data | 145 | 0.38 | 2.2 |
| Feeding operations | Continuous forage, poor. Concentration factors from prior regional data | 145 | 0.38 | 2.2 |
| Woods or Forest Land | Deciduous forest (woods, good). Concentration factors from prior regional data | 90 | 0.1 | 1.5 |
| Woods or Forest Land | Evergreen forest (woods, good condition). <br> Concentration factors from prior regional data. | 90 | 0.1 | 1.5 |
| Woods or Forest Land | Mixed forest (woods, good). Concentration factors from prior regional data. | 90 | 0.1 | 1.5 |
| Farmsteads | Agricultural buildings, breeding and training facilities | 163 | 0.38 | 2.2 |
| Brush | Herbaceous, fair. | 90 | 0.38 | 1.5 |
| Extractive |  | 1000 | 0.38 | 1.5 |
| Wetlands |  | 0 | 0.38 | 1.5 |
| Beaches |  | 0 | 0.1 | 1.5 |
| Bare ground |  | 1000 | 0.38 | 1.5 |

The largest TSS concentrations were observed in extractive activities and for bare ground, or exposed soil sites. Land uses where the intensive use of fertilizers is most frequent had the largest total phosphorus and total nitrogen concentrations. The lowest nutrient concentrations were observed in forested areas.

Figure 60 shows the area distributions and the relative contributions for major sources of runoff, total suspended solids, total nitrogen and total phosphorus for sites located in Anne Arundel County in Maryland. Forested and urban land use areas represent almost $80 \%$ of the total land uses in the county. About $15 \%$ of the area is agricultural and the remaining of $5 \%$ is associated with other activities.

Urban land use areas produce almost $65 \%$ of the total runoff volume for the county, followed by agricultural and forested areas (about $15 \%$ each). As expected, impervious surfaces in urban land use areas were responsible for most of the total discharged runoff volume. Agricultural land uses produce almost half of the total TSS discharges, although they make up only about $15 \%$ of the county area. Urban land uses are the second major source of TSS in the county, contributing about $28 \%$ of the total annual TSS discharges in the county. Forested areas and other land uses contribute the smallest fractions of the total load, with almost $11 \%$ each. Urban and agricultural areas combined produced almost $90 \%$ of the total phosphorus load, in about equal percentages. Forested areas and other land uses contribute about $10 \%$ of the total countywide phosphorus load. Finally, urban land uses contributed almost $60 \%$ of the total nitrogen load for the county, followed by agricultural activities ( $25 \%$ ), forested areas ( $13 \%$ ) and other land uses (2\%).


Figure 60. Sources of runoff, TSS and nutrients for different sources in Anne Arundel County

## Statistical Analyses Performed

The following discussion describes the statistical analyses performed to identify the different groups in TSS, total nitrogen and total phosphorus by season and precipitations depth for the Chesapeake Bay region data. The objective of the statistical tests was to identify significantly distinct categories of the Chesapeake Bay regional data. Specifically, land use, season, precipitation, percent imperviousness, and watershed drainage area, were considered potentially important factors that would affect the concentration values. In addition, variations of reported concentrations with time were also examined. After appropriate normalization of the data, three-way and two-way ANOVA tests were used to identify the significant factors and interactions between these potentially important factors, while one-way ANOVA tests, along with parametric and nonparametric comparison tests, were used to identify groupings within the range of any one factor. As an example, one-way ANOVA analyses were performed to identify any ranges of percentage of imperviousness that produce different distributions of stormwater constituent concentrations from other ranges, while two-way ANOVA analyses were used to identify any seasonal-total precipitation interactions in the distribution of the stormwater constituent concentrations

Before ANOVA analyses can be conducted, the first step is to examine the data to ensure that it fits a normal probability distribution. If not, the data needs to be transformed. Prior tests (reported in this report) found that most all of the stormwater constituents in the NSQD fit lognormal distributions. In this case, the base 10 logarithm of the original observations adequately followed a normal distribution. Therefore, data from the same population group will fall along the same straight line. Groups in either tail that do not fall on the line can be considered different. This procedure was used in the ANOVA analyses to identify if the concentration values were statistically different for different levels of the factor, or factors, being examined. For example, if the expected values are different for different levels of imperviousness, or different seasons, then those data groupings will not follow the main probability distribution, and the ANOVA test results will indicate a likely significantly different data population. The significant ANOVA coefficients were then used to create a model to predict the concentration values for the different groups. All of the observed conditions within each group will have the same expected concentration value. Once the groups were identified, the mean and standard deviations were calculated from the original observations in the database for each observation in each group, and the data for each group are plotted on probability and box and whisker plots. The following discussion is a detailed description of the tests conducted using the Chesapeake data for total suspended solids.

## Residential Area Total Suspended Solids Analyses

The ANOVA tests did not identify any significant groupings for either drainage area, or percentage imperviousness variations. Trends with time since the last rain, and for time since the initiation of the watershed monitoring were also examined, but these analyses did not identify any apparent, or significant, trends for any of the test sites. Initial data evaluations indicated a possible significant variation due to the level of imperviousness in the test watersheds, but when evaluated in conjunction with season and rain depth, these other factors were found to be the only significant factors to describe the variations in TSS concentrations in residential areas. Obviously, the percentage imperviousness values will have a large effect on the amount of runoff volume expected, so the imperviousness will be very important in affecting the mass of pollutants discharged. This is similar to data evaluations for other regions. The Maryland and Virginia data provided a great opportunity to test this hypothesized effect, because there were 13 residential area test watersheds having imperviousness values ranging from 7 to $65 \%$ (although most of the data were represented in six watersheds ranging in imperviousness from 20 to $50 \%$ ). The statistical tests identified two distinct groups of residential TSS data, as represented in the following plots and tables: small summer rains (in the range of 0.1 to 0.35 inches) which had an average TSS concentration of about $143 \mathrm{mg} / \mathrm{L}$, and all other residential conditions which had an average TSS concentration of about $58 \mathrm{mg} / \mathrm{L}$. The following plots and data summaries describe these two data groupings.


- Summer, $0.1<P<0.35$ in

Figure 61. Residential TSS distributions by groups

Residential Total Suspended Solids Concentrations by Groups


Figure 62. Residential TSS box and whiskers plot distribution by groups

Table 62. Results for Residential TSS (mg/L)


## Residential Area Total Phosphorus Data Analyses

The statistical tests of the residential total phosphorus data indicated a significant effect associated with the amount of imperviousness cover in the monitored watersheds. Sites having small amounts of impervious cover (7 to $25 \%$ ) had significantly higher total phosphorus concentrations than sites having larger amounts of impervious cover (29 to $65 \%$ ). Winter rains had lower total phosphorus concentrations in each group (all winter rains in the first group, and small winter rains of less than 0.1 inch in the second group). The following plots and data summaries describe these data groupings, separated by the two impervious cover categories.


- All Others
- Winter

Figure 63. Residential total phosphorus concentrations for sites having < 27\% impervious surfaces


Figure 64. Box and whiskers plot for residential total phosphorus concentrations for sites having < 27\% impervious surfaces

Table 63. Results for Residential Total Phosphorus (Impervious < 27\%)


Residential Total Phosphorus Concentrations by Group


Figure 65. Residential total phosphorus concentrations for sites having $\mathbf{>} \mathbf{2 7 \%}$ impervious surfaces

## Residential Total Phosphorus Concentrations for Impervious Cover Greater than 27\%



Figure 66. Box and whiskers plot for residential total phosphorus concentrations for sites having > 27\% impervious surfaces

Table 64. Results for Residential Total Phosphorus (Impervious > 27\%)


## Residential Area Total Nitrogen Data Analyses

The statistical analysis of the residential total nitrogen data identified several important interactions between season and rain depth. There were no significant factors associated with drainage area, percent imperviousness, or trend with time. Five significant groups were identified for residential total nitrogen concentrations:

1) Fall rains $<0.1$ and $>1$ inch
2) Winter rains between 0.35 and 1 inch
3) Fall for rains between 0.35 and 1 inch, and winter rains between
0.1 and 0.35 inches
4) All other conditions
5) Spring and summer rains between 0.1 and 0.35 inches

The following plots and data summaries describe these five data groupings for residential area total nitrogen concentrations.


Figure 67. Residential total nitrogen concentration groups

## Residential Total Nitrogen Concentrations



Figure 68. Box and whiskers plot for residential total nitrogen

Table 65. Results for Residential Total Nitrogen

| Descriptive statistics of residential TSS (mg/L) by groups: |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Groups | N | Mean | Median | StDev | C |  |
| Fall, $\mathrm{P}<0.1$ and $\mathrm{P}>1$ | 22 | 1.4 | 1.3 | 0.77 |  |  |
| Winter, $0.35<\mathrm{P}<1$ | 42 | 1.5 | 1.5 | 0.46 |  |  |
| Fall, $0.35<\mathrm{P}<1$; |  |  |  |  |  |  |
| WI, $0.1<\mathrm{P}<0.35$ | 43 | 1.9 | 1.6 | 0.95 |  |  |
| All other conditions | 112 | 2.4 | 2.1 | 1.5 |  |  |
| Sp and $\mathrm{Su}, 0.1<\mathrm{P}<0.35$ | 40 | 2.6 | 2.6 | 1.0 |  |  |
| Groups | SE Mean | Minimum | Max | num | Q1 | Q3 |
| Fall, $\mathrm{P}<0.1$ and $\mathrm{P}>1$ | 0.17 | 0.44 | 4.1 |  | 0.93 | 1.7 |
| Winter, $0.35<\mathrm{P}<1$ | 0.072 | 0.72 | 3.0 |  | 1.23 | 1.7 |
| Fall, $0.35<\mathrm{P}<1$; |  |  |  |  |  |  |
| WI, $0.1<\mathrm{P}<0.35$ | 0.14 | 0.68 | 5.7 |  | 1.3 | 2.3 |
| All other conditions | 0.14 | 0.74 | 13 |  | 1.5 | 2.9 |
| Sp and $\mathrm{Su}, 0.1<\mathrm{P}<0.35$ | 0.16 | 0.62 | 5.4 |  | 1.9 | 3.4 |


| Groups |
| :--- |
| Fall, rains <0.1 inches and rains >1 inch: |
| Overall $95 \%$ confidence interval of all observed data: 0.45 to $3.2 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| $95 \%$ confidence interval of reported median: 0.98 to $1.5 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| Winter, rains between 0.35 and 1 inch: |
| Overall $95 \%$ confidence interval of all observed data: 0.85 to $2.5 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| 95\% confidence interval of reported median: 1.3 to $1.6 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| Fall, rains between 0.35 and 1 inch; and Winter rains between 0.1 and 0.35 inches: |
| Overall $95 \%$ confidence interval of all observed data: 0.68 to $4.1 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| 95\% confidence interval of reported median: 1.5 to $1.9 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| Spring and Summer rains between 0.1 and 0.35 inches: |
| Overall $95 \%$ confidence interval of all observed data: 1.1 to $5.7 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| $95 \%$ confidence interval of reported median: 2.1 to $2.8 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| All other conditions: |
| Overall $95 \%$ confidence interval of all observed data: 0.82 to $5.9 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |
| 95\% confidence interval of reported median: 2.0 to $2.3 \mathrm{mg} / \mathrm{L}$ (from fitted probability distribution) |

## Commercial Area Total Suspended Solids Analyses

The commercial area total solids data appears to be affected by season and rain depth interactions, plus season and rain depth main factors. No affects associated with drainage area, or trends with time or interevent period, were identified.

## Commercial Area Total Phosphorus Analyses

The commercial area total phosphorus data appears to be affected by season and rain depth interactions, plus season main factors. No affects associated with drainage area, or trends with time or interevent period, were identified.

## Commercial Area Total Nitrogen Analyses

The commercial area total nitrogen data appears to be affected by season and rain depth interactions alone. No affects associated with drainage area, or trends with time or interevent period, were identified.

## Industrial Area Total Suspended Solids Analyses

The industrial area total suspended solids data appears to be affected by season main factors alone. No affects associated with rain depth, drainage area, or trends with time or interevent period, were identified.

## Industrial Area Total Phosphorus Analyses

The industrial area total phosphorus solids data appears to be affected by rain depth main factors alone. No affects associated with season, drainage area, or trends with time or interevent period, were identified.

## Industrial Area Total Nitrogen Analyses

The commercial area total nitrogen solids data does not appear to be affected by any of the factors, or interactions examined. No affects associated with rain depth, season, drainage area, or trends with time or interevent period, were identified.

## Summary

In this chapter, the NSQD was used to estimate the expected total suspended solids and nutrient mass discharges from urban, agricultural and forested sources in Anne Arundel County, Maryland, in a typical year. The parameters used in Schueler's simple method are the mean concentrations from each of these sources, the areas associated with each source, the volumetric runoff coefficient for these sources, and the rain depth associated with the period of calculation. The NSQD includes several catchments and more than 1,000 storm events in the Chesapeake Bay area which were used to determine the most appropriate urban area mean stormwater concentrations for residential, commercial and industrial land uses.

The effects associated with different seasons and rain depths on the urban area concentrations of solids and nutrients were also addressed for these regional data in this chapter. ANOVA analyses indicated that there are significant differences in the concentrations according to the seasonal period when the samples were collected, the total precipitation depth, and the interaction between these two factors for some of these pollutants and urban land uses. A stronger influence of the interactions between these factors was observed in residential areas compared with commercial or industrial land use areas.

The data summaries indicated that solids concentrations from forested and urban areas are similar, however the total runoff volume produced in forest areas is very small compared with the urban areas. For this reason, annual mass discharges from forested areas are less than half of the annual mass discharges produced from urban areas, even though the areas for these two main land use categories are similar.

Annual agricultural mass discharges of suspended solids are almost twice those calculated for urban areas. In urban areas, lower TSS concentrations occur, but a much larger fraction of the precipitation is transformed to runoff. Total urban area nitrogen mass discharges are expected to be almost twice the loads discharged from agricultural areas.

# Chapter 9: Findings and Conclusions 

## Introduction

The purpose of this report was to examine several commonly accepted assumptions concerning stormwater characteristics (and associated management decisions) by stormwater managers and researchers. These included assumptions relating to the existence of "first flushes;" the effect of the abundance of impervious areas and the length of antecedent dry period on stormwater constituent concentrations; the influences of non-detected observations on stormwater characteristics; among others. These assumptions were evaluated using information contained in the National Stormwater Quality Database (NSQD). More than 3,765 events were monitored at 360 sites throughout the U.S. and the monitored water quality and associated information was included in the first version of the database. Most of the data were collected from residential, commercial and industrial land use areas in the eastern and southern parts of the U.S. (according to the original study design), although most geographical areas are represented.

## Major Findings, as Reported in Report Chapters <br> Findings from Chapter 2: The National Stormwater Quality Database (NSQD) Description

- Drainage Areas by Land Use. Drainage areas for each outfall varied for different land uses, with freeways having the smallest drainage areas and open space having the largest drainage areas. Generally, the median drainage areas ranged from about 40 to 110 acres, excluding the freeway sites which were only about 1.5 acres in size.



## Land Use

- Impervious Areas in each Land Use. The percentage of impervious areas in each drainage area is obviously related to the land uses. Open space, and the open space mixed areas have the lowest fraction of impervious areas (at close to zero and about $20 \%$ respectively), while freeways and commercial land uses have the largest fractions of impervious areas (close to $100 \%$ and $85 \%$, respectively). Residential areas have about $40 \%$ impervious surfaces. The database is not able to distinguish the directly connected vs. the partially connected impervious areas.

- Runoff Coefficients and Impervious Cover. The reported volumetric runoff coefficients were closely related to the percentage of impervious cover. Again, the database cannot separate the directly connected impervious areas from the partially connected areas, so there is some expected variation in this relationship. This relationship significantly affects the mass discharges of pollutants. As noted later in these findings, very few significant relationships were found between the impervious covers and runoff concentrations.

- Reported Monitoring Problems. About 58\% of the communities described problems during the monitoring process:
- One of the basic sampling requirements was to collect three samples every year for each of the land use stations. These samples were to be collected at least one month apart during rains having at least 0.1 inch rains, and with at least 72 hours from the previous 0.1 -inch storm event. It was also required (when feasible), that the variance in the duration of the event and the total rainfall not exceeded the median rainfall for the area. About $47 \%$ of the communities reported problems meeting these requirements. In many areas of the country, it was difficult to have three storm events per year having these characteristics.
- The second most frequent problem, reported by $26 \%$ of the communities, concerned backwater tidal influences during sampling, or the outfall became submerged during the event. In other cases, it was observed that there was flow under the pipe (flowing outside of the pipe, in the backfill material, likely groundwater), or sometimes there was not flow at all.
- About $12 \%$ of the communities described errors related to malfunctions of the sampling equipment. When reported, the equipment failures were due to incompatibility between the software and the equipment, clogging of the rain gauges, and obstruction in the sampling or bubbler lines. Memory losses in the equipment recording data were also periodically reported. Other reported problems were associated with lighting, false starts of the automatic sampler before the runoff started, and operator error due to misinterpretation of the equipment configuration manual.


## - Suggested Changes in Monitoring Requirements:

- Base flows can commonly occur in separate storm drainage systems for a variety of reasons and they may be more important during some seasons than during others. In many cases, they cannot be avoided and should be included in the monitoring program, and their effects need to be recognized as an important flow phase.
- The rain gauges need to be placed close to the monitored watersheds. In the NSQD, about 7\% of the events had site precipitation estimated using a rain gauge located at the city airport. About $16 \%$ of the events had precipitation depths estimated using their own monitoring network. Some communities had precipitation networks that were used for flood control purposes for the surrounding area. These networks can be considered better than the single airport rain gauge, but should at least be supplemented with a rain gauge located in the monitored watershed. Large watersheds cannot be represented with a single rain gauge at the monitoring station; in those cases the monitoring networks will be a better approach..
- Many of the monitoring stations lacked flow monitoring instrumentation, or did not properly evaluate the flow data. Accurate flow monitoring can be difficult, but it greatly adds to the value of the expensive water quality data.
- The three hour monitoring period that most used may have resulted in some bias in the reported water quality data. For example, it is unlikely that manual samplers were able to initiate sampling near the beginning of the events, unless they were deployed in anticipation of an event later in the day. A more cost-effective and reliable option would be to have semi-permanent monitoring stations located at the monitoring locations and sampling equipment installed in anticipation of a monitored event. Most monitoring agencies operated three to five land use stations at one time. This number of samplers, and flow equipment, could have been deployed in anticipation of an acceptable event and would not need to be continuously installed in the field at all sampling locations.
- Some of the site descriptions lacked important information and local personnel sometimes did not have the needed information. This was especially critical for watershed delineations on maps of the area. Also, few of the watershed descriptions adequately described how the impervious areas were connected to the drainage system, one of the most important factors affecting urban hydrologic analyses. In most cases, information concerning local stormwater controls was able to be determined from a variety of sources, but it was not clearly described in the annual reports.
- Comparisons of Stormwater Databases. The NSQD can be compared to the older NURP database:

Comparison of NURP and NSQD Stormwater Databases

| Constituent | Units | Source | Event Mean Concentrations Mean Median |  | Number of events |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Total Suspended Solids | mg/L | NURP | 174 | 113 | 2000 |
|  |  | NSQD | 79 | 50 | 3404 |
| Biochemical Oxygen Demand | mg/L | NURP | 10 | 8.4 | 474 |
|  |  | NSQD | 11 | 8.6 | 2973 |
| Chemical Oxygen Demand | mg/L | NURP | 66 | 55 | 1538 |
|  |  | NSQD | 71 | 55 | 2699 |
| Total Phosphorus | mg/L | NURP | 0.34 | 0.27 | 1902 |
|  |  | NSQD | 0.37 | 0.29 | 3162 |
| Dissolved Phosphorus | mg/L | NURP | 0.10 | 0.078 | 767 |
|  |  | NSQD | 0.11 | 0.078 | 2093 |
| Total Kjeldahl Nitrogen | mg/L | NURP | 1.7 | 1.4 | 1601 |
|  |  | NSQD | 1.7 | 1.4 | 3034 |
| Nitrite and Nitrate | mg/L | NURP | 0.84 | 0.67 | 1234 |
|  |  | NSQD | 0.77 | 0.61 | 2983 |
| Copper | $\mu \mathrm{g} / \mathrm{L}$ | NURP | 67 | 55 | 849 |
|  |  | NSQD | 18 | 14 | 2356 |
| Lead | $\mu \mathrm{g} / \mathrm{L}$ | NURP | 175 | 131 | 1579 |
|  |  | NSQD | 24 | 17 | 2250 |
| Zinc | $\mu \mathrm{g} / \mathrm{L}$ | NURP | 176 | 140 | 1281 |
|  |  | NSQD | 110 | 88 | 2888 |

- The nutrient, COD , and $\mathrm{BOD}_{5}$ means and medians are very close in both databases, while the suspended solids and metals are much smaller in the NSQD than in the NURP database. As part of their MS4 Phase I application, Denver and Milwaukee (Milwaukee data not yet included in the NSQD) both returned to some of their earlier sampled monitoring stations used during their local NURP projects (EPA 1983). In the time period between the early 1980s (NURP) and the early 1990s (MS4 permit applications), they did not detect any significant differences, except for large decreases in lead concentrations, as shown in the figure below for a Milwaukee site.


Comparison of pollutant concentrations collected during NURP (1981) to MS4 application data
$(1990)$ at the same location (personal communication, Roger Bannerman, WI DNR)

- The differences found in both the NURP and the NSQD databases are therefore most likely due to differences in geographical areas emphasized by each database. Half of the events included in the NSQD database were collected in EPA Rain Zone 2 (Maryland, Virginia, North Carolina, Kentucky and Tennessee), while half of the events contained in the NURP database were collected in EPA Rain Zone 1 (Minnesota, Wisconsin, Michigan, Illinois, New York, Massachusetts and New Hampshire). The NSQD best represents the coastal states and the southern states, while NURP best represents the upper Midwest and northeast states.


## Findings from Chapter 3: QA/QC Procedures

- $\boldsymbol{Q A} / \mathbf{Q C} \boldsymbol{E f f o r t}$ QA/QC takes a great deal of time and effort to ensure that the database content is correct and accurate. During this project, about 6 months were spent in collecting the majority of the information from the communities, while more than 9 months were spent in reviewing the accuracy of the data. All data was compared against original information, if at all possible, and all transcribed data was carefully compared to the source data. In addition, the behavior of the data was also carefully reviewed to identify unusual data observations. "Outliers" were not casually eliminated from the dataset unless errors were likely that could not be corrected. Comparisons to associated data and to likely data levels were the most important methods used to identify errors. In addition, unusually high and low observations were all verified.
- Non-Detected Analyses. Left-censored data refers to observations that are reported as below the limits of detection, while right-censored data refers to over-range observations. Unfortunately, many important stormwater measurements (such as for filtered heavy metals) have large fractions of undetected values. These missing data greatly hinder many statistical tests. To estimate the problems associated with censored values, it is important to identify the probability distributions of the data in the dataset and the level of censoring. Most of the constituents in
the NSQD follow a lognormal distribution. When the frequencies of the censored observations were lower than $5 \%$, the means, standard deviations and coefficients of variation were almost identical when the censored observations were replaced by half of the detection limit, or estimated using Cohen's Method. As the percentage of non-detected values increases, replacing the censored observation by half of the detection limit instead of estimating them using the Cohen's maximum likelihood method produced lower means and larger standard deviations. Replacing the censored observations by half of the detection limit is not recommended for levels of censoring larger than $15 \%$. The censored observations in the database were replaced using estimated values using Cohen's maximum likelihood method for each site before the statistical tests in this report. Because this method uses the detected observations to estimate the non-detected values, it is not very accurate, and therefore not recommended, when the percentage of censored observations is larger than $40 \%$.
- Selection of Analytical Methods. The best method to eliminate problems associated with left-censored data is to use an appropriate analytical method. By keeping the non-detectable level below $5 \%$, there are many fewer statistical analysis problems and the value of the datasets can be fully realized. The following table summarizes the recommended minimum detection limits for various stormwater constituents to obtain manageable non-detection frequencies ( $<5 \%$ ). Some of the open space stormwater measurements (lead, and oil and grease, for example), would likely have greater than $5 \%$ non-detects, even with the detection limits shown. The detection limits for filtered heavy metals should also be substantially less than shown on this table.


## Suggested Analytical Detection Limits for Stormwater Monitoring Programs to Obtain <5\% Nondetects

|  | Residential, commercial, <br> industrial, freeway | Open Space |
| :--- | :--- | :--- |
| Conductivity | $20 \mu \mathrm{~S} / \mathrm{cm}$ | $20 \mu \mathrm{~S} / \mathrm{cm}$ |
| Hardness | $10 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| Oil and grease | $0.5 \mathrm{mg} / \mathrm{L}$ | $0.5 \mathrm{mg} / \mathrm{L}$ |
| TDS | $10 \mathrm{mg} / \mathrm{L}$ | $10 \mathrm{mg} / \mathrm{L}$ |
| TSS | $5 \mathrm{mg} / \mathrm{L}$ | $1 \mathrm{mg} / \mathrm{L}$ |
| $\mathrm{BOD}_{5}$ | $2 \mathrm{mg} / \mathrm{L}$ | $1 \mathrm{mg} / \mathrm{L}$ |
| $\mathrm{COD}_{\mathrm{Ammonia}}$ | $10 \mathrm{mg} / \mathrm{L}$ | $5 \mathrm{mg} / \mathrm{L}$ |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3}$ | $0.05 \mathrm{mg} / \mathrm{L}$ | $0.01 \mathrm{mg} / \mathrm{L}$ |
| $\mathrm{TKN}^{\text {Dissolved P }}$ | $0.1 \mathrm{mg} / \mathrm{L}$ | $0.05 \mathrm{mg} / \mathrm{L}$ |
| Total P | $0.2 \mathrm{mg} / \mathrm{L}$ | $0.2 \mathrm{mg} / \mathrm{L}$ |
| Total Cu | $0.02 \mathrm{mg} / \mathrm{L}$ | $0.01 \mathrm{mg} / \mathrm{L}$ |
| Total Pb | $0.05 \mathrm{mg} / \mathrm{L}$ | $0.02 \mathrm{mg} / \mathrm{L}$ |
| Total Ni | $2 \mu \mathrm{~g} / \mathrm{L}$ | $2 \mu \mathrm{~g} / \mathrm{L}$ |
| Total Zn | $3 \mu \mathrm{~g} / \mathrm{L}$ (residential $1 \mu \mathrm{~g} / \mathrm{L})$ | $1 \mu \mathrm{~g} / \mathrm{L}$ |

## Findings from Chapter 4: Stormwater Quality Descriptions Using the Three Parameter Lognormal Distribution

- Statistical Distributions. Knowing the statistical distributions of stormwater concentrations is a critical step in data analysis. The selection of the correct statistical analyses tools is dependent on the data distribution, and many QA/QC operations depend on examining the distribution behavior. However, much data is needed for accurate determinations of the statistical distributions of the data, especially when examining unusual behavior. The comparison of probability distributions between different data subsets is also a fundamental method to identify important factors affecting data observations.
- Log-Normal Statistical Distribution. Most of the stormwater constituents in the NSQD can be assumed to follow a lognormal distribution with little error. The use of the third parameter does not show a significant improvement in estimating the empirical distribution compared with the 2-parameter lognormal distribution. When the number of samples is very large per category (approximately more than 400 samples) the maximum likelihood and the 2parameter lognormal distribution better fit the empirical distribution. For large sample sizes, the L-moments method usually unacceptably truncates the distribution in the lower tail. When the sample size is small ( $<100$ samples), the
use of the third parameter does not improve the fit with the empirical distribution and the 2-parameter lognormal distribution produces a better fit than the other two methods.
- Effects of Data Errors. Incorrect data observations can have a great effect on the characteristics of the dataset. For example, when only $0.5 \%$ of the sample is affected by a factor of a thousand, the coefficient of variation increases almost 12 times more than the correct value. An error of a factor of a thousand occurs periodically, especially for heavy metal values when the concentrations are reported in $\mathrm{mg} / \mathrm{L}$ units when they are actually in $\mu \mathrm{g} / \mathrm{L}$ units.
- Data Observations Needed. Determining the number of data observations needed to compare two datasets with known, and similar distributions, can be readily determined. The following plot shows the power of the D test for $1 \%, 5 \%$, and $10 \%$ levels of confidence of the test for samples size larger than 35 . For example, assume that the maximum distance between the alternative cumulative and the estimated cumulative probability distributions is 0.2 (approximately a $20 \%$ difference in the concentrations in the datasets), and we want an $80 \%$ power ( 0.8 ) against the alternative at a $5 \%$ level of confidence. To calculate the number of required samples, we read that $\Delta(\mathrm{N})^{0.5}$ is 1.8 for a power of 0.8 and $5 \%$ level of confidence. Solving for $N=(1.8 / 0.2)^{2}=81$ samples. If we want to calculate the number of samples when the difference between the alternative cumulative and the estimated cumulative probability function is 0.05 (a difference of only $5 \%$ ), with the same power and level of confidence, then 1,296 samples would be required. When the lines are very close together, it is obviously very difficult to statistically show that they are different, and many samples are needed.

Lower Bounds for the Power of the $D$ test


| Difference <br> between datasets <br> to be detected $(\boldsymbol{\Delta})$ | Percentage <br> difference <br> between datasets | Number of samples <br> needed, $\mathbf{5} \%$ confidence, <br> $\mathbf{8 0 \%} \%$ power $\left[\mathbf{N}=(\mathbf{1 . 8} / \boldsymbol{\Delta})^{\mathbf{2}}\right]$ |
| :--- | :--- | :--- |
| 0.05 | 5 | 1,300 |
| 0.10 | 10 | 320 |
| 0.25 | 25 | 13 |
| 0.50 | 50 | 3 |
| 1.00 | 100 |  |

- Obviously, a careful decision has to be made between monitoring budgets and data quality objectives. The sample needs increase dramatically as the difference between datasets become small. Typically, a difference of about $25 \%$ (requiring about 50 sample pairs) is a reasonable objective for most stormwater projects. This is especially important when monitoring programs attempt to distinguish test and control conditions associated with stormwater control practices. It is easy to confirm significant differences between influent and effluent conditions at wet detention ponds, as they have relatively high removal rates. Less effective controls are much more difficult to verify, as the sampling program requirements become very expensive.


## Findings from Chapter 5: Identification of Significant Factors Affecting Stormwater Quality Using the NSQD

- Manual vs. Automatic Sampling. About $80 \%$ of the NSQD samples were collected using automatic samplers. It was observed that manual sampling can result in lower TSS concentrations compared to automatic sampling procedures. This may occur, for example, if the manual sampling team arrives after the start of runoff and therefore misses an elevated first flush (if it exists for the site), resulting in reduced event mean concentrations. The following figure contains box and whisker plots comparing resultant sample concentrations when the samples were collected by automatic versus manual sampling methods, for residential land uses in EPA Rain Zone 2.


- The decision to use automatic or manual sampling methods is not always clear. There were statistical differences found between both methods in residential areas for several constituents. Most communities calculate their EMC values using flow-composited sample analyses. If first flush effects are present, manual sampling may likely miss these more concentrated flows due to delays in arriving at the site to initiate sampling. If the first flush is for a very short duration, time-composited samples may overly emphasize these higher flows. Flow compositing produces more accurate EMC values than time composite analyses. An automatic sampler with flow-weighted samples, in conjunction with a bed load sampler, is likely the most accurate sampling alternative.
- Sample Compositing Methods. Time and flow-weighted composite options. were also evaluated in residential, commercial, and industrial land uses in EPA Rain Zone 2 and in industrial land uses in EPA Rain Zone 3.
- No significant differences were observed for $\mathrm{BOD}_{5}$ concentrations using either of the compositing schemes for any of the four categories. A similar result was observed for COD except for commercial land uses in EPA Rain Zone 2, where not enough samples were collected to detect a significant difference. TSS and total lead median concentrations in EPA Rain Zone 2 were two to five times higher in concentration when time-compositing was used instead of flow-compositing.
- Nutrients in EPA Rain Zone 2 collected in residential, commercial and industrial areas showed no significant differences using either compositing method. The only exceptions were for ammonia in residential and commercial land use areas and total phosphorus in residential areas where time-composite samples had higher concentrations. Metals were higher when time-compositing was used in residential and commercial land use areas. No differences were observed in industrial land use areas, except for lead.
- Stormwater Controls. The following figure shows the observed TSS concentrations in residential areas for EPA Rain Zone 2, for different drainage area stormwater controls (Channel weir: a flow measurement weir in an open channel that forms a small pool; Dry pond: a dry detention pond that drains completely between each storm event; Wet pond: a wet detention pond that retains water between events, forming a small lake or pond; Detention storage: Oversize pipes with small outlet orifices, usually under parking lots).

- There is a significant reduction in TSS, nitrite-nitrate, total phosphorus, total copper, and total zinc concentration at sites having wet detention ponds, the control practice having the largest concentration reductions. No reductions in TKN concentrations were found using wet ponds, but TKN seems to be reduced by dry ponds. Locations with detention storage facilities had smaller reductions of TSS, $\mathrm{BOD}_{5}$, COD , total lead and total zinc concentrations compared to wet pond sites. Unfortunately, there were few sites in the database having grass swales that could be compared with data from sites having curbs and gutters.
- Concentration Effects Associated with Impervious Cover Amounts. The following plot shows no apparent trend in TSS concentration that can be explained by impervious cover differences. However, it is very likely that a significant and important trend does exist between percent effective imperviousness and the pollutant mass that is discharged. While the relationship between imperviousness and concentration is not clear, the relationship between effective imperviousness and total runoff volume is much stronger (as noted previously) and more obvious as the non-paved areas can infiltrate much water.
- There is a certain amount of redundancy (self-correlation) between land use and the percentage of impervious areas, as each land use category generally has a defined narrow range of paved and roof areas. Therefore, it is not possible to test the hypothesis that different levels of impervious (surface coverage) are more important than differences in land use (activities within the area). Residential land uses cover only the lower range of imperviousness, while commercial sites have imperviousness amounts larger than $50 \%$. In order to perform a valid comparison test, the range of imperviousness needs to be similar for both test cases.



## TSS concentrations by impervious cover and single land use

- Seasonal Effects. Another factor that may affect stormwater quality is the season when the sample was obtained. If the few samples collected for a single site were all collected in the same season, the results may not be representative of the whole year. The NPDES sampling protocols were designed to minimize this effect by requiring the three samples per year to be separated by at least 1 month. The few samples still could be collected within a single season, but at least not within the same week. Seasonal variations for residential fecal coliform data are shown in the following figure for all residential areas. The bacteria levels are lowest during the winter season and highest during the summer and fall (a similar conclusion was obtained during the NURP, EPA 1983, data evaluations). The database does not contain any snowmelt data, so all of the data corresponds to rain-related runoff only. No other seasonal trends in stormwater quality were identified.

- Rain Depth Effects. The following figure contains several scatter plots showing concentrations plotted against rain depth. There are no obvious trends of concentration associated with rain depth for the NSQD data..




## Examples of scatter plots by precipitation depth

- No effects on concentration were observed according to precipitation depth. Rainfall energy determines erosion and washoff of particulates, but sufficient runoff volume is needed to carry the particulate pollutants to the outfalls. Different travel times from different locations in the drainage areas results in these materials arriving at different times, plus periods of high rainfall intensity (that increase pollutant washoff and movement) occur randomly throughout the storm. The resulting outfall stormwater concentration patterns for a large area having various surfaces is therefore complex and rain depth is just one of the factors involved. Chapter 6 examines time delivery of pollutants in more detail.
- Interevent Period Effects. The following figure shows box and whisker plots of the number of days having no rain before the monitored events by each EPA Rain Zone. Antecedent dry periods before sampling was found to have a significant effect for $\mathrm{BOD}_{5}, \mathrm{COD}$, ammonia, nitrates, TKN, dissolved, and total phosphorus concentrations at residential land use sites. As the number of days increased, there was an increase in the concentrations of the stormwater constituents. This relationship was not observed for freeway sites.
- Only EPA Rain Zone 2 has enough observations to evaluate possible effects of the antecedent dry period on the concentration of stormwater pollutants. In residential land uses, 7 out of 12 constituents indicated that antecedent dry period had a significant effect on the median concentrations. As the number of days having no rain increased, the concentrations also increased.

- Except for nitrates, all the nutrients have positive regressions inside the $95 \%$ confidence interval. In commercial land uses, the effects of antecedent dry periods on the median concentrations were less important. Only total phosphorus and total lead had significant regression results. As in the residential case, phosphorus has a positive coefficient with a small coefficient of determination. However, lead decreases with the number of dry days before the storm.
- Trends with Time. The following plots show likely decreasing lead and COD concentrations with time. Statistically however, the trend lines are not significant due to the large variation in observed concentrations.



## Findings from Chapter 6: Comparisons of First 30-minute Samples to 3-hour Composite Samples

- First-Flush Effects. Sample collection conducted for some of the NPDES MS4 Phase I permits required both a grab and a composite sample for each event. A grab sample was to be taken during the first 30 minutes of discharge, and a flow-weighted composite sample for the entire time of discharge (up to three hours). The initial grab sample was used for the analysis of the "first flush effect," which assumes that more of the pollutants are discharged during the first period of runoff than during later periods. The composite sample was obtained with aliquots collected about every 15 to 20 minutes for at least 3 hours, or until the event ended.
- About 400 paired sets of 30-minute and 3-hour samples were available for comparisons. The following table shows the results of the analyses.

Significant First Flushes Ratios (first flush to composite median concentration)

| Parameter | Commercial |  |  |  | Industrial |  |  |  | Institutional |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | SC | R | ratio | n | Sc | R | ratio | n | Sc | R | ratio |
| Turbidity, NTU | 11 | 11 | = | 1.32 |  |  | X |  |  |  | X |  |
| pH, S.U. | 17 | 17 | = | 1.03 | 16 | 16 | = | 1.00 |  |  | X |  |
| COD, mg/L | 91 | 91 | \# | 2.29 | 84 | 84 | \# | 1.43 | 18 | 18 | \# | 2.73 |
| TSS, mg/L | 90 | 90 | \# | 1.85 | 83 | 83 | = | 0.97 | 18 | 18 | \# | 2.12 |
| $\mathrm{BOD}_{5}, \mathrm{mg} / \mathrm{L}$ | 83 | 83 | \# | 1.77 | 80 | 80 | \# | 1.58 | 18 | 18 | \# | 1.67 |
| TDS, mg/L | 82 | 82 | \# | 1.83 | 82 | 81 | \# | 1.32 | 18 | 18 | \# | 2.66 |
| O\&G, mg/L | 10 | 10 | \# | 1.54 |  |  | X |  |  |  | X |  |
| Fecal Coliform, col/100mL | 12 | 12 | = | 0.87 |  |  | X |  |  |  | X |  |
| Fecal Streptococcus, col/ 100 mL | 12 | 11 | = | 1.05 |  |  | X |  |  |  | X |  |
| Ammonia, mg/L | 70 | 52 | \# | 2.11 | 40 | 33 | = | 1.08 | 18 | 16 | \# | 1.66 |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3,} \mathrm{mg} / \mathrm{L}$ | 84 | 82 | \# | 1.73 | 72 | 71 | \# | 1.31 | 18 | 18 | \# | 1.70 |
| N Total, mg/L | 19 | 19 | = | 1.35 | 19 | 16 | = | 1.79 |  |  | X |  |
| TKN, mg/L | 93 | 86 | \# | 1.71 | 77 | 76 | \# | 1.35 |  |  | X |  |
| P Total, mg/L | 89 | 77 | \# | 1.44 | 84 | 71 | = | 1.42 | 17 | 17 | = | 1.24 |
| P Dissolved, mg/L | 91 | 69 | = | 1.23 | 77 | 50 | = | 1.04 | 18 | 14 | = | 1.05 |
| Ortho-P, mg/L |  |  | X |  | 6 | 6 | = | 1.55 |  |  | X |  |
| Cadmium Total, $\mu \mathrm{g} / \mathrm{L}$ | 74 | 48 | \# | 2.15 | 80 | 41 | = | 1.00 |  |  | X |  |
| Chromium Total, $\mu \mathrm{g} / \mathrm{L}$ | 47 | 22 | \# | 1.67 | 54 | 25 | = | 1.36 |  |  | X |  |
| Copper Total, $\mu \mathrm{g} / \mathrm{L}$ | 92 | 82 | \# | 1.62 | 84 | 76 | \# | 1.24 | 18 | 7 | = | 0.94 |
| Lead Total, $\mu \mathrm{g} / \mathrm{L}$ | 89 | 83 | \# | 1.65 | 84 | 71 | \# | 1.41 | 18 | 13 | \# | 2.28 |
| Nickel, $\mu \mathrm{g} / \mathrm{L}$ | 47 | 23 | \# | 2.40 | 51 | 22 | = | 1.00 |  |  | X |  |
| Zinc, $\mu \mathrm{g} / \mathrm{L}$ | 90 | 90 | \# | 1.93 | 83 | 83 | \# | 1.54 | 18 | 18 | \# | 2.48 |
| Turbidity, NTU |  |  | X |  | 12 | 12 | = | 1.24 | 26 | 26 | = | 1.26 |
| pH, S.U. |  |  | X |  | 26 | 26 | = | 1.01 | 63 | 63 | = | 1.01 |
| COD, mg/L | 28 | 28 | = | 0.67 | 140 | 140 | \# | 1.63 | 363 | 363 | \# | 1.71 |
| TSS, mg/L | 32 | 32 | = | 0.95 | 144 | 144 | \# | 1.84 | 372 | 372 | \# | 1.60 |
| BOD5, mg/L | 28 | 28 | = | 1.07 | 133 | 133 | \# | 1.67 | 344 | 344 | \# | 1.67 |
| TDS, mg/L | 31 | 30 | = | 1.07 | 137 | 133 | \# | 1.52 | 354 | 342 | \# | 1.55 |
| O\&G, mg/L |  |  | X |  |  |  | X |  | 18 | 14 | \# | 1.60 |
| Fecal Coliform, col/100mL |  |  | X |  | 10 | 9 | = | 0.98 | 22 | 21 | $=$ | 1.21 |
| Fecal Streptococcus, col/100 mL |  |  | X |  | 11 | 8 | = | 1.30 | 26 | 22 | = | 1.11 |
| Ammonia, mg/L |  |  | X |  | 119 | 86 | \# | 1.36 | 269 | 190 | \# | 1.54 |
| $\mathrm{NO}_{2}+\mathrm{NO}_{3,} \mathrm{mg} / \mathrm{L}$ | 30 | 21 | $=$ | 0.96 | 121 | 118 | \# | 1.66 | 324 | 310 | \# | 1.50 |
| N Total, mg/L | 6 | 6 | $=$ | 1.53 | 31 | 30 | $=$ | 0.88 | 77 | 73 | = | 1.22 |
| TKN, mg/L | 32 | 14 | = | 1.28 | 131 | 123 | \# | 1.65 | 335 | 301 | \# | 1.60 |
| P Total, mg/L | 32 | 20 | $=$ | 1.05 | 140 | 128 | \# | 1.46 | 363 | 313 | \# | 1.45 |
| P Dissolved, mg/L | 32 | 14 | = | 0.69 | 130 | 105 | \# | 1.24 | 350 | 254 | $=$ | 1.07 |


| Ortho-P, $\mathrm{mg} / \mathrm{L}$ |  |  | X |  | 14 | 14 | $=$ | 0.95 | 22 | 22 | $=$ | 1.30 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cadmium Total, $\mu \mathrm{g} / \mathrm{L}$ | 30 | 15 | $=$ | 1.30 | 123 | 33 | $\neq$ | 2.00 | 325 | 139 | $\neq$ | 1.62 |
| Chromium Total, $\mu \mathrm{g} / \mathrm{L}$ | 16 | 4 | $=$ | 1.70 | 86 | 31 | $=$ | 1.24 | 218 | 82 | $\neq$ | 1.47 |
| Copper Total, $\mu \mathrm{g} / \mathrm{L}$ | 30 | 22 | $=$ | 0.78 | 144 | 108 | $\neq$ | 1.33 | 368 | 295 | $\neq$ | 1.33 |
| Lead Total, $\mu \mathrm{g} / \mathrm{L}$ | 31 | 16 | $=$ | 0.90 | 140 | 93 | $\neq$ | 1.48 | 364 | 278 | $\neq$ | 1.50 |
| Nickel, $\mu \mathrm{g} / \mathrm{L}$ |  |  | X |  | 83 | 18 | $=$ | 1.20 | 213 | 64 | $\neq$ | 1.50 |
| Zinc, $\mu \mathrm{g} / \mathrm{L}$ | 21 | 21 | $=$ | 1.25 | 136 | 136 | $\neq$ | 1.58 | 350 | 350 | $\neq$ | 1.59 |

Note: $\mathrm{n}=$ number of total possible events. $\mathrm{sc}=$ number of selected events with detected values. $\mathrm{R}=$ result. Not enough data (X); not enough evidence to conclude that median values are different $(=)$; median values are different $(\neq)$.

- Generally, a statistically significant first flush is associated with a median concentration ratio of about 1.4, or greater (the exceptions are where the number of samples in a specific category is much smaller). The largest ratios are about 2.5 , indicating that for these conditions, the first flush sample concentrations are about 2.5 times greater than the composite sample concentrations. More of the larger ratios are found for the commercial and institutional land use categories, areas where larger paved areas are likely to be found. The smallest ratios are associated with the residential, industrial, and open spaces land uses, locations where there may be larger areas of unpaved surfaces. Approximately $70 \%$ of the constituents in the commercial land use category had elevated first flush concentrations, about $60 \%$ of the constituents in the residential, institutional and the mixed (mostly commercial and residential) land use categories had elevated first flushes, and only $45 \%$ of the constituents in the industrial land use category had elevated first flushes. In contrast, no constituents were found to have elevated first flushes in the open space category.
- $\mathrm{COD}, \mathrm{BOD}_{5}$, TDS, TKN and Zn all had first flushes in all areas (except for the open space category). In contrast, turbidity, pH , fecal coliform, fecal streptococcus, total N , dissolved and ortho-P never showed a statistically significant first flush in any category.
- This investigation of first flush conditions indicated that a first flush effect was not present for all the land use categories, and certainly not for all constituents. Commercial and residential areas were more likely to show this phenomenon, especially if the peak rainfall occurred near the beginning of the event. It is expected that this effect will be more likely to occur in a watershed with a high level of imperviousness, but even so, the data indicated first flushes for less than $50 \%$ of the samples for the most impervious areas. This reduced frequency of observed first flushes in these areas most likely to have first flushes is likely associated with the varying rain conditions during the different events, including composite samples that did not represent the complete runoff durations.
- Groups of constituents showed different behaviors for different land uses. All the heavy metals evaluated showed higher concentrations at the beginning of the event in the commercial land use category. Similarly, all the nutrients showed higher initial concentrations in residential land use areas, except for total nitrogen and ortho-phosphorus. This phenomenon was not found in the bacteria analyses. None of the land uses showed a higher population of bacteria at the beginning of the event. Conventional constituents showed elevated concentrations in commercial, residential and institutional land uses.


## Findings from Chapter 7: Effects of Land Use and Geographical Location on Stormwater Quality

- ANOVA for land use and geographical location. All land uses in EPA Rain Zone two (except for freeways) have reduced TSS values when compared with the overall NSQD average. On the other hand, conditions in EPA Rain Zones 4, 6 and 9 have higher TSS values for the land uses noted. Industrial and freeway land uses increase the TSS concentrations compared with the other land uses, as expected from the one-way ANOVA tests. Of the 45 possible EPA Rain Zone and land use interactions, 21 have significantly different coefficients and resultant TSS concentrations. All of these possible TSS concentrations, based on this model, are shown in the following table.

TSS Concentrations (mg/L) for Different Land Uses and Rain Zones (if values not shown, use 60 $\mathrm{mg} / \mathrm{L}$ )

|  | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Open space |  | 40 |  | 139 |  |  |  |  |  |
| Residential |  | 40 |  | 84 |  | 92 | 35 |  |  |
| Commercial |  | 43 | 33 | 90 | 30 | 120 |  |  | 152 |
| Industrial | 25 |  |  | 103 | 147 | 206 | 121 |  |  |
| Freeways |  | 215 |  |  |  | 86 | 99 |  | 409 |

- Grouping of TSS Data. The following table shows the combined groups that had statistically similar TSS concentrations. The figure also indicates that about half of the TSS single land use data in the NSQD database were in the first group ( $52 \%$ ). Most of this data are from residential areas and EPA Rain Zone 2. Twenty-four percent of the observations were not affected by the land use - EPA Rain Zone interaction. Only $1.5 \%$ of the data are present in groups 4 and 5 . These groups are significantly different than groups 1 and 2 . Overall, there are three main levels of TSS concentrations in stormwater: Low (1), Medium (2) and High (3). Other minor categories correspond to groups 4 and 5 and contain the unusually high values.

Five TSS Concentration Categories in NSQD

|  | Land use*rain zone interactions (Rain Zone: land uses) | Concentrations (mean $\pm$ st. dev. in mg/L) | Range (mean; mean - st. dev. and mean + st. dev., $\mathrm{mg} / \mathrm{L}$ ) | Number of single land use TSS observations in category in NSQD |
| :---: | :---: | :---: | :---: | :---: |
| Low | 1: residential <br> 2: open space; residential; commercial <br> 3: commercial <br> 5: commercial <br> 7: residential | $3.69 \pm 1.12$ | 40 (13-123) | 1056 |
| Medium | All others not noted elsewhere | $4.02 \pm 1.11$ | 56 (18-169) | 478 |
| High | 4: residential; commercial; industrial; open space <br> 5: industrial <br> 6: freeways; residential; commercial <br> 7: freeways; industrial <br> 9: commercial | $4.60 \pm 1.20$ | 99 (30-330) | 460 |
| Unusually high 1 | 2: freeways 6: industrial |  |  | 22 |
| Unusually high 2 | 9: industrial |  |  | 9 |

TSS mg/L for single landuses using NSQD


Box plots of five groups also showing $5^{\text {th }}$ and $95^{\text {th }}$ percentiles.


Probability plot of the TSS data in the NSQD using the 5 main sample groups.

- Land Use and Geographical Area Interactions. When examining the detailed land use and seasonal interactions, it is clear that some of the constituents do not have many significant interactions in these factors, or that there are too few observations (or sites) represented in the NSQD. The constituents that have few, if any clear geographical area/land use interactions include: pH (I6), temperature, hardness, oil and grease (I5 and I7), TDS (C2), ammonia (C7), and dissolved P (R2 and R5). The values in the parentheses are the significant interaction terms (the land use and the EPA Rain Zone). If individual land use/geographical interaction cell values are not available, the overall land use, or overall data base summary values should be used:
- Constituents that should clearly be separated by land use: copper, lead, and zinc
- Constituents that clearly did not have any significant differences for different land use categories, therefore use overall values: pH , temperature (obvious seasonal effects), TDS, and TKN
- Constituents where residential data should be separated from commercial plus industrial area data: TSS (possible) and nitrates plus nitrites
- Constituents where it is not clear; conflicts in p values when comparing different combinations of land uses: hardness, oil and grease, $\mathrm{BOD}_{5}, \mathrm{COD}$, ammonia, total P , and dissolved P


## Findings from Chapter 8: Example Application of the National Stormwater Quality Database (TSS and Nutrient Export Calculations for Chesapeake Bay Watersheds)

- Mass Discharge Calculations. This chapter demonstrates how the NSQD information can be used to make mass discharge calculations for large drainage areas. This is an example for Maryland's Anne Arundel County, an important tributary of Chesapeake Bay. TSS and nutrient concentrations for the urban land uses in the county were calculated using NSQD data for Maryland and Virginia. Various factors were found to influence these concentrations using ANOVA analyses. Specifically, season, rain depth, and impervious cover were examined for each land use category. The resulting coefficients of variation were all significantly reduced with these categories of data, as shown on the following table.


## Average Concentrations by Land Use

| Land Use | Constituent | Conditions | Average (COV) |
| :---: | :---: | :---: | :---: |
| Urban - Residential | TSS | Summer rains (between 0.1 and 0.35 inches in depth) | 143 (0.71) |
|  |  | All other rains | 58 (0.70) |
|  | TP | Sites having <27\% impervious cover: |  |
|  |  | Winter rains | 0.28 (0.59) |
|  |  | All other rains | 0.41 (0.65) |
|  |  | Sites having >27\% impervious cover: |  |
|  |  | Winter rains (less than 0.1 inches in depth) | 0.16 (0.86) |
|  |  | All other rains | 0.30 (0.63) |
|  | TN | Fall rains (less than 0.1 and greater than 1 inch in depth) | 1.4 (0.57) |
|  |  | Winter rains (0.35 and 1 inch in depth) | 1.5 (0.30) |
|  |  | Fall rains ( 0.35 and 1 inch) and Winter rains (between 0.1 and 0.35 inches in depth) | 1.9 (0.51) |
|  |  | All other rains | 2.4 (0.62) |
|  |  | Spring and summer rains (between 0.1 and 0.35 inches in depth) | 2.6 (0.38) |
| Urban - Commercial | TSS and TN | See tables 7.4 and 7.5 |  |
|  | TP | Summer rains $>1$ inch and fall rains between 0.1 and 0.35 inch | 0.46 (0.36) |
|  |  | All other rains | 0.23 (0.71) |
| Urban - Industrial | TSS | Fall, spring, and summer | 77 (1.48) |
|  |  | Winter | 81 (0.93) |
|  | TP | Rains less than 0.35 inches | 0.29 (0.81) |
|  |  | Rains greater than 0.35 inches | 0.22 (1.05) |
|  | TN | All conditions | 2.1 (0.79) |
| Rural - Agricultural | Sediment |  | $\begin{aligned} & 1115 \mathrm{lb} / \mathrm{ac} / \mathrm{yr} \\ & \text { (unreliable estimate) } \end{aligned}$ |
|  | TN |  | $40 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |
|  | TP |  | $5.4 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |
| Rural - Forest | Sediment |  | $\begin{aligned} & \hline 4500 \mathrm{lb} / \mathrm{ac} / \mathrm{yr} \\ & \text { (unreliable estimate) } \end{aligned}$ |
|  | TN |  | $0 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |
|  | TP |  | $0 \mathrm{lb} / \mathrm{ac} / \mathrm{yr}$ |

Total Suspended Solids Concentrations and Volumetric Runoff Coefficients for Land Use Categories in Anne Arundel County, Maryland

| Land Use Description | \# of acres in 2000 | $\begin{aligned} & \text { TSS } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | Concentration reliability? | Rv | $R_{V}$ reliability? |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Large lot subdivision (1 unit/ 5-10 ac) | 0 | 60 | excellent | 0.09 | excellent |
| Low-density residential (1 unit/ 5 acres to 2 units/acre) | 33,337 | 60 | excellent | 0.14 | excellent |
| Medium-density residential (2 to 8 units/acre) | 33,791 | 60 | excellent | 0.23 | excellent |
| High-density residential (8+ units/acre) | 6,274 | 60 | excellent | 0.34 | excellent |
| Commercial | 11,670 | 58 | excellent | 0.72 | excellent |
| Industrial | 3,249 | 80 | excellent | 0.52 | excellent |
| Institutional (schools, churches, military institutions, etc.) | 9,813 | 58 | moderate | 0.49 | excellent |
| Open urban land | 4,139 | 50 | moderate | 0.08 | excellent |
| Transportation | 1,557 | 99 | moderate | 0.41 | excellent |
| Extractive | 1,686 | 350 | poor | 0.3 | moderate |
| Deciduous forest | 43,901 | 90 | moderate | 0.08 | moderate |
| Evergreen forest | 4,891 | 90 | moderate | 0.08 | moderate |
| Mixed forest | 56,621 | 90 | moderate | 0.08 | moderate |
| Brush | 2,565 | 90 | poor | 0.08 | moderate |
| Wetlands | 1,643 | 0 | poor | 0.65 | moderate |
| Beaches | 29 | 0 | poor | 0.1 | moderate |
| Bare ground | 224 | 1000 | poor | 0.3 | moderate |
| Row and garden crops | 300 | 357 | very good | 0.2 | poor |
| Cropland | 42,368 | 357 | very good | 0.2 | poor |
| Orchards / vineyards / horticulture | 63 | 357 | very good | 0.15 | poor |
| Pasture | 4,690 | 145 | very good | 0.08 | moderate |
| Feeding operations | 49 | 145 | very good | 0.2 | poor |
| Agricultural building, breeding and training facilities | 163 | 145 | very good | 0.5 | poor |



Calculated Sources of TSS for Anne Arundel County, Maryland

## Research Hypotheses

The main hypothesis for this research was that commonly accepted assumptions concerning stormwater characteristics are correct and applicable for a wide range of conditions, including different land uses, precipitation depths, seasons, watershed area and geographic locations throughout the U.S. This assumption was evaluated by testing the following hypotheses:

## Research Hypothesis 1. Lognormal distributions are robust descriptions of stormwater quality data and a few unusual values have little effect on dataset summary statistical descriptions.

A total of 25 constituents in 5 land uses were evaluated using the NSQD database. In $71 \%$ of the cases, lognormal distributions better described the stormwater constituent concentrations compared with gamma and exponential distributions. These last two distributions better represented $10 \%$ and $4 \%$ of the cases, respectively. In $15 \%$ of the cases, lognormal, gamma and exponential distributions did not adequately represent stormwater constituent concentrations. Constituents that mostly were not well described by any of these three distributions included: $\mathrm{BOD}_{5}$, total arsenic, total cadmium and total copper in residential, commercial and industrial land uses.

Gamma and exponential distributions better described bacteria and nutrient concentrations in open space land use areas.

The use of the 3-parameter lognormal distribution, did not improve the description information compared with the simpler 2-parameter distribution. The 2-parameter lognormal distribution is therefore recommended for those constituents were the use of lognormal distributions produced a better fit of the data.

Unusually elevated values have a significant effect in the mean, median, standard deviation and coefficient of variation of the sample distribution. As an example, when $0.5 \%$ of the data are affected by a factor of a thousand (such as may occur when heavy metals are incorrectly expressed with $\mathrm{mg} / \mathrm{L}$ units when they should be $\mu \mathrm{g} / \mathrm{L}$ units), the coefficient of variation will be increased almost 15 times compared to the value when the extreme observations are not present. The effect on the coefficient of variation is larger as the percentage of extreme samples is reduced.

Unusually low values do not have a significant effect on the mean, median, standard deviation, or coefficient of variation, unless the percentage of samples having the low values is higher than $25 \%$.

## Research Hypothesis 2. Censored data can be adequately adjusted by substituting half of the detection limit, with little resulting effects on the mean and variance of stormwater datasets.

Replacing non-detected observations by half of the detection limit is appropriate when the percentage of left censored observations (those having concentrations lower than the detection limit) is lower than $15 \%$ of the total data set. Replacing the non-detected values with zero will have more extreme effects on these distribution summary values.

Ignoring the non-detected values will result in higher means, medians and standard deviations, and lower coefficients of variation than the true values for the distributions.

The use of the Cohen's maximum likelihood method is recommended to replace the censored observations for those constituents that have lognormal distributions. This is an appropriate method when non-paired statistical tests are to be performed, as the assignment of replacement values for specific tests in not important. However, no replacements are suitable when paired comparison tests are to be conducted, and these tests should only be conducted on the data sets having complete pairs (not using pairs where one or both parts of the pair are below detection). When calculating percentage reductions, or other comparison tests, non-detected effluent concentrations can be used, without substitutions, in calculations to determine the lower limit of removal.

When the number of non-detected observations exceed about $40 \%$ of the total number of observations, no substitution method (neither the maximum likelihood method or half of the detection limit) is suitable.

## Research Hypothesis 3. Different levels of imperviousness are more important than differences in land use categories when predicting stormwater constituent concentrations.

The use of the impervious area information alone did not reduce the uncertainty about the variability of stormwater constituents. One of the main factors associated with land use concerns the activities that occur in the land use. It is expected that the use of both factors (land use and information about the surface covers in the area, such as the percentage of impervious areas) will reduce the variability of the stormwater concentrations observed, rather than when only one of these factors is considered. However, these tests were only conducted on stormwater concentrations, not on mass discharges. Increases in impervious cover are directly associated with increases in runoff volumes, and therefore in pollutant mass discharges.

When only residential area data from EPA Rain Zone 2 were used, the percentage of impervious areas was found to have a significant effect on the concentration of nitrates. The concentrations of nitrates were reduced as the percentage of impervious cover increased. This is an expected finding; when the impervious areas increase, less landscaping is likely (a major source of nutrient discharges).

No significant relationships were observed between the amount of impervious cover and any stormwater constituent concentration that was examined for industrial and commercial land use areas.

## Research Hypothesis 4. Antecedent dry periods have a significant effect on stormwater constituent concentrations.

Antecedent dry periods are not the same for all the EPA Rain Zones in the country. Longer antecedent dry periods occur at west coast sites compared to other locations.

The antecedent dry periods had a positive and significant $(\alpha=5 \%)$ effect on the concentration of 7 of 13 constituents examined: nutrients (ammonia, nitrates, TKN, total and dissolved phosphorus), COD and $\mathrm{BOD}_{5}$ at
residential sites located in EPA Rain Zone 2. It was not significant in oil and grease, TDS, TSS, total copper, total lead and total zinc.

Only total phosphorus and total lead concentrations were affected by the antecedent dry period at commercial sites located in EPA Rain Zone 2. Total phosphorus concentrations increased with increasing days before the sampled storm. An opposite relation was observed for total lead at commercial sites.

Only TSS was affected by the antecedent dry period at industrial sites located in EPA Rain Zone 2. A positive relationship was observed, with TSS concentrations increasing as the number of antecedent dry days increased.

## Research Hypothesis 5. Outfall samples collected during the "first flush" periods of storms have significantly greater concentrations than total storm composite samples.

The first flush effect was not present for all constituents and all land uses. The phenomenon was most likely to occur in commercial and high density residential land uses, watersheds having high percentages of impervious areas. It was not observed in open space areas, watersheds having low percentages of impervious areas.

TSS, COD, TDS, total copper, total lead, total zinc and TKN had observed flush concentrations that were significantly higher than the composite sample concentrations in those areas where the "first flush" was most likely to occur. pH was the only constituent that did not indicate a first flush effect. Observed elevated first flush concentrations were less than 3 times higher than the corresponding storm composite concentrations.

## Recommendations for Future Stormwater Permit Monitoring Activities

- The NSQD is an important tool for the analysis of stormwater discharges at outfalls. About a fourth of the total existing information from the NPDES Phase I program is included in the database. Most of the analyses in this research were performed for residential, commercial and industrial land uses in EPA Rain Zone 2 (the area of emphasis according to the terms of the EPA funded research). Much more data are available from other stormwater permit holders that were not included in this database. Acquiring this additional data for inclusion in the NSQD is a recommended and cost-effective activity and should be accomplished as additional data are also being obtained from on-going monitoring projects.
- The use of automatic samplers, coupled with bedload samplers, is preferred over manual sampling procedures. In addition, flow monitoring and on-site rainfall monitoring needs to be included as part of all stormwater characterization monitoring. The additional information associated with flow and rainfall data will greatly enhance the usefulness of the much more expensive water quality monitoring. Flow monitoring must also be correctly conducted, with adequate verification and correct base-flow subtraction methods applied. A related issue frequently mentioned by the monitoring agencies is the lack of on-site rainfall information for many of the sites. Using regional rainfall data from locations distant from the monitoring location is likely to be a major source of error when rainfall factors are being investigated.
- Many of the stormwater permits also only required monitoring during the first three hours of the rain event. This may have influenced the event mean concentrations if the rain event continued much beyond this time. Flowweighted composite monitoring should continue for the complete rain duration. Monitoring only three events per year from each monitoring location requires many years before statistically adequate numbers of observations are obtained. In addition, it is much more difficult to ensure that such a small fraction of the total number of annual events is representative. Also, there is minimal value in obtaining continued data from an area after sufficient information is obtained. It is recommended that a more concentrated monitoring program be conducted for a two or three year period, with a total of about 30 events monitored for each site, covering a wide range of rain conditions. Periodic checks can be made in future years, such as repeating concentrated monitored every 10 years, or so (and for only 15 events during the follow-up surveys).
- Finally, better watershed area descriptions, especially accurate drainage area delineations, are needed for all monitored sites. While the data contained in the NSQD is extremely useful, it is believed that future monitoring information obtained as part of the stormwater permit program would be greatly enhanced with these additional considerations.


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## Appendix A: Sites Included in the Database

The following table shows the number of samples, land use and community for each site, along with the site ID.
Table A1. Site Name and Land Use

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| AL | ALHUCHIP | ID | City of Huntsville | Chase Industrial Park |
| AL | ALHUDRAV | RE | City of Huntsville | Drake Avenue |
| AL | ALHUHURI | RE | City of Huntsville | Hunters Ridge |
| AL | ALHUMASM | CO | City of Huntsville | Madison Square Mall |
| AL | ALHUWERP | CO | City of Huntsville | Western Research Park |
| AL | ALJC004L | CO FW ID RE IS | Jefferson County | C004L |
| AL | ALJC004R | CO FW ID RE IS | Jefferson County | C004R |
| AL | ALJCC001 | FW ID CO | Jefferson County | C001 |
| AL | ALJCC002 | RE OP ID CO | Jefferson County | C002 |
| AL | ALJCC009 | RE FW | Jefferson County | C009 |
| AL | ALJCC010 | RE FW | Jefferson County | C010 |
| AL | ALJCC012 | CO FW RE OP | Jefferson County | C012 |
| AL | ALMOCREO | RE | City of Mobile | Creola |
| AL | ALMODAPH | CO | City of Mobile | Daphne |
| AL | ALMOSARA | RE | City of Mobile | Saraland |
| AL | ALMOSIIV | ID | City of Mobile | Mobile Site IV |
| AL | ALMOSITV | CO | City of Mobile | Mobile Site V |
| AL | ALMOSIVI | RE | City of Mobile | Mobile Site VI |
| AL | ALMOTHEO | ID | City of Mobile | Theodore |
| AZ | AZMCA001 | ID | Maricopa Cnty | 48th Street Drain |
| AZ | AZMCA002 | OP | Maricopa Cnty | South Mountain Park |
| AZ | AZMCA003 | ID | Maricopa Cnty | 27 th Ave at Salt River |
| AZ | AZMCA004 | RE CO | Maricopa Cnty | Aqua Fria at Youngtown |
| AZ | AZMCA005 | CO | Maricopa Cnty | 43rd Ave at Peoria |
| AZ | AZMCA006 | RE | Maricopa Cnty | 67th Ave Olive Ave at Glendale |
| AZ | AZTUA001 | RE | Tucson | Grant Road and Wilson Avenue |
| AZ | AZTUA002 | RE | Greenlee Road |  |
|  | Tucson |  |  |  |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| AZ | AZTUA003 | CO | Tucson | El Con Mall |
| AZ | AZTUA004 | ID | Tucson | 17th Street |
| CA | CAALAL03 | ID RE | Alameda County | Woods Street |
| CA | CAALAL04 | RE CO FW | Alameda County | Alice Street and 4th |
| CA | CAALAL07 | CORE | Alameda County | Cotter Way |
| CA | CAALAL09 | ID | Alameda County | Pacific Street |
| CA | CAALAL10 | ID CO RE | Alameda County | 37TH ST 8TH AVE |
| CA | CACTA001 | FW | Caltrans | 307 Sacramento |
| CA | CACTA002 | FW | Caltrans | 435 Solano |
| CA | CACTA003 | FW | Caltrans | 6205 Fresno |
| CA | CACTA004 | FW | Caltrans | 6209 Fresno |
| CA | CACTA005 | FW | Caltrans | 701 Los Angeles |
| CA | CACTA006 | FW | Caltrans | 7127 Los Angeles |
| CA | CACTA007 | FW | Caltrans | 7128 Los Angeles |
| CA | CACTA008 | FW | Caltrans | 7201 Los Angeles |
| CA | CACTA009 | FW | Caltrans | 7202 Los Angeles |
| CA | CACTA010 | FW | Caltrans | 7203 Los Angeles |
| CA | CACTA011 | FW | Caltrans | 801 Riverside |
| CA | CACTA012 | FW | Caltrans | 802 San Bernardino |
| CA | CACTA013 | FW | Caltrans | 803 Riverside |
| CA | CACTA014 | FW | Caltrans | 1201 Orange |
| CA | CACTA015 | FW | Caltrans | 1202 Orange |
| CO | COCSA001 | CO OP RE | Colorado Springs | Sixteenth Hole Valley Hi Golf Course |
| CO | COCSA002 | ID OP | Colorado Springs | Chestnut Street at Douglas Creek |
| CO | COCSA003 | ID CO | Colorado Springs | Beacon Street at Buchanan Street |
| CO | COCSA004 | RE OP | Colorado Springs | Wasatch Street at Cross Lane |
| CO | COCSA005 | OP CO ID | Colorado Springs | Wal-Mart at Eighth Street |
| CO | CODEA001 | CO | Denver Metro | Cherry Creek Storm Drain at Colfax Ave |
| CO | CODEA002 | CO | Denver Metro | Cherry Creek Storm Drain at University Blvd |
| CO | CODEA003 | RE | Denver Metro | North Sanderson Gulch Tributary at Lakewood |
| CO | CODEA004 | ID | Denver Metro | Sand Creek Tributary at 34th and Havana |
| CO | CODEA005 | RE | Denver Metro | Shop Creek at Parker Road |
| CO | CODEA006 | ID | Denver Metro | South Platte River Storm Drain at 54th and Steele |
| CO | CODEA007 | ID | Denver Metro | South Platte River Storm Drain at 7th Ave |
| CO | CODEA008 | CO | Denver Metro | Villa Italia Storm Drain at Lakewood |
| GA | GAATAT01 | ID | City of Atlanta | Ellsworth Industrial Drive |
| GA | GAATAT02 | RE | City of Atlanta | Beverly Road Doncaster Drive |
| GA | GACLCOSI | ID | Clayton County | Southridge Industrial Park |
| GA | GACLCOTR | RE | Clayton County | Tara Road |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| GA | GACOC1A2 | RE CO | Cobb County | Cobb Long Term 1 Pebble Creek |
| GA | GACOC1A3 | RE | Cobb County | Cobb Long Term 1 Sewell Mill Creek Roswell Road |
| GA | GACOCOL2 | RE CO | Cobb County | Cobb Long Term 2 Worley Rd Noonday Creek |
| GA | GADKCOTD | ID CO | Dekalb County | Truman Drive |
| GA | GAFUCOS1 | RE OP | Fulton County | Johns Creek Buice Road |
| GA | GAFUCOS2 | ID OP | Fulton County | Boat Road Blvd Grange Blvd |
| GA | GAFUCOS3 | RE CO | Fulton County | Long Island Creek Northside Drive |
| ID | IDADA001 | COID | Ada County Highway District | Koppels Site |
| ID | IDADA002 | RE | Ada County Highway District | Lucky Drive Site |
| ID | IDADA003 | RE FW | Ada County Highway District | Franklin Road Site |
| ID | IDADA004 | COID | Ada County Highway District | Production Avenue Site |
| KA | KATOATWO | RE | City of Topeka | Atwood |
| KA | KATOBROO | RE | City of Topeka | Brookfield |
| KA | KATOJACK | CO | City of Topeka | Jackson |
| KA | KATOSTFE | ID | City of Topeka | Santee |
| KA | KAWIHUNT | RE | City of Wichita | Huntington |
| KA | KAWIMCLE | ID | City of Wichita | McLean |
| KA | KAWISBWY | RE CO | City of Wichita | Broadway |
| KA | KAWITOWN | CO | City of Wichita | Towne East |
| KY | KYLOTSR1 | RE | City of Louisville | Buechel |
| KY | KYLOTSR2 | ID | City of Louisville | Obannon |
| KY | KYLOTSR3 | RE | City of Louisville | St Matthews |
| KY | KYLOTSR4 | ID | City of Louisville | Okolona |
| KY | KYLOTSR5 | RE CO | City of Louisville | Pleasure Ridge Park |
| KY | KYLOTSR6 | RE CO | City of Louisville | Hurstbourne Acres |
| KY | KYLXEHL4 | OP | City of Lexington | Lakeside golf |
| KY | KYLXEHL5 | OP | City of Lexington | Walnut Hill Chilesburg |
| KY | KYLXEHL6 | FW | City of Lexington | Alumni ManOwar |
| KY | KYLXEHL7 | RE | City of Lexington | Squires Road |
| KY | KYLXNEL1 | RE | City of Lexington | Greenbrier East |
| KY | KYLXNEL2 | RE OP | City of Lexington | Greenbrier |
| KY | KYLXNEL3 | CO | City of Lexington | Eastland |
| KY | KYLXTBL1 | RE | City of Lexington | Mt Vernon |
| KY | KYLXTBL2 | ID | City of Lexington | Leestown |
| KY | KYLXTBL3 | OP | City of Lexington | Viley Road |
| KY | KYLXWHL1 | CO | City of Lexington | Wilhite Drive |
| MA | MABOA001 | RE OP | Boston | Charlestown $29 . J 212$ |
| MA | MABOA002 | RE | Boston | West Roxbury 13D077 078 |
| MA | MABOA003 | CO | Boston | Dorchester 8J102 |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| MA | MABOA004 | ID | Boston | Brighton 25E037 |
| MA | MABOA005 | FW ID | Boston | Hyde Park 2F120 |
| MA | MABOA006 | RE | Boston | Mount Vernon 26K099 |
| MA | MABOA007 | OP | Boston | Wesley G Ross 6G108 |
| MD | MDAACOMW | ID | Anne Arundel | Midway industrial park MW |
| MD | MDAACOOD | RE | Anne Arundel County | Odenton OD |
| MD | MDAACOPP | CO | Anne Arundel | Parole Plaza PP |
| MD | MDAACORK | RE | Anne Arundel | Rolling Knolls RK |
| MD | MDAACOSC | COID | Anne Arundel | Science Drive SC |
| MD | MDBACOBC | ID | Baltimore County | Brien Run BC |
| MD | MDBACOLC | CO RE OP | Baltimore County | Long Quarter Branch LC |
| MD | MDBACOSC | RE | Baltimore County | Spring Branch SC |
| MD | MDBACOTC | ID | Baltimore County | Tobasoo creek TC |
| MD | MDBACOWC | RE | Baltimore County | White Marsh Run WC |
| MD | MDBCTYBO | ID | Baltimore City | BO |
| MD | MDBCTYFM | ID | Baltimore City | FM |
| MD | MDBCTYHA | RE CO | Baltimore City | Hamilton HA |
| MD | MDBCTYHO | RE | Baltimore City | Home land HO |
| MD | MDBCTYHR | RE | Baltimore City | Herring Run HR |
| MD | MDBCTYKO | CO | Baltimore City | Coppers Avenue KO |
| MD | MDCHCOIP | ID | Charles County | IP |
| MD | MDCHCOPA | RE | Charles County | PA |
| MD | MDCHCOPF | RE | Charles County | PF |
| MD | MDCHCOTG | CO ID | Charles County | TG |
| MD | MDCLCOBP | COID | Carroll County | Route 97 airport industrial BP |
| MD | MDCLCOCE | RE | Carroll County | Candice estates CE |
| MD | MDCLCOJS | CO | Carroll County | John street JS |
| MD | MDCLCOKW | OP RE | Carroll County | Kate Wagner KW |
| MD | MDCLCOSD | RE ID | Carroll County | Sunset Drive SD |
| MD | MDHACFBA | XX | Harford County | FBA |
| MD | MDHACOBP | RE | Harford County | Brentwood Park Woodland Hills |
| MD | MDHACOCF | CO | Harford County | Constant Friendship CF |
| MD | MDHACOCS | RE | Harford County | Cool Spring CS |
| MD | MDHACOGR | RE | Harford County | Green Ridge-II GR |
| MD | MDHACOIC | ID | Harford County | Greater Harford industrial centre IC |
| MD | MDHOCODC | CO | Howard County | Dobbin center DC |
| MD | MDHOCOFM | ID | Howard County | Food market FM |
| MD | MDHOCOGM | RE | Howard County | Green Moon GM |
| MD | MDHOCOMH | RE | Howard County | Murray Hill MH |
| MD | MDHOCOOC | ID | Howard County | Oak land centre OC |
| MD | MDMOCOBC | CO | Montgomery County | Burtons ville crossing BC |
| MD | MDMOCOCV | ID | Montgomery County | Coles villeCV |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| MD | MDMOCONV | RE | Montgomery County | Venture V |
| MD | MDMOCOQA | RE | $\begin{gathered} \text { Montgomery } \\ \text { County } \end{gathered}$ | Quaint Acres QA |
| MD | MDMOCOSL | ID | Montgomery County | Southlawn lane SL |
| MD | MDMOCOWP | CO | Montgomery County | Wheaten plaza WP |
| MD | MDPGCOS1 | CO | Prince Georges County | Brightseat Rd S1 |
| MD | MDPGCOS2 | RE | Prince Georges County | Flagstaff Street S2 |
| MD | MDPGCOS3 | CO ID | Prince Georges County | Maryland 50 industrial park S3 |
| MD | MDPGCOS4 | RE | Prince Georges County | Wayne Place S4 |
| MD | MDPGCOS5 | ID | Prince Georges County | John Hanson S5 |
| MD | MDPGCOS6 | ID | Prince Georges County | Pennsy Dr N3 |
| MD | MDSHDTDV | OP ID | State Highway | DV |
| MD | MDSHDTPS | OP ID | State Highway | PS |
| MN | MNMISD01 | RE | City of Minneapolis | E Harriet Pkwy W44 St |
| MN | MNMISD02 | RE | City of Minneapolis | Luella St Orange Ave |
| MN | MNMISD03 | ID | City of Minneapolis | Vandalia st |
| MN | MNMISD04 | RE CO | City of Minneapolis | Charles Ave |
| MN | MNMISD05 | RE CO | City of Minneapolis | E 29 St 31 Ave S |
| NC | NCCHBREV | ID | City of Charlotte | Brevi1 |
| NC | NCCHHIDD | RE | City of Charlotte | Hiddr2 |
| NC | NCCHHOSK | ID | City of Charlotte | Hoski2 |
| NC | NCCHNANC | RE | City of Charlotte | Nancr1 |
| NC | NCCHROSE | RE ID OP CO | City of Charlotte | Rosem1 |
| NC | NCCHSHEF | OP RE | City of Charlotte | Shefo1 |
| NC | NCCHSIMS | RE | City of Charlotte | Simsr3 |
| NC | NCCHSTAR | CO | City of Charlotte | Starc1 |
| NC | NCCHYARD | CORE | City of Charlotte | Yardc2 |
| NC | NCFV71ST | IS | City of Fayetteville | 71 ST High School 100ft NE Raeford SR1409 |
| NC | NCFVCLEA | RE | City of Fayetteville | 3606 Clearwater Drive |
| NC | NCFVELMS | CO | City of Fayetteville | ELM Street Eutaw Shopping Center |
| NC | NCFVROSE | RE OP CO | City of Fayetteville | Rose Apartments 225 Tiffany Court |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| NC | NCFVSTRK | OP | City of <br> Fayetteville | Strickland Bridge Road |
| NC | NCFVTRYO | RE | City of <br> Fayetteville | 1740 Tryon Rd |
| NC | NCFVWINS | ID | City of <br> Fayetteville | Winslow Pine Railroad tracks |
| NC | NCGRATHE | CO | City of <br> Greensboro | Athena |
| NC | NCGRCOUN | OP | City of <br> Greensboro | Country Park |
| NC | NCGRHUST | ID | City of <br> Greensboro | Husbands Street |
| NC | NCGRMERR | CO | City of <br> Greensboro | Merrit Drive |
| NC | NCGRRAND | RE | City of <br> Greensboro | Randlem Road |
| NC | NCGRUNIO | ID CO IS RE | City of <br> Greensboro | Union Street |
| NC | NCGRWILL | RE | City of <br> Greensboro | Willoughby |
| NC | NCRASIT1 | OP RE | City of Raleigh | I40 400ft east S State Street |
| NC | NCRASIT2 | RE CO | City of Raleigh | Williamson Drive Pineview Street |
| NC | NCRASIT3 | RE CO OP | City of Raleigh | I40 Dandridge Drive Bunche Drive |
| NC | NCRASIT4 | CO RE | City of Raleigh | Williamson Drive Wade Avenue |
| NC | NCRASIT5 | ID OP | City of Raleigh | Pylon Drive 100ft North Hutton Street |
| NC | NCRASIT6 | ID RE | City of Raleigh | South Wilmington Street City Farm |
| NC | NCRASIT7 | CO RE ID OP | City of Raleigh | 50ft east N West Street Peace Street |
| Dortch Street |  |  |  |  |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| OR | ORODA003 | FW | ODOT | Eugene |
| OR | ORODA004 | FW | ODOT | Neskowin |
| OR | ORODA005 | FW | ODOT | Portland |
| OR | ORPOA001 | CO | City of Portland | C 1 Jantzen Beach |
| OR | ORPOA002 | CO RE | City of Portland | C 2 Salmon Street |
| OR | ORPOA003 | ID | City of Portland | 11 Yeon Ave 35th Ave |
| OR | ORPOA004 | ID | City of Portland | 12 Swan Island |
| OR | ORPOA005 | RE CO | City of Portland | M 1 Columbia Slough |
| OR | ORPOA006 | RE | City of Portland | R 2 Sandy Boulevard |
| OR | ORPOA007 | FW | City of Portland | T 1 |
| OR | ORSAA001 | CO RE | City of Salem | Commercial |
| OR | ORSAA002 | CO | City of Salem | Cottage |
| OR | ORSAA003 | ID | City of Salem | Edgewater |
| OR | ORSAA004 | RE | City of Salem | Redleaf |
| PA | PAPH0864 | RE | Philadelphia | Cresheim Creek |
| PA | PAPH0891 | RE | Philadelphia | Tacony Creek |
| PA | PAPH1014 | RE | Philadelphia | Byberry Creek |
| PA | PAPH1051 | RE CO | Philadelphia | Wooden Bridge Run |
| PA | PAPH1182 | OP RE | Philadelphia | North Byberry Creek |
| TN | TNKXTYAP | ID OP RE | City of Knoxville | Acker Place |
| TN | TNKXTYFC | RE FW | City of Knoxville | First Creek |
| TN | TNKXTYGV | RE OP | City of Knoxville | Gallaher View |
| TN | TNKXTYTC | OP ID IS RE | City of Knoxville | Third Creek |
| TN | TNKXTYWE | CO RE IS OP | City of Knoxville | Wellington Drive |
| TN | TNMET207 | OP | City of Memphis | 207 Walnut Grove |
| TN | TNMET211 | ID | City of Memphis | 211 Warford |
| TN | TNMET231 | RE | City of Memphis | 231 Raleigh Lagrange |
| TN | TNMET260 | CO RE | City of Memphis | 260 Austin Peay |
| TN | TNMET410 | RE | City of Memphis | 410 Whitehaven |
| TX | TXARA001 | CO | City of Arlington | The Parks mall AC603 |
| TX | TXARA002 | RE | City of Arlington | R Legacy PK AR602 |
| TX | TXARA003 | RE | City of Arlington | Trib to W FK Tri AR601 |
| TX | TXARA004 | RE | City of Arlington | Trib To Johnson Creek AI604 |
| TX | TXDAA001 | ID | City of Dallas | Joes Cr 138 |
| TX | TXDAA002 | ID | City of Dallas | Bastille St 325 |
| TX | TXDAA003 | RE ID CO | City of Dallas | Knights Branch 34 |
| TX | TXDAA004 | RE | City of Dallas | White Rock Creek 86 |
| TX | TXDAA005 | RE | City of Dallas | Ash Creek 55 |
| TX | TXDAA006 | RE OP | City of Dallas | Newton Creek 189 |
| TX | TXDCA001 | OP FW | TXDOT Dallas | Mountain Creek DH902 |
| TX | TXDCA002 | OP FW | TXDOT Dallas | Bachman Branch DH901 |
| TX | TXFWA001 | OP RE | City of Fort Worth | Clear FK Trin R TRI STG1 |
| TX | TXFWA002 | OP IS | City of Fort Worth | Pylon St PY1 |
| TX | TXFWA003 | CO RE | City of Fort Worth | West Fk Trinity R BEL1 |

Table A1.Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| TX | TXFWA004 | ID | City of Fort Worth | Dry Branch CRA1 |
| TX | TXFWA005 | RE CO OP | City of Fort Worth | Estrn Hills HS EH1 |
| TX | TXGAA001 | RE CO ID OP | City of Garland | Mills Branch Tributary GM404 |
| TX | TXGAA002 | ID OP | City of Garland | Trib to Duck Creek Gl401 |
| TX | TXGAA003 | RE CO | City of Garland | Sleepy Hollow St GR402 |
| TX | TXGAA004 | CO RE | City of Garland | 1635 Outfall at CE GC603 |
| TX | TXHCA001 | RE | Harris County | Overbluff |
| TX | TXHCA002 | RE OP | Harris County | Cypress Trace Station |
| TX | TXHCA003 | CO | Harris County | Steeplechase |
| TX | TXHCA004 | ID | Harris County | Bayport |
| TX | TXHCA005 | CO | Harris County | WillowBrook Mall |
| TX | TXHCA006 | RE | Harris County | Little Cypress Creek |
| TX | TXHCA007 | OP | Harris County | Hadden Road |
| TX | TXHOA001 | OP | City of Houston | Briar Forest |
| TX | TXHOA002 | ID | City of Houston | Eleventh Street |
| TX | TXHOA003 | RE | City of Houston | Lazybrook |
| TX | TXHOA004 | CO | City of Houston | Memorial City Mall |
| TX | TXHOA005 | RE | City of Houston | Tanglewilde |
| TX | TXIRA001 | RE | City of Irving | Bear Cr IR501 |
| TX | TXIRA002 | ID RE OP | City of Irving | Cottonwood Branch Trib IM504 |
| TX | TXIRA003 | ID CO | City of Irving | Hereford Rd II503 |
| TX | TXIRA004 | ID CO | City of Irving | Trib to ELM FK 11502 |
| TX | TXMEA001 | CO FW | City of Mesquite | South mesquite I635 MC801 |
| TX | TXMEA002 | RE | City of Mesquite | South Mesquite South Parkway MC802 |
| TX | TXMEA003 | RE | City of Mesquite | South Mesquite Bruton Road MC803 |
| TX | TXPLA001 | RE CO | City of Plano | Rowlett Cr PR701 |
| TX | TXPLA002 | OP FW | City of Plano | Beck Brach PU704 |
| TX | TXPLA003 | CO OP | City of Plano | Spring Creek PC702 |
| TX | TXPLA004 | CO ID RE | City of Plano | Spring Creek PI703 |
| TX | TXTCA001 | FW OP ID | TXDOT Tarrant County | Deer Creek TH904 |
| VA | VAARLCV2 | RE | Arlington | Colonial Village CV2 |
| VA | VAARLLP1 | RE | Arlington | Little Pimmet LP1 |
| VA | VAARLRS3 | CO | Arlington | Randolph Street RS3 |
| VA | VAARLTC4 | ID | Arlington | Trades Center TC4 |
| VA | VACHCCC4 | CO | Chesterfield County | CoverLeaf Mall CC4 |
| VA | VACHCCC5 | RE | Chesterfield County | Buck Rub Drive CC5 |
| VA | VACHCN1A | RE | Chesterfield County | Gates bluff 1A |
| VA | VACHCN2A | RE | Chesterfield County | Helmsley road 2A |
| VA | VACHCOF1 | ID | Chesterfield County | unnamed OF1 |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| VA | VACHCOF2 | OP RE | Chesterfield County | Oak river drive OF2 |
| VA | VACHCOF3 | RE | Chesterfield County | Kings mill road OF3 |
| VA | VACHCOF4 | RE | Chesterfield County | OF4 |
| VA | VACHCOF5 | RE | Chesterfield County | Laurel oak road OF5 |
| VA | VACPTC1A | RE | Chesapeake | Briarfield Drive C1A |
| VA | VACPTSF2 | RE | Chesapeake | Woodards Mill SF2 |
| VA | VACPTYC1 | RE | Chesapeake | Etheridge rd Mt Pleasant Rd C1 |
| VA | VACPTYC2 | RE OP | Chesapeake | Hunningdon Lakes C2 |
| VA | VACPTYC3 | RE | Chesapeake | Horse Run Ditch C3 |
| VA | VACPTYC4 | CO | Chesapeake | Woodford Square Along Battlefield Blvd C4 |
| VA | VACPTYC5 | ID | Chesapeake | Cavalier Industrial Park C5 |
| VA | VACPTYO1 | ID | Chesapeake | Paramount Avenue O1 |
| VA | VAFFCOF1 | RE | Fairfax County | Apple Ridge Road |
| VA | VAFFCOF2 | RE CO ID | Fairfax County | Sunset Hills Road |
| VA | VAFFCOF3 | RE | Fairfax County | Onley Road |
| VA | VAFFCOF4 | CO | Fairfax County | Green Look Place |
| VA | VAFFCOF5 | RE | Fairfax County | Oakton Terrace Road |
| VA | VAFFCOF6 | CO | Fairfax County | Fairview Park Drive |
| VA | VAFFCOF7 | RE | Fairfax County | Lakeview Drive |
| VA | VAFFCOF8 | RE | Fairfax County | Pumphrey Drive |
| VA | VAFFCOF9 | RE | Fairfax County | Rock Ridge Road |
| VA | VAFFOF10 | ID | Fairfax County | Boston Boulevard |
| VA | VAFFOF11 | ID | Fairfax County | Prosperity Avenue |
| VA | VAHAHMS2 | ID | Hampton | Copeland Industrial Park HMS2 |
| VA | VAHAHMS5 | RE OP | Hampton | Grays Landing HMS5 |
| VA | VAHATYH1 | CO | Hampton | Commerce Drive H1 |
| VA | VAHATYH2 | ID | Hampton | Mingee Drive H2 |
| VA | VAHATYH3 | RE | Hampton | Hampton Club H3 |
| VA | VAHATYH4 | RE | Hampton | Bay Avenue H4 |
| VA | VAHATYH5 | RE | Hampton | Willow Oaks Boulevard H5 |
| VA | VAHCCOC1 | CO | Henrico County | Dickens Place C1 |
| VA | VAHCCOC2 | CO | Henrico County | Carousel Lane C2 |
| VA | VAHCCON1 | ID | Henrico County | Tomlyn Street N1 |
| VA | VAHCCON2 | ID | Henrico County | Impala Drive and Galaxy Road N2 |
| VA | VAHCCOR1 | RE | Henrico County | Prestwick Circle R1 |
| VA | VAHCCOR2 | RE | Henrico County | Westbury Drive R2 |
| VA | VANFTMS5 | CO | Norfolk | Village avenue MS5 |
| VA | VANFTMS6 | RE | Norfolk | Robin hood road MS6 |
| VA | VANFTMS8 | CO | Norfolk | North Hampton MS8 |
| VA | VANFTMS9 | CO | Norfolk | Bay side road MS9 |
| VA | VANFTYN1 | COID | Norfolk | Armistead Avenue N1 |

Table A1. Site Name and Land Use - Continued

| State | LOCATION ID | Land use | Jurisdiction | Site ID |
| :---: | :---: | :---: | :---: | :---: |
| VA | VANFTYN2 | RE | Norfolk | Modoc Avenue N2 |
| VA | VANFTYN3 | RE | Norfolk | Little creek road N3 |
| VA | VANFTYN4 | CO | Norfolk | Military circle N4 |
| VA | VANFTYN5 | RE | Norfolk | Sewel's point N5 |
| VA | VANNTMF1 | RE | Newport News | Marshall Avenue MF1 |
| VA | VANNTMF4 | RE | Newport News | Chesapeake Bay Apartments MF4 |
| VA | VANNTNN1 | RE | Newport News | Glendale Road NN1 |
| VA | VANNTNN2 | RE OP | Newport News | Shields Road NN2 |
| VA | VANNTNN3 | CO | Newport News | Patrick Henry Mall NN3 |
| VA | VANNTNN4 | CO ID | Newport News | Oyster Point Park Jefferson Ave NN4 |
| VA | VANNTNN5 | CO RE ID | Newport News | Oyster Point Park Thimble Shoals |
| VA | VANNTSF4 | RE | Newport News | Central Parkway SF4 |
| VA | VANNTSF6 | RE | Newport News | Jefferson Avenue SF6 |
| VA | VANNTYI2 | ID OP | Newport News | City Line Rd I2 |
| VA | VAPMTYP1 | CO | Portsmouth | Cradock Shopping center P1 |
| VA | VAPMTYP2 | RE | Portsmouth | West park homes P2 |
| VA | VAPMTYP3 | RE CO | Portsmouth | Church land shopping center P3 |
| VA | VAPMTYP4 | RE | Portsmouth | Edgefield apartmentsP4 |
| VA | VAPMTYP5 | RE | Portsmouth | South Hampton P5 |
| VA | VAVBTYA1 | OP ID | Virginia Beach | Morris Neck Road A1 |
| VA | VAVBTYI1 | ID | Virginia Beach | Airport Industrial Park I1 |
| VA | VAVBTYM2 | RE CO OP ID | Virginia Beach | Ketlam Road M2 |
| VA | VAVBTYR1 | RE | Virginia Beach | Homestead Drive R1 |
| VA | VAVBTYV1 | RE | Virginia Beach | Bow creek V1 |
| VA | VAVBTYV2 | RE | Virginia Beach | Salem Road V2 |
| VA | VAVBTYV3 | CO OP | Virginia Beach | Haygood V3 |
| VA | VAVBTYV4 | ID | Virginia Beach | Viking Drive V4 |
| VA | VAVBTYV5 | RE OP | Virginia Beach | Holland road V5 |

Table A2. Site Characteristics

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | \% Impervious | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ALHUCHIP | 19.5 | 3 |  |  | GS |  | 08/27/92 | 09/12/00 | 8 |
| ALHUDRAV | 20 | 3 |  |  | GS |  | 09/19/01 | 09/19/01 | 1 |
| ALHUHURI | 78.3 | 3 |  |  | CG |  | 08/27/92 | 09/19/01 | 9 |
| ALHUMASM | 87 | 3 |  |  | CG |  | 08/27/92 | 09/19/01 | 9 |
| ALHUWERP | 130 | 3 |  |  | CG |  | 08/27/92 | 09/19/01 | 9 |
| ALJC004L | 2564 | 3 |  |  | CG |  | 09/19/01 | 01/19/02 | 2 |
| ALJC004R | 1047 | 3 |  |  | CG |  | 09/20/01 | 01/19/02 | 2 |
| ALJCC001 | 336 | 3 |  |  | CG |  | 11/27/01 | 01/19/02 | 2 |
| ALJCC002 | 750 | 3 |  |  | CG |  | 11/27/01 | 03/20/02 | 2 |
| ALJCC009 | 112 | 3 |  |  | CG |  | 08/31/01 | 03/09/02 | 2 |
| ALJCC010 | 167 | 3 |  |  | CG |  | 08/31/01 | 03/09/02 | 2 |
| ALJCC012 | 244 | 3 |  |  | CG |  | 12/17/01 | 12/17/01 | 1 |
| ALMOCREO | 74 | 3 |  |  | GS |  | 02/10/93 | 04/15/93 | 3 |
| ALMODAPH | 14 | 3 |  |  |  |  | 02/16/93 | 04/20/93 | 3 |
| ALMOSARA | 64 | 3 |  |  | GS |  | 01/24/93 | 04/04/93 | 3 |
| ALMOSIIV | 450 | 3 |  |  |  |  | 02/11/93 | 04/15/93 | 3 |
| ALMOSITV | 304 | 3 |  |  | CG |  | 01/24/93 | 04/04/93 | 3 |
| ALMOSIVI | 194 | 3 |  |  | CG |  | 01/24/93 | 04/04/93 | 3 |
| ALMOTHEO | 27 | 3 |  |  | GS |  | 01/24/93 | 03/30/93 | 3 |
| AZMCA001 | 39 | 6 | 80 |  |  |  | 11/10/91 | 07/22/98 | 27 |
| AZMCA002 | 1120 | 6 | 1 |  | GS |  | 01/12/92 | 02/07/92 | 2 |
| AZMCA003 | 45 | 6 | 15 |  |  |  | 12/10/91 | 07/22/98 | 27 |
| AZMCA004 | 81 | 6 | 33 |  | CG |  | 10/27/91 | 08/22/92 | 6 |
| AZMCA005 | 3.4 | 6 | 94 |  | CG |  | 12/04/92 | 08/07/98 | 26 |
| AZMCA006 | 17.8 | 6 | 60 |  | CG |  | 03/07/94 | 09/11/98 | 20 |
| AZTUA001 | 103 | 6 |  |  | CG |  | 07/25/96 | 12/04/01 | 13 |
| AZTUA002 | 48.3 | 6 |  |  | CG |  | 08/26/96 | 12/04/01 | 12 |
| AZTUA003 | 29 | 6 |  |  | CG |  | 08/14/96 | 12/04/01 | 11 |
| AZTUA004 | 83 | 6 |  |  | CG |  | 09/24/96 | 12/11/01 | 11 |
| CAALAL03 | 168 | 6 |  |  | CG |  | 02/15/90 | 03/25/93 | 20 |
| CAALAL04 | 20 | 6 |  |  | CG |  | 03/02/90 | 02/27/91 | 5 |
| CAALAL07 | 78 | 6 |  |  | CG |  | 01/13/90 | 03/17/91 | 5 |
| CAALAL09 | 260 | 6 |  |  | CG |  | 01/13/90 | 03/17/91 | 9 |
| CAALAL10 | 144 | 6 |  |  | CG |  | 03/02/90 | 03/19/91 | 8 |
| CACTA001 | 0.69 | 6 | 95 |  | CG |  | 01/23/01 | 03/10/02 | 14 |
| CACTA002 | 1.61 | 6 | 100 |  | CG |  | 10/28/00 | 03/06/02 | 16 |
| CACTA003 | 1.85 | 6 | 70 |  | CG |  | 01/23/01 | 03/23/02 | 10 |
| CACTA004 | 0.44 | 6 | 70 |  | CG |  | 01/23/01 | 03/23/02 | 11 |
| CACTA005 | 0.99 | 6 | 100 |  | CG |  | 11/26/97 | 03/25/99 | 8 |
| CACTA006 | 0.99 | 6 | 100 |  | CG |  | 01/25/99 | 03/20/99 | 3 |
| CACTA007 | 0.99 | 6 | 100 |  | CG |  | 01/25/99 | 03/25/99 | 4 |

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | \% Impervious | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| CACTA008 | 3.16 | 6 | 80 |  | CG |  | 01/17/00 | 01/27/02 | 19 |
| CACTA009 | 4.18 | 6 | 80 |  | CG |  | 11/20/99 | 03/17/02 | 24 |
| CACTA010 | 0.96 | 6 | 80 |  | CG |  | 11/08/99 | 03/17/02 | 26 |
| CACTA011 | 0.4 | 6 | 100 |  | CG |  | 11/10/97 | 02/09/99 | 4 |
| CACTA012 | 0.99 | 6 | 100 |  | CG |  | 11/13/97 | 02/09/99 | 4 |
| CACTA013 | 1.41 | 6 | 100 |  | CG |  | 11/10/97 | 02/09/99 | 3 |
| CACTA014 | 0.99 | 6 | 100 |  | CG |  | 11/13/97 | 01/25/99 | 4 |
| CACTA015 | 0.99 | 6 | 100 |  | CG |  | 11/13/97 | 01/25/99 | 4 |
| COCSA001 | 80 | 9 | 58.1 |  | CG |  | 06/03/92 | 11/21/92 | 7 |
| COCSA002 | 105.6 | 9 | 37.5 |  | CG |  | 05/31/92 | 07/29/02 | 7 |
| COCSA003 | 110.72 | 9 | 55.9 |  | CG |  | 06/05/92 | 11/21/92 | 7 |
| COCSA004 | 209.28 | 9 | 34.2 |  | CG |  | 05/26/92 | 11/21/92 | 7 |
| COCSA005 | 31.36 | 9 | 40.1 |  | CG |  | 06/10/92 | 12/06/92 | 7 |
| CODEA001 | 150 | 9 | 83 |  | CG |  | 06/05/92 | 07/12/92 | 3 |
| CODEA002 | 55 | 9 | 83 |  | CG |  | 04/14/92 | 07/12/92 | 3 |
| CODEA003 | 269 | 9 | 20 |  | CG |  | 03/22/92 | 08/23/92 | 4 |
| CODEA004 | 498 | 9 | 85 |  | CG |  | 05/21/92 | 07/10/92 | 3 |
| CODEA005 | 495 | 9 | 44 |  | CG |  | 06/06/92 | 08/23/92 | 3 |
| CODEA006 | 636 | 9 | 85 |  | CG |  | 06/08/92 | 07/10/92 | 3 |
| CODEA007 | 56 | 9 | 85 |  | CG |  | 03/28/92 | 07/02/92 | 3 |
| CODEA008 | 146 | 9 | 83 |  | CG |  | 03/28/92 | 05/31/92 | 3 |
| GAATAT01 | 28 | 3 |  |  | CG |  | 03/08/95 | 02/16/97 | 10 |
| GAATAT02 | 95 | 3 |  |  | CG |  | 10/04/95 | 02/16/97 | 9 |
| GACLCOSI | 18 | 3 |  |  | CG |  | 11/29/95 | 03/19/00 | 20 |
| GACLCOTR | 125 | 3 |  |  | CG |  | 05/01/95 | 03/16/00 | 24 |
| GACOC1A2 | 63.6 | 3 |  |  |  |  | 01/27/96 | 02/28/00 | 17 |
| GACOC1A3 | 7590.4 | 3 |  |  |  |  | 08/24/00 | 03/19/01 | 6 |
| GACOCOL2 | 2947 | 3 |  |  |  |  | 01/19/95 | 03/12/01 | 22 |
| GADKCOTD | 115 | 3 |  |  | CG | WP | 12/13/93 | 06/06/00 | 25 |
| GAFUCOS1 | 10339 | 3 |  |  | GS |  | 11/10/94 | 04/25/01 | 22 |
| GAFUCOS2 | 3915 | 3 |  |  | CG |  | 10/30/94 | 04/25/01 | 19 |
| GAFUCOS3 | 6257 | 3 |  |  | GS |  | 01/06/95 | 04/25/01 | 22 |
| IDADA001 | 10.9 | 8 |  |  | CG |  | 08/11/99 | 04/19/01 | 7 |
| IDADA002 | 105 | 8 |  |  | CG |  | 04/29/99 | 04/11/01 | 7 |
| IDADA003 | 17 | 8 |  |  | CG |  | 04/29/99 | 07/30/01 | 9 |
| IDADA004 | 18 | 8 |  |  |  |  | 04/11/01 | 04/11/01 | 1 |
| KATOATWO | 38 | 4 | 55 |  | CG |  | 04/27/98 | 09/13/02 | 15 |
| KATOBROO | 18.5 | 4 | 25 |  | CG |  | 04/27/98 | 09/13/02 | 16 |
| KATOJACK | 218 | 4 | 65 |  | CG |  | 04/27/98 | 09/13/02 | 16 |
| KATOSTFE | 39.5 | 4 | 75 |  | CG |  | 04/27/98 | 08/16/02 | 17 |
| KAWIHUNT | 36 | 4 | 50 |  |  |  | 02/09/98 | 10/05/01 | 16 |

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | \% Impervious | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| KAWIMCLE | 30 | 4 | 65 |  |  |  | 02/09/98 | 10/05/01 | 16 |
| KAWISBWY | 250 | 4 | 60 |  |  |  | 02/09/98 | 10/05/01 | 16 |
| KAWITOWN | 40 | 4 | 90 |  |  |  | 02/09/98 | 10/05/01 | 16 |
| KYLOTSR1 | 96.64 | 2 | 39.6 |  | CG |  | 01/15/91 | 10/05/91 | 3 |
| KYLOTSR2 | 108.16 | 2 | 20.6 |  | GS |  | 02/05/91 | 10/05/91 | 3 |
| KYLOTSR3 | 134.4 | 2 | 35 |  | CG |  | 03/01/91 | 12/12/91 | 3 |
| KYLOTSR4 | 43.52 | 2 | 45.5 |  | GS |  | 03/12/91 | 04/15/92 | 4 |
| KYLOTSR5 | 84.48 | 2 | 68.9 |  | CG |  | 03/27/91 | 05/12/92 | 5 |
| KYLOTSR6 | 180.48 | 2 | 63.5 |  | CG |  | 04/04/91 | 12/12/91 | 3 |
| KYLXEHL4 | 13 | 2 | 10 |  | GS |  | 10/07/98 | 10/07/98 | 1 |
| KYLXEHL5 | 550 | 2 |  |  | GS |  | 07/30/98 | 07/30/98 | 1 |
| KYLXEHL6 | 1.3 | 2 |  |  | CG |  | 10/24/97 | 01/05/98 | 3 |
| KYLXEHL7 | 4.8 | 2 |  |  | CG |  | 10/24/97 | 01/05/98 | 3 |
| KYLXNEL1 | 32 | 2 |  |  | GS |  | 10/07/98 | 10/07/98 | 1 |
| KYLXNEL2 | 580 | 2 |  |  | CG | WP | 07/30/98 | 10/07/98 | 2 |
| KYLXNEL3 | 73 | 2 |  |  | CG |  | 06/03/92 | 09/27/96 | 12 |
| KYLXTBL1 | 71 | 2 |  |  | CG |  | 06/03/92 | 09/27/96 | 12 |
| KYLXTBL2 | 94 | 2 |  |  | GS |  | 06/30/92 | 09/27/96 | 12 |
| KYLXTBL3 | 205 | 2 |  |  | GS |  | 06/19/92 | 09/21/96 | 5 |
| KYLXWHL1 | 38 | 2 |  |  | CG |  | 06/03/92 | 09/27/96 | 13 |
| MABOA001 | 40.4 | 1 | 74 |  | CG |  | 04/11/92 | 08/14/92 | 5 |
| MABOA002 | 86.7 | 1 | 52 |  | CG |  | 04/17/92 | 06/24/92 | 3 |
| MABOA003 | 5 | 1 | 55 |  | CG |  | 04/11/92 | 06/24/92 | 3 |
| MABOA004 | 32 | 1 | 97 |  | CG |  | 04/11/92 | 06/24/92 | 3 |
| MABOA005 | 102.7 | 1 | 38 |  | CG |  | 04/17/92 | 06/24/92 | 3 |
| MABOA006 | 3.3 | 1 | 74 |  | CG |  | 06/02/01 | 07/17/01 | 3 |
| MABOA007 | 12.2 | 1 |  |  | GS |  | 09/25/01 | 09/25/01 | 1 |
| MDAACOMW | 5 | 2 | 94 | E | CG |  | 07/31/92 | 09/25/92 | 3 |
| MDAACOOD | 28 | 2 | 41 | E | CG |  | 08/11/92 | 10/09/92 | 3 |
| MDAACOPP | 25 | 2 | 85 | E | CG |  | 08/11/92 | 11/14/00 | 26 |
| MDAACORK | 12 | 2 | 41 | E | CG |  | 08/28/92 | 10/30/92 | 3 |
| MDAACOSC | 26 | 2 | 41 | E | CG |  | 08/11/92 | 10/09/92 | 3 |
| MDBACOBC | 25.3 | 2 | 60 |  |  |  | 12/15/93 | 03/08/94 | 3 |
| MDBACOLC | 225 | 2 | 70 |  |  |  | 10/20/93 | 01/15/98 | 19 |
| MDBACOSC | 83.5 | 2 | 30 |  |  |  | 12/15/93 | 06/19/98 | 26 |
| MDBACOTC | 144.06 | 2 |  |  |  |  | 01/12/94 | 04/07/94 | 3 |
| MDBACOWC | 73 | 2 | 7 |  |  |  | 01/12/94 | 03/21/94 | 3 |
| MDBCTYBO | 48.43 | 2 |  |  |  |  | 04/16/93 | 03/21/94 | 3 |
| MDBCTYFM | 45.96 | 2 |  |  |  |  | 05/30/92 | 11/04/93 | 3 |
| MDBCTYHA | 104.4 | 2 | 32 | E |  |  | 05/17/95 | 12/14/00 | 66 |
| MDBCTYHO | 354.09 | 2 |  |  |  |  | 06/05/92 | 09/25/92 | 3 |

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | $\%$ Impervious | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| MDBCTYHR | 38.8 | 2 | 54 |  |  |  | 05/30/92 | 10/20/93 | 3 |
| MDBCTYKO | 54.36 | 2 |  |  |  |  | 04/26/93 | 03/21/94 | 3 |
| MDCHCOIP | 11 | 2 |  |  | CG |  | 10/04/95 | 01/02/96 | 3 |
| MDCHCOPA | 10 | 2 |  |  | CG |  | 09/22/95 | 01/24/96 | 3 |
| MDCHCOPF | 50 | 2 |  |  | GS |  | 01/02/96 | 04/30/96 | 3 |
| MDCHCOTG | 10 | 2 |  |  | CG |  | 10/14/95 | 10/14/95 | 1 |
| MDCLCOBP | 15 | 2 |  |  |  |  | 07/21/94 | 12/10/94 | 3 |
| MDCLCOCE | 22.35 | 2 | 26 | E |  |  | 12/10/93 | 11/21/94 | 3 |
| MDCLCOJS | 20 | 2 | 91 | E |  |  | 12/21/93 | 11/01/94 | 3 |
| MDCLCOKW | 66.2 | 2 | 11 | E |  |  | 03/21/94 | 09/22/94 | 3 |
| MDCLCOSD | 36 | 2 | 49 | E |  |  | 02/21/94 | 11/10/94 | 3 |
| MDHACFBA |  | 2 |  |  |  |  | 11/05/93 | 11/10/94 | 3 |
| MDHACOBP | 69.7 | 2 | 16 | E |  | WP | 02/17/99 | 12/16/00 | 18 |
| MDHACOCF | 14.4 | 2 |  |  |  |  | 01/05/93 | 04/10/93 | 2 |
| MDHACOCS | 51 | 2 |  |  |  |  | 04/21/93 | 09/22/94 | 3 |
| MDHACOGR | 80 | 2 |  |  |  |  | 04/26/93 | 04/10/94 | 2 |
| MDHACOIC | 10 | 2 |  |  |  |  | 08/06/93 | 09/22/94 | 2 |
| MDHOCODC | 7.5 | 2 | 90 |  | CG | WP | 12/15/93 | 04/13/94 | 3 |
| MDHOCOFM | 3.5 | 2 | 77 |  | CG | GS | 12/15/93 | 11/01/94 | 3 |
| MDHOCOGM | 29.5 | 2 | 38 |  | CG | WP | 12/10/93 | 11/01/94 | 3 |
| MDHOCOMH | 19 | 2 | 65 |  | CG | WP | 12/10/93 | 04/13/94 | 3 |
| MDHOCOOC | 11.7 | 2 | 49 |  | CG | WP | 11/17/93 | 03/21/94 | 3 |
| MDMOCOBC | 14.2 | 2 | 83 | E | CG |  | 05/25/94 | 09/22/95 | 3 |
| MDMOCOCV | 11.5 | 2 | 55 | E |  | WP | 08/13/96 | 09/25/00 | 37 |
| MDMOCONV | 75.4 | 2 | 57 | E | CG |  | 05/04/94 | 03/08/95 | 3 |
| MDMOCOQA | 34.2 | 2 | 45 | E | CG |  | 05/04/94 | 10/27/95 | 3 |
| MDMOCOSL | 81 | 2 | 92 | E | CG | OT | 09/22/94 | 09/22/95 | 3 |
| MDMOCOWP | 70 | 2 | 96 | E | CG | OT | 05/25/94 | 10/27/95 | 3 |
| MDPGCOS1 | 19.7 | 2 | 47 | E | CG |  | 08/11/92 | 01/22/97 | 26 |
| MDPGCOS2 | 57.3 | 2 | 45 | E | CG |  | 08/11/92 | 09/25/00 | 63 |
| MDPGCOS3 | 34.4 | 2 | 96 | E | CG |  | 08/11/92 | 03/04/93 | 3 |
| MDPGCOS4 | 102.5 | 2 | 33 | E | CG |  | 08/11/92 | 03/04/93 | 3 |
| MDPGCOS5 | 41.3 | 2 | 83 | E | CG |  | 08/11/92 | 03/04/93 | 3 |
| MDPGCOS6 | 42.4 | 2 |  |  | GS |  | 10/23/94 | 08/20/97 | 28 |
| MDSHDTDV | 4 | 2 |  |  | CG |  | 06/14/99 | 06/06/00 | 8 |
| MDSHDTPS | 20 | 2 |  |  | GS |  | 02/11/98 | 06/21/00 | 13 |
| MNMISD01 | 143 | 1 |  |  | CG |  | 05/06/01 | 10/13/01 | 10 |
| MNMISD02 | 95 | 1 |  |  | CG |  | 05/06/01 | 10/13/01 | 9 |
| MNMISD03 | 80 | 1 |  |  | CG |  | 05/20/01 | 11/12/01 | 10 |
| MNMISD04 | 63 | 1 |  |  | CG |  | 05/06/01 | 11/12/01 | 9 |
| MNMISD05 | 100 | 1 |  |  | CG |  | 05/20/01 | 11/12/01 | 10 |

Table A2. Site Characteristics - Continued
$\left.\begin{array}{|c|c|c|c|c|c|c|c|c|c|}\hline \text { LOCATION ID } & \begin{array}{c}\text { Area } \\ \text { (acres) }\end{array} & \begin{array}{c}\text { EPA } \\ \text { Rain } \\ \text { Zone }\end{array} & \begin{array}{c}\text { Impervious }\end{array} & \mathbf{Q} & \begin{array}{c}\text { Convey } \\ \text { ance }\end{array} & \text { Control } & \begin{array}{c}\text { First } \\ \text { Sample }\end{array} & \begin{array}{c}\text { Last } \\ \text { Sample }\end{array} & \begin{array}{c}\text { Number of } \\ \text { Samples }\end{array} \\ \hline \text { NCCHBREV } & 15.1 & 2 & 75 & & \text { CG } & & 05 / 13 / 92 & 03 / 03 / 93 & 4 \\ \hline \text { NCCHHIDD } & 20 & 2 & 30 & & \text { CG } & & 05 / 13 / 92 & 12 / 10 / 93 & 5 \\ \hline \text { NCCHHOSK } & 17.4 & 2 & 71.83 & & & & 05 / 13 / 92 & 03 / 03 / 93 & 4 \\ \hline \text { NCCHNANC } & 10.9 & 2 & 20 & & \text { GS } & & 08 / 27 / 92 & 02 / 21 / 94 & 4 \\ \hline \text { NCCHROSE } & 78.87 & 2 & 42.55 & & \text { CG } & & 05 / 13 / 92 & 08 / 12 / 92 & 3 \\ \hline \text { NCCHSHEF } & 42.5 & 2 & 20.68 & & & & 06 / 04 / 92 & 02 / 21 / 94 & 3 \\ \hline \text { NCCHSIMS } & 6.8 & 2 & 50 & & \text { CG } & & 05 / 13 / 92 & 03 / 03 / 93 & 4 \\ \hline \text { NCCHSTAR } & 14.1 & 2 & 70 & & \text { GS } & & 05 / 13 / 92 & 03 / 03 / 93 & 4 \\ \hline \text { NCCHYARD } & 88.6 & 2 & 68.21 & & \text { CG } & & 05 / 13 / 92 & 03 / 03 / 93 & 4 \\ \hline \text { NCFV71ST } & 36 & 2 & 45 & & \text { CG } & & 01 / 21 / 93 & 06 / 15 / 99 & 18 \\ \hline \text { NCFVCLEA } & 12 & 2 & 20 & & & & 01 / 04 / 93 & 04 / 01 / 99 & 14 \\ \hline \text { NCFVELMS } & 40 & 2 & 90 & & \text { CG } & & 01 / 04 / 93 & 04 / 01 / 99 & 18 \\ \hline \text { NCFVROSE } & 39.27 & 2 & 50 & & \text { CG } & & 12 / 17 / 92 & 06 / 15 / 99 & 14 \\ \hline \text { NCFVSTRK } & 85 & 2 & 1 & & & & 02 / 07 / 93 & 06 / 16 / 96 & 6 \\ \hline \text { NCFVTRYO } & 25 & 2 & 50 & & & & 01 / 21 / 93 & 04 / 01 / 99 & 18 \\ \hline \text { NCFVWINS } & 12 & 2 & 75 & & & & & 01 / 04 / 93 & 06 / 15 / 99\end{array}\right] 189$

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | \% Impervious | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ORGRA004 | 64 | 7 |  |  | CG |  | 03/02/93 | 04/11/96 | 6 |
| ORODA001 | 22.4 | 7 |  |  | CG |  | 12/04/95 | 12/04/95 | 1 |
| ORODA002 | 1.2 | 7 |  |  | CG |  | 10/02/95 | 10/02/95 | 1 |
| ORODA003 | 18.2 | 7 |  |  | CG |  | 01/07/95 | 01/18/96 | 5 |
| ORODA004 | 3.6 | 7 |  |  | CG |  | 10/12/95 | 10/12/95 | 1 |
| ORODA005 | 23.1 | 7 |  |  | CG |  | 01/07/95 | 01/18/96 | 6 |
| ORPOA001 | 35 | 7 |  |  | CG |  | 05/07/91 | 10/25/95 | 13 |
| ORPOA002 | 75 | 7 |  |  | CG |  | 05/07/91 | 03/03/96 | 16 |
| ORPOA003 | 46 | 7 |  |  | CG |  | 05/07/91 | 03/03/96 | 14 |
| ORPOA004 | 49 | 7 |  |  | CG |  | 08/09/91 | 10/25/95 | 13 |
| ORPOA005 | 91 | 7 |  |  | CG |  | 05/07/91 | 03/03/96 | 14 |
| ORPOA006 | 85 | 7 |  |  | CG |  | 05/07/91 | 03/03/96 | 13 |
| ORPOA007 | 10 | 7 |  |  | CG |  | 05/07/91 | 03/03/96 | 14 |
| ORSAA001 | 31 | 7 |  |  | CG |  | 01/07/95 | 01/14/96 | 6 |
| ORSAA002 | 40 | 7 |  |  | CG |  | 01/07/95 | 01/14/96 | 6 |
| ORSAA003 | 35 | 7 |  |  | CG |  | 01/07/95 | 01/14/96 | 6 |
| ORSAA004 | 72 | 7 |  |  | CG |  | 01/07/95 | 01/14/96 | 6 |
| PAPH0864 | 22 | 2 | 84 | E | CG |  | 09/10/92 | 09/25/92 | 2 |
| PAPH0891 | 35 | 2 | 83 | E | CG |  | 09/22/92 | 10/09/92 | 2 |
| PAPH1014 | 22 | 2 | 82 | E | CG |  | 09/10/92 | 10/09/92 | 3 |
| PAPH1051 | 223 | 2 | 87 | E | CG |  | 09/22/92 | 10/09/92 | 2 |
| PAPH1182 | 31 | 2 | 57 | E | CG |  | 09/10/92 | 10/09/92 | 3 |
| TNKXTYAP | 582.4 | 2 | 44 |  | GS | WP | 03/27/91 | 06/30/01 | 63 |
| TNKXTYFC | 2880 | 2 | 40 |  |  | WP | 03/06/92 | 06/07/01 | 47 |
| TNKXTYGV | 224 | 2 | 37 |  | GS |  | 08/14/91 | 08/25/99 | 39 |
| TNKXTYTC | 352 | 2 | 34 |  |  |  | 02/13/92 | 04/11/00 | 54 |
| TNKXTYWE | 364.8 | 2 | 60 |  |  |  | 04/08/91 | 05/03/00 | 51 |
| TNMET207 | 157 | 2 |  |  | GS | WP | 06/21/00 | 04/23/01 | 5 |
| TNMET211 | 45 | 2 |  |  | CG |  | 01/11/00 | 04/23/01 | 4 |
| TNMET231 | 26 | 2 |  |  | CG |  | 01/11/00 | 04/23/01 | 4 |
| TNMET260 | 294 | 2 |  |  | CG |  | 07/20/00 | 05/17/01 | 4 |
| TNMET410 | 154 | 2 |  |  | CG |  | 06/21/00 | 04/23/01 | 4 |
| TXARA001 | 38.8 | 5 | 76.2 |  | CG |  | 10/28/92 | 03/08/01 | 22 |
| TXARA002 | 160.6 | 5 | 47.4 |  | CG |  | 10/28/92 | 03/08/01 | 21 |
| TXARA003 | 77 | 5 | 89 |  | CG |  | 10/29/92 | 04/14/93 | 7 |
| TXARA004 | 85.5 | 5 | 80.9 |  |  | WP | 12/09/92 | 03/28/93 | 7 |
| TXDAA001 | 9 | 5 | 80 |  | CG |  | 03/03/92 | 09/21/92 | 7 |
| TXDAA002 | 49.5 | 5 | 80 |  | CG |  | 03/03/92 | 03/08/01 | 19 |
| TXDAA003 | 486.7 | 5 |  |  | CG |  | 12/02/97 | 05/04/01 | 21 |
| TXDAA004 | 59.1 | 5 | 84.5 |  | CG |  | 02/22/92 | 04/11/01 | 20 |
| TXDAA005 | 71.3 | 5 | 50 |  | CG |  | 02/12/92 | 09/21/92 | 7 |

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | \% Impervious | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TXDAA006 | 38.9 | 5 | 44.9 |  | CG |  | 02/22/92 | 03/24/01 | 20 |
| TXDCA001 | 115.36 | 5 | 10 |  |  |  | 09/03/97 | 03/08/01 | 17 |
| TXDCA002 | 12.05 | 5 | 33 |  |  |  | 01/11/98 | 03/08/01 | 9 |
| TXFWA001 | 61.7 | 5 | 21.9 |  |  |  | 02/22/92 | 08/12/92 | 7 |
| TXFWA002 | 151.6 | 5 | 27.7 | * I | CG |  | 02/03/92 | 03/08/01 | 21 |
| TXFWA003 | 136 | 5 | 66.5 |  | CG |  | 03/09/92 | 10/28/92 | 7 |
| TXFWA004 | 73.7 | 5 | 79.3 |  |  |  | 03/24/92 | 03/08/01 | 21 |
| TXFWA005 | 150.8 | 5 | 61.4 |  |  |  | 04/17/92 | 03/24/01 | 23 |
| TXGAA001 | 268 | 5 |  |  | CG |  | 12/02/97 | 03/27/01 | 23 |
| TXGAA002 | 33.9 | 5 | 67.3 |  |  |  | 06/20/92 | 03/27/01 | 22 |
| TXGAA003 | 67.3 | 5 | 55.4 |  |  |  | 09/01/92 | 01/23/93 | 7 |
| TXGAA004 | 36.2 | 5 | 84.6 |  | CG |  | 09/01/92 | 01/23/93 | 7 |
| TXHCA001 | 560 | 4 |  |  | GS |  | 01/29/99 | 03/27/01 | 8 |
| TXHCA002 | 95 | 4 | 65 |  |  |  | 07/20/92 | 04/16/01 | 14 |
| TXHCA003 | 32 | 4 |  |  | CG |  | 02/11/99 | 06/22/01 | 8 |
| TXHCA004 | 99 | 4 | 71.25 |  |  |  | 04/07/93 | 03/08/01 | 14 |
| TXHCA005 | 81 | 4 | 95 |  | CG |  | 06/30/92 | 04/07/93 | 6 |
| TXHCA006 | 401 | 4 | 45 |  |  |  | 06/30/92 | 04/29/93 | 6 |
| TXHCA007 | 872 | 4 |  |  |  | WP | 07/15/92 | 11/09/93 | 6 |
| TXHOA001 | 44 | 4 | 5.7 |  | GS |  | 06/30/92 | 11/19/92 | 7 |
| TXHOA002 | 232 | 4 | 76.5 |  |  |  | 06/22/92 | 05/31/01 | 16 |
| TXHOA003 | 65 | 4 | 45 |  | GS |  | 06/22/92 | 03/27/01 | 14 |
| TXHOA004 | 24 | 4 | 98 |  | CG |  | 07/19/92 | 11/22/99 | 12 |
| TXHOA005 | 38 | 4 | 65 |  | CG |  | 06/22/92 | 07/19/01 | 16 |
| TXIRA001 | 65.3 | 5 | 41.9 |  | CG |  | 09/03/92 | 03/24/01 | 22 |
| TXIRA002 | 127.7 | 5 |  |  | CG |  | 03/18/99 | 05/28/01 | 22 |
| TXIRA003 | 43.4 | 5 | 77.3 |  |  | WP | 08/24/92 | 01/09/93 | 7 |
| TXIRA004 | 43.9 | 5 | 77.8 |  |  |  | 09/21/92 | 01/28/93 | 7 |
| TXMEA001 | 45.9 | 5 | 89.4 |  | CG |  | 02/24/93 | 03/24/01 | 22 |
| TXMEA002 | 45.4 | 5 | 49.8 |  | CG |  | 03/11/93 | 06/25/93 | 7 |
| TXMEA003 | 46.2 | 5 | 49.9 |  | CG | WP | 02/10/93 | 05/23/93 | 7 |
| TXPLA001 | 51.4 | 5 | 54.3 |  | CG |  | 12/09/92 | 04/14/93 | 7 |
| TXPLA002 | 73.5 | 5 |  |  |  |  | 11/09/98 | 04/11/01 | 22 |
| TXPLA003 | 22.7 | 5 | 73.5 |  | CG |  | 12/09/92 | 05/04/01 | 25 |
| TXPLA004 | 49 | 5 | 81.6 |  |  |  | 01/09/93 | 06/09/93 | 7 |
| TXTCA001 | 63.13 | 5 | 27 |  |  |  | 02/06/97 | 03/24/01 | 15 |
| VAARLCV2 | 24.7 | 2 | 35 |  |  | DS | 02/11/98 | 01/19/01 | 9 |
| VAARLLP1 | 38.7 | 2 | 35 |  |  |  | 10/20/99 | 03/04/01 | 8 |
| VAARLRS3 | 14 | 2 | 74 |  |  | DS | 09/21/99 | 01/19/01 | 8 |
| VAARLTC4 | 36 | 2 | 39 |  |  |  | 02/03/98 | 06/01/01 | 13 |
| VACHCCC4 | 60 | 2 | 80 |  | CG |  | 08/12/96 | 12/10/01 | 13 |

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | $\begin{gathered} \text { \% } \\ \text { Impervious } \end{gathered}$ | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VACHCCC5 | 10 | 2 | 50 |  | CG |  | 08/12/96 | 12/10/01 | 13 |
| VACHCN1A | 10 | 2 | 10 |  | CG |  | 08/19/99 | 01/08/01 | 4 |
| VACHCN2A | 60 | 2 | 20 |  | CG |  | 08/19/99 | 01/08/01 | 4 |
| VACHCOF1 | 22.5 | 2 |  |  |  |  | 04/16/93 | 10/26/93 | 3 |
| VACHCOF2 | 19.05 | 2 | 10 |  | CG |  | 04/16/93 | 10/08/98 | 8 |
| VACHCOF3 | 13.5 | 2 | 20 |  | CG |  | 04/16/93 | 02/01/99 | 11 |
| VACHCOF4 | 38.5 | 2 |  |  |  |  | 04/16/93 | 12/15/93 | 3 |
| VACHCOF5 | 55.6 | 2 | 50 |  | CG |  | 04/16/93 | 12/10/01 | 16 |
| VACPTC1A | 130 | 2 | 25 | E | CG GS |  | 11/01/97 | 06/15/99 | 8 |
| VACPTSF2 | 91 | 2 | 10 | E | GS |  | 04/16/93 | 10/26/93 | 3 |
| VACPTYC1 | 57 | 2 | 25 | E | CG |  | 02/26/93 | 12/05/96 | 7 |
| VACPTYC2 | 188 | 2 | 25 | E | CG | WP | 02/26/93 | 01/24/99 | 15 |
| VACPTYC3 | 32 | 2 | 50 | E | CG |  | 02/26/93 | 02/18/99 | 15 |
| VACPTYC4 | 28 | 2 | 85 | E | CG |  | 04/16/93 | 02/02/99 | 14 |
| VACPTYC5 | 16 | 2 | 57 | E | CG |  | 03/27/93 | 01/15/99 | 15 |
| VACPTYO1 | 14 | 2 | 50 | E | CG |  | 02/26/93 | 04/16/93 | 3 |
| VAFFCOF1 | 32.3 | 2 |  |  | CG |  | 03/18/92 | 08/11/92 | 3 |
| VAFFCOF2 | 20.1 | 2 | 50 | E | CG | DP | 07/03/92 | 08/01/00 | 14 |
| VAFFCOF3 | 63.9 | 2 |  |  | CG |  | 06/18/92 | 09/02/92 | 3 |
| VAFFCOF4 | 108.8 | 2 | 70 | E | CG | WP | 04/21/92 | 09/03/00 | 13 |
| VAFFCOF5 | 39.7 | 2 |  |  | CG | DP | 04/16/92 | 09/22/92 | 3 |
| VAFFCOF6 | 213.4 | 2 | 21 | E | CG | WP | 07/12/92 | 11/10/00 | 14 |
| VAFFCOF7 | 49.9 | 2 | 25 | E | CG |  | 06/24/92 | 11/29/00 | 15 |
| VAFFCOF8 | 57.5 | 2 |  |  | CG |  | 04/21/92 | 09/02/92 | 3 |
| VAFFCOF9 | 63.8 | 2 | 50 | E | CG | WP | 07/21/92 | 09/02/00 | 13 |
| VAFFOF10 | 82 | 2 |  |  | CG |  | 04/21/92 | 08/11/92 | 3 |
| VAFFOF11 | 37.9 | 2 | 66 | E | CG |  | 06/26/97 | 11/29/00 | 11 |
| VAHAHMS2 | 793 | 2 | 67 | E | CG |  | 11/26/92 | 01/21/93 | 3 |
| VAHAHMS5 | 53 | 2 | 28 | E |  |  | 11/12/92 | 02/12/93 | 3 |
| VAHATYH1 | 115 | 2 | 80 | E | CG |  | 11/12/92 | 05/14/99 | 18 |
| VAHATYH2 | 47 | 2 | 70 | E | CG |  | 11/26/92 | 04/24/99 | 19 |
| VAHATYH3 | 18 | 2 | 40 | E | CG |  | 11/12/92 | 06/20/99 | 17 |
| VAHATYH4 | 134 | 2 | 25 | E | CG |  | 11/12/92 | 04/24/99 | 17 |
| VAHATYH5 | 35 | 2 | 25 | E | CG |  | 11/12/92 | 04/24/99 | 17 |
| VAHCCOC1 | 65 | 2 | 89 | E | CG |  | 11/13/92 | 12/20/92 | 2 |
| VAHCCOC2 | 70 | 2 | 87 | E | CG |  | 10/30/92 | 01/05/93 | 3 |
| VAHCCON1 | 75 | 2 | 89 | E | CG |  | 12/18/92 | 01/22/93 | 2 |
| VAHCCON2 | 23 | 2 | 89 | E | CG |  | 11/22/92 | 01/22/93 | 3 |
| VAHCCOR1 | 40 | 2 | 61 | E | CG |  | 11/03/92 | 01/05/93 | 3 |
| VAHCCOR2 | 70 | 2 | 57 | E | CG |  | 11/03/92 | 01/05/93 | 3 |
| VANFTMS5 | 56 | 2 |  |  | CG |  | 04/22/92 | 07/27/92 | 3 |

Table A2. Site Characteristics - Continued

| LOCATION ID | Area (acres) | EPA <br> Rain <br> Zone | $\begin{gathered} \text { \% } \\ \text { Impervious } \end{gathered}$ | Q | Convey ance | Control | First Sample | Last Sample | Number of Samples |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| VANFTMS6 | 68 | 2 |  |  |  |  | 05/05/92 | 07/27/92 | 3 |
| VANFTMS8 | 65 | 2 |  |  |  |  | 04/22/92 | 07/27/92 | 3 |
| VANFTMS9 | 40 | 2 |  |  |  |  | 05/05/92 | 08/27/92 | 3 |
| VANFTYN1 | 43 | 2 | 47 |  | CG |  | 04/22/92 | 02/12/00 | 28 |
| VANFTYN2 | 97 | 2 | 25 |  | CG |  | 05/30/92 | 12/14/99 | 30 |
| VANFTYN3 | 27 | 2 | 37 |  | CG |  | 04/22/92 | 12/14/99 | 28 |
| VANFTYN4 | 43 | 2 | 70 |  | CG |  | 04/22/92 | 12/14/99 | 28 |
| VANFTYN5 | 39 | 2 | 25 |  | CG |  | 06/09/92 | 02/18/00 | 28 |
| VANNTMF1 | 39 | 2 | 50 |  | CG |  | 10/04/92 | 01/21/93 | 3 |
| VANNTMF4 | 12 | 2 | 73 |  | CG |  | 12/28/92 | 03/13/93 | 3 |
| VANNTNN1 | 75 | 2 | 40 |  | OT |  | 10/31/92 | 04/02/99 | 12 |
| VANNTNN2 | 397 | 2 | 24 |  | CG | DP | 12/10/92 | 04/16/99 | 15 |
| VANNTNN3 | 24 | 2 | 85 |  | CG | WP | 10/04/92 | 04/16/99 | 15 |
| VANNTNN4 | 294 | 2 | 58 |  | CG | WP | 10/04/92 | 04/16/99 | 16 |
| VANNTNN5 | 83 | 2 | 62 |  | OT |  | 12/28/92 | 04/02/99 | 11 |
| VANNTSF4 | 111 | 2 | 30 |  | GS |  | 12/10/92 | 02/26/93 | 3 |
| VANNTSF6 | 207 | 2 | 37 |  | GS |  | 10/04/92 | 03/03/93 | 4 |
| VANNTYI2 | 49 | 2 | 73 |  | GS |  | 10/04/92 | 01/21/93 | 3 |
| VAPMTYP1 | 27.2 | 2 | 68 |  | CG |  | 01/16/93 | 05/14/99 | 18 |
| VAPMTYP2 | 101.1 | 2 | 36 |  | CG |  | 02/26/93 | 06/20/99 | 17 |
| VAPMTYP3 | 46 | 2 |  |  | CG |  | 01/16/93 | 05/14/99 | 17 |
| VAPMTYP4 | 35.3 | 2 | 39 |  | CG |  | 12/20/92 | 06/20/99 | 17 |
| VAPMTYP5 | 53.5 | 2 | 14 | E | CG |  | 12/20/92 | 05/23/99 | 17 |
| VAVBTYA1 | 225 | 2 | 7 |  |  |  | 07/01/92 | 10/30/92 | 5 |
| VAVBTYI1 | 8 | 2 | 90 |  |  |  | 06/09/92 | 10/04/92 | 3 |
| VAVBTYM2 | 310 | 2 | 35 |  |  |  | 10/04/90 | 10/30/92 | 4 |
| VAVBTYR1 | 49 | 2 | 25 |  |  |  | 05/07/92 | 09/19/92 | 5 |
| VAVBTYV1 | 63 | 2 | 29 |  | OT |  | 03/26/92 | 02/28/99 | 27 |
| VAVBTYV2 | 260 | 2 | 29 |  | OT | WP | 05/07/92 | 02/18/99 | 30 |
| VAVBTYV3 | 25 | 2 | 25 |  | CG |  | 04/12/92 | 02/28/99 | 33 |
| VAVBTYV4 | 29 | 2 | 55 |  | CG |  | 04/12/92 | 03/14/99 | 30 |
| VAVBTYV5 | 882 | 2 | 47 |  | OT | WP | 05/07/92 | 03/14/99 | 28 |

## Appendix B: Modified Values in the Database

## Description

The following table indicates the values that were modified in the database. The column "Order" corresponds to the row number in the table. The column "Problem" indicates the reason why the value was deleted or modified. In the case that the information available can solve the problem, the action was described in the column "action". The last column indicates the community where the event was located.

Table B1. Modified Values in the NSQD

| Order | Constituent | Original <br> value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1373 | TSS | 10100 | High | Delete | Acker Place |
| 890 | TSS | 53000 | High | Delete | Philadelphia |
| 1909 | TP | 80.1 | High - Ortho very Low | Delete | Louisville |
| 1707 | TP | 35 | High | Delete | Lexington |
| 1629 | TP | 15.4 | High | Delete |  |
| 1907 | Ortho P | 60.1 | High | Delete | Louisville |
| 3135 | Dis Zn / Tot Zn |  | High Ratio | Delete Total | Boston |
| 3118 | TDS | 17900 | High TDS | Deleted | Boston |
| 2893 | Dis Cu / Tot Cu |  | High Ratio | Delete Dissolved <br> Copper | Portland |
| 2883 | Dis Cu / Tot Cu |  | High Ratio | Delete Dissolved <br> Copper | Ada County |
| 561 | Dis P / Tot P |  | Wrong Dissolved <br> Values | corrected | Portsmouth |
| 562 | Dis P / Tot P |  | Wrong Dissolved <br> Values | corrected | Portsmouth |
| 563 | Dis P / Tot P |  | Wrong Dissolved <br> Values | corrected | Portsmouth |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 221 | Dis $\mathrm{P} /$ Tot P |  | Wrong Dissolved Values | Corrected | Hampton |
| 223 | Dis $\mathrm{P} / \mathrm{Tot} \mathrm{P}$ |  | Wrong Dissolved Values | Corrected | Hampton |
| 1707 | Dis $\mathrm{P} /$ Tot P |  | High Values | Deleted | Lexington |
| 1999 | Dis $\mathrm{P} /$ Tot P |  | High Dissolved Value | Deleted | Cobb |
| 2301 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Fayetteville |
| 2268 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Fayetteville |
| 2293 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Fayetteville |
| 4315 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Raleigh |
| 4306 | Dis P/Tot P |  | Low Total | Delete both | Raleigh |
| 4351 | Dis P/Tot P |  | Low Total | Delete both | Raleigh |
| 4342 | Dis P/Tot P |  | Low Total | Delete both | Raleigh |
| 4055 | Dis P/Tot P |  | Low Total | Delete both | Greensboro |
| 4070 | Dis P/Tot P |  | Inverted | Corrected | Greensboro |
| 4197 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 4249 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 4085 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 4217 | Dis P/Tot P |  | Low Total | Delete both | Greensboro |
| 4068 | Dis $\mathrm{P} /$ Tot P |  | Wrong values | Corrected | Greensboro |
| 4038 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 4134 | Dis P/Tot P |  | Low Total | Delete both | Greensboro |
| 4233 | Dis P/Tot P |  | Low Total | Delete both | Greensboro |
| 4149 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 3698 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 4024 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Greensboro |
| 2150 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Fulton |
| 1449 | Dis P/Tot P |  | Low Total | Delete both | Knoxville |
| 1617 | Dis P/Tot P |  | Low Total | Delete both | Knoxville |
| 1596 | Dis P/Tot P |  | Low Total | Delete both | Knoxville |
| 1616 | Dis $\mathrm{P} /$ Tot P |  | Low Total | Delete both | Knoxville |
| 1460 | Dis P/Tot P |  | Low Total | Delete both | Knoxville |
| 1707 | TKN | 290 | High | Deleted | Lexington |
| 1000 | TKN | 250 | High | Deleted | Baltimore City |
| 4149 | TKN | 147 | High | Deleted | Greensboro |
| 2699 | TKN | 120 | High | Deleted | Maricopa |
| 3136 | NO2 NO3 | 1690 | High | Deleted | Atlanta |
| 3281 | NO2 NO3 | 50 | High | Deleted | Houston 06/30/92 |
| 3331 | NO2 NO3 | 48 | High | Deleted | Houston 06/30/92 |
| 3289 | NO2 NO3 | 32.1 | High | Deleted | Houston 06/30/92 |
| 3305 | NO2 NO3 | 28 | High | Deleted | Houston 06/30/92 |
| 48 | COD BOD | < 5 | COD low | Deleted | $\begin{gathered} \hline \text { Chesterfield } \\ 02 / 03 / 98 \\ \hline \end{gathered}$ |
| 737 | COD | 5050 |  | Deleted | Bow creek V1 |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 97 | BOD |  | Weird Value <30 | Deleted | Chesterfield 09/27/99 |
| 1895 | BOD | 610 |  | Deleted | $\begin{gathered} \text { Pleasure_Ridge_P } \\ \text { ark } \end{gathered}$ |
| 2676 | COD BOD | 4300 |  | Deleted | Maricopa |
| 3299 | COD BOD | 1260 |  | Retyped | Houston |
| 4087 | BOD | 545 |  | Deleted | Husbands_Street |
| 4343 | COD BOD |  |  | Deleted | Williamson_Drive_ Wade_Avenue |
| 4399 | COD BOD |  |  | Deleted | 50 ft east N West Street_Peace_Stre et_Dortch_Street |
| 96 | Cadmium |  | Weird Values Dissolved Higher | Deleted | $\begin{gathered} \text { Chesterfield } \\ 08 / 19 / 99 \\ \hline \end{gathered}$ |
| 97 | Cooper |  | Weird Values Dissolved Higher | Deleted | $\begin{gathered} \text { Chesterfield } \\ 09 / 27 / 99 \\ \hline \end{gathered}$ |
| 85 | Cooper |  | Weird Values Dissolved Higher | Deleted | $\begin{gathered} \text { Chesterfield } \\ 09 / 27 / 99 \\ \hline \end{gathered}$ |
| 110 | Lead |  | Weird Values Dissolved Higher | Deleted | $\begin{aligned} & \text { Chesterfield } \\ & 09 / 27 / 99 \\ & \hline \end{aligned}$ |
| 76 | Zinc |  | Weird Values Dissolved Higher | Deleted | $\begin{gathered} \text { Chesterfield } \\ 09 / 27 / 99 \\ \hline \end{gathered}$ |
| 110 | Zinc |  | Weird Values Dissolved Higher | Deleted | $\begin{aligned} & \text { Chesterfield } \\ & 09 / 27 / 99 \\ & \hline \end{aligned}$ |
| 3288 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3304 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3330 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3281 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3289 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3305 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3331 | Antimony |  | Elevated values for the same set of samples | Deleted | Houston |
| 3276 | Oil \& Grease |  | Detection Limit is different | Replace as a detected value | Harris County |
| 2836 | Conductivity |  | Elevated Value. Two samples | Use mean value | Colorado Springs |
| 4077 | Turbidity |  | NT in cell | Move to qualifier | Greensboro |
| 446 | TDS <46 |  | Wrong Qualifier | Delete Qualifier.. | Norfolk N2 |
| 2128 | TDS | 0.065 | Factor of a thousand. | Change value from 0.065 to 65 | Fulton County |
| 2257 | TDS |  | Wrong value | corrected ( $32 \mathrm{mg} / \mathrm{L}$ ) | Fayetteville |
| 3136 | TDS |  | Value not clear in hardcopy | Delete value <31 $\mathrm{mg} / \mathrm{L}$ | Atlanta |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2649 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2658 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2660 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2664 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2665 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2672 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2686 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2699 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2704 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2706 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2709 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2713 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2714 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2715 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2716 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2717 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2719 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2720 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2721 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2725 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2726 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2729 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2730 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2734 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2740 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2742 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2750 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 2751 | TDS |  | TDS>TS | Delete pair | Maricopa |
| 3321 | TSS |  | Wrong Detection Limit | Change for value | Houston |
| 3154 | TSS |  | Wrong Detection Limit | Change for <1 | Atlanta |
| 1117 | TSS |  | Wrong Detection Limit | Deleted | Montgomery County |
| 2543 | TSS |  | First Flush was compared with TDS | Change for 160 | Los Angeles |
| 32 | COD |  | Weird Detection Limit | Delete Qualifier.. | Arlington |
| 5 | COD |  | Weird Detection Limit | Delete Qualifier. Quantification limit $=5 \mathrm{mg} / \mathrm{L}$ | Arlington |
| 16 | COD |  | Weird Detection Limit | Delete Qualifier. Quantification limit $=5 \mathrm{mg} / \mathrm{L}$ | Arlington |
| 859 | Fecal Streptococcus |  | Atypical Growth | Delete value >6000 | Fairfax |
| 1401 | Ammonia |  | Typo in detection limit | $\begin{gathered} \text { Change }<2 \text { to }<0.2 \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | Knoxville |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2299 | Ammonia |  | Detection Limit is different | Change for $<0.05$ (other sites) | Fayetteville |
| 2886 | Ammonia |  | Different Detection Limit | $\begin{gathered} \text { Change }<0.14 \text { to } \\ <0.20 \end{gathered}$ | Portland |
| 2899 | Ammonia |  | Different Detection Limit | $\begin{gathered} \text { Change }<0.14 \text { to } \\ <0.20 \end{gathered}$ | Portland |
| 2915 | Ammonia |  | Different Detection Limit | $\begin{gathered} \text { Change }<0.14 \text { to } \\ <0.20 \end{gathered}$ | Portland |
| 2942 | Ammonia |  | Different Detection Limit | $\begin{gathered} \hline \text { Change }<0.14 \text { to } \\ <0.20 \end{gathered}$ | Portland |
| 2956 | Ammonia |  | Different Detection Limit | $\begin{gathered} \hline \text { Change }<0.14 \text { to } \\ <0.20 \end{gathered}$ | Portland |
| 1999 | TKN |  | Detection Limit is Weird | Delete <0.6 | Cobb County |
| 2097 | TKN |  | Typo in Detection Limit | $\begin{aligned} & \text { Delete }<0.0 \text { for } \\ & <0.2 \end{aligned}$ | Fulton County |
| 2257 | TKN |  | Rows seem to be wrong | Change 46 for 1.2 | Fayetteville |
| 2257 | TDS |  | Rows seem to be wrong | Change 1.2 for 46 | Fayetteville |
| 2300 | Nitrogen Total Organic |  | Value $=0, \mathrm{Grab}=1.46$ | Delete value | Fayetteville |
| 2336 | Nitrogen Total Organic |  | Value $=0, \mathrm{Grab}=1.14$ | Delete value | Fayetteville |
| 13 | Phosphorus Dissolved |  | Low Value | $\begin{gathered} \text { Change } 0.009 \text { by } \\ 0.09 \end{gathered}$ | Arlington |
| 1488 | Phosphorus Dissolved |  | Values lower than DL | Change by $<0.02$ | Knoxville |
| 1527 | Phosphorus Dissolved |  | Values lower than DL | Change by 0.02 | Knoxville |
| 1580 | Phosphorus Dissolved |  | Values lower than DL | Change by 0.02 | Knoxville |
| 4079 | Beryllium |  | Detection limit | $\begin{gathered} \text { Change }<0.6 \text { by } \\ <0.06 \end{gathered}$ | Greensboro |
| 4245 | Cadmium |  | Detection limit | $\begin{gathered} \text { Change }<0.4 \text { by } \\ <0.04 \end{gathered}$ | Greensboro |
| 2150 | Cadmium |  | Wrong Columns | Copy 16000 in TotCol 230 Fec | Fulton |
| 2128 | Cadmium |  | Wrong Columns | Cd, Tot col and Fec Col in correct columns | Fulton |
| 1107 | Cadmium |  | LD in cell | Replace by detection limit | Howard County |
| 1110 | Cadmium |  | LD in cell | Replace by detection limit | Howard County |
| 2864 | Cadmium |  | Detection limit | $\begin{gathered} \hline \text { Replace }<2.5 \text { by } \\ <0.5 \\ \hline \end{gathered}$ | Ada County |
| 1107 | Chromium |  | LD in cell | Replace by detection limit | Howard County |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1110 | Chromium |  | LD in cell | Replace by detection limit | Howard County |
| $\begin{gathered} 2935- \\ 3419 \\ \hline \end{gathered}$ | Cyanide |  | Factor of a thousand. | Multiply by 1000 | Texas |
| 18 | Cyanide |  | Wrong Detection Limit | Changed | Arlington |
| 29 | Cyanide |  | Wrong Detection Limit | Changed | Arlington |
| 35 | Cyanide |  | Wrong Detection Limit | Changed | Arlington |
| 2460 | Conductivity | 1 | The value is the detection limit | Deleted | CAALA001 |
| 2871 | Conductivity | 2.5 | Conductivity was collected in grab sample | Deleted | IDADA002 |
| 2662 | DO | 11.6 | Evaluated by temperature | Deleted | AZMCA001 |
| 3016 | DO | 10.2 | Evaluated by temperature | Deleted | ORODA005 |
| 3076 | DO | 15 | Evaluated by temperature | Deleted | OREUA001 |
| 3077 | DO | 12.2 | Evaluated by temperature | Deleted | OREUA001 |
| 3078 | DO | 17 | Evaluated by temperature | Deleted | OREUA001 |
| 3092 | DO | 12.1 | Evaluated by temperature | Deleted | OREUA002 |
| 3093 | DO | 18.4 | Evaluated by temperature | Deleted | OREUA002 |
| 3097 | DO | 14 | Evaluated by temperature | Deleted | OREUA002 |
| 3107 | DO | 11.5 | Evaluated by temperature | Deleted | OREUA003 |
| 3108 | DO | 19.2 | Evaluated by temperature | Deleted | OREUA003 |
| 3115 | DO | 16.3 | Evaluated by temperature | Deleted | MABOA001 |
| 3120 | DO | 21.8 | Evaluated by temperature | Deleted | MABOA002 |
| 3122 | DO | 10.2 | Evaluated by temperature | Deleted | MABOA002 |
| 3126 | DO | 15.4 | Evaluated by temperature | Deleted | MABOA004 |
| 3129 | DO | 14.6 | Evaluated by temperature | Deleted | MABOA005 |
| 3020 | HARDNESS | <1 | Weight of evidence compared with conductivity and TDS | Deleted | ORODA005 |
| 4065 | TSS | 66 | Turbidity high but TSS low, checked with other parameters | TSS to 660 | NCGRHUST |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2619 | Turbidity | 5.2 | Grab samples higher than composite | Calculate time composite | CACTA010 |
| 2620 | Turbidity | 4.3 | Grab samples higher than composite | Calculate time composite | CACTA010 |
| $\begin{gathered} 1350- \\ 1351 \end{gathered}$ | Various |  | Values don't seem correct. The community changed lab after this samples | Delete event | GACLCOSI |
| 65 | TDS | 1 | Low value | Use <5 | VACHCOF5 |
| 875 | TDS | 2 | Low value | Use <5 | VAFFOF10 |
| 1611 | TDS | 2 | Low value | Use < 5 | TNKXTYWE |
| 1645 | TDS | 5406 | Elevated value without support | Deleted | TNMET410 |
| 2102 | TDS | 4000 | Elevated value without support | Deleted | GAFUCOS1 |
| 2122 | TDS | 4100 | Elevated value without support | Deleted | GAFUCOS2 |
| 2144 | TDS | 4200 | Elevated value without support | Deleted | GAFUCOS3 |
| $\begin{gathered} 2128, \\ 2150 \end{gathered}$ | Various |  | Values don't seem correct. The community changed lab after this samples | Delete event | GAFUCOS2,GAFU COS3 |
| 2155 | TDS | < | Missing detection limit | Deleted | GAFUCOS3 |
| 2699 | TDS | 1290 | Elevated value without support | Deleted | AZMCA003 |
| 2965 | TDS | 3 | Low Value | Use<5 | ORPOA006 |
| 2942 | TDS | 4 | Low Value | Use < 5 | ORPOA005 |
| 3691 | TDS | 1 | Low Value | Use < 5 | TXIRA002 |
| 3772 | TDS | 1 | Low Value | Use < 5 | TXIRA002 |
| 3119 | TDS | 17900 | Elevated value, but other samples support it. | Keep with ? | MABOA001 |
| 16 | TSS | 2 | Low Value | Use <5 | VAARLCV2 |
| 19 | TSS | 1 | Low Value | Use<5 | VAARLRS3 |
| 48 | TSS | 2 | Low Value | Use <5 | VACHCOF2 |
| 65 | TSS | 1 | Low Value | Use < 5 | VACHCOF5 |
| 76 | TSS | 2 | Low Value | Use < 5 | VACHCOF5 |
| 266 | TSS | 2 | Low Value | Use <5 | VAHATYH3 |
| 498 | TSS | 2 | Low Value | Use < 5 | VANFTYN4 |
| 812 | TSS | 1 | Low Value | Use < 5 | VAFFCOF3 |
| 842 | TSS | 2.5 | Low Value | Use < 5 | VAFFCOF6 |
| 935 | TSS | 2 | Low Value | Use < 5 | MDAACORK |
| 965 | TSS | 1.8 | Low Value | Use < 5 | MDBACOSC |
| 1160 | TSS | 2.87 | Low Value | Use <5 | MDMOCOCV |
| 1441 | TSS | 1 | Low Value | Use < 5 | TNKXTYFC |
| 1482 | TSS | 1 | Low Value | Use < 5 | TNKXTYGV |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2014 | TSS | 2 | Low Value | Use <5 | GACOC1A2 |
| 2028 | TSS | 2.2 | Low Value | Use < 5 | GACOCOL2 |
| 2143 | TSS | 1 | Low Value | Use<5 | GAFUCOS3 |
| 2235 | TSS | 2 | Low Value | Use <5 | NCFVTRYO |
| 3143 | TSS | 2.99 | Low Value | Use < 5 | GAATAT02 |
| 3149 | TSS | 1.83 | Low Value | Use<5 | GAATAT01 |
| 3150 | TSS | 1.98 | Low Value | Use < 5 | GAATAT01 |
| 3265 | TSS | 2 | Low Value | Use < 5 | TXHCA005 |
| 3451 | TSS | 0.5 | Low Value | Use < 5 | TXDAA004 |
| 3453 | TSS | 0.5 | Low Value | Use < 5 | TXDAA004 |
| 3552 | TSS | 0.5 | Low Value | Use<5 | TXFWA003 |
| 3647 | TSS | 0.5 | Low Value | Use < 5 | TXGAA003 |
| 3775 | TSS | 2 | Low Value | Use < 5 | TXPLA002 |
| 3776 | TSS | 2 | Low Value | Use < 5 | TXPLA002 |
| 3781 | TSS | 1 | Low Value | Use <5 | TXPLA002 |
| 1314 | BOD | 0.73 | Low Value | Use <1 | MDSHDTPS |
| 1317 | BOD | 0.41 | Low Value | Use <1 | MDSHDTPS |
| 1322 | BOD | 0.91 | Low Value | Use<1 | MDSHDTPS |
| 3868 | BOD | 0.7 | Low Value | Use <1 | NCCHSHEF |
| 32 | COD | <150 | Unusual Detection Limit | Deleted | VAARLTC4 |
| 2250 | COD | 1500 | Unusual elevated value, no evidence | Deleted | NCFVTRYO |
| 3479 | COD | 1300 | Unusual elevated value, no evidence | Deleted | TXDAA006 |
| 1897 | Ammonia | 60.3 | Unusual elevated value, no evidence | Deleted | KYLOTSR5 |
| 1907 | Ammonia | 60.5 | Unusual elevated value, no evidence | Deleted | KYLOTSR6 |
| 1909 | Ammonia | 30.4 | Unusual elevated value, no evidence | Deleted | KYLOTSR6 |
| 2699 | Ammonia | 64 | Unusual elevated value, no evidence | Deleted | AZMCA003 |
| 8 | NO2 NO3 | 13 | Unusual elevated value, no evidence | Deleted | VAARLLP1 |
| 1314 | NO2 NO3 | 7.05 | Unusual elevated value, no evidence | Deleted | MDSHDTPS |
| 2011 | NO2 NO3 | 6.3 | Unusual elevated value, no evidence | Deleted | GACOC1A2 |
| 2030 | NO2 NO3 | >0.2 | Unusual Detection Limit | Deleted | GACOCOL2 |
| 2140 | NO2 NO3 | 9.3 | Unusual elevated value, no evidence | Deleted | GAFUCOS3 |
| 2966 | NO2 NO3 | 6.5 | Unusual elevated value, no evidence | Deleted | ORPOA006 |
| 1905 | TN | 0.39 | TN < NH3 | Deleted both | KYLOTSR6 |
| 1907 | TN | 0.9 | TN < NH3 | Deleted both | KYLOTSR6 |
| 1909 | TN | 3 | TN < NH3 | Deleted both | KYLOTSR6 |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1600 | TN | 1.5 | TN < NH3 | Deleted both | TNKXTYWE |
| 4324 | TN | 3.42 | TN < TKN | Deleted both | NCRASIT3 |
| 4387 | TN | 6.65 | TN < TKN | Deleted both | NCRASIT6 |
| 3281 | TN | 50.2 | Unusual elevated value, no evidence | Deleted | TXHOA001 |
| 3289 | TN | 33 | Unusual elevated value, no evidence | Deleted | TXHOA002 |
| 3305 | TN | 28.9 | Unusual elevated value, no evidence | Deleted | TXHOA003 |
| 3331 | TN | 49.7 | Unusual elevated value, no evidence | Deleted | TXHOA005 |
| 1907 | Nitrogen Nitrite | 40 | Unusual elevated value, no evidence | Deleted | KYLOTSR6 |
| 1907 | Phosphate Ortho | 60.1 | Unusual elevated value, no evidence | Deleted | KYLOTSR6 |
| 2978 | Phosphate Ortho | 0.8 | Unusual elevated value, no evidence | Deleted | ORPOA007 |
| 3419 | Phosphorus Total | 0.005 | Low value | Deleted | TXDAA002 |
| 4073 | Antimony | 0.02 | Grab sample | Deleted | NCGRHUST |
| 2006 | Cadmium Total | 40.16 | Unusual elevated value, no evidence | Deleted | GACOC1A2 |
| 2007 | Cadmium Total | 42.7 | Unusual elevated value, no evidence | Deleted | GACOC1A2 |
| 2036 | Cadmium Total | 122 | Unusual elevated value, no evidence | Deleted | GACOCOL2 |
| 2128 | Cadmium Total | 16000 | Unusual elevated value, no evidence, seems to be wrong columns | Deleted | GAFUCOS2 |
| 1131 | Chromium Total | 120 | Unusual elevated value, no evidence | Deleted | MDMOCOWP |
| 797 | Copper Total | 396 | Unusual elevated value, no evidence | Deleted | VAFFCOF2 |
| 889-900 | Mercury Total | 0 |  | Deleted | PAPH |
| 1790 | Nickel Total | 200 | Unusual elevated value, Detection limit 20 | Change by 20 | KYLXWHL1 |
| 3299 | Nickel Total | 325 | Unusual elevated value, no evidence | Deleted | TXHOA002 |
| 3321 | Nickel Total | 720 | Unusual elevated value, no evidence | Deleted | TXHOA004 |
| 3504 | Nickel Total | 0.013 | Low value | Deleted | TXDCA001 |
| 3515 | Nickel Total | 0.01 | Low value | Deleted | TXDCA002 |
| 2456 | Selenium Total | 0.3 | Low value | Deleted | CAALA001 |
| 2457 | Selenium Total | 0.4 | Low value | Deleted | CAALA001 |
| 2458 | Selenium Total | 0.068 | Low value | Deleted | CAALA001 |
| 2459 | Selenium Total | 0.2 | Low value | Deleted | CAALA001 |
| 2460 | Selenium Total | 0.059 | Low value | Deleted | CAALA001 |
| 2461 | Selenium Total | 0.13 | Low value | Deleted | CAALA001 |

Table B1. Modified Values in the NSQD - Continued

| Order | Constituent | Original <br> value | Problem | Action | Location_ID |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2462 | Selenium Total | 0.095 | Low value | Deleted | CAALA001 |
| 1082 | Silver Total | 290 | Unusual elevated <br> value, no evidence | Deleted | MDCLCOJS |
| 1262 | Silver Total | 90 | Unusual elevated <br> value, no evidence | Deleted | MDPGCOS4 |
| 2006 | Zinc Total | 0.11 | Low value | Deleted | GACOC1A2 |
| 2007 | Zinc Total | 0.19 | Low value | Deleted | GACOC1A2 |
| 3704 | Zinc Total | 1 | Low value | Deleted | TXIRA002 |
| 3777 | Zinc Total | 1 | Low value | Deleted | TXPLA002 |
| 3514 | Runoff | 2.296 | High Value | Deleted | TXDCA002 |
| 3515 | Runoff | 0.909 | High Value | Deleted | TXDCA002 |
| 3201 | Runoff | 3.25 | High Value | Deleted | KATOBROO |
| 1364 | Runoff | 1.318 | High Value | Deleted | GACLCOSI |
| 2401 | Runoff | 1.73 | High Value | Deleted | NCFVWINS |

# Appendix C: Methods to Estimate Non-Detected Values in Stormwater Datasets 

## Introduction

A few large stormwater quality databases have been prepared in the past 20 years (EPA 1983; Smullen 2002, for example). The data collected generally shows that there are important variabilities in stormwater pollutant concentrations for different land uses. Other factors that some researchers have found to be important include: imperviousness, slope, and size of the watershed. However, these databases include numerous instances where the laboratory results are reported to be "below detection." Statistical analyses can be greatly affected by these uncertain values, depending on their number and percentage of occurrence. There are several schemes that have generally been used to overcome the problems associated with these non-detected values.

The NSQD database has collected data representing more than 3,700 storm events in the U.S., including information about the location of the monitoring station, watershed characteristics, hydrology, and chemical constituents. Each community has the flexibility to choose the equipment and analytical methods to detect the constituents in the stormwater. Chemical constituents in this database had been preliminary analyzed for different land uses (Pitt, et al. 2003). It has been observed while preparing the NSQD database, that different methods and procedures had been used for the analyses of the samples. The use of different methods generates different detection limits in the database for the same constituent.

Datasets containing values below the detection limits (censored data) complicate the statistical analyses, even including the basic calculations of the means and variance. Most of the time, "left-censored" data are of concern (observations below the detection limit). However, there are situations where "right-censored" data may occur, especially for bacteria analyses, when the observations are greater than the upper limit of the dilution. Three main approaches to the analysis of censored data can be found in the literature: substitution, statistical estimation, and graphical methods. In this chapter, these methods will be presented using different data sets.

## Analysis of Multiple Censored Data

Estimation methods for single censored data have been widely discussed in the literature. However, in the case of multiple censored data (datasets affected by several different detection limits), the situation is not the same. Helsel and Cohn (1988) continued the previous work of Guilliom and Helsel, but for multiple censored data.

Eight methods were studied in the multiple censored cases:

1) ZE : Censored data are assumed to equal zero.
2) DL: Censored data are assumed to equal the detection limit.
3) HA: Censored data are assumed to equal half the detection limit.
4) LR: Entire data set is log transformed and is assumed to be normally distributed. Censored data is estimated using least squares regression.
5) MR: Plotting positions are calculated using equations given by Hirsch and Stedinger (1987).
6) LM: Concentrations are assumed log normally distributed with parameters using the Cohen method. The mean and standard deviation of the untransformed values were estimated using the equations given by Aitchison and Brown (1969).
7) MM: This method uses the maximum likelihood method, but for the case of multiple censored data. Cohen (1976)
8) AM: Adjusted maximum likelihood procedure of Cohn (1988). The AM method is the same as the MM but makes a first order correction in the bias.

When the LR and LM methods were used, all the points below the highest of the censoring thresholds were treated as less than that censoring level. This will simplify the problem as a single censored occurrence. In the last three methods listed above, it is assumed that the data is log-normally distributed.

The results indicate that the MM and MR methods are improvements compared to the results obtained with the single threshold assumption. The MR, MM and AM methods were also compared. A higher RMSE (root mean squared error) for the moments was estimated by the MM method. The AM method present lower errors than the plotting position method (MR), but it is less robust for distributions different than log-normal. The substitution methods present a higher error than the MR or the AM methods.

One of the main problems using these methods was to assume that that the data is lognormal. There is no certainty that water quality follows this distribution. For that reason, robust methods are considered very important in water quality analysis. When data depart from the lognormal distribution, the RMSE of the mean and standard deviation values, when using the MM and AM methods, can be larger than $1000 \%$. Helsel and Cohn (1988) indicate that in water quality data the lognormal distribution and the gamma with a coefficient of variation of two are very common. The MR model present better results when the distribution is not known.

They also evaluate the plotting position using the Weibull, Blom, and Hazel equations. There really is not an effect in the results when any of these equations are used.

If the distribution is unknown, the MR method should be chosen. If there is certainty that the distribution is lognormal, the AM method is recommended. The previous methods were evaluated with copper observations in commercial areas during the fall. Table C 1 shows the original observations, and Table C 2 show the log-transformed observations.

Table C1. Copper Observations in Commercial Areas

| 2 | 2 | 3 | 5 | 5 | 5 | 5 | 5.2 | 5.4 | 5.5 | 6 | 6 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 6 | 6.5 | 6.5 | 7 | 8 | 8 | 8.1 | 8.4 | 9 | 9 | 9 | 10 |
| 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 11 | 11 | 13 |
| 13 | 13.4 | 14 | 14 | 14 | 14 | 14 | 14 | 14.4 | 14.5 | 15 | 15 |
| 15 | 17 | 17 | 17 | 17 | 17.1 | 18 | 19 | 19 | 20 | 20 | 20 |
| 20 | 20 | 20 | 20 | 20 | 20 | 21 | 21 | 21 | 22 | 22 | 22.4 |
| 23 | 24 | 24 | 26 | 26.6 | 26.9 | 29 | 29 | 30 | 30 | 33 | 33.7 |
| 36 | 37 | 37 | 40 | 41.3 | 42 | 50 | 50 | 50.5 | 50.7 | 59.4 | 60 |
| 60 | 61 | 62 | 70 | 100 | 130 | 130 | 175 | $<10$ | $<10$ | $<10$ | $<10$ |
| $<10$ | $<10$ | $<20$ |  |  |  |  |  |  |  |  |  |

Table C2. Copper Observations in Commercial Areas - (Log Values)

| 0.3 | 0.3 | 0.5 | 0.7 | 0.7 | 0.7 | 0.7 | 0.7 | 0.73 | 0.74 | 0.78 | 0.78 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0.8 | 0.81 | 0.8 | 0.9 | 0.9 | 0.9 | 0.9 | 0.9 | 0.95 | 0.95 | 0.95 | 1 |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1.04 | 1.04 | 1.11 |
| 1.1 | 1.13 | 1.2 | 1.2 | 1.15 | 1.15 | 1.2 | 1.2 | 1.16 | 1.16 | 1.18 | 1.18 |
| 1.2 | 1.23 | 1.2 | 1.2 | 1.23 | 1.23 | 1.3 | 1.3 | 1.28 | 1.3 | 1.3 | 1.3 |
| 1.3 | 1.3 | 1.3 | 1.3 | 1.3 | 1.3 | 1.3 | 1.3 | 1.32 | 1.34 | 1.34 | 1.35 |
| 1.4 | 1.38 | 1.4 | 1.4 | 1.43 | 1.43 | 1.5 | 1.5 | 1.48 | 1.48 | 1.52 | 1.53 |
| 1.6 | 1.57 | 1.6 | 1.6 | 1.62 | 1.62 | 1.7 | 1.7 | 1.7 | 1.7 | 1.77 | 1.78 |
| 1.8 | 1.79 | 1.8 | 1.9 | 2 | 2.11 | 2.1 | 2.2 | $<1$ | $<1$ | $<1$ | $<1$ |
| $<1$ | $<1$ | $<1.30$ |  |  |  |  |  |  |  |  |  |

Figure C 1 shows the probability plots when censored data was deleted, replaced by the detection limit, replaced by half of the detection limit, and estimating the values below the highest detection limit using the LR method. The results indicated that there is a bias in the mean value when the censored data is deleted or replaced by the detection limit. When data is replaced by half of the detection limit, or is estimated by the LR methods, the results are very similar. Notice that in the LR method, all the values below the highest detection limit are considered censored. This assumption changes the level of censoring from $6 \%$ to $58 \%$, but even at this level of censoring, the results are very close to those obtained with the substitution methods. Because the transformed data seems to follows a log-normal distribution, it is possible to estimate the moments using only the upper side of the line.


Figure C1. Probability plot for different replacements of the censored data for copper

The MR method was proposed by Hirsch and Stedinger (1987). They define a variable Aj as the number of uncensored data greater than the detection limit $j$ and below the next detection limit. In the copper data, there are 47 uncensored observations above the highest detection limit ( $A 3=47, D L=20$ ). There are 34 observations between the detection limits of 20 and $10 \mu \mathrm{~g} / \mathrm{L},(\mathrm{A} 2=34, \mathrm{DL}=10)$. Finally, there are 23 uncensored observations above the minimum detection limit, zero $(\mathrm{A} 1=23, \mathrm{DL}=0)$. This assumption must be made in the case that the smallest value in the dataset is not a censored value. The parameter Bj is defined as the number of censored and uncensored observations below the j detection limit. In the example case, there are 64 observations below the second detection limit $(B 3=64)$, 29 observations below the first detection limit $(B 2=29)$, and zero observations below the detection limit zero $(B 1=0)$. The method uses the probability of exceeding the jth detection limit pe,j to calculate the probability position of each observation.

$$
\begin{equation*}
p_{e, j}=p_{e, j+1}+\frac{A_{j}}{A_{j}+B_{j}}\left(1-p_{e, j+1}\right) \tag{C.1}
\end{equation*}
$$

The calculations are easier going from higher to lower values. In the copper example, there are three detection limits; by definition, the probability of exceeding a fourth detection limit is zero. The probability of exceeding the third, second and first detection limits are $0.423,0.735$ and 1 , respectively.

The Weibull formula was used to calculate the plotting position of the censored data in the range between the probabilities of exceeding boundaries.

$$
\begin{equation*}
p(i)=\left(1-p_{e, j}\right)+\left(p_{e, j}-p_{e, j+1}\right) \frac{i}{A_{j}+1} \tag{C.2}
\end{equation*}
$$

This formula indicates that the values observed between the j and $\mathrm{j}+1$ range are distributed according to the Weibull formula. The plotting position for the censored data follows the same concept; distribute the censored data between the limits using the Weibull formula. For the censored observations, the plotting positions can be calculated as:

$$
\begin{equation*}
p c(i)=\left(1-p_{e, j}\right) \frac{i}{C_{j}+1} \tag{C.3}
\end{equation*}
$$

The formula calculates the position of the ith censored observation, among the C tied observations, in the j th detection limit. The probability plot is shown in Figure C2.


Figure C2. Probability plot using the MR method.

The MM method presented by Cohen (1976) uses the maximum likelihood method for the three parameter lognormal distribution.

$$
\begin{equation*}
f(x ; \mu, \sigma, \gamma)=\frac{1}{\sigma \sqrt{2 \pi}(x-\gamma)} \exp \left[-\frac{(\ln (x-\gamma)-\mu)^{2}}{2 \sigma^{2}}\right] \tag{C.4}
\end{equation*}
$$

$$
\gamma<x<\infty
$$

In this distribution, X is a random variable lognormal. The method assumes that there are k censored observations and n noncensored observations. For a censored level Tj , the transformed yj will be:

$$
\begin{equation*}
y_{i}=\ln \left(T_{j}-\gamma\right) \tag{C.5}
\end{equation*}
$$

and,

$$
\begin{equation*}
\xi_{j}=\frac{y_{i}-\mu}{\sigma} \tag{C.6}
\end{equation*}
$$

Cohen created a new variable Zj that is used to solve the maximum likelihood estimators.

$$
\begin{equation*}
Z_{j}=Z\left(\xi_{j}\right)=\frac{\varphi\left(\xi_{j}\right)}{1-\Phi\left(\xi_{j}\right)} \tag{C.7}
\end{equation*}
$$

Where:
$\phi\left(\xi_{\mathrm{j}}\right)=$ normal density function $\mathrm{N}(0,1)$
$\Phi\left(\xi_{\mathrm{j}}\right)=$ normal cumulative density function $\mathrm{N}(0,1)$
Three simultaneous equations can be solved to estimate the parameters $\mu$, $\sigma$, and $\gamma$.

$$
\begin{align*}
& \sum_{i=1}^{n}\left[\ln \left(x_{i}-\gamma\right)-\mu\right]+\sigma \sum_{j=1}^{k} Z_{j}=0  \tag{C.8}\\
& \sum_{i=1}^{n}\left[\ln \left(x_{i}-\gamma\right)-\mu\right]^{2}+\sigma^{2}\left[\sum_{j=1}^{k} \xi_{j} Z_{j}-n\right]=0 \tag{C.9}
\end{align*}
$$

Cohn suggested assuming a $\gamma$ value and solve for $\mu$ and $\sigma$ from the first two equations. After that, the $\gamma$ parameter can be recalculated using the third equation. In some cases, the parameter g does not converge. In that case the following approximation must be used.

$$
\begin{equation*}
\gamma=x_{k}-\exp \left(\mu+\sigma \xi_{k}\right) \tag{C.10}
\end{equation*}
$$

Where $\mathrm{x}_{\mathrm{k}}$ is the $\mathrm{k}^{\text {th }}$ order statistic in the sample.
The AM method is the best alternative in the case that the distribution is lognormal. In any other case, it was found that elevated bias and rmse are obtained in the mean and the variance $(>1000 \%)$ if the distribution is different.

The National Council of the Paper Industry for Air and Stream Improvement had created seven technical bulletins about statistical methods used with environmental data sets (NCASI 1995). One of the reports presents a decision tree to select the appropriate statistical method and a description of the Cohen's multilevel MLE procedure. This method was recommended after compare it with other methods, such as replacement/deletion, D-log procedure, regression of normal order statistics balancing techniques, and graphical techniques.

The diagram indicates that in some cases of multiple detection limits, the problem can be solved using single censoring point (SCP) methods, for example when all the non-detected values are smaller than the detected values. In other situations, the simplification cannot be done and multiple censoring point (MCP) methods must be used. The Cohen's maximum likelihood method obtained the mean and variance estimates from the logarithm of the likelihood function and obtaining the partial derivate in respect to the mean and the variance. The following equation defines the log likelihood function:

$$
\begin{equation*}
L(S)=-n \ln (\sigma)-\frac{1}{2} \sum_{1}^{n}\left(\frac{x_{i}-\mu}{\sigma}\right)^{2}+\sum_{1}^{k} r_{i} \ln \left[F_{i}\right]-n \ln [\sqrt{2 \pi}] \tag{C.11}
\end{equation*}
$$

Where:
$\mathrm{S}=$ sample set containing a total of N censored observations and fully quantified values
$\mathrm{x}_{\mathrm{i}}=\mathrm{i}^{\text {th }}$ fully quantified value
$\mu=$ population mean
$\sigma=$ population standard deviation
$k=$ number of censored levels
$r_{i}=$ number of censored values at each censored level i
$\mathrm{n}=$ number of noncensored observations
$\mathrm{F}_{\mathrm{i}}=\mathrm{F}\left(\xi_{\mathrm{i}}\right)=$ area under standard normal curve at f
$\xi_{i}=\left(T_{i}-\mu\right) / \sigma$, standard normal variate for the $\mathrm{i}^{\text {th }}$ censoring level
$\phi(t)=(2 \pi)^{-1} \exp \left[-\mathrm{t}^{2} / 2\right]$, ordinate value of normal variate, $\mathrm{f} /\left(\xi_{\mathrm{i}}\right)$
$\mathrm{T}_{\mathrm{i}}=$ the limit of detection of the $\mathrm{i}^{\text {th }}$ level of censoring.
The derivates are:

$$
\begin{equation*}
\frac{\partial L}{\partial \mu}=\frac{n}{\sigma}\left[\frac{(\bar{x}-\mu)}{\sigma}-\sum_{1}^{k} \frac{r_{i}}{n} Z_{i}\right]=0 \tag{C.12}
\end{equation*}
$$

(C.13)

$$
\frac{\partial L}{\partial \sigma}=\frac{n}{\sigma}\left[\frac{s^{2}+(\bar{x}-\mu)^{2}}{\sigma^{2}}-1-\sum_{1}^{k} \frac{r_{i}}{n} \xi_{i} Z_{i}\right]=0
$$

$\mathrm{Z}_{\mathrm{i}}=$ the hazard function $\phi_{\mathrm{i}} / \mathrm{F}_{\mathrm{i}}$.
$s^{2}=$ Sample variance.
NCASI (1995) includes the program source code for using Cohen's method, in FORTRAN and SAS. The procedure and the code presented in the technical bulletin No. 703 were used to estimate the censored observations for this NSQD research.

## Appendix D: Unusual Sites Identified Using Xbar Plots

This appendix describes sites having unusual stormwater concentrations for all land uses, besides the residential areas that were described in Chapter 3.

## Evaluation of the Methods Selected to Estimate Non-Detected Observations

Three methods were used to estimate appropriate substitution values for the non-detected observations: delete them ("ignore"), replace them by half of the detection limit ("HD") or estimate them using the Cohen's maximum likelihood method (an extrapolation of the probability plot of the data) ("estimate"), as presented in the preceding appendix. The following discusses the analyses for each constituent for each land use category.

## Hardness

Total hardness was detected in all samples, except in industrial land use areas where less than $2 \%$ of the samples were not detected. Changes in the average, median, standard deviation and coefficient of variation were not significant if the non-detected values were ignored, estimated, or replaced by half of the detection limit. Table D1 shows that there are no important differences in the industrial land descriptions using any of these three methods.

Table D1. Summary Statistics for Estimated Observations for Total Hardness (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 250 | 250 | 250 | 139 | 139 | 139 | 138 | 138 | 138 |
| \% Detected | 100.00 |  |  | 100.00 |  |  | 96.38 |  |  |
| Minimum | 3.00 | 3.00 | 3.00 | 1.90 | 1.90 | 1.90 | 5.50 | 5.00 | 5.00 |
| Maximum | 401.00 | 401.00 | 401.00 | 356.00 | 356.00 | 356.00 | 888.00 | 888.00 | 888.00 |
| Average | 43.32 | 43.32 | 43.32 | 62.03 | 62.03 | 62.03 | 68.83 | 66.52 | 66.52 |
| Median | 32.00 | 32.00 | 32.00 | 38.90 | 38.90 | 38.90 | 39.00 | 38.50 | 38.50 |
| Standard Dev. | 44.87 | 44.87 | 44.87 | 65.17 | 65.17 | 65.17 | 104.55 | 103.32 | 103.32 |
| Coeff. of Var. | 1.04 | 1.04 | 1.04 | 1.05 | 1.05 | 1.05 | 1.52 | 1.55 | 1.55 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 8 | 8 | 8 | 127 | 127 | 127 |
| \% Detected | 100.00 |  |  | 100.00 |  |  |
| Minimum | 11.00 | 11.00 | 11.00 | 5.00 | 5.00 | 5.00 |
| Maximum | 270.00 | 270.00 | 270.00 | 1000.00 | 1000.00 | 1000.00 |
| Average | 145.25 | 145.25 | 145.25 | 57.19 | 57.19 | 57.19 |
| Median | 150.00 | 150.00 | 150.00 | 34.00 | 34.00 | 34.00 |
| Standard Dev. | 85.12 | 85.12 | 85.12 | 105.95 | 105.95 | 105.95 |
| Coeff. of Var. | 0.59 | 0.59 | 0.59 | 1.85 | 1.85 | 1.85 |

Figure A1 shows probability plots for industrial land use hardness values. The plot indicates that the mean value is smaller when the non-detected values are either estimated or replaced by half of the detection limit. The lower $40 \%$ of the distribution is displaced to the left. All the non-detected values were observed at $10 \mathrm{mg} / \mathrm{L}$. The upper $60 \%$ of the distribution is not affected by the non-detected values.


Figure D1. Estimated hardness distributions in industrial land use areas

## Oil and Grease

Oil and grease had censored data for $37 \%$ and $72 \%$ of the observations. Table D2 shows the differences in the descriptive statistics using the three methods. The greatest change occurred in the coefficient of variation values for freeway sites. The mean oil and grease values increased in a range of $30 \%$ to $60 \%$ when the censored observations were ignored. The difference was below $4 \%$ when the censored observations were replaced using Cohen's maximum likelihood method, or replaced by half of the detection limit.

Table D2. Summary Statistics for Estimated Observations for Oil and Grease (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 533 | 533 | 533 | 308 | 308 | 308 | 327 | 327 | 327 |
| \% Detected | 57.79 |  |  | 70.78 |  |  | 65.14 |  |  |
| Minimum | 0.20 | 0.02 | 0.20 | 0.80 | 0.03 | 0.25 | 0.50 | 0.00 | 0.25 |
| Maximum | 2980 | 2980 | 2980 | 359 | 359 | 359 | 11000 | 11000 | 11000 |
| Average | 22.85 | 13.87 | 13.89 | 12.63 | 9.42 | 9.39 | 62.87 | 41.40 | 41.39 |
| Median | 3.85 | 2.50 | 2.50 | 4.70 | 3.00 | 3.00 | 5.00 | 2.50 | 2.60 |
| Standard Dev. | 175.53 | 133.76 | 133.76 | 39.75 | 33.80 | 33.81 | 753.77 | 608.56 | 608.56 |
| Coeff. of Var. | 7.68 | 9.65 | 9.63 | 3.15 | 3.59 | 3.60 | 11.99 | 14.70 | 14.70 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 19 | 19 | 19 | 60 | 60 | 60 |
| \% Detected | 36.84 |  |  | 71.67 |  |  |
| Minimum | 0.50 | 0.50 | 0.50 | 3.00 | 0.50 | 0.25 |
| Maximum | 3.70 | 3.70 | 3.70 | 30.00 | 30.00 | 30.00 |
| Average | 1.53 | 1.09 | 1.09 | 8.49 | 6.57 | 6.45 |
| Median | 1.30 | 0.50 | 0.50 | 8.00 | 4.65 | 4.65 |
| Standard Dev. | 1.07 | 0.93 | 0.93 | 5.28 | 5.42 | 5.52 |
| Coeff. of Var. | 0.70 | 0.85 | 0.85 | 0.62 | 0.83 | 0.86 |

The probability plot in residential land use areas indicates that the lower tail is better described with the Cohen estimated method (Figure D2). The upper tail was the same for the estimated and the half detection limit method. About $40 \%$ of the non-detected values were at the $<1 \mathrm{mg} / \mathrm{L}$ level, and another $40 \%$ were at the $<5 \mathrm{mg} / \mathrm{L}$ level. The estimated values better describe the lower tail, however there was no significant differences in the means, standard deviations and coefficients of variation. This case is very important because the level of censoring was large ( $42.2 \%$ ). Ignoring the non-detected values increased the mean value by more than $64 \%$ and the standard deviation by more than $30 \%$, and reduces the coefficient of variation in $20 \%$.

The analyses for commercial land use data resulted in a similar trend as observed for the residential land use areas (Figure D3). There is a better description of the lower tail, but the mean, standard deviation and coefficient of variation values are almost the same if the censored data are replaced by half of the detection limit, or if they are estimated. The most frequent reported level of non-detected values was $<5 \mathrm{mg} / \mathrm{L}$, followed by $<1 \mathrm{mg} / \mathrm{L}$. The average was increased by $34 \%$, and the standard deviation by $18 \%$, when the censored data was ignored, and the coefficient of variation was reduced about $12 \%$ when the non-detected values were ignored.

Figure D4 shows the probability plot for oil and grease data at industrial land use areas and illustrates the case when an unusual value was present in the dataset. The maximum observation was larger by a factor of 2,200 compared with the median value of the distribution. This generates a coefficient of variation of 12 when the censored data are ignored, or 14.7 in the case when they are estimated or replaced by half of the detection limit.


Figure D2. Estimated oil and grease distributions in residential land use areas


Figure D3. Estimated oil and grease distributions in commercial land use areas

The percentage of detected values for oil and grease in open space areas was very low (only 7 of 19 observations were detected) (Figure D5). Almost all of the non-detected values were at $<1 \mathrm{mg} / \mathrm{L}$. It was not possible to use the Cohen's maximum likelihood method in this case because the percentage of non-detected values was too high. Ignoring the non-detected values will increase the mean value by almost $40 \%$ compared when the non-detected values were replaced with half of the detection limit.

The probability plot for freeway oil and grease values indicate that estimating or replacing the censored observations for half of the detection limit does not cause a significant difference in the coefficient of variation (Figure D6). The coefficient of variation was $3 \%$ larger when half of the detection limit was used instead of Cohen's method. A different situation occurs when the non-detected values were ignored. In this case, the coefficient of variation was reduced by $30 \%$ compared with the estimated method.


Figure D4. Estimated oil and grease distributions in industrial land use areas


Figure D5. Estimated oil and grease distributions in industrial land use areas


Figure D6. Estimated oil and grease distributions in freeway land use areas

## Total Dissolved Solids (TDS)

In all the land use categories, the percentages of non-detected TDS values were very low. The lowest percentage was observed in open space areas, with $2 \%$ not detected. No important differences were observed in the means, standard deviations and coefficients of variation when the non-detected values were ignored, estimated using with the Cohen method, or substituting with half the detection limit. Descriptive statistics for each of the three methods are shown in Table D3.

Table D3. Summary Statistics for Estimated Observations for TDS (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 861 | 861 | 861 | 399 | 399 | 399 | 412 | 412 | 412 |
| \% Detected | 99.19 |  |  | 99.50 |  |  | 99.51 |  |  |
| Minimum | 3.00 | 3.00 | 0.50 | 4.00 | 4.00 | 4.00 | 4.50 | 1.78 | 2.50 |
| Maximum | 1700 | 1700 | 1700 | 3860 | 3860 | 3860 | 11200 | 11200 | 11200 |
| Average | 96.26 | 95.54 | 95.50 | 109.94 | 109.44 | 109.42 | 161.99 | 161.23 | 161.22 |
| Median | 72.00 | 70.50 | 70.50 | 74.00 | 74.00 | 74.00 | 91.00 | 89.50 | 89.50 |
| Standard Dev. | 102.45 | 102.35 | 102.38 | 208.76 | 208.36 | 208.37 | 582.40 | 581.09 | 581.09 |
| Coeff. of Var. | 1.06 | 1.07 | 1.07 | 1.90 | 1.90 | 1.90 | 3.60 | 3.60 | 3.60 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 45 | 45 | 45 | 97 | 97 | 97 |
| \% Detected | 97.78 |  |  | 98.97 |  |  |
| Minimum | 32.00 | 10.79 | 2.50 | 12.00 | 5.85 | 0.50 |
| Maximum | 542 | 542 | 542 | 470 | 470 | 470 |
| Average | 151.41 | 148.28 | 148.10 | 95.31 | 94.39 | 94.34 |
| Median | 124.50 | 119.00 | 119.00 | 77.50 | 77.00 | 77.00 |
| Standard Dev. | 109.83 | 110.58 | 110.82 | 76.38 | 76.52 | 76.59 |
| Coeff. of Var. | 0.73 | 0.75 | 0.75 | 0.80 | 0.81 | 0.81 |

Figure D7 shows the probability plots for residential land use TDS concentrations. The plot indicates that using half of the detection limit lowers values compared to the Cohen's maximum likelihood method. The upper $95 \%$ of the distributions are identical for the three cases. The probability plots don't indicate significant differences among the three methods for the remaining land uses. For example, Figure D8 shows the probability plots for commercial areas. The three lines overlap, except for a small fraction in the lower tail of the distribution.


Figure D7. Estimated TDS distributions in residential land use areas


Figure D8. Estimated TDS distributions in commercial land use

## Total Suspended Solids (TSS)

The results for TSS were similar to above described results for TDS, the maximum level of non-detected values was observed in open space areas, where about $5 \%$ of the observations were censored. Table D4 indicates that there are not any relevant differences in means, standard deviations or coefficients of variation for any of the three methods.

Table D4. Summary Statistics for Estimated Observations for TSS (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 991 | 990 | 991 | 458 | 457 | 458 | 427 | 426 | 427 |
| \% Detected | 98.59 |  |  | 98.25 |  |  | 99.06 |  |  |
| Minimum | 3.00 | 0.63 | 0.25 | 3.00 | 1.56 | 0.25 | 3.00 | 0.43 | 0.50 |
| Maximum | 2462 | 2462 | 2462 | 2385 | 2385 | 2385 | 2490 | 2490 | 2490 |
| Average | 99.84 | 98.53 | 98.46 | 110.06 | 108.45 | 108.18 | 142.44 | 141.36 | 141.12 |
| Median | 49.00 | 48.00 | 48.00 | 42.00 | 41.00 | 41.00 | 78.00 | 76.36 | 76.00 |
| Standard Dev. | 179.12 | 178.29 | 178.22 | 218.51 | 217.22 | 217.05 | 218.76 | 218.35 | 218.15 |
| Coeff. of Var. | 1.79 | 1.81 | 1.81 | 1.99 | 2.00 | 2.01 | 1.54 | 1.54 | 1.55 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 44 | 44 | 44 | 134 | 134 | 134 |
| \% Detected | 95.45 |  |  | 99.25 |  |  |
| Minimum | 3.00 | 1.22 | 0.50 | 3.00 | 3.00 | 0.50 |
| Maximum | 980 | 980 | 980 | 4800 | 4800 | 4800 |
| Average | 176.88 | 168.98 | 168.91 | 173.39 | 172.13 | 172.10 |
| Median | 48.50 | 39.00 | 39.00 | 99.00 | 98.50 | 98.50 |
| Standard Dev. | 263.04 | 259.44 | 259.49 | 448.85 | 447.39 | 447.41 |
| Coeff. of Var. | 1.49 | 1.54 | 1.54 | 2.59 | 2.60 | 2.60 |

The probability plots indicate that the lower values were better estimated using half of the detection limit, rather than the Cohen's method. This indicate that with large numbers observations and small percentages of non-detected values, replacing the missing data by half of the detection limit will produce similar means compared to those obtained when using the maximum likelihood method. Figure A9 shows the probability plot for TSS concentrations for residential land use areas. The three curves overlap, indicating than the three methods will produce practically the same result.

The probability plot for open space has the lower number of observations among the five land uses. In this case, the pattern observed in the three methods was almost the same. The coefficient of variation increases only $3 \%$ when the censored data was estimated with the Cohen method, or replaced by half of the detection limit.


Figure D9. Estimated TSS distributions in residential land use areas

## Biochemical Oxygen Demand (BOD ${ }_{5}$ )

The percentage of non-detected values for $\mathrm{BOD}_{5}$ was higher in open space and freeway areas compared with the other land uses (Table D5). The lowest concentrations were observed in open space areas with a median $\mathrm{BOD}_{5}$ value of $4 \mathrm{mg} / \mathrm{L}$. Freeways, commercial and residential land use areas have similar concentrations, with $15 \mathrm{mg} / \mathrm{L}$ average $\mathrm{BOD}_{5}$ values. The highest $\mathrm{BOD}_{5}$ concentration was observed at an industrial land use site, however a single unusual $\mathrm{BOD}_{5}$ observation of $6,920 \mathrm{mg} / \mathrm{L}$ had a large effect on the mean, standard deviation and coefficient of variation values.

The lognormal probability plot for industrial land use areas showed one unusual $\mathrm{BOD}_{5}$ observation. This $\mathrm{BOD}_{5}$ concentration was 35 times larger than the second highest observation. This unusual value increased the standard deviation almost 18 times compared with the other land uses. Figure D10 shows the probability plot for $\mathrm{BOD}_{5}$ concentrations at industrial land use areas.

Table A5. Summary Statistics for Estimated Observations for $\mathrm{BOD}_{5}$ (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 941 | 941 | 941 | 432 | 432 | 432 | 406 | 406 | 406 |
| \% Detected | 97.56 |  |  | 97.45 |  |  | 95.32 |  |  |
| Minimum | 1.00 | 1.00 | 0.50 | 2.00 | 0.75 | 0.50 | 1.00 | 0.55 | 0.50 |
| Maximum | 350 | 350 | 350 | 150 | 220 | 150 | 6920 | 6920 | 6920 |
| Average | 15.05 | 14.97 | 14.84 | 18.16 | 18.58 | 18.14 | 35.92 | 34.65 | 34.47 |
| Median | 9.00 | 9.00 | 9.00 | 11.00 | 11.00 | 11.00 | 9.00 | 9.00 | 9.00 |
| Standard Dev. | 22.25 | 22.34 | 22.11 | 20.25 | 22.59 | 20.63 | 351.89 | 343.62 | 343.61 |
| Coeff. of Var. | 1.48 | 1.49 | 1.49 | 1.12 | 1.22 | 1.14 | 9.80 | 9.92 | 9.97 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 44 | 43 | 44 | 26 | 26 | 26 |
| \% Detected | 86.36 |  |  | 84.62 |  |  |
| Minimum | 1.00 | 0.62 | 0.50 | 2.0 | 1.5 | 1.5 |
| Maximum | 20 | 20 | 20 | 89 | 89 | 89 |
| Average | 6.25 | 5.68 | 5.74 | 14.86 | 13.06 | 12.88 |
| Median | 5.40 | 4.00 | 4.00 | 8.0 | 6.5 | 6.5 |
| Standard Dev. | 4.30 | 4.34 | 4.38 | 18.68 | 17.67 | 17.76 |
| Coeff. of Var. | 0.69 | 0.76 | 0.76 | 1.26 | 1.35 | 1.38 |

Open space and freeway areas had the largest level of non-detected $\mathrm{BOD}_{5}$ values. The mean value for open space areas increased by $10 \%$ when the censored data were ignored. No significance difference was observed for the variance values (Figure D11). Estimating the non-detected value using Cohen's method, or replacing the nondetected values by half of the detection limit results in almost the same means, standard deviations and coefficients of variation values.


Figure $\mathbf{A 1 0}$. Estimated $\mathrm{BOD}_{5}$ distributions in industrial land use areas


Figure D11. Estimated $\mathrm{BOD}_{5}$ distributions for open space land use areas

## Chemical Oxygen Demand (COD)

Differences in the means, averages and coefficients of variation for COD concentrations between the different methods for replacing the censored data were not important, except for the open space land use area where the level of non-detected observations was high (close to $25 \%$ ) (Table D6). In the remaining land use areas, the frequency of non-detected values was smaller than $2 \%$.

Table D6. Summary Statistics for Estimated Observations for COD (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 796 | 796 | 796 | 373 | 373 | 373 | 361 | 361 | 361 |
| \% Detected | 98.87 |  |  | 98.39 |  |  | 98.89 |  |  |
| Minimum | 5.00 | 1.74 | 0.50 | 4.00 | 1.96 | 0.50 | 2.00 | 2.00 | 2.00 |
| Maximum | 620 | 620 | 620 | 635 | 635 | 635 | 1260 | 1260 | 1260 |
| Average | 74.34 | 73.55 | 73.52 | 94.11 | 92.70 | 92.63 | 103.23 | 102.26 | 102.17 |
| Median | 55.00 | 53.60 | 53.60 | 60.00 | 59.00 | 59.00 | 60.00 | 59.00 | 59.00 |
| Standard Dev. | 69.12 | 69.12 | 69.15 | 94.39 | 94.28 | 94.34 | 127.35 | 126.97 | 127.03 |
| Coeff. of Var. | 0.93 | 0.94 | 0.94 | 1.00 | 1.02 | 1.02 | 1.23 | 1.24 | 1.24 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 44 | 44 | 44 | 67 | 67 | 67 |
| \% Detected | 75.00 |  |  | 98.51 |  |  |
| Minimum | 8.00 | 3.70 | 5.00 | 2.44 | 2.44 | 2.44 |
| Maximum | 476 | 476 | 476 | 1012.82 | 1012.82 | 1012.82 |
| Average | 51.47 | 40.93 | 40.76 | 140.99 | 139.10 | 138.96 |
| Median | 42.10 | 24.85 | 24.85 | 100.00 | 100.00 | 100.00 |
| Standard Dev. | 79.11 | 70.73 | 70.78 | 148.89 | 148.56 | 148.69 |
| Coeff. of Var. | 1.54 | 1.73 | 1.74 | 1.06 | 1.07 | 1.07 |

One characteristic of the COD probability plot is that the lower tail does not follow the trend showed by the rest of the distribution. Figure D12 shows an example COD distribution for residential land use areas. This effect is increased when the censored data is estimated or replaced by half of the detection limit.

The mean value in open space land use areas was increased by $25 \%$ when the censored data was ignored (Figure D13). In contrast, the coefficient of variation was reduced by almost $12 \%$ when the non-detected values were ignored. No significant differences can be observed when the censored data was estimated using Cohen's method or replaced with half of the detection limit.


Figure D12. Estimated COD distributions in residential land use areas


Figure D13. Estimated COD distributions in open space land use areas

## Ammonia $\left(\mathbf{N H}_{3}\right)$

Ammonia had one of the largest levels of censored observations of the common stormwater constituents examined in detail. The percentage of non-detected observations was about $20 \%$, except for open space areas where it is more than $80 \%$. The highest ammonia concentrations were observed at the freeway sites. Ignoring the censored observations increased the mean values by about $15 \%$, while ignoring the non-detected values increased the coefficients of variation by almost $15 \%$.

Table A7. Summary Statistics for Estimated Observations for Ammonia (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 595 | 595 | 595 | 299 | 299 | 299 | 253 | 252 | 253 |
| \% Detected | 81.51 |  |  | 83.28 |  |  | 83.40 |  |  |
| Minimum | 0.01 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 |
| Maximum | 5.60 | 5.60 | 5.60 | 7.80 | 7.80 | 7.80 | 9.84 | 9.84 | 9.84 |
| Average | 0.47 | 0.40 | 0.39 | 0.85 | 0.73 | 0.73 | 0.78 | 0.68 | 0.68 |
| Median | 0.32 | 0.27 | 0.25 | 0.50 | 0.41 | 0.40 | 0.47 | 0.38 | 0.36 |
| Standard Dev. | 0.51 | 0.48 | 0.48 | 1.02 | 0.97 | 0.97 | 0.96 | 0.91 | 0.91 |
| Coeff. of Var. | 1.09 | 1.20 | 1.22 | 1.20 | 1.32 | 1.33 | 1.23 | 1.35 | 1.35 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 32 | 32 | 32 | 79 | 79 | 79 |
| \% Detected | 18.75 |  |  | 87.34 |  |  |
| Minimum | 0.07 | 0.02 | 0.01 | 0.08 | 0.08 | 0.08 |
| Maximum | 1.80 | 1.80 | 1.80 | 11.87 | 11.87 | 11.87 |
| Average | 0.64 | 0.27 | 0.26 | 1.73 | 1.53 | 1.52 |
| Median | 0.18 | 0.25 | 0.25 | 1.07 | 0.90 | 0.90 |
| Standard Dev. | 0.79 | 0.38 | 0.38 | 2.24 | 2.16 | 2.16 |
| Coeff. of Var. | 1.24 | 1.43 | 1.44 | 1.30 | 1.41 | 1.42 |

The probability plots showed that replacing the non-detected values by half of the detection limit resulted in lower values than if the Cohen's method was used. The Anderson Darling statistic for normality increased when the censored data was estimated, indicating a better fit to a normal distribution. Figure D14 shows the probability plot for ammonia for commercial land use areas. In open space areas, the estimated values don't seem to fit the log normal distribution (Figure D15). Estimating the censored observations using Cohen's method when more than 80\% of the observations were below the detection limit is certainly not recommended.


Figure D14. Estimated ammonia distributions in commercial land use areas


Figure D15. Estimated Ammonia distributions in open space land use areas

## Nitrite and Nitrate ( $\mathrm{NO}_{2}+\mathrm{NO}_{3}$ )

The percentages of non-detected values was smaller than $5 \%$ in all the land uses for nitrites plus nitrates, except for open space areas where the level of censored values was higher than $15 \%$. There were no significant differences in the means, standard deviations and coefficients of variation, except for the open space data set, when the alternative substitution methods were used.

Table D8. Summary Statistics for Estimated Observations for $\mathrm{NO}_{\mathbf{2}} \mathbf{+} \mathrm{NO}_{\mathbf{3}}(\mathrm{mg} / \mathrm{L})$

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 927 | 927 | 927 | 425 | 425 | 425 | 417 | 417 | 417 |
| \% Detected | 97.41 |  |  | 98.12 |  |  | 96.16 |  |  |
| Minimum | 0.01 | 0.01 | 0.01 | 0.03 | 0.02 | 0.01 | 0.02 | 0.02 | 0.01 |
| Maximum | 18.00 | 18.00 | 18.00 | 8.21 | 8.21 | 8.21 | 8.40 | 8.40 | 8.40 |
| Average | 0.76 | 0.75 | 0.74 | 0.86 | 0.85 | 0.85 | 0.98 | 0.95 | 0.94 |
| Median | 0.59 | 0.58 | 0.58 | 0.61 | 0.60 | 0.60 | 0.73 | 0.72 | 0.70 |
| Standard Dev. | 0.87 | 0.86 | 0.86 | 0.91 | 0.91 | 0.91 | 0.87 | 0.86 | 0.86 |
| Coeff. of Var. | 1.14 | 1.15 | 1.16 | 1.06 | 1.08 | 1.08 | 0.89 | 0.91 | 0.91 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 44 | 44 | 44 | 25 | 25 | 25 |
| \% Detected | 84.09 |  |  | 96.00 |  |  |
| Minimum | 0.09 | 0.02 | 0.05 | 0.10 | 0.10 | 0.10 |
| Maximum | 3.33 | 3.33 | 3.33 | 3.00 | 3.00 | 3.00 |
| Average | 0.99 | 0.84 | 0.84 | 0.51 | 0.50 | 0.50 |
| Median | 0.59 | 0.50 | 0.50 | 0.28 | 0.28 | 0.26 |
| Standard Dev. | 0.88 | 0.88 | 0.88 | 0.63 | 0.62 | 0.62 |
| Coeff. of Var. | 0.89 | 1.04 | 1.04 | 1.23 | 1.23 | 1.25 |

The probability plots for residential, commercial and industrial land use areas show a different trend for the lower tail of the distribution up to the $10^{\text {th }}$ percentile for the different methods. The departures from normality are more evident in the case when the censored observations are replaced by half of the detection limit (Figure D16). In open space areas, when the censored data was estimated or replaced, the coefficient of variation increased almost $17 \%$ due the elevated level of censoring (Figure D17). There were no observed differences in the means, standard deviations and coefficients of variation when the censored values were replaced by half of the detection limit or estimated using Cohen's method.


Figure D16. Estimated nitrate - nitrite distributions in commercial land use areas


Figure D17. Estimated nitrate - nitrite distributions in open space land use areas

## Total Kjeldahl Nitrogen (TKN)

The level of censoring for TKN was smaller than $4 \%$ for all land use areas except for open space areas. The highest TKN concentrations were observed in freeway areas, and the lowest TKN concentrations were observed in open space areas (Table D9). Large changes in the coefficient of variation were observed in open space areas when using Cohen's method (an increase of $15 \%$ ) and when replacing the censored values by half of the detection limit (increases of $22 \%$ ).

Table D9. Summary Statistics for Estimated Observations for TKN (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 957 | 957 | 957 | 449 | 449 | 449 | 439 | 439 | 439 |
| \% Detected | 96.76 |  |  | 97.33 |  |  | 95.90 |  |  |
| Minimum | 0.05 | 0.00 | 0.01 | 0.05 | 0.02 | 0.01 | 0.05 | 0.01 | 0.01 |
| Maximum | 36.00 | 36.00 | 36.00 | 15.00 | 15.00 | 15.00 | 25.00 | 25.00 | 25.00 |
| Average | 1.96 | 1.91 | 1.90 | 2.23 | 2.18 | 2.17 | 2.23 | 2.17 | 2.16 |
| Median | 1.43 | 1.40 | 1.40 | 1.59 | 1.55 | 1.55 | 1.40 | 1.37 | 1.37 |
| Standard Dev. | 2.05 | 2.04 | 2.04 | 2.08 | 2.07 | 2.08 | 2.56 | 2.53 | 2.54 |
| Coeff. of Var. | 1.05 | 1.07 | 1.07 | 0.93 | 0.95 | 0.96 | 1.15 | 1.17 | 1.18 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 45 | 45 | 45 | 125 | 125 | 125 |
| \% Detected | 71.11 |  |  | 96.80 |  |  |
| Minimum | 0.20 | 0.20 | 0.20 | 0.20 | 0.19 | 0.05 |
| Maximum | 4.70 | 4.70 | 4.70 | 36.15 | 36.15 | 36.15 |
| Average | 1.35 | 1.08 | 1.03 | 3.29 | 3.20 | 3.19 |
| Median | 0.74 | 0.50 | 0.50 | 2.00 | 1.93 | 1.93 |
| Standard Dev. | 1.20 | 1.10 | 1.13 | 4.49 | 4.44 | 4.45 |
| Coeff. of Var. | 0.89 | 1.02 | 1.09 | 1.37 | 1.39 | 1.39 |

The lognormal probability plot follows a straight line, except for the lower tail up to the 5th percentile (Figure D18). The effect on the Anderson Darling statistic is increased when the censored data is estimated. The effect is higher when the non-detected values are replaced by half of the detection limit, instead of being estimated using Cohen's maximum likelihood estimator. In open space areas when the level of censoring is elevated and the number of observations is low, the Cohen's estimated method did not follow a lognormal distribution. In Figure A19, two groups seem to exist, but it is important to mention that more than $44 \%$ of the total TKN observations were lower than $0.5 \mathrm{mg} / \mathrm{L}$. All the censored values in this land use were located at $0.5 \mathrm{mg} / \mathrm{L}$ TKN.


Figure D18. Estimated TKN distributions in residential land use areas


Figure D19. Estimated TKN distributions in open space land use areas

## Total Phosphorus

Total phosphorus has low level of censored observations (less than 5\%) at all land use areas, except for open space (where it is close to $15 \%$ ) (Table D10). Variations in the coefficient of variation were not significant, except in open space areas where ignoring the censored observations reduces the coefficient of variation by almost $7 \%$.

Table D10. Summary Statistics for Estimated Observations for Total Phosphorus (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 963 | 963 | 963 | 446 | 446 | 446 | 434 | 434 | 434 |
| \% Detected | 96.88 |  |  | 95.74 |  |  | 95.85 |  |  |
| Minimum | 0.01 | 0.00 | 0.01 | 0.02 | 0.01 | 0.01 | 0.02 | 0.01 | 0.01 |
| Maximum | 6.90 | 6.90 | 6.90 | 3.35 | 3.35 | 3.35 | 7.90 | 7.90 | 7.90 |
| Average | 0.42 | 0.41 | 0.41 | 0.35 | 0.34 | 0.34 | 0.46 | 0.45 | 0.45 |
| Median | 0.30 | 0.30 | 0.30 | 0.22 | 0.22 | 0.22 | 0.26 | 0.25 | 0.25 |
| Standard Dev. | 0.47 | 0.47 | 0.47 | 0.40 | 0.39 | 0.39 | 0.64 | 0.63 | 0.63 |
| Coeff. of Var. | 1.13 | 1.14 | 1.14 | 1.16 | 1.16 | 1.16 | 1.39 | 1.41 | 1.40 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 46 | 46 | 46 | 128 | 128 | 128 |
| \% Detected | 84.78 |  |  | 99.22 |  |  |
| Minimum | 0.02 | 0.01 | 0.01 | 0.06 | 0.05 | 0.02 |
| Maximum | 15.40 | 15.40 | 15.40 | 7.19 | 7.19 | 7.19 |
| Average | 0.68 | 0.59 | 0.60 | 0.43 | 0.43 | 0.43 |
| Median | 0.31 | 0.22 | 0.25 | 0.25 | 0.25 | 0.25 |
| Standard Dev. | 2.43 | 2.24 | 2.24 | 0.76 | 0.76 | 0.76 |
| Coeff. of Var. | 3.54 | 3.77 | 3.74 | 1.76 | 1.77 | 1.77 |

When the censored data is ignored, the observations followed a lognormal distribution. However, if the non-detected values are replaced by half of the detection limit or estimated using the Cohen method, the lower tail has lower values than expected.

There is an unusual observation 20 times higher than the second highest observation for the open space data (Figure D20). The most frequent non-detected observation was $<0.5 \mathrm{mg} / \mathrm{L}$. Replacing the censored observations by half of the detection limit produces values smaller than those estimated by Cohen's method. In the freeway plot, it was observed that the higher observations are higher than the lognormal trend. The upper 20th percentile has a different slope than the remaining observations shown on the distribution.


Figure D20. Estimated total phosphorus distributions in open space land use areas

## Dissolved Phosphorus

Dissolved phosphorus has a large amount of non-detected values in all the land use areas (about 13 to 20\%), except for freeways where only $5 \%$ of the observations were censored. In general, ignoring the non-detected values increased the means and standard deviations and reduced the coefficients of variation. Table D11 shows the descriptive statistics for dissolved phosphorus.

Table D11. Summary Statistics for Estimated Observations Dissolved Phosphorus (mg/L)

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 738 | 738 | 738 | 323 | 323 | 323 | 325 | 325 | 325 |
| \% Detected | 84.15 |  |  | 81.11 |  |  | 87.38 |  |  |
| Minimum | 0.009 | 0.001 | 0.005 | 0.01 | 0.00 | 0.01 | 0.003 | 0.003 | 0.003 |
| Maximum | 1.69 | 1.69 | 1.69 | 1.60 | 1.60 | 1.60 | 1.60 | 1.60 | 1.60 |
| Average | 0.23 | 0.20 | 0.20 | 0.21 | 0.18 | 0.19 | 0.17 | 0.16 | 0.16 |
| Median | 0.17 | 0.14 | 0.14 | 0.11 | 0.09 | 0.09 | 0.11 | 0.10 | 0.10 |
| Standard Dev. | 0.21 | 0.21 | 0.21 | 0.27 | 0.25 | 0.25 | 0.20 | 0.19 | 0.19 |
| Coeff. of Var. | 0.94 | 1.04 | 1.05 | 1.24 | 1.35 | 1.34 | 1.18 | 1.23 | 1.23 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 44 | 44 | 44 | 22 | 22 | 22 |
| \% Detected | 79.55 |  |  | 95.45 |  |  |
| Minimum | 0.010 | 0.003 | 0.005 | 0.06 | 0.01 | 0.01 |
| Maximum | 0.52 | 0.52 | 0.52 | 6.97 | 6.97 | 6.97 |
| Average | 0.18 | 0.16 | 0.17 | 0.78 | 0.75 | 0.75 |
| Median | 0.13 | 0.09 | 0.14 | 0.20 | 0.20 | 0.20 |
| Standard Dev. | 0.16 | 0.15 | 0.15 | 1.66 | 1.63 | 1.63 |
| Coeff. of Var. | 0.89 | 0.95 | 0.87 | 2.13 | 2.18 | 2.18 |

As in the previous cases, ignoring the censored observations results in larger mean values. There were no observed practical differences between the maximum likelihood method and replacing the non-detected values with half of the detection limit (Figure D21). , Dissolved phosphorus had the lowest level of censoring at freeway sites. The probability plot indicates that the distribution is heavy in the tails; the slope between the 20th and 60th percentiles is higher than in the tails (Figure D22).


Figure D21. Estimated dissolved phosphorus distributions in industrial land use areas


Figure D22. Estimated dissolved phosphorus distributions in freeways land use areas

## Total Cooper (Cu)

The levels of censoring for copper vary from 1 to $15 \%$ among the different land uses. When the non-detected values are estimated or replaced by half of the detection limit, the coefficients of variation increased between $1 \%$ and $6 \%$, in addition there is a reduction in the means and standard deviations. Table D12 shows the descriptive statistics for each method by land use.

Table D12. Summary Statistics for Estimated Observations for Total Cooper ( $\mu \mathrm{g} / \mathrm{L}$ )

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 799 | 799 | 799 | 387 | 387 | 387 | 415 | 415 | 415 |
| \% Detected | 83.60 |  |  | 92.76 |  |  | 89.64 |  |  |
| Minimum | 1.00 | 0.25 | 0.23 | 1.50 | 1.50 | 1.00 | 1.97 | 1.77 | 1.00 |
| Maximum | 590 | 590 | 590 | 384 | 384 | 384 | 1360 | 1360 | 1360 |
| Average | 21.06 | 18.54 | 18.51 | 29.02 | 27.47 | 27.30 | 47.00 | 43.37 | 42.98 |
| Median | 12.00 | 10.00 | 10.00 | 17.00 | 15.60 | 15.00 | 21.88 | 20.00 | 20.00 |
| Standard Dev. | 38.51 | 35.70 | 35.69 | 42.92 | 41.73 | 41.79 | 93.81 | 89.47 | 89.60 |
| Coeff. of Var. | 1.83 | 1.93 | 1.93 | 1.48 | 1.52 | 1.53 | 2.00 | 2.06 | 2.08 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 39 | 39 | 39 | 97 | 97 | 97 |
| \% Detected | 74.36 |  |  | 98.97 |  |  |
| Minimum | 2.00 | 2.00 | 2.00 | 5.00 | 5.00 | 5.00 |
| Maximum | 210 | 210 | 210 | 244 | 244 | 244 |
| Average | 19.15 | 15.79 | 15.65 | 48.29 | 47.86 | 47.85 |
| Median | 10.00 | 5.30 | 5.00 | 34.70 | 33.40 | 33.40 |
| Standard Dev. | 38.97 | 33.98 | 34.00 | 45.91 | 45.87 | 45.89 |
| Coeff. of Var. | 2.04 | 2.15 | 2.17 | 0.95 | 0.96 | 0.96 |

The lognormal probability plots for residential and commercial land use areas indicate that the upper 5th percentile of the copper concentrations have higher values than expected if the distribution was lognormal. This observation is important because the upper tail of the distribution has an important effect in the mean and standard deviation values of the dataset.

In open space areas, replacing the non-detected values by the Cohen's method or replacing the non-detected values by half of the detection limit, reduce the means and standard deviations of the distribution by $18 \%$ and $13 \%$, respectively. The probability plot for freeway areas is almost a perfect lognormal trend. In this case, the level of non-detected values was only $1 \%$, and the difference in the coefficients of variations was also $1 \%$.

## Lognormal Probability Plot for Total Cooper in Open Space



Figure D23. Estimated total cooper distributions in open space land use areas

## Total Lead

The level of non-detected values for lead varied from 0 to $58 \%$. All the observations at the freeway sites indicate a presence of lead, in addition to the highest concentration among the land uses. Open land use areas had the highest level of non-detected lead values. There was about a $10 \%$ reduction in the coefficient of variation when the censored data were ignored. Table D13 shows the descriptive statistics for each method.

The probability plots indicate that when replacing the censored data by half of the detection limit, the values are smaller than when using Cohen's method (Figure D24). Estimating the censored values reduces the Anderson Darling statistic. In open space areas, most of the censored values were observed at $<40 \mathrm{mg} / \mathrm{L},<50 \mathrm{mg} / \mathrm{L}$ and $<100$ $\mathrm{mg} / \mathrm{L}$. In all land use areas, almost $80 \%$ of the lead observations were smaller than $50 \mathrm{mg} / \mathrm{L}$. In open space areas, the estimated means, standard deviations and coefficients of variation are dubious because most of the censored observations were located in the upper part of the distribution (the frequency of non-detectable observations was quite high, at about $58 \%$ ).

Table D13. Summary Statistics for Estimated Observations for Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ )

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 788 | 723 | 788 | 377 | 355 | 377 | 411 | 377 | 411 |
| \% Detected | 71.32 |  |  | 85.41 |  |  | 76.40 |  |  |
| Minimum | 0.50 | 0.03 | 0.10 | 1.00 | 0.21 | 0.35 | 1.00 | 0.21 | 0.50 |
| Maximum | 585 | 585 | 585 | 689.07 | 689.07 | 689.07 | 1200 | 1200 | 1200 |
| Average | 26.00 | 21.03 | 22.08 | 37.42 | 34.27 | 33.84 | 70.10 | 59.52 | 57.49 |
| Median | 12.00 | 8.20 | 10.00 | 18.00 | 17.00 | 17.00 | 25.00 | 20.00 | 20.00 |
| Standard Dev. | 48.98 | 44.21 | 43.17 | 59.53 | 57.56 | 56.07 | 128.57 | 119.79 | 115.57 |
| Coeff. of Var. | 1.88 | 2.10 | 1.96 | 1.59 | 1.68 | 1.66 | 1.83 | 2.01 | 2.01 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 45 | 29 | 45 | 107 | 107 | 107 |
| \% Detected | 42.22 |  |  | 100.00 |  |  |
| Minimum | 0.20 | 0.08 | 0.10 | 1.60 | 1.60 | 1.60 |
| Maximum | 150 | 150 | 150 | 450 | 450 | 450 |
| Average | 28.39 | 19.21 | 23.98 | 48.77 | 48.77 | 48.77 |
| Median | 10.00 | 3.16 | 10.00 | 25.00 | 25.00 | 25.00 |
| Standard Dev. | 47.36 | 40.10 | 33.70 | 70.74 | 70.74 | 70.74 |
| Coeff. of Var. | 1.67 | 2.09 | 1.41 | 1.45 | 1.45 | 1.45 |

## Total Zinc

The percentage of non-detected zinc values was smaller than $4 \%$, except for open space areas where it was close to $30 \%$ (Table D14). No important changes in the coefficient of variations were observed, except for open space areas where ignoring the censored values reduced the coefficients of variation by $13 \%$.

Lognormal Probability Plot for Total Lead in Industrial Land Use


Figure D24. Estimated total lead distributions in industrial land use areas

Table D14. Summary Statistics for Estimated Observations for Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ )

|  | RESIDENTIAL |  |  | COMMERCIAL |  |  | INDUSTRIAL |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 810 | 810 | 810 | 392 | 392 | 392 | 432 | 432 | 432 |
| \% Detected | 96.42 |  |  | 98.98 |  |  | 98.61 |  |  |
| Minimum | 3.00 | 0.48 | 0.30 | 5.00 | 5.00 | 5.00 | 5.77 | 3.05 | 2.00 |
| Maximum | 1580 | 1580 | 1580 | 3050 | 3050 | 3050 | 8100 | 8100 | 8100 |
| Average | 116.70 | 113.53 | 113.23 | 225.32 | 224.06 | 223.55 | 318.25 | 315.02 | 314.34 |
| Median | 73.00 | 70.00 | 70.00 | 150.00 | 150.00 | 150.00 | 209.50 | 204.50 | 201.00 |
| Standard Dev. | 151.81 | 150.25 | 150.24 | 275.81 | 274.74 | 274.96 | 474.36 | 471.89 | 472.21 |
| Coeff. of Var. | 1.30 | 1.32 | 1.33 | 1.22 | 1.23 | 1.23 | 1.49 | 1.50 | 1.50 |


|  | OPEN SPACE |  |  | FREEWAY |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Land use | Ignore | Estimate | HD | Ignore | Estimate | HD |
| Observations | 45 | 45 | 45 | 93 | 93 | 93 |
| \% Detected | 71.11 |  |  | 96.77 |  |  |
| Minimum | 5.00 | 2.00 | 2.50 | 6.00 | 6.00 | 2.50 |
| Maximum | 390 | 390 | 390 | 1829 | 1829 | 1829 |
| Average | 72.44 | 55.90 | 55.62 | 279.43 | 271.63 | 271.52 |
| Median | 40.00 | 20.00 | 20.00 | 200.00 | 194.49 | 194.49 |
| Standard Dev. | 96.88 | 85.85 | 85.99 | 281.16 | 279.87 | 279.98 |
| Coeff. of Var. | 1.34 | 1.54 | 1.55 | 1.01 | 1.03 | 1.03 |

The probability plot indicates that in the lower tail, replacing the non-detected observations by half of the detection limit will create smaller values than when estimating them using Cohen's method (Figure D25). In open space areas, if the censored data are estimated using Cohen's method, there is a reduction in the mean and variance of the dataset of $23 \%$ and $12 \%$, respectively, however the coefficients of variation increased by $15 \%$ (Figure D26).

## Lognormal Probability Plot for Total Zinc in Residential Land Use <br> ML Estimates



NON DETECTED

- IGNORE
- estimate
- HALF DETECTION

Goodness of Fit AD*
2.524 5.012 5.698

Figure D25. Estimated total zinc distributions in residential land use areas

Lognormal Probability Plot for Total Zinc in Open Space Land Use
ML Estimates


Figure D26. Estimated total zinc distributions in open space land use areas

## Sites with Unusual TSS Concentrations for Different Land Uses

This section presents the continuation of the example presented in Chapter 3, where sites having unusual conditions were identified and examined more carefully to try to understand the reasons for these values. A similar procedure was followed in this appendix for the commercial, industrial and mixed land use areas to complement the Chapter 3 analyses, which were conducted for residential areas only.

## Residential and Mixed Residential Locations

The box and whisker plot (Figure D27 shows TSS concentrations by rain zone and location) indicates that there is only one site that seems to have a different TSS concentration probability distribution compared to the remaining sites in this group. The site of interest is located in a residential-commercial area in Wooden Bridge Run, Philadelphia (PAPH1051), and has much lower concentrations that the other sites. Only two samples were collected at this site, and both were below $15 \mathrm{mg} / \mathrm{L}$. The few samples available reduce the significance of this observation, however.

The results from the Xbar S chart analyses for mixed residential land uses are presented in Table D15. These analyses consider the numbers of samples and the variability of the data from each site, compared to the complete data set in the category being examined.

Table D15. Sites failing Xbar and S Chart Tests in Mixed Residential Land Use Areas

| EPA Rain Zone | Sites Failing Xbar Chart Test | Sites Failing S Chart Test |
| :---: | :---: | :---: |
| ALL | 9COCSA004(H) 2NCFVROSE (L) 7ORPOA005 (H) <br> 2TNKXTYGV(H) 5TXFWA005(H) 2VAVBTYV5(L) | GAFUCOS3 (H) |
| 1 | None | None |
| 2 | TNKXTYGV (H) VAVBTYV5 (L) | None |
| 3 | None | None |
| 4 | None | None |
| 5 | None | None |
| 6 | None | None |
| 7 | None | None |
| 8 | None | None |
| 9 | None | None |

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When 40 mixed residential sites were examined, only six sites were designated as being "out of control." These sites had unusual concentrations that were outside a band described by three standard deviations from the mean values. Two sites with means (log) below the lower control limit were located in EPA Rain Zone 2 (North Carolina and Virginia). The site located at Long Island Creek in Fulton County, Georgia, has the largest standard deviation among the mixed residential sites examined. However, the S chart indicates that this site is in control compared to other sites in EPA Rain Zone 3.

There are 21 sites located in mixed residential land use areas and in EPA Rain Zone 2, 17 sites have more than one observation each. Two sites, one above the upper control limit and one below the control limit, were observed in the Xbar chart. The site with the high median (log) value is located in Gallaher view, Knoxville, Tennessee (TNKXTYGV, 38 observations, median TSS $=105 \mathrm{mg} / \mathrm{L}$ ). This site information included construction activity in the north part of the watershed, and a self-storage business, north and east of Cedar Hills apartments. The site located in Holland road, Virginia Beach (VAVBTYV5, 26 observations, median TSS $=32 \mathrm{mg} / \mathrm{L}$ ) has wet ponds in the watershed that seem to control high concentrations, but the average value is the same as the other mixed residential sites.

The ANOVA analyses indicate that there is at least one EPA Rain Zone with TSS concentrations different than the other EPA Rain Zone with a p-value smaller than $1 \%$. The Dunnett's comparison test at a family error rate of $5 \%$ indicates that EPA Rain Zones 5 and 7 have higher concentrations than those observed in EPA Rain Zone 2. In summary, at a family error rate of $5 \%$, higher concentrations occurred in EPA Rain Zones 5 (six sites, median TSS $=108 \mathrm{mg} / \mathrm{L}$ ) and 7 (two sites, TSS $=175 \mathrm{mg} / \mathrm{L}$ ) compared with EPA Rain Zone $2(21$ sites, TSS = $59 \mathrm{mg} / \mathrm{L})$. The Kurskal-Wallis test indicates that there is a significant difference in the TSS median concentrations (with a pvalue close to zero). Site TNKXTYGV has higher characteristics than the other residential mixed sites, most likely due to the noted construction activity close to the outfall location.

## Commercial and Mixed Commercial Locations

Box plots Xbar and S charts and ANOVA tests were used for commercial land use data. Figure D28 identifies a site with high TSS concentration in EPA Rain Zone 4 (KATOJACK, 15 observations, median TSS $=603 \mathrm{mg} / \mathrm{L}$ ). In general, it seems that sites in EPA Rain Zone 7 and 9 have higher concentrations than the other EPA Rain Zones. No other trend or variation among EPA Rain Zone was identified from the box plot.
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EPA Rain Zone and Location_ID
Figure D28. Box and whiskers plots for TSS concentrations at commercial land use areas

The second approach was to identify unusual sites by EPA Rain Zone using Xbar S charts. The results for commercial land uses are presented in Table D16.

Table D16. Sites failing Xbar and S Chart Tests in Commercial Land Use Areas

| EPA Rain Zone | Sites Failing Xbar Chart Test | Sites Failing S Chart Test |
| :---: | :---: | :---: |
| ALL | 2MDPGCOS1 (H) 2VACHCCC4 (L) 3ALHUWERP (L) <br> 4KATOJACK (H) 4TXHCA005 (L) 4TXHOA004 (L) <br> 9KAWITOWN (H)   | None |
| 1 | None | None |
| 2 | 2MDPGCOS1 (H) 2VACHCCC4 (L) | None |
| 3 | None | None |
| 4 | 4KATOJACK ( H ) | None |
| 5 | None | None |
| 6 | None | None |
| 7 | None | None |
| 9 | None | None |

The Xbar S plot did not indicate any trend by geographical region for the 45 sites. Sites with low concentrations were observed in EPA Rain Zones 2, 3 and 4. There were three sites identified with concentrations above the control limit, one in EPA Rain Zone 2, another in EPA Rain Zone 9, and a site identified by the box plot located in EPA Rain Zone 4.

In EPA Rain Zone 2, two sites were found outside the control limits. MDPGCOS1 is located in a shopping center in Arena Plaza, Price Georges County, Maryland. 26 samples were collected at this location. The median TSS concentration for this site is $158 \mathrm{mg} / \mathrm{L}$. No reason was given for the high observed TSS concentrations. The second site is located at Clover Leaf Mall in Chesterfield County, Virginia (VACHCCC4, 12 observations, 60 acres, median $\mathrm{TSS}=14 \mathrm{mg} / \mathrm{L}$ ). There is no clear reason that explains the low concentrations found at this location. No sites outside the control limits were found in other EPA Rain Zones except for EPA Rain Zone 4. This outfall is located in Jackson Street in Topeka, Kansas. The high TSS concentrations may have been affected by tracking of sediment from a sand quarry close to the watershed. There were collected 16 samples collected between April 1998 and Septembers 2002.

The ANOVA test indicated that there was a significant difference among EPA Rain Zones ( P -value $=0$ ). The Dunnett's comparison test, with a family error of $5 \%$, indicates that TSS concentrations compared with EPA Rain Zone $2($ median TSS $=48 \mathrm{mg} / \mathrm{L}$ ) are larger in EPA Rain Zones 4 (median TSS $=82 \mathrm{mg} / \mathrm{L}$ ) and $9($ median TSS $=128$ $\mathrm{mg} / \mathrm{L}$ ). The median TSS concentrations at the remaining EPA Rain Zones are not statistically different than those observed in EPA Rain Zone 2.

There are 24 sites located in mixed commercial land use areas with more than one observation. EPA Rain Zone 2 has the largest number of sites ( 10 sites), followed by EPA Rain Zone 5 ( 5 sites). Figure D29 shows the box plots for mixed commercial land uses by EPA Rain Zone.


The box plot indicates that there is a mixed commercial site located in Plano, Texas, and a site in Colorado with higher concentrations than the other sites in this category. Because the low number of sites sampled by geographical region, it is not possible to identify any trend by EPA Rain Zone. Table D17 lists those sites outside the control limits by EPA Rain Zone for all of the mixed commercial sites.

| EPA Rain Zone | Sites Failing Xbar Chart Test |  | Sites Failing S Chart Test |
| :---: | :---: | :---: | :---: |
| ALL | 2TNKXTYWE (L) 5TXPLA004 (H) | $\begin{aligned} & \text { 2VANFTYN1 (H) 2VAVBTYV3(L) } \\ & \text { 9COCSA001 (H) } \end{aligned}$ | None |
| 1 | None |  | None |
| 2 | 2VANFTYN1 ( H ) | 2VAVBTYV3(L) | None |
| 3 | None |  | None |
| 4 | None |  | None |
| 5 | 5TXPLA004 (H) |  | None |
| 6 | None |  | None |
| 7 | None |  | None |
| 9 | None |  | None |

The Xbar chart for all mixed commercial observations indicates that sites with high TSS concentrations occurred in EPA Rain Zones 5 and 9. In EPA Rain Zone 2, three sites were outside of the control limits, two below the lower control limit and one above the upper control limit. As in the commercial site analyses, EPA Rain Zone 9 seems to have higher TSS concentrations than the other EPA Rain Zones.

The analysis by EPA Rain Zone indicates that only EPA Rain Zones 2 and 5 have sites outside the control limits. In EPA Rain Zone 2, the site with high concentrations (VANFTYN1) is located at Armistead Avenue in Norfolk, Virginia. A total of 28 observations were collected at this site. The median TSS for this location was $117 \mathrm{mg} / \mathrm{L}$. The site having unusually low median TSS concentration was at Haygood, Virginia Beach, Virginia (VAVBTYV3). A total of 33 storms were sampled at this site. The median TSS concentration at this location was $26 \mathrm{mg} / \mathrm{L}$. This site is $79 \%$ commercial and $13 \%$ open space. The site having unusually high TSS concentrations in EPA Rain Zone 5 is located at Spring Creek, Plano, Texas (TXPLA004). There are 7 events from this site in the database. The median TSS concentration is $575 \mathrm{mg} / \mathrm{L}$. No information was found to explain the elevated concentrations. Another site that appears to be outside the control limits compared to all the sites, but not in its group. It is located in Sixteenth Hole Valley, Colorado Springs, Colorado. The median concentration for this site was $251 \mathrm{mg} / \mathrm{L}$. This site has two automobile dealerships and a gas station, along with evidence of erosion observed in the aerial photograph. The ANOVA analysis indicates that there are significant differences among EPA Rain Zones (P-value $=0$ ) in mixed commercial land uses. The Dunnett's comparison test, with a family error of $5 \%$, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS $=46 \mathrm{mg} / \mathrm{L}$ ) are larger in EPA Rain Zones 5 (median TSS $=72 \mathrm{mg} / \mathrm{L}$ ) and 9 (median TSS $=254 \mathrm{mg} / \mathrm{L}$ ). The median TSS values in the remaining EPA Rain Zones are not statistically different than those observed in EPA Rain Zone 2.

## Industrial and Mixed Industrial Locations

Box plots, Xbar, S charts, and ANOVA tests were used to examine the observations from sites located in industrial land use areas. Figure D30 shows the box plots by EPA Rain Zone and location. Sites located in EPA Rain Zones 6 and 9 seem to have higher concentrations than the remaining industrial sites. A site with two unusually low concentrations was located in Boston, Massachusetts.


Figure D30. Box and whiskers plots for TSS concentrations at industrial land use areas

Table D18 shows those industrial sites that are outside the control limits of the pooled dataset and by each EPA Rain Zone.

Table D18. Sites failing Xbar and S Chart Tests in Industrial Land Use Areas

| EPA Rain Zone | Sites Failing Xbar Chart Test |  |  | Sites Failing S Chart Test |
| :---: | :---: | :---: | :---: | :---: |
| ALL | 1MABOA004 (L) | 2VACPTYC5 (L) | 2VAVBTYV4 (L) | None |
|  | 3GAATAT01 (L) | 5TXFWA004 (H) | 6AZMCA003 (H) |  |
|  | 6AZTUA004 (H) |  |  |  |
| 1 | None |  |  | None |
| 2 | MDPGCOS 6 (H) <br> VACPTYC5 (L) <br> VAVBTYV4 (L) |  |  | None |
|  |  |  |  |  |
|  |  |  |  |  |
| 3 | None |  |  | None |
| 4 | None |  |  | None |
| 5 | TXFWA004 (H) |  |  | None |
| 6 | AZMCA003 (H) |  |  | None |
| 7 | None |  |  | None |
| 9 | None |  |  | None |

As in the other land uses, sites with concentrations below the control limit were observed in EPA Rain Zones 1, 2 and 3. Sites with median concentrations larger than the upper control limit were located in EPA Rain Zones 5 and 6. Three sites were outside the control limits in EPA Rain Zone 2, one in EPA Rain Zone 5, and one in EPA Rain Zone 6. The two sites in EPA Rain Zone 2 with low concentrations were located in Virginia, and the site with high concentrations was located in Maryland. One of the sites located in Virginia was located in Cavalier Industrial Park in the city of Chesapeake (VACPTYC5). This 16 acres site is $92 \%$ industrial, with the remaining $8 \%$ open space. A total of 15 samples were collected from this site during the period 1993 to 1999. The median TSS concentration for this site is $13 \mathrm{mg} / \mathrm{L}$. No additional information was observed in the aerial photos that might explain the low concentrations.

The second site was located in Viking Drive, Virginia Beach (VAVBTYV5). This 29-acre site was comprised of 55 percent impervious surfaces. There are 30 samples from this site in the database. The samples were collected between 1992 and 1999. The median TSS concentration is $29 \mathrm{mg} / \mathrm{L}$.

The site with elevated concentrations in EPA Rain Zone 2 is located in Pennsy Drive in Riverdale, Prince George County, Maryland (MDPGCOS6). This 42.4-acre size site has a grass swale drainage system. There are 30 samples in the database from this location. The samples were collected between 1994 and 1997. The median TSS concentration is $98 \mathrm{mg} / \mathrm{L}$. The site is located next to Glenridge Elementary School. The aerial photo shows construction activity in the northwest part of the watershed.

The site with high TSS concentrations in EPA Rain Zone 5 is located at Dry Branch, in Fort Worth, Texas (TXFWA004). A total of 21 samples were obtained at this site. The median TSS for this location is $288 \mathrm{mg} / \mathrm{L}$. Several bare ground open space areas were observed in the aerial photograph. The site located in EPA Rain Zone 6 is at $27^{\text {th }}$ Avenue at Salt River in Maricopa County, Arizona (AZMCA003). There are 27 samples from this location. The median TSS concentration is $660 \mathrm{mg} / \mathrm{L}$. The scarce vegetation and the type of soils may be the reason of this elevated median value.

The ANOVA analysis indicates that there are significant differences among EPA Rain Zones $(\mathrm{P}$-value $=0)$ for industrial land uses. The Dunnett's comparison test with a family error of 5\%, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS $=53 \mathrm{mg} / \mathrm{L}$ ) are larger for EPA Rain Zones 4 (median TSS = 92 $\mathrm{mg} / \mathrm{L}), 5($ median TSS $=147 \mathrm{mg} / \mathrm{L}), 6($ median $\mathrm{TSS}=288 \mathrm{mg} / \mathrm{L}), 7($ median TSS $=120 \mathrm{mg} / \mathrm{L})$, and $9($ median TSS
$=170 \mathrm{mg} / \mathrm{L})$. The median TSS concentrations in EPA Rain Zones 1 and 3 are not statistically different from those observed in EPA Rain Zone 2.

The box plots in mixed industrial land uses are shown in Figure D31. Most of the box plots have the same median except for those located in EPA Rain Zone 9. The sites that fail the quality control charts are shown in Table D19. Three sites are outside the control limits for mixed industrial land uses. Two sites in Colorado and one site in North Carolina are out of control. This result is similar to those observed in the other land uses. When each EPA Rain Zone was analyzed individually, no sites were found to be out of control.


Figure D31. Box and whiskers plots for TSS concentrations at industrial land use areas

Table D19. Sites failing Xbar and S Chart Tests in Mixed Industrial Land Use Areas

| EPA Rain Zone | Sites Failing Xbar Chart Test |  |  | Sites Failing S Chart Test |
| :---: | :---: | :---: | :---: | :---: |
| ALL | 9COCSA002 (H) | 9CODEA006 (H) | 2NCGRUNIO (L) | None |
| 2 | None |  |  | None |
| 3 | None |  |  | None |
| 5 | None |  |  | None |
| 6 | None |  |  | None |
| 7 | None |  |  | None |
| 9 | None |  |  | None |

The ANOVA analysis indicates that there are significant differences among EPA Rain Zones $(P$-value $=0)$ at mixed industrial land use sites. The Dunnett's comparison test with a family error of $5 \%$, indicates that TSS concentrations compared with EPA Rain Zone 2 (median TSS $=82 \mathrm{mg} / \mathrm{L}$ ) are larger only for EPA Rain Zone 9 (median TSS = 341 $\mathrm{mg} / \mathrm{L}$ ). The median TSS concentrations in EPA Rain Zones 3, 5, 6, and 7 are not statistically different from the median TSS concentrations found in EPA Rain Zone 2.

## Appendix E: First Flush Tables

## Description

The following table shows the summary statistic for each constituent included in the database.

Table E1. Results of Preliminary Statistical Analysis for Total Suspended Solids (TSS)

| TSS (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE Skew (Log) | Test Norm. (Log) p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 90 | 90 | 54.95 | 1.740 | 0.106 | 0.168 | 0.254 | 0.730 | 1.85 |
| Commercial First Flush | 90 | 90 | 101.86 | 2.008 | 0.200 | -0.508 | 0.254 | 0.016 |  |
| Industrial Composite | 83 | 83 | 66.07 | 1.820 | 0.186 | -0.021 | 0.264 | 0.336 | 0.97 |
| Industrial First Flush | 83 | 83 | 63.97 | 1.806 | 0.374 | -0.157 | 0.264 | 0.055 |  |
| Institutional Composite | 18 | 18 | 16.48 | 1.217 | 0.110 | -0.176 | 0.536 | 0.122 | 2.12 |
| Institutional First Flush | 18 | 18 | 34.99 | 1.544 | 0.145 | -0.164 | 0.536 | 0.846 |  |
| Open Space Composite | 32 | 32 | 21.98 | 1.342 | 0.424 | -0.526 | 0.414 | 0.511 | 0.95 |
| Open Space First Flush | 32 | 32 | 20.89 | 1.320 | 0.563 | -0.126 | 0.414 | 0.847 |  |
| Residential Composite | 144 | 144 | 37.50 | 1.574 | 0.217 | -0.033 | 0.202 | 0.282 | 1.84 |
| Residential First Flush | 144 | 144 | 69.02 | 1.839 | 0.302 | -0.267 | 0.202 | 0.533 |  |
| All Land Uses Composite | 372 | 372 | 44.36 | 1.647 | 0.226 | -0.381 | 0.126 | 0.008 | 1.60 |
| All Land Uses First Flush | 372 | 372 | 70.96 | 1.851 | 0.335 | 0.457 | 0.126 | 0 |  |

Table E2. Results of Preliminary Test Analysis for Total Suspended Solids (TSS)

| TSS (mg/L | Mann Wittn. <br> p-value | Fligner Policello | Normality for t-Test p-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :--- | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=5.345 ; \mathrm{P}=0$ | 0.014 | $\mathrm{~N} / \mathrm{A}$ | Different (first flush) |
| Industrial | 0.627 | $\mathrm{U}=0.483 ; \mathrm{P}=0.31$ | 0.222 | 0.432 | Same (no first flush) |
| Institutional | 0.007 | $\mathrm{U}=3.095 ; \mathrm{P}=0$ | 0.309 | 0.001 | Different |
| Open Space | 0.706 | $\mathrm{U}=0.39 ; \mathrm{P}=0.35$ | 0.183 | 0.614 | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.89 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=6.65 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E3. Results of Preliminary Statistical Analysis for Turbidity (NTU)

| Turbidity (NTU) | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial <br> Composite | 11 | 11 | 19.68 | 1.294 | 0.062 | -0.984 | 0.661 | 0.186 |  |
| Commercial First <br> Flush | 11 | 11 | 26.00 | 1.415 | 0.078 | 0.523 | 0.661 | 0.564 | 1.32 |
| Residential <br> Composite | 12 | 12 | 23.44 | 1.370 | 0.163 | 0.213 | 0.637 | 0.721 | 1.24 |
| Residential First Flush | 12 | 12 | 28.97 | 1.462 | 0.148 | 1.407 | 0.637 | 0.168 | 0.406 |
| All Land Uses <br> Composite | 26 | 26 | 21.73 | 1.337 | 0.109 | 0.204 | 0.456 | 0.406 |  |
| All Land Uses First <br> Flush | 26 | 26 | 27.48 | 1.439 | 0.105 | 1.197 | 0.456 | 0.108 |  |

Table E4. Results of Preliminary Test Analysis for Turbidity (NTU)

| Turbidity (NTU) | Mann Wittn. <br> $\mathbf{p}$-value | Fligner Policello | Normality <br> for $\mathbf{t - T e s t}$ <br> $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :--- | :---: | :---: | :---: |
| Commercial | 0.224 | $\mathrm{U}=1.26 ; \mathrm{P}>0.1$ | 0.652 | 0.219 | Same (no first flush) |
| Residential | 0.418 | $\mathrm{U}=0.853 ; \mathrm{P}>0.1$ | 0.240 | 0.021 | Same |
| All Land Uses | 0.124 | $\mathrm{U}=0.673 ; \mathrm{P}=0.25$ | 0.134 | 0 | Same |

Table E5. Results of Preliminary Statistical Analysis for pH

| $\mathbf{c \|} \mathbf{p H}$ | Total <br> Events | Selected <br> Cases | Median | Var <br> $(\mathbf{L o g})$ | Skew | SE <br> Skew | Test <br> Norm. p- <br> value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial <br> Composite | 17 | 17 | 7.4 | 0.368 | -0.299 | 0.550 | 0.527 | 1.03 |
| Commercial First Flush | 17 | 17 | 7.6 | 0.509 | 0.788 | 0.550 | 0.351 |  |
| Industrial Composite | 16 | 16 | 6.755 | 0.194 | 0.482 | 0.564 | 0.179 | 1.00 |
| Industrial First Flush | 16 | 16 | 6.750 | 0.388 | -0.854 | 0.564 | 0.307 |  |
| Residential Composite | 26 | 26 | 7.213 | 0.195 | -0.520 | 0.456 | 0.447 | 1.01 |
| Residential First Flush | 26 | 26 | 7.250 | 0.212 | -0.283 | 0.456 | 0.408 |  |
| All Composite | 63 | 63 | 7.2 | 0.302 | 0.102 | 0.302 | 0.562 | 1.01 |
| All First Flush | 63 | 63 | 7.3 | 0.437 | 0.036 | 0.302 | 0.110 |  |

Table E6. Results of Preliminary Test Analysis for pH

| pH | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test p-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :--- | :---: | :---: | :---: |
| Commercial | 0.208 | $\mathrm{U}=1.28 ; \mathrm{P}=0.10$ | 0.007 | $\mathrm{~N} / \mathrm{A}$ | Same |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.428 ; \mathrm{P}=0.33$ | 0.341 | 0.828 | Same |
| Residential | 0.308 | $\mathrm{U}=1.32 ; \mathrm{P}=0.09$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same |
| All Land Uses | 0.219 | $\mathrm{U}=1.68 ; \mathrm{P}=0.05$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same |

Table E7. Results of Preliminary Statistical Analysis for $\mathrm{BOD}_{5}$

| $B O D_{5}(\mathrm{mg} / \mathrm{L})$ | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 83 | 83 | 15.21 | 1.182 | 0.125 | 0.263 | 0.264 | 0.513 | 1.77 |
| Commercial First Flush | 83 | 83 | 26.98 | 1.431 | 0.153 | -0.241 | 0.264 | 0.390 |  |
| Industrial Composite | 80 | 80 | 15.14 | 1.18 | 0.188 | 0.190 | 0.269 | 0.013 | 1.58 |
| Industrial First Flush | 80 | 80 | 23.99 | 1.38 | 0.180 | -0.502 | 0.269 | 0.044 |  |
| Institutional Composite | 18 | 18 | 7.48 | 0.874 | 0.151 | -0.737 | 0.536 | 0.247 | 1.67 |
| Institutional First Flush | 18 | 18 | 12.47 | 1.096 | 0.173 | -0.732 | 0.536 | 0.281 |  |
| Open Space Composite | 28 | 28 | 3.79 | 0.579 | 0.148 | 0.523 | 0.441 | 0.242 | 1.07 |
| Open Space First Flush | 28 | 28 | 4.05 | 0.607 | 0.197 | 0.449 | 0.441 | 0.077 |  |
| Residential Composite | 133 | 133 | 12.59 | 1.100 | 0.154 | 0.314 | 0.210 | 0.137 | 1.67 |
| Residential First Flush | 133 | 133 | 20.99 | 1.322 | 0.220 | -0.150 | 0.210 | 0.010 |  |
| All Land uses Composite | 344 | 344 | 12.53 | 1.098 | 0.184 | 0.073 | 0.131 | 0.003 | 1.67 |
| All Land Uses First Flush | 344 | 344 | 20.89 | 1.320 | 0.233 | -0.385 | 0.131 | 0 |  |

Table E8. Results of Preliminary Test Analysis for $\mathrm{BOD}_{5}$

| $\mathrm{BOD}_{5}$ | Mann <br> Wittn. <br> $\mathbf{p - v a l u e ~}$ | Fligner Policello | Normality for <br> $\mathbf{t - T e s t ~ p - v a l u e ~}$ | Paired <br> $\mathbf{t}-\mathbf{T e s t}$ | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | 0 | $\mathrm{U}=4.85 ; \mathrm{P}=0$ | 0.013 | $\mathrm{~N} / \mathrm{A}$ | Different (first flush) |
| Industrial | 0.007 | $\mathrm{U}=2.76 ; \mathrm{P}=0$ | 0.434 | 0.012 | Different |
| Institutional | 0.027 | $\mathrm{U}=2.46 ; \mathrm{P}=0.01$ | 0.056 | 0.001 | Different |
| Open Space | 0.706 | $\mathrm{U}=0.39 ; \mathrm{P}=0.35$ | 0.183 | 0.614 | Same (no first flush) |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.89 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=6.65 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E9. Results of Preliminary Statistical Analysis for COD

| COD (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE Skew (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite) | 91 | 91 | 71.94 | 1.857 | 0.075 | 0.261 | 0.253 | 0.022 | 2.29 |
| Commercial First Flush | 91 | 91 | 164.82 | 2.217 | 0.119 | -0.201 | 0.253 | 0.877 |  |
| Industrial Composite | 84 | 84 | 75.34 | 1.877 | 0.100 | 0.167 | 0.263 | 0.014 | 1.43 |
| Industrial First Flush | 84 | 84 | 107.40 | 2.031 | 0.151 | -0.141 | 0.263 | 0.804 |  |
| Institutional Composite | 18 | 18 | 43.85 | 1.642 | 0.220 | -0.456 | 0.536 | 0.567 | 2.73 |
| Institutional First Flush | 18 | 18 | 119.67 | 2.078 | 0.151 | -0.969 | 0.536 | 0.105 |  |
| Open Space Composite | 28 | 28 | 20.00 | 1.301 | 0.130 | 0.441 | 0.441 | 0.084 | 0.67 |
| Open Space First Flush | 28 | 28 | 13.43 | 1.128 | 0.211 | 0.731 | 0.441 | 0.013 |  |
| Residential Composite | 140 | 140 | 67.92 | 1.832 | 0.095 | 0.271 | 0.205 | 0.008 | 1.63 |
| Residential First Flush | 140 | 140 | 110.41 | 2.043 | 0.138 | -0.831 | 0.205 | 0.005 |  |
| All Land Uses Composite | 363 | 363 | 65.92 | 1.819 | 0.123 | -0.293 | 0.128 | 0 | 1.71 |
| All Land Uses First Flush | 363 | 363 | 112.98 | 2.053 | 0.194 | -0.710 | 0.128 | 0 |  |

Table E10. Results of Preliminary Test Analysis for COD

| COD | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> t-Test p-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :--- | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.83 ; \mathrm{P}=0$ | 0.269 | 0 | Different (first flush) |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.67 ; \mathrm{P}=0.05$ | 0.691 | 0.01 | Different |
| Institutional | 0.01 | $\mathrm{U}=2.94 ; \mathrm{P}=0$ | 0.677 | 0 | Different |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.269 ; \mathrm{P}=0.39$ | 0.004 | $\mathrm{~N} / \mathrm{A}$ | Same (no first flush) |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=6.715 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=9.19 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E11. Results of Preliminary Statistical Analysis for Total Dissolved Solids (TDS)

| TDS (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew <br> (Log) | SE Skew (Log) | Test Norm. (Log) $p$-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 82 | 82 | 73.28 | 1.865 | 0.064 | -0.338 | 0.266 | 0.263 | 1.83 |
| Commercial First Flush | 82 | 82 | 133.97 | 2.127 | 0.065 | -0.219 | 0.266 | 0.115 |  |
| Industrial Composite | 82 | 81 | 97.72 | 1.990 | 0.093 | -0.482 | 0.267 | 0.341 | 1.32 |
| Industrial First Flush | 82 | 81 | 128.82 | 2.110 | 0.126 | -0.513 | 0.267 | 0.109 |  |
| Institutional Composite | 18 | 18 | 52.48 | 1.720 | 0.068 | -0.034 | 0.536 | 0.360 | 2.66 |
| Institutional First Flush | 18 | 18 | 139.64 | 2.145 | 0.090 | -0.303 | 0.536 | 0.158 |  |
| Open Space Composite | 31 | 30 | 69.98 | 1.845 | 0.051 | 0.617 | 0.427 | 0.376 | 1.07 |
| Open Space First Flush | 31 | 30 | 74.99 | 1.875 | 0.104 | -1.483 | 0.427 | 0.005 |  |
| Residential Composite | 137 | 133 | 70.31 | 1.870 | 0.119 | -0.245 | 0.210 | 0.041 | 1.52 |
| Residential First Flush | 137 | 133 | 107.15 | 2.030 | 0.125 | 0.500 | 0.210 | 0.167 |  |
| All Land Uses Composite | 354 | 342 | 77.62 | 1.890 | 0.083 | 0.188 | 0.132 | 0.334 | 1.55 |
| All Land Uses First Flush | 354 | 342 | 120.23 | 2.080 | 0.104 | 0.225 | 0.132 | 0.126 |  |

Table E12. Results of Preliminary Test Analysis for Total Dissolved Solids (TDS)

| TDS (mg/L) | Mann Wittn. <br> $\mathbf{p}$-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :--- | :---: | :---: | :---: |
| Commercial | 0 | $\mathrm{U}=7.33 ; \mathrm{P}=0$ | 0.160 | 0 | Different (first flush) |
| Industrial | 0.0245 | $\mathrm{U}=2.28 ; \mathrm{P}=0.01$ | 0.070 | 0.003 | Different |
| Institutional | 0.0118 | $\mathrm{U}=2.945 ; \mathrm{P}=0$ | 0.544 | 0 | Different |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.161 ; \mathrm{P}=0.44$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same (no first flush) |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.89 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | 0 | $\mathrm{U}=7.58 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E13. Results of Preliminary Statistical Analysis for O\&G

| O\&G (mg/L) | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 10 | 10 | 5.19 | 0.715 | 0.068 | -0.976 | 0.687 | 0.016 | 1.54 |
| Commercial First Flush | 10 | 10 | 8.00 | 0.027 | 0.903 | 1.641 | 0,687 | 0.019 |  |
| Residential Composite | 8 | 4 | 5.00 | 0.699 | 0.066 | 1.985 | 1.014 | 0.013 | 2.05 |
| Residential First Flush | 8 | 4 | 10.23 | 1.010 | 0.134 | 0.003 | 1.014 | 0.056 |  |
| All Land Uses <br> Composite | 18 | 14 | 5.00 | 0.699 | 0.073 | -0.370 | 0.597 | 0.015 | 1.60 |
| All Land Uses First <br> Flush | 18 | 14 | 8.00 | 0.903 | 0.051 | 0.890 | 0.597 | 0.011 |  |

## Table E14. Results of Preliminary Test Analysis for O\&G

| O\&G (mg/L) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | N/A | $\mathrm{U}=6.198 ; \mathrm{P}<0.01$ | 0.222 | 0.004 | Different |
| Residential | N/A | $\mathrm{U}=1.069 ; \mathrm{P}>0.1$ | 0.049 | 0.306 | Same |
| All Land Uses | N/A | $\mathrm{U}=4.072 ; \mathrm{P}=0$ | 0.036 | N/A | Different |

Table E15. Results of Preliminary Statistical Analysis for Fecal Coliforms

| Fecal Coliforms <br> (mpn/100 mL) | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 12 | 12 | 67764 | 4.831 | 1.099 | -0.691 | 0.637 | 0.627 | 0.87 |
| Commercial First Flush | 12 | 12 | 58884 | 4.770 | 1.732 | -0.388 | 0.637 | 0.228 | 0.87 |
| Residential Composite | 10 | 9 | 41976 | 4.623 | 0.292 | 0.485 | 0.717 | 0.276 | 0.98 |
| Residential First Flush | 10 | 9 | 41020 | 4.643 | 0.685 | 0.247 | 0.717 | 0.799 | 0.9 |
| All Land Uses <br> Composite | 22 | 21 | 46238 | 4.665 | 0.745 | -0.886 | -0.515 | 0.511 | 1.21 |
| All Land Uses First <br> Flush | 22 | 21 | 55976 | 4.748 | 1.269 | 0.501 | 0.501 | 0.391 |  |

Table E16. Results of Preliminary Test Analysis for Fecal Coliforms

| Fecal Coliforms <br> (mpn/100 $\mathbf{m L}$ ) | Mann Wittn. <br> $\mathbf{p}$-value | Fligner Policello | Normality <br> for <br> $\mathbf{t}$-Test $\mathbf{p}-$ <br> value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | N/A | $\mathrm{U}=0 ; \mathrm{P}>0.10$ | 0.833 | 0.583 | Same |
| Residential | N/A | $\mathrm{U}=0.289 ; \mathrm{P}>0.1$ | 0.016 | 0.973 | Same |
| All Land Uses | N/A | $\mathrm{U}=0.181 ; \mathrm{P}=0.43$ | 0.086 | 0.665 | Same |

Table E17. Results of Preliminary Statistical Analysis for Fecal Streptococcus

| Fecal Streptococcus <br> (mpn/100 $\mathbf{m L}$ ) | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 12 | 11 | 37153 | 4.570 | 0.780 | -0.255 | 0.661 | 0.948 | 1.05 |
| Commercial First Flush | 12 | 11 | 38904 | 4.590 | 1.094 | 0.009 | 0.661 | 0.722 | 1.05 |
| Residential Composite | 11 | 8 | 77625 | 4.890 | 0.231 | -0.223 | 0.752 | 0.426 | 1.30 |
| Residential First Flush | 11 | 8 | 101158 | 5.005 | 0.327 | -0.659 | 0.752 | 0.319 |  |
| All Land Uses <br> Composite | 26 | 22 | 43651 | 4.640 | 0.536 | -0.513 | 0.491 | 0.713 | 1.11 |
| All Land Uses First <br> Flush | 26 | 22 | 48417 | 4.685 | 0.705 | -0.188 | 0.491 | 0.802 |  |

Table E18. Results of Preliminary Test Analysis for Fecal Streptococcus

| Fecal <br> Streptococcus <br> (mpn/100mL) | Mann Wittn. <br> $\mathbf{p}$-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$-Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | N/A | $\mathrm{U}=0.281 ; \mathrm{P}>0.10$ | 0.027 | $\mathrm{~N} / \mathrm{A}$ | Same (no first flush) |
| Residential | N/A | $\mathrm{U}=0.344 ; \mathrm{P}>0.10$ | 0.109 | 0.905 | Same |
| All Land Uses | N/A | $\mathrm{U}=0.309 ; \mathrm{P}=0.38$ | 0.033 | N/A | Same |

Table E19. Results of Preliminary Statistical Analysis for Ammonia

| Ammonia (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE Skew (Log) | Test Norm. (Log) p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 70 | 52 | 0.76 | -0.122 | 0.147 | -0.245 | 0.330 | 0.237 | 2.11 |
| Commercial First Flush | 70 | 52 | 1.60 | 0.204 | 0.117 | -0.718 | 0.330 | 0.027 |  |
| Industrial Composite | 40 | 33 | 0.62 | -0.208 | 0.166 | -0.399 | 0.409 | 0.284 | 1.08 |
| Industrial First Flush | 40 | 33 | 0.67 | -0.174 | 0.201 | -0.535 | 0.409 | 0.046 |  |
| Institutional Composite | 18 | 16 | 0.31 | -0.509 | 0.058 | -0.038 | 0.564 | 0.273 | 1.66 |
| Institutional First Flush | 18 | 16 | 0.51 | -0.290 | 0.077 | 0.284 | 0.564 | 0.384 |  |
| Residential Composite | 119 | 86 | 0.50 | -0.301 | 0.370 | 0.779 | 0.260 | 0.001 | 1.36 |
| Residential First Flush | 119 | 86 | 0.68 | -0.168 | 0.172 | 0.195 | 0.260 | 0.519 |  |
| All Land Uses Composite | 269 | 190 | 0.52 | -0.284 | 0.251 | 0.501 | 0.176 | 0.002 | 1.54 |
| All Land Uses First Flush | 269 | 190 | 0.80 | -0.097 | 0.176 | -0.197 | 0.176 | 0.713 |  |

[^0]Table E20. Results of Preliminary Test Analysis for Ammonia

| Ammonia (mg/L) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> t-Test $\mathbf{p}$-value | Paired <br> t - Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.467 ; \mathrm{P}=0$ | 0.028 | $\mathrm{~N} / \mathrm{A}$ | Different |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.113 ; \mathrm{P}=0.46$ | 0.262 | 0.985 | Same |
| Institutional | 0.0287 | $\mathrm{U}=2.484 ; \mathrm{P}=0.01$ | 0.254 | 0 | Different |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=2.283 ; \mathrm{P}=0.01$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.092 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E21. Results of Preliminary Statistical Analysis for $\mathrm{NO}_{\mathbf{2}} \mathbf{+ N O}_{\mathbf{3}}$

| $\mathrm{NO}_{2}+\mathrm{NO}_{3}(\mathrm{mg} / \mathrm{L})$ | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE Skew (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 84 | 82 | 0.75 | -0.125 | 0.095 | -0.092 | 0.266 | 0.188 | 1.73 |
| Commercial First Flush | 84 | 82 | 1.30 | 0.114 | 0.166 | -0.790 | 0.266 | 0.007 |  |
| Industrial Composite | 72 | 71 | 0.90 | -0.046 | 0.073 | -0.240 | 0.285 | 0.807 | 1.31 |
| Industrial First Flush | 72 | 71 | 1.18 | 0.072 | 0.116 | -0.839 | 0.285 | 0.030 |  |
| Institutional | 18 | 18 | 0.60 | -0.222 | 0.122 | -0.714 | 0.536 | 0.117 | 1.70 |
| Institutional First Flush | 18 | 18 | 1.02 | 0.009 | 0.151 | 0.268 | 0.536 | 0.381 |  |
| Open Space Composite | 30 | 21 | 0.24 | -0.620 | 0.290 | 0.468 | 0.501 | 0.141 | 0.96 |
| Open Space First Flush | 30 | 21 | 0.23 | -0.638 | 0.356 | 0.823 | 0.501 | 0.030 |  |
| Residential Composite | 121 | 118 | 0.60 | -0.222 | 0.104 | -0.196 | 0.223 | 0.504 | 1.66 |
| Residential First Flush | 121 | 118 | 1.00 | -0.002 | 0.125 | -0.292 | 0.223 | 0.102 |  |
| All Land Uses Composite | 324 | 310 | 0.70 | -0.155 | 0.124 | -0.497 | 0.138 | 0 | 1.50 |
| All Land Uses First Flush | 324 | 310 | 1.05 | 0.021 | 0.162 | -0.584 | 0.138 | 0 |  |

Table E22. Results of Preliminary Test Analysis for $\mathrm{NO}_{\mathbf{2}}+\mathrm{NO}_{\mathbf{3}}$

| $\mathbf{N O}_{\mathbf{2}}+\mathbf{N O}_{\mathbf{3}}$ | Mann Wittn. p- <br> value | Fligner Policello | Normality for <br> t-Test p-value | Paired <br> $\mathbf{t}-\mathbf{T e s t}$ | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=3.286 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different (first flush) |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.836 ; \mathrm{P}=0.03$ | 0.941 | 0.034 | Different |
| Institutional | 0.043 | $\mathrm{U}=2.242 ; \mathrm{P}=0.01$ | 0.026 | $\mathrm{~N} / \mathrm{A}$ | Different |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.209 ; \mathrm{P}=0.42$ | 0.023 | $\mathrm{~N} / \mathrm{A}$ | Same (no first flush) |
| Residential | 0 | $\mathrm{U}=4.769 ; \mathrm{P}=0$ | 0.023 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=5.834 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E23. Results of Preliminary Statistical Analysis for Total Nitrogen

| Total N (mg/L) | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 19 | 19 | 1.42 | 0.152 | 0.180 | -0.133 | 0.524 | 0.215 | 1.35 |
| Commercial First Flush | 19 | 19 | 1.91 | 0.281 | 0.203 | -0.617 | 0.524 | 0.337 |  |
| Industrial Composite | 19 | 16 | 2.01 | 0.303 | 0.286 | -0.306 | 0.564 | 0.431 | 1.79 |
| Industrial First Flush | 19 | 16 | 3.61 | 0.557 | 0.349 | -0.452 | 0.564 | 0.029 |  |
| Open Space Composite | 6 | 6 | 1.39 | 0.142 | 0.112 | -0.150 | 0.845 | 1.53 |  |
| Open Space First Flush | 6 | 6 | 2.12 | 0.326 | 0.248 | -0.100 | 0.845 | 0.221 | 0.009 |
| Residential Composite | 31 | 30 | 1.67 | 0.222 | 0.325 | 1.22 | 0.427 | 0.88 |  |
| Residential First Flush | 31 | 30 | 1.47 | 0.166 | 0.447 | -0.587 | 0.427 | 0.367 | 0.136 |
| All Land Uses <br> Composite | 77 | 73 | 1.60 | 0.204 | 0.253 | 0.769 | 0.281 | 1.22 |  |
| All Land Uses First <br> Flush | 77 | 73 | 1.95 | 0.290 | 0.331 | 0.599 | 0.281 | 0.071 |  |

Table E24. Results of Preliminary Test Analysis for Total Nitrogen

| Total N(mg/L) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> t-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-\mathbf{T e s t}$ | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | 0.220 | $\mathrm{U}=1.234 ; \mathrm{P}=0.11$ | 0.329 | 0.013 | Same |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.460 ; \mathrm{P}=0.32$ | 0.759 | 0.161 | Same |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0 ; \mathrm{P}>0.104$ | 0.339 | 0.703 | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.106 ; \mathrm{P}=0.46$ | 0.002 | $\mathrm{~N} / \mathrm{A}$ | Same |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.919 ; \mathrm{P}=0.18$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same |

Table E25. Results of Preliminary Statistical Analysis for TKN

| TKN (mg/L) | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 93 | 86 | 1.63 | 0.213 | 0.085 | -0.275 | 0.260 | 0.003 | 1.71 |
| Commercial First Flush | 93 | 86 | 2.80 | 0.447 | 0.120 | -0.117 | 0.260 | 0.714 |  |
| Industrial Composite | 77 | 76 | 1.69 | 0.227 | 0.116 | 1.157 | 0.276 | 0 | 1.35 |
| Industrial First Flush | 77 | 76 | 2.27 | 0.356 | 0.130 | 0.536 | 0.276 | 0.232 | 1.35 |
| Open Space Composite | 32 | 14 | 0.61 | -0.215 | 0.142 | 0.585 | 0.597 | 0.109 | 1.28 |
| Open Space First Flush | 32 | 14 | 0.78 | -0.107 | 0.269 | 0.948 | 0.597 | 0.139 | 0 |
| Residential Composite | 131 | 123 | 1.40 | 0.146 | 0.110 | 1.752 | 0.218 | 1.65 |  |
| Residential First Flush | 131 | 123 | 2.31 | 0.364 | 0.115 | 0.309 | 0.218 | 0.076 | 0 |
| All Land Uses <br> Composite | 335 | 301 | 1.50 | 0.176 | 0.114 | 0.856 | 0.140 | 0 | 1.60 |
| All Land Uses First <br> Flush | 335 | 301 | 2.40 | 0.380 | 0.139 | 0.088 | 0.140 | 0 |  |

Table E26. Results of Preliminary Test Analysis for TKN

| TKN (mg/L) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | N/A | $\mathrm{U}=6.499 ; \mathrm{P}=0$ | 0.126 | 0 | Different (first flush) |
| Industrial | N/A | $\mathrm{U}=1.698 ; \mathrm{P}=0.04$ | 0.054 | 0.063 | Different |
| Open Space | N/A | $\mathrm{U}=0.374 ; \mathrm{P}=0.35$ | 0.116 | 0.364 | Same (no first flush) |
| Residential | N/A | $\mathrm{U}=6.079 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | N/A | $\mathrm{U}=7.68 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E27. Results of Preliminary Statistical Analysis for Total Phosphorus

| Total P (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | $\begin{gathered} \text { SE } \\ \text { Skew } \\ \text { (Log) } \end{gathered}$ | Test Norm. (Log) p -value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 89 | 77 | 0.34 | -0.469 | 0.160 | -0.454 | 0.274 | 0.129 | 1.44 |
| Commercial First Flush | 89 | 77 | 0.49 | -0.310 | 0.205 | 0.033 | 0.274 | 0.035 |  |
| Industrial Composite | 84 | 71 | 0.29 | -0.538 | 0.130 | 0.495 | 0.285 | 0.003 | 1.42 |
| Industrial First Flush | 84 | 71 | 0.41 | -0.387 | 0.257 | -0.441 | 0.285 | 0.397 |  |
| Institutional Composite | 17 | 17 | 0.17 | -0.770 | 0.203 | -0.736 | 0.550 | 0.374 | 1.24 |
| Institutional First Flush | 17 | 17 | 0.21 | -0.678 | 0.066 | -0.177 | 0.550 | 0.704 |  |
| Open Space Composite | 32 | 20 | 0.09 | -1.023 | 0.147 | 0.613 | 0.512 | 0.218 | 1.05 |
| Open Space First Flush | 32 | 20 | 0.10 | -1.000 | 0.381 | 0.833 | 0.512 | 0.288 |  |
| Residential Composite | 140 | 128 | 0.28 | -0.553 | 0.252 | 1.232 | 0.214 | 0 | 1.46 |
| Residential First Flush | 140 | 128 | 0.41 | -0.389 | 0.188 | -0.335 | 0.214 | 0.042 |  |
| All Land Uses Composite | 363 | 313 | 0.28 | -0.553 | 0.209 | 0.605 | 0.138 | 0 | 1.45 |
| All Land Uses First Flush | 363 | 313 | 0.41 | -0.391 | 0.238 | -0.258 | 0.138 | 0.003 |  |

Table E28. Results of Preliminary Test Analysis for Total Phosphorus

| Total P (mg/L) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> t-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-\mathbf{T e s t}$ | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=3.089 ; \mathrm{P}=0$ | 0.594 | 0 | Different (first flush) |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.864 ; \mathrm{P}=0.19$ | 0.194 | 0.667 | Same (no first flush) |
| Institutional | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.774 ; \mathrm{P}=0.22$ | 0.044 | $\mathrm{~N} / \mathrm{A}$ | Same |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.142 ; \mathrm{P}=0.44$ | 0.091 | 0.527 | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=2.671 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=3.641 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E29. Results of Preliminary Statistical Analysis for Dissolved Phosphorus

| Dissolved P (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew <br> (Log) | SE Skew (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 91 | 69 | 0.16 | -0.788 | 0.152 | 0.467 | 0.289 | 0 | 1.23 |
| Commercial First Flush | 91 | 69 | 0.20 | -0.699 | 0.212 | 0.904 | 0.289 | 0.005 |  |
| Industrial Composite | 77 | 50 | 0.14 | -0.854 | 0.142 | 1.248 | 0.337 | 0.093 | 1.04 |
| Industrial First Flush | 77 | 50 | 0.14 | -0.839 | 0.160 | 0.406 | 0.337 | 0.043 |  |
| Institutional Composite | 18 | 14 | 0.13 | -0.891 | 0.066 | -0.114 | 0.597 | 0.563 | 1.05 |
| Institutional First Flush | 18 | 14 | 0.13 | -0.870 | 0.095 | -0.770 | 0.597 | 0.122 |  |
| Open Space Composite | 32 | 14 | 0.05 | -1.301 | 0.111 | -0.073 | 0.597 | 0.601 | 0.69 |
| Open Space First Flush | 32 | 14 | 0.03 | -1.460 | 0.087 | 1.061 | 0.597 | 0.017 |  |
| Residential Composite | 130 | 105 | 0.17 | -0.770 | 0.117 | 0.152 | 0.236 | 0.458 | 1.24 |
| Residential First Flush | 130 | 105 | 0.21 | -0.678 | 0.170 | 0.121 | 0.236 | 0.044 |  |
| All Land Uses Composite | 350 | 254 | 0.15 | -0.824 | 0.143 | 0.353 | 0.153 | 0.051 | 1.07 |
| All Land Uses First Flush | 350 | 254 | 0.16 | -0.796 | 0.200 | 0.401 | 0.153 | 0.001 |  |

Table E30. Results of Preliminary Test Analysis for Dissolved Phosphorus

| Dissolved P (mg/L | Mann Wittn. <br> $\mathbf{p}$-value | Fligner Policello | Normality for <br> $\mathbf{t - T e s t ~} \mathbf{p}$-value | Paired <br> $\mathbf{t}-\mathbf{T e s t}$ | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.582 ; \mathrm{P}=0.06$ | 0.046 | $\mathrm{~N} / \mathrm{A}$ | Same |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.051 ; \mathrm{P}=0.48$ | 0.063 | 0.881 | Same |
| Institutional | 0.549 | $\mathrm{U}=0.605 ; \mathrm{P}=0.27$ | 0.015 | $\mathrm{~N} / \mathrm{A}$ | Same |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.760 ; \mathrm{P}=0.22$ | 0.018 | $\mathrm{~N} / \mathrm{A}$ | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.702 ; \mathrm{P}=0.04$ | 0.039 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.657 ; \mathrm{P}=0.05$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same |

Table E31. Results of Preliminary Statistical Analysis for Orthophosphate

| Orthophosphate (mg/L) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE Skew (Log) | Test Norm. (Log) p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Industrial Composite | 6 | 6 | 0.16 | -0.797 | 0.287 | -0.047 | 0.845 | 0.838 | 1.55 |
| Industrial First Flush | 6 | 6 | 0.25 | -0.607 | 0.356 | -0.106 | 0.845 | 0.720 |  |
| Residential Composite | 14 | 14 | 0.19 | -0.714 | 0.554 | 2.557 | 0.597 | 0.001 | 0.95 |
| Residential First Flush | 14 | 14 | 0.18 | -0.737 | 0.214 | 0.708 | 0.597 | 0.362 |  |
| All Land Uses Composite | 22 | 22 | 0.19 | -0.714 | 0.423 | 2.270 | 0.491 | 0.004 | 1.30 |
| All Land Uses First Flush | 22 | 22 | 0.25 | -0.600 | 0.222 | 0.260 | 0.491 | 0.503 |  |

Table E32. Results of Preliminary Test Analysis for Orthophosphate

| Orthophosphate <br> (mg/L) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> $\mathbf{t - T e s t ~ p - v a l u e ~}$ | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Industrial | 0.471 | $\mathrm{U}=0.772 ; \mathrm{P}>0.104$ | 0.071 | 0.611 | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.022 ; \mathrm{P}=0.49$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.460 ; \mathrm{P}=0.32$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Same |

Table E33. Results of Preliminary Statistical Analysis for Total Cadmium

| Total Cadmium <br> ( $\mu \mathrm{g} / \mathrm{L})$ | Total <br> Events | Selected <br> Cases | Median | Median <br> (Log) | Var <br> (Log) | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median <br> Ratio |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial <br> Composite | 74 | 48 | 0.56 | -0.253 | 0.246 | -0.325 | 0.343 | 0 |  |
| Commercial First <br> Flush | 74 | 48 | 1.20 | 0.079 | 0.261 | 0.080 | 0.343 | 0.089 | 2.15 |
| Industrial <br> Composite | 80 | 41 | 1 | 0 | 0.124 | -0.015 | 0.369 | 0.008 |  |
| Industrial First <br> Flush | 80 | 41 | 1 | 0 | 0.130 | 0.261 | 0.369 | 0.065 | 1.00 |
| Open Space <br> Composite | 30 | 15 | 0.23 | -0.638 | 0.282 | 1.074 | 0.580 | 0.183 | 1.30 |
| Open Space First <br> Flush | 30 | 15 | 0.30 | -0.523 | 0.325 | 0.465 | 0.580 | 0.402 | 0.002 |
| Residential <br> Composite | 123 | 33 | 0.28 | -0.553 | 0.359 | 0.693 | 0.409 | 0.00 |  |
| Residential First <br> Flush | 123 | 33 | 0.56 | -0.252 | 0.264 | 0.512 | 0.409 | 0.061 | 2.00 |
| All Land Uses <br> Composite | 325 | 139 | 0.60 | -0.222 | 0.269 | -0.065 | 0.206 | 0.071 | 0.241 |
| All Land Uses First <br> Flush | 325 | 139 | 0.97 | -0.013 | 0.249 | 0.041 | 0.206 | 0.24 | 1.62 |

Table E34. Results of Preliminary Test Analysis for Total Cadmium

| Total Cadmium <br> $(\boldsymbol{\mu g} / \mathbf{L})$ | Mann Wittn. <br> $\mathbf{p}$-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | 0.006 | $\mathrm{U}=2.797 ; \mathrm{P}=0$ | 0.009 | $\mathrm{~N} / \mathrm{A}$ | Different (first flush) |
| Industrial | 0.922 | $\mathrm{U}=0.100 ; \mathrm{P}=0.46$ | 0.118 | 0.529 | Same (no first flush) |
| Open Space | 0.442 | $\mathrm{U}=0.765 ; \mathrm{P}=0.22$ | 0.292 | 0.191 | Same |
| Residential | 0.038 | $\mathrm{U}=2.131 ; \mathrm{P}=0.02$ | 0.015 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | 0.005 | $\mathrm{U}=2.839 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E35. Results of Preliminary Statistical Analysis for Total Chromium

| Total Chromium ( $\mu \mathrm{g} / \mathrm{L}$ ) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew <br> (Log) | SE <br> Skew <br> (Log) | Test Norm. (Log) p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 47 | 22 | 6.81 | 0.833 | 0.086 | -0.051 | 0.491 | 0.911 | 1.67 |
| Commercial First Flush | 47 | 22 | 11.40 | 1.057 | 0.134 | -0.796 | 0.491 | 0.121 |  |
| Industrial Composite | 54 | 25 | 8.79 | 0.944 | 0.111 | 0.338 | 0.464 | 0.456 | 1.36 |
| Industrial First Flush | 54 | 25 | 11.99 | 1.079 | 0.155 | -0.307 | 0.464 | 0.784 |  |
| Open Space Composite | 16 | 4 | 2.64 | 0.422 | 0.169 | -0.556 | 1.014 | 0.492 | 1.70 |
| Open Space First Flush | 16 | 4 | 4.50 | 0.653 | 0.015 | 1.291 | 1.014 | 0.355 |  |
| Residential Composite | 86 | 31 | 8.00 | 0.903 | 0.169 | -0.077 | 0.421 | 0.612 | 1.24 |
| Residential First Flush | 86 | 31 | 9.91 | 0.996 | 0.137 | 0.326 | 0.421 | 0.904 |  |
| All Land Uses Composite | 218 | 82 | 7.50 | 0.875 | 0.140 | -0.104 | 0.266 | 0.591 | 1.47 |
| All Land Uses First Flush | 218 | 82 | 10.99 | 1.041 | 0.141 | -0.056 | 0.266 | 0.803 |  |

Table E36. Results of Preliminary Test Analysis for Total Chromium

| Total Chromium <br> $(\mu \mathrm{g} / \mathrm{L})$ | Mann Wittn. p- <br> value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | 0.0513 | $\mathrm{U}=2.024 ; \mathrm{P}=0.02$ | 0.283 | 0.036 | Different |
| Industrial | 0.3032 | $\mathrm{U}=1.023 ; \mathrm{P}=0.15$ | 0.216 | 0.320 | Same |
| Open Space | 0.3032 | $\mathrm{U}=1.586 ; \mathrm{P}=0.10$ | 0.160 | 0.199 | Same |
| Residential | 0.6023 | $\mathrm{U}=0.519 ; \mathrm{P}=0.30$ | 0.007 | $\mathrm{~N} / \mathrm{A}$ | Same |
| All Land Uses | 0.0547 | $\mathrm{U}=1.939 ; \mathrm{P}=0.03$ | 0.001 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E37. Results of Preliminary Statistical Analysis for Total Copper

| Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) | Total Events | Selected Cases | Media n | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | $\begin{aligned} & \text { Skew } \\ & \text { (Log) } \end{aligned}$ | SE Skew (Log) | Test Norm. (Log) p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 92 | 82 | 16.98 | 1.230 | 0.083 | -0.038 | 0.266 | 0.117 | 1.62 |
| Commercial First Flush | 92 | 82 | 27.48 | 1.439 | 0.120 | 0.343 | 0.266 | 0.035 |  |
| Industrial Composite | 84 | 76 | 25.00 | 1.398 | 0.079 | 0.184 | 0.276 | 0.344 | 1.24 |
| Industrial First Flush | 84 | 76 | 30.97 | 1.491 | 0.166 | -0.014 | 0.276 | 0.007 |  |
| Institutional Composite | 18 | 7 | 16.98 | 1.230 | 0.083 | -0.228 | 0.794 | 0.167 | 0.94 |
| Institutional First Flush | 18 | 7 | 16.00 | 1.204 | 0.047 | 0.954 | 0.794 | 0.555 |  |
| Open Space Composite | 30 | 22 | 5.14 | 0.711 | 0.103 | 0.085 | 0.491 | 0.252 | 0.78 |
| Open Space First Flush | 30 | 22 | 4.00 | 0.602 | 0.120 | 1.005 | 0.491 | 0.015 |  |
| Residential Composite | 144 | 108 | 11.99 | 1.079 | 0.082 | -0.677 | 0.233 | 0 | 1.33 |
| Residential First Flush | 144 | 108 | 16.00 | 1.204 | 0.087 | 0.023 | 0.233 | 0.256 |  |
| All Land Uses Composite | 368 | 295 | 15.00 | 1.176 | 0.116 | -0.268 | 0.142 | 0 | 1.33 |
| All Land Uses First Flush | 368 | 295 | 20.00 | 1.301 | 0.167 | 0.009 | 0.142 | 0 |  |

Table E38. Results of Preliminary Test Analysis for Total Copper

| Total Copper <br> $(\boldsymbol{\mu g} / \mathrm{L})$ | Mann Wittn. <br> p -value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test p value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=5.160 ; \mathrm{P}=0$ | 0.001 | $\mathrm{~N} / \mathrm{A}$ | Different (first flush) |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.864 ; \mathrm{P}=0.03$ | 0.329 | 0.012 | Different |
| Institutional | 0.5224 | $\mathrm{U}=0.665 ; \mathrm{P}>0.099$ | 0.318 | 0.029 | Same (no first flush) |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.846 ; \mathrm{P}=0.19$ | 0.074 | 0.337 | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.029 ; \mathrm{P}=0$ | 0.292 | 0 | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=5.146 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E39. Results of Preliminary Statistical Analysis for Total Lead

| Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) | Total Events | Selected Cases | Median | Media <br> (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew <br> (Log) | SE Skew (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 89 | 83 | 16.98 | 1.230 | 0.062 | 0.075 | 0.264 | 0.824 | 1.65 |
| Commercial First Flush | 89 | 83 | 27.99 | 1.447 | 0.123 | 0.070 | 0.264 | 0.476 |  |
| Industrial Composite | 84 | 71 | 16.98 | 1.230 | 0.160 | 0.527 | 0.285 | 0.081 | 1.41 |
| Industrial First Flush | 84 | 71 | 23.99 | 1.380 | 0.240 | 0.319 | 0.285 | 0.608 |  |
| Institutional Composite | 18 | 13 | 7.00 | 0.845 | 0.082 | 0.675 | 0.616 | 0.158 | 2.28 |
| Institutional First Flush | 18 | 13 | 15.96 | 1.203 | 0.051 | 0.128 | 0.616 | 0.228 |  |
| Open Space Composite | 31 | 16 | 5.00 | 0.699 | 0.381 | -0.303 | 0.564 | 0.199 | 0.90 |
| Open Space First Flush | 31 | 16 | 4.48 | 0.651 | 0.346 | -0.466 | 0.564 | 0.563 |  |
| Residential Composite | 140 | 93 | 8.79 | 0.944 | 0.231 | 0.084 | 0.250 | 0.884 | 1.48 |
| Residential First Flush | 140 | 93 | 13.00 | 1.114 | 0.204 | 0.130 | 0.250 | 0.105 |  |
| All Land Uses Composite | 364 | 278 | 13.00 | 1.114 | 0.198 | -0.365 | 0.146 | 0.006 | 1.50 |
| All Land Uses First Flush | 364 | 278 | 19.50 | 1.290 | 0.239 | -0.307 | 0.146 | 0.401 |  |

Table E40. Results of Preliminary Test Analysis for Total Lead

| Total Lead ( $\boldsymbol{\mu} \mathbf{g / L}$ ) | Mann Wittn. <br> p-value | Fligner Policello | Normality for <br> $\mathbf{t - T e s t ~ p - v a l u e ~}$ | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | 0 | $\mathrm{U}=5.256 ; \mathrm{P}=0$ | 0.794 | 0 | Different |
| Industrial | 0.083 | $\mathrm{U}=1.742 ; \mathrm{P}=0.04$ | 0.167 | 0.016 | Different |
| Institutional | 0.004 | $\mathrm{U}=3.973 ; \mathrm{P}=0$ | 0.680 | 0.000 | Different |
| Open Space | 0.771 | $\mathrm{U}=0.292 ; \mathrm{P}=0.39$ | 0.008 | 0.578 | Same |
| Residential | 0.012 | $\mathrm{U}=2.59 ; \mathrm{P}=0$ | 0.014 | $\mathrm{~N} / \mathrm{A}$ | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.77 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

Table E41. Results of Preliminary Statistical Analysis for Total Nickel

| Total Nickel ( $\mu \mathrm{g} / \mathrm{L}$ ) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew <br> (Log) | SE Skew (Log) | Test <br> Norm. <br> (Log) p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 47 | 23 | 5.00 | 0.699 | 0.094 | 0.660 | 0.481 | 0.254 | 2.40 |
| Commercial First Flush | 47 | 23 | 11.99 | 1.079 | 0.134 | -0.606 | 0.481 | 0.523 |  |
| Industrial Composite | 51 | 22 | 7.00 | 0.845 | 0.106 | -0.293 | 0.491 | 0.229 | 1.00 |
| Industrial First Flush | 51 | 22 | 7.00 | 0.845 | 0.197 | 0.605 | 0.491 | 0.228 |  |
| Residential Composite | 83 | 18 | 7.48 | 0.874 | 0.094 | 0.152 | 0.536 | 0.814 | 1.20 |
| Residential First Flush | 83 | 18 | 8.99 | 0.954 | 0.115 | 1.551 | 0.536 | 0.048 |  |
| All Land Uses Composite | 213 | 64 | 6.00 | 0.778 | 0.104 | 0.146 | 0.299 | 0.161 | 1.50 |
| All Land Uses First Flush | 213 | 64 | 8.99 | 0.954 | 0.147 | 0.322 | 0.299 | 0.443 |  |

Table E42. Results of Preliminary Test Analysis for Total Nickel

| Total Nickel <br> $(\boldsymbol{\mu g} / \mathbf{L}$ | Mann Wittn. <br> $\mathbf{p}-$-value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-\mathbf{T e s t}$ | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | 0.006 | $\mathrm{U}=3.005 ; \mathrm{P}=0$ | 0.128 | 0.002 | Different (first flush) |
| Industrial | 0.715 | $\mathrm{U}=0.365 ; \mathrm{P}=0.36$ | 0.203 | 0.484 | Same (no first flush) |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=1.143 ; \mathrm{P}=0.13$ | 0.512 | 0.098 | Same |
| All Land Uses | 0.014 | $\mathrm{U}=2.539 ; \mathrm{P}=0.01$ | 0.367 | 0.001 | Different |

Table E43. Results of Preliminary Statistical Analysis for Total Zinc

| Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) | Total Events | Selected Cases | Median | Median (Log) | $\begin{aligned} & \text { Var } \\ & \text { (Log) } \end{aligned}$ | Skew (Log) | SE <br> Skew <br> (Log) | Test <br> Norm. <br> (Log) <br> p-value | Median Ratio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Commercial Composite | 90 | 90 | 149.97 | 2.176 | 0.089 | -1.359 | 0.254 | 0 | 1.93 |
| Commercial First Flush | 90 | 90 | 289.07 | 2.461 | 0.139 | -0.374 | 0.254 | 0.647 |  |
| Industrial Composite | 83 | 83 | 225.94 | 2.354 | 0.184 | 0.828 | 0.264 | 0 | 1.54 |
| Industrial First Flush | 83 | 83 | 348.34 | 2.542 | 0.135 | -0.181 | 0.264 | 0.930 |  |
| Institutional Composite | 18 | 18 | 304.79 | 2.484 | 0.114 | -0.227 | 0.536 | 0.878 | 2.48 |
| Institutional First Flush | 18 | 18 | 755.09 | 2.878 | 0.133 | -0.696 | 0.536 | 0.055 |  |
| Open Space Composite | 21 | 21 | 20.00 | 1.301 | 0.165 | 0.081 | 0.501 | 0.073 | 1.25 |
| Open Space First Flush | 21 | 21 | 25.00 | 1.398 | 0.075 | -0.242 | 0.501 | 0.295 |  |
| Residential Composite | 136 | 136 | 69.34 | 1.841 | 0.114 | 0.824 | 0.208 | 0.003 | 1.58 |
| Residential First Flush | 136 | 136 | 109.90 | 2.041 | 0.200 | -0.232 | 0.208 | 0.014 |  |
| All Land Uses Composite | 350 | 350 | 125.89 | 2.100 | 0.216 | 0.121 | 0.130 | 0.001 | 1.59 |
| All Land Uses First Flush | 350 | 350 | 199.99 | 2.301 | 0.268 | 0.437 | 0.130 | 0.020 |  |

Table E44. Results of Preliminary Test Analysis for Total Zinc

| Total Zinc( $\boldsymbol{\mu g} / \mathrm{L}$ ) | Mann Wittn. <br> p -value | Fligner Policello | Normality for <br> $\mathbf{t}$-Test $\mathbf{p}$-value | Paired <br> $\mathbf{t}-$ Test | Result |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Commercial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=6.156 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |
| Industrial | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=2.087 ; \mathrm{P}=0.02$ | 0.006 | $\mathrm{~N} / \mathrm{A}$ | Different |
| Institutional | 0.007 | $\mathrm{U}=3.1 ; \mathrm{P}=0$ | 0.498 | 0 | Different |
| Open Space | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=0.023 ; \mathrm{P}=0.49$ | 0.667 | 0.977 | Same |
| Residential | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=4.329 ; \mathrm{P}=0$ | 0 | N/A | Different |
| All Land Uses | $\mathrm{N} / \mathrm{A}$ | $\mathrm{U}=5.374 ; \mathrm{P}=0$ | 0 | $\mathrm{~N} / \mathrm{A}$ | Different |

## Appendix F: Detailed Statistical Test Results to Identify Significant Land Use and Geographical Interactions

pH
Summary
Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  | Minimum | Maximum |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | :--- |
| 2 | 0 | $123^{*}$ | $*$ | $*$ | $*$ |  |  |
| 3 | 0 | $6^{*}$ | $*$ | $*$ | $*$ |  |  |
| 4 | 0 | $16^{*}$ | $*$ | $*$ | $*$ |  |  |
| 5 | 41 | 1 | 7.7039 | 0.438 | 6.9 | 9.1 |  |
| 6 | 27 | 7 | 6.737 | 0.632 | 5.7 | 8.6 |  |
| 7 | 27 | 14 | 7.363 | 0.662 | 5.3 | 8.38 |  |

Industrial

| EPA_Rain_Zone | N det/est |  | N ND/NZ Mean |  | StDev Minimum |  | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 2 | 18 | 91 | 7.319 | 0.533 | 6.4 | 8.65 |
|  | 3 | 9 | 7 | 6.6111 | 0.2205 | 6.5 | 7 |
|  | 4 | 0 | 17* |  | * | * | * |
|  | 5 | 46 | 1 | 7.9822 | 0.504 | 6.6 | 9 |
|  | 6 | 53 | 17 | 7.6836 | 0.7219 | 5.9 | 9.3 |
|  | 7 | 5 | 28 | 7.34 | 0.358 | 6.8 | 7.7 |

Residential

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  | Minimum | Maximum |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 14 | 317 | 7.364 | 0.389 | 6.7 | 7.87 |  |
| 3 | 8 | 10 | 6.788 | 0.416 | 6.5 | 7.4 |  |
| 4 | 0 | $31^{*}$ | ${ }^{*}$ |  | $*$ | $*$ |  |
| 5 | 71 | 0 | 7.5179 | 0.5816 | 6.2 | 9.9 |  |
| 6 | 22 | 16 | 7.341 | 0.559 | 6.3 | 8.3 |  |
| 7 | 25 | 15 | 7.319 | 0.713 | 6.2 | 9.1 |  |

Highlighted cells indicated groups with enough observations to calculate mean and standard deviation.

1. First Analysis. EPA Rain zones 5,6, and 7 Landuse: Residential Commercial Industrial



Sites with one observation or less will not be included in the analyses.

| Variable | Rainloc | N | $N^{*}$ | Mean | Variance | Minimum | Maximum |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| pH | 5_TXARA001 | 22 | 0 | 7.5073 | 0.1132 | 6.9000 | 8.2000 |
|  | 5_TXARA002 | 21 | 0 | 7.563 | 0.220 | 6.700 | 8.600 |


| 5_TXARA003 | 7 | 0 | 7.529 | 0.449 | 6.200 | 8.300 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5_TXDAA001 | 7 | 0 | 7.9286 | 0.0190 | 7.7000 | 8.1000 |
| 5_TXDAA002 | 19 | 0 | 8.079 | 0.306 | 6.900 | 9.000 |
| 5_TXDAA004 | 19 | 1 | 7.932 | 0.192 | 7.200 | 9.100 |
| 5-TXDAA005 | 7 | 0 | 7.700 | 1.270 | 6.400 | 9.900 |
| 5-TXFWA004 | 20 | 1 | 7.909 | 0.289 | 6.600 | 8.800 |
| 5_TXIRA001 | 22 | 0 | 7.3795 | 0.2017 | 6.8000 | 8.2000 |
| 5_TXMEA002 | 7 | 0 | 7.900 | 0.0933 | 7.400 | 8.300 |
| 5_TXMEA003 | 7 | 0 | 7.243 | 0.320 | 6.400 | 8.000 |
| 6_AZMCA001 | 26 | 1 | 7.277 | 0.331 | 5.900 | 8.300 |
| 6_AZMCA003 | 26 | 1 | 8.050 | 0.398 | 6.900 | 9.300 |
| 6_AZMCA005 | 26 | 0 | 6.665 | 0.271 | 5.700 | 8.100 |
| 6_AZMCA006 | 20 | 0 | 7.290 | 0.313 | 6.300 | 8.300 |
| 6_AZTUA001 | 1 | 9 | 7.7000 | * | 7.7000 | 7.7000 |
| 6_AZTUA002 | 1 | 7 | 8.0000 | * | 8.0000 | 8.0000 |
| 6_AZTUA003 | 1 | 7 | 8.6000 | * | 8.6000 | 8.6000 |
| 6_AZTUA004 | 1 | 6 | 8.7300 | * | 8.7300 | 8.7300 |
| 6-CAALAL09 | 0 | 9 | * | * | * | * |
| 7-OREUA001 | 16 | 0 | 7.437 | 0.633 | 5.300 | 8.300 |
| 7-OREUA003 | 14 | 1 | 7.502 | 0.452 | 6.600 | 9.100 |
| 7-ORGRA003 | 5 | 1 | 7.450 | 0.681 | 6.800 | 8.450 |
| 7-ORGRA004 | 5 | 1 | 7.280 | 0.393 | 6.800 | 8.380 |
| 7-ORPOA001 | 0 | 13 | * | * | * | * |
| 7-ORPOA003 | 0 | 14 | * | * | * | * |
| 7-ORPOA004 | 0 | 13 | * | * | * | * |
| 7-ORPOA006 | 0 | 13 | * | * | * | * |
| 7-ORSAA002 | 6 | 0 | 7.2333 | 0.0227 | 7.1000 | 7.4000 |
| 7_ORSAA003 | 5 | 1 | 7.340 | 0.128 | 6.800 | 7.700 |
| 7-ORSAA004 | 6 | 0 | 6.783 | 0.266 | 6.200 | 7.400 |

The following two plots will identify unusual sites among all the sites with pH observations. These sites will be included in the analysis, however they must be analyzed to identify potential conditions that produce these unusual observations.


## Xbar-R Chart of pH



The GLM will be used to identify if there is a significant difference in PH among land use and EPA rain zone. GLM were used instead of the ANOVA model because the number of observations is not the same in each combination land use - EPA rain zone.

RESULTS:
General Linear Model: pH versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels | Values |
| :--- | :--- | ---: | :--- |
| Landuse | fixed | 3 | CO, ID, RE |
| EPA_Rain_Zone | fixed | 3 | $5,6,7$ |

Analysis of Variance for pH, using Adjusted SS for Tests

| Source | DF | Seq SS | AdjSS | Adj MS | F | P |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Landuse | 2 | 12.1467 | 4.6160 | 2.3080 | 6.64 | 0.001 |
| EPA_Rain_Zone | 2 | 14.0505 | 16.1575 | 8.0788 | 23.26 | 0.000 |
| Landuse*EPA_Rain_Zone | 4 | 7.9699 | 7.9699 | 1.9925 | 5.74 | 0.000 |
| Error | 304 | 105.5977 | 105.5977 | 0.3474 |  |  |
| Total | 312 | 139.7647 |  |  |  |  |

In this case the main factors and interaction term are considered significant. Now simultaneous tests will be used to evaluate if there is a significant difference among the land uses or the EPA rain zones. Bonferroni is the most conservative method (conservative means "true error rate is less than the stated one").

```
Bonferroni Simultaneous Tests
Response Variable pH
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
```



| ID | 0.4179 | 0.11470 | 3.644 | 0.0009 |
| :--- | ---: | ---: | ---: | ---: |
| RE | 0.1317 | 0.08873 | 1.485 | 0.4160 |
|  |  |  |  |  |
| Landuse $=$ | ID $\quad$ subtracted from: |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | Of Means | Difference | T-Value | P-Value |
| RE | -0.2862 | 0.1154 | -2.480 | 0.0411 |

Tukey Simultaneous Tests
Response Variable pH
All Pairwise Comparisons among Levels of Landuse
Landuse $=$ CO subtracted from:

|  | Difference <br> of Means | SE of | Difference | T-Value |
| :--- | ---: | ---: | ---: | ---: | | Adjusted |
| ---: |
| P-Value |

Landuse $=$ ID subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.2862 | 0.1154 | -2.480 | 0.0351 |

pH in industrial land use is significantly different than in residential and commercial land uses. pH in commercial and residential land uses is not significantly different.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable pH |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA Rain Zone | $=5$ subtrac | ted from: |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | -0.5284 | 0.08050 | -6.564 | 0.0000 |
| 7 | -0.3941 | 0.11409 | -3.454 | 0.0019 |
| EPA_Rain_Zone = 6 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | 0.1343 | 0.1219 | 1.102 | 0.8137 |


| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable pH |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA Rain Zone |  |  |  |  |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | -0.5284 | 0.08050 | -6.564 | 0.0000 |
| 7 | -0.3941 | 0.11409 | -3.454 | 0.0016 |

EPA_Rain_Zone $=6$ subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: | ---: |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | 0.1343 | 0.1219 | 1.102 | 0.5127 |

pH in EPA rain zone 5 is significantly different than zones 6 and $7 . \mathrm{pH}$ in rain zones 6 and 7 is not significantly different.


The last analysis is to inspect the residual plots to confirm that they are normally distributed around zero and there is not a specific pattern. In this case the plots indicate normality and no trend against observation order.

The final plot is the interaction plot. In this plot it is possible to identify if there in a difference in the pattern of the levels when change from one level to another. In this case, for example, it was observed that pH in region 7 is not affected by land use. In the other two rain zones, when the land use changes from residential or commercial to industrial, there is an increase in the pH .

## Interaction Plot (fitted means) for pH



2. Second Analysis. EPA Rain zones 2, 3, 5, 6, and 7 Landuse: Residential and Industrial

Areas in Yellow are not included in the analysis

| Variable | Rainloc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| pH | 2_KYLOTSR1 | 3 | 0 | 7.100 | 0.0900 | 6.800 | 7.400 |
|  | 2_KYLOTSR2 | 3 | 0 | 6.967 | 0.0633 | 6.700 | 7.200 |
|  | 2_KYLOTSR3 | 3 | 0 | 7.367 | 0.343 | 6.700 | 7.800 |
|  | 2_KYLOTSR4 | 4 | 0 | 7.025 | 0.422 | 6.400 | 7.900 |
|  | 2_MDAACOMW | 0 | 3 | * | * | * | * |
|  | 2_MDAACOOD | 0 | 3 | * | * | * | * |
|  | 2_MDAACORK | 0 | 3 | * | * | * | * |
|  | 2_MDBACOBC | 0 | 3 | * | * | * | * |
|  | 2_MDBACOSC | 0 | 26 | * | * | * | * |
|  | 2_MDBACOTC | 0 | 3 | * | * | * | * |
|  | 2_MDBACOWC | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYBO | 0 | 3 | * | * | * | * |
|  | 2-MDBCTYFM | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHO | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHR | 0 | 3 | * | * | * | * |
|  | 2_VAARLLP1 | 8 | 0 | 7.461 | 0.116 | 7.040 | 7.870 |
|  | 2-VAARLTC4 | 11 | 2 | 7.523 | 0.226 | 6.830 | 8.650 |
|  | 2_VACPTC1A | 0 | 8 | * | * | * | * |
|  | 2_VACPTSF2 | 0 | 3 | * | * | * | * |
|  | 2_VACPTYC1 | 0 | 7 | * | * | * | * |
|  | 2-VACPTYC3 | 0 | 14 | * | * | * | * |
|  | 2_VACPTYC5 | 0 | 14 | $\star$ | * | * | * |
|  | 2_VACPTYO1 | 0 | 3 | * | * | * | * |
|  | 2_VAHATYH2 | 0 | 19 | * | * | * | * |
|  | 2_VAHATYH3 | 0 | 17 | * | * | * | * |
|  | 2-VAHATYH4 | 0 | 17 | * | * | * | * |
|  | 2_VAHATYH5 | 0 | 17 | * | * | * | * |
|  | 2_VAHCCON1 | 0 | 2 | * | * | * | * |


| 2 VAHCCON2 | 0 | 3 | * | * | * | * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2-VAHCCOR1 | 0 | 3 | * | * | * | * |
| 2-VAHCCOR2 | 0 | 3 | * | * | * | * |
| 2-VANFTMS 6 | 0 | 3 | * | * | * | * |
| 2_VANFTYN2 | 0 | 29 | * | * | * | * |
| 2_VANFTYN3 | 0 | 27 | * | * | * | * |
| 2 VANFTYN5 | 0 | 27 | * | * | * | * |
| 2-VANNTMF1 | 0 | 1 | * | * | * | * |
| 2-VANNTMF4 | 0 | 3 | * | * | * | * |
| 2-VANNTNN1 | 0 | 10 | * | * | * | * |
| 2-VANNTSF4 | 0 | 2 | * | * | * | * |
| 2-VANNTSF6 | 0 | 2 | * | * | * | * |
| 2_VAPMTYP2 | 0 | 17 | * | * | * | * |
| 2-VAPMTYP4 | 0 | 17 | * | * | * | * |
| 2_VAPMTYP5 | 0 | 17 | * | * | * | * |
| 2-VAVBTYI1 | 0 | 3 | * | * | * | * |
| 2 VAVBTYR1 | 0 | 5 | * | * | * | * |
| 2-VAVBTYV1 | 0 | 27 | * | * | * | * |
| 2_VAVBTYV4 | 0 | 30 | * | * | * | * |
| 3-ALMOCREO | 0 | 3 | * | * | * | * |
| 3 ALMOSARA | 0 | 3 | * | * | * | * |
| 3-ALMOSIIV | 0 | 3 | * | * | * | $\star$ |
| 3_ALMOSIVI | 0 | 3 | * | * | * | * |
| 3_ALMOTHEO | 0 | 3 | * | * | * | * |
| 3_GAATAT01 | 9 | 1 | 6.6111 | 0.0486 | 6.5000 | 7.0000 |
| 3_GAATAT02 | 8 | 1 | 6.788 | 0.173 | 6.500 | 7.400 |
| 5_TXARA002 | 21 | 0 | 7.563 | 0.220 | 6.700 | 8.600 |
| 5_TXARA003 | 7 | 0 | 7.529 | 0.449 | 6.200 | 8.300 |
| 5_TXDAA001 | 7 | 0 | 7.9286 | 0.0190 | 7.7000 | 8.1000 |
| 5 TXDAA002 | 19 | 0 | 8.079 | 0.306 | 6.900 | 9.000 |
| 5_TXDAA005 | 7 | 0 | 7.700 | 1.270 | 6.400 | 9.900 |
| 5_TXFWA004 | 20 | 1 | 7.909 | 0.289 | 6.600 | 8.800 |
| 5_TXIRA001 | 22 | 0 | 7.3795 | 0.2017 | 6.8000 | 8.2000 |
| 5_TXMEA002 | 7 | 0 | 7.900 | 0.0933 | 7.400 | 8.300 |
| 5 TXMEA003 | 7 | 0 | 7.243 | 0.320 | 6.400 | 8.000 |
| 6_AZMCA001 | 26 | 1 | 7.277 | 0.331 | 5.900 | 8.300 |
| 6_AZMCA003 | 26 | 1 | 8.050 | 0.398 | 6.900 | 9.300 |
| 6_AZMCA006 | 20 | 0 | 7.290 | 0.313 | 6.300 | 8.300 |
| 6_AZTUA001 | 1 | 9 | 7.7000 | * | 7.7000 | 7.7000 |
| 6 AZTUA002 | 1 | 7 | 8.0000 | * | 8.0000 | 8.0000 |
| 6_AZTUA004 | 1 | 6 | 8.7300 | * | 8.7300 | 8.7300 |
| 6_CAALAL0 9 | 0 | 9 | * | * | * | * |
| 7 OREUA003 | 14 | 1 | 7.502 | 0.452 | 6.600 | 9.100 |
| 7 ORGRA003 | 5 | 1 | 7.450 | 0.681 | 6.800 | 8.450 |
| 7 ORPOA003 | 0 | 14 | * | * | * | * |
| 7-ORPOA004 | 0 | 13 | * | * | * | * |
| 7-ORPOA006 | 0 | 13 | * | * | * | * |
| 7_ORSAA003 | 5 | 1 | 7.340 | 0.128 | 6.800 | 7.700 |
| 7_ORSAA004 | 6 | 0 | 6.783 | 0.266 | 6.200 | 7.400 |


pH about 7.5 but some sites can have mean pH as low as 6.5 or high as 8.0
Results from general linear model.

## General Linear Model: pH versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 2 | ID, RE |  |  |  |  |
| EPA_Rain_Zone fixed | 5 | 2, 3, | 6, 7 |  |  |  |
| Analysis of Variance for | for pH, | using Ad | usted SS | for Te |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 1 | 4.2658 | 0.5749 | 0.5749 | 1.68 | 0.196 |
| EPA Rain Zone | 4 | 18.5915 | 19.0338 | 4.7584 | 13.93 | 0.000 |
| Lan̄${ }^{\text {duse*EPA_Rain_Zone }}$ | 4 | 3.0489 | 3.0489 | 0.7622 | 2.23 | 0.066 |
| Error | 258 | 88.1302 | 88.1302 | 0.3416 |  |  |
| Total | 267 | 114.0363 |  |  |  |  |
| $S=0.584457 \quad \mathrm{R}-\mathrm{Sq}=$ | 22.72\% | R-Sq(adj) $=20.02 \%$ |  |  |  |  |

No significant differences by land use or interactions. Significant differences by EPA rain zone.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable pH |  |  |  |  |
| All Pairwise Comparisons among Levels of Landus Landuse $=$ ID subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.1276 | 0.09837 | -1.297 | 0.1957 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable pH |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.1276 | 0.09837 | -1.297 | 0.1945 |

No significant differences in pH between industrial and residential land uses

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable pH |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.6422 | 0.1761 | -3.647 | 0.0032 |
| 5 | 0.4085 | 0.1179 | 3.465 | 0.0062 |
| 6 | 0.1352 | 0.1294 | 1.045 | 1.0000 |
| 7 | -0.0119 | 0.1770 | -0.067 | 1.0000 |
| EPA_Rain_Zone = 3 subtracted from: |  |  |  |  |
| EPA_Rain_Zone | Difference | SE of | Adjusted |  |
|  | of Means | Difference | T-Value | P-Value |
| 5 | 1.0507 | 0.1524 | 6.895 | 0.0000 |
| 6 | 0.7774 | 0.1615 | 4.814 | 0.0000 |
| 7 | 0.6303 | 0.2016 | 3.126 | 0.0198 |
| EPA_Rain_Zone | 5 subtrac | ed from: |  |  |
|  | Difference | SE of |  | Adjusted |


| EPA_Rain_Zone | Of Means | Difference | T-Value | P-Value |
| :--- | ---: | ---: | ---: | ---: |
| 6 | -0.2733 | 0.09472 | -2.885 | 0.0424 |
| 7 | -0.4204 | 0.15348 | -2.739 | 0.0658 |
|  |  |  |  |  |
| EPA_Rain_Zone $=$ | subtracted from: |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.1471 | 0.1625 | -0.9054 | 1.000 |

pH in rain zones 3 and 5 is different than in region 2
No significant differences in pH between region 2 and 6 and 2 and 7
pH in rain zones 5, 6 and 7 is different than in region 3
No significant differences in pH between region 6 and 7 compared with region 5
No significant differences in pH between regions 6 and 7

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable pH |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.6422 | 0.1761 | -3.647 | 0.0025 |
| 5 | 0.4085 | 0.1179 | 3.465 | 0.0048 |
| 6 | 0.1352 | 0.1294 | 1.045 | 0.8345 |
| 7 | -0.0119 | 0.1770 | -0.067 | 1.0000 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 1.0507 | 0.1524 | 6.895 | 0.0000 |
| 6 | 0.7774 | 0.1615 | 4.814 | 0.0000 |
| 7 | 0.6303 | 0.2016 | 3.126 | 0.0153 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | -0.2733 | 0.09472 | -2.885 | 0.0320 |
| 7 | -0.4204 | 0.15348 | -2.739 | 0.0484 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.1471 | 0.1625 | -0.9054 | 0.8949 |

pH in rain zones 3 and 5 is different than in region 2
No significant differences in pH between region 2 and 6 and 2 and 7
pH in rain zones 5,6 and 7 is different than in region 3
pH in rain zones 6 and 7 is different than in region 5
No significant differences in pH between regions 6 and 7

## Residual Plots for pH



Normality of the residuals, except for three or four unusual observations. No unusual patterns in the residuals

Interaction Plot (fitted means) for $\mathbf{~ p H}$


## Temperature

Summary
Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  | Minimum | Maximum |
| :---: | :---: | ---: | ---: | ---: | :--- | ---: | :--- | :--- |
| 2 | 0 | $123^{*}$ | $*$ | $*$ | $*$ |  |
| 3 | 0 | $6^{*}$ | $*$ | $*$ | $*$ |  |
| 4 | 0 | $16^{*}$ | $*$ | $*$ | $*$ |  |
| 5 | 41 | 1 | 16.229 | 4.762 | 6.5 | 24 |
| 6 | 26 | 8 | 20.87 | 7.13 | 11 | 30 |
| 7 | 21 | 20 | 12.724 | 4.54 | 5 | 21 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  | Minimum | Maximum |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 0 | $109^{*}$ | $*$ | $*$ | $*$ |  |  |
| 3 | 0 | $16^{*}$ | $*$ | $*$ | $*$ |  |  |
| 4 | 0 | $17^{*}$ | $*$ | $*$ | $*$ |  |  |
| 5 | 47 | 0 | 19.47 | 5.837 | 7 | 30 |  |
| 6 | 52 | 18 | 20.213 | 6.395 | 11.5 | 31.5 |  |
| 7 | 5 | 28 | 9.9 | 2.7 | 7 | 14 |  |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev | Minimum | Maximum |  |
| :---: | :---: | ---: | :--- | :--- | :--- | :--- | :--- |
| 2 | 0 | $331^{*}$ | $*$ | $*$ | $*$ |  |
| 3 | 0 | $18^{*}$ | $*$ | $*$ | $*$ |  |
| 4 | 0 | $31^{*}$ | $*$ | $*$ | $*$ |  |
| 5 | 71 | 0 | 18.099 | 5.191 | 7.5 | 29 |
| 6 | 20 | 18 | 19.54 | 7.12 | 11 | 28.5 |
| 7 | 20 | 20 | 11.345 | 4.413 | 5 | 23 |

Analysis: Rain zone 5, 6 and 7. Land use RE, CO and ID

## Descriptive Statistics: Temperature (C)

| Variable | Rainloc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Temperature (C) | 5_TXARA001 | 22 | 0 | 16.65 | 29.37 | 7.00 | 24.00 |
|  | 5-TXARA002 | 21 | 0 | 16.58 | 29.91 | 7.50 | 24.00 |
|  | 5-TXARA003 | 7 | 0 | 15.36 | 14.73 | 10.50 | 21.00 |
|  | 5_TXDAA001 | 7 | 0 | 22.64 | 16.81 | 14.50 | 27.00 |
|  | 5_TXDAA002 | 19 | 0 | 19.97 | 35.33 | 7.50 | 30.00 |
|  | 5_TXDAA004 | 19 | 1 | 15.737 | 15.654 | 6.500 | 23.000 |
|  | 5_TXDAA005 | 7 | 0 | 19.00 | 11.17 | 15.00 | 23.50 |
|  | 5-TXFWA004 | 21 | 0 | 17.96 | 35.39 | 7.00 | 27.10 |
|  | 5_TXIRA001 | 22 | 0 | 19.90 | 31.47 | 9.00 | 29.00 |
|  | 5-TXMEA002 | 7 | 0 | 20.14 | 27.73 | 13.00 | 26.00 |
|  | 5-TXMEA003 | 7 | 0 | 16.79 | 14.40 | 12.00 | 23.00 |
|  | 6_AZMCA001 | 25 | 2 | 20.74 | 46.94 | 11.50 | 31.50 |
|  | 6_AZMCA003 | 26 | 1 | 20.04 | 35.10 | 12.50 | 29.00 |
|  | 6_AZMCA005 | 26 | 0 | 20.87 | 50.79 | 11.00 | 30.00 |
|  | 6_AZMCA006 | 19 | 1 | 19.95 | 49.94 | 11.00 | 28.50 |
|  | 6_AZTUA001 | 1 | 9 | 11.800 | * | 11.800 | 11.800 |
|  | 6_AZTUA002 | 0 | 8 | * | * | * | * |
|  | 6_AZTUA003 | 0 | 8 | * | * | * | * |


| 6_AZTUA004 | 1 | 6 | 11.600 | * | 11.600 | 11.600 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6-CAALAL09 | 0 | 9 | * | * | * | * |
| 7 OREUA001 | 15 | 1 | 13.95 | 21.28 | 5.00 | 21.00 |
| 7-OREUA003 | 14 | 1 | 12.06 | 22.35 | 5.30 | 23.00 |
| 7-ORGRA003 | 0 | 6 | * | * | * | * |
| 7-ORGRA004 | 0 | 6 | * | * | * | * |
| 7-ORPOA001 | 0 | 13 | * | * | * | * |
| 7-ORPOA003 | 0 | 14 | * | * | * | * |
| 7-ORPOA004 | 0 | 13 | * | * | * | * |
| 7_ORPOA006 | 0 | 13 | * | * | * | * |
| 7_ORSAA002 | 6 | 0 | 9.67 | 7.17 | 6.00 | 14.00 |
| 7-ORSAA003 | 5 | 1 | 9.90 | 7.30 | 7.00 | 14.00 |
| 7-ORSAA004 | 6 | 0 | 9.67 | 11.07 | 5.00 | 14.00 |



Samples in Rain zone 7 are colder than in the rain zones 5 and 6


## Rainloc

## General Linear Model: Temperature (C) versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID, | RE |  |  |  |
| EPA_Rain_Zone fixed | 3 | 5, 6, 7 |  |  |  |  |
| Analysis of Variance for Temperature (C), using Adjusted SS for Tests |  |  |  |  |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 418.05 | 0.87 | 0.44 | 0.01 | 0.986 |
| EPA_Rain_Zone | 2 | 1955.86 | 1815.42 | 907.71 | 28.34 | 0.000 |
| Landuse*EPA_Rain_Zone | 4 | 222.71 | 222.71 | 55.68 | 1.74 | 0.141 |
| Error | 292 | 9351.72 | 9351.72 | 32.03 |  |  |
| Total | 300 | 11948.35 |  |  |  |  |

Land use and interaction are not significant. EPA rain zone is significant.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of LanduseLanduse = CO subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.0220 | 1.1182 | -0.0196 | 1.000 |
| RE | -0.1425 | 0.8991 | -0.1585 | 1.000 |
| Landuse = ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.1205 | 1.128 | -0.1069 | 1.000 |

No differences among the three land uses

| Response Variable Temperature (C) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.0220 | 1.1182 | -0.0196 | 0.9998 |
| RE | -0.1425 | 0.8991 | -0.1585 | 0.9862 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.1205 | 1.128 | -0.1069 | 0.9937 |

No differences among the three land uses

```
Bonferroni Simultaneous Tests
Response Variable Temperature (C)
All Pairwise Comparisons among Levels of EPA_Rain_Zone
EPA_Rain_Zone = 5 subtracted from:
\begin{tabular}{lrrrrr} 
& Difference & SE of \\
EPA_Rain_Zone & of Means & Difference & T-Value & Adjusted \\
6 & 2.466 & 0.7788 & 3.166 & 0.0051 \\
7 & & -6.610 & 1.1277 & -5.861 & 0.0000 \\
EPA_Rain_Zone \(=6\) & subtracted from: & &
\end{tabular}
EPA_Rain_Zone \begin{tabular}{rrr} 
Difference & SE of & \\
Of Means
\end{tabular}\(\quad\) Difference \begin{tabular}{rl} 
& Adjusted
\end{tabular}
```

Temperature in rain zone 5 is different than in rain zones 6 and 7 . Temperature in zone 6 is different than in rain zone 7.

```
Tukey Simultaneous Tests
Response Variable Temperature (C)
All Pairwise Comparisons among Levels of EPA Rain Zone
EPA_Rain_Zone = 5 subtracted from:
\begin{tabular}{lrrrrr} 
& Difference & SE of & & Adjusted \\
EPA_Rain_Zone & of Means & Difference & T-Value & P-Value \\
6 & 2.466 & 0.7788 & 3.166 & 0.0044 \\
7 & -6.610 & 1.1277 & -5.861 & 0.0000
\end{tabular}
EPA_Rain_Zone = 6 subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
EPA_Rain_Zone & of Means & Difference & T-Value & P-Value \\
7 & -9.075 & 1.205 & -7.529 & 0.0000
\end{tabular}
```

Temperature in rain zone 5 is different than in rain zones 6 and 7 . Temperature in zone 6 is different than in rain zone 7.

Bi modal probability plot. Normality assumption might not be valid. No specific pattern in the residuals. Season influence can be important.


## Hardness

Summary statistics in LOG base 10 scale
Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 0 | 123* |  | * | * | * |
|  | 3 | 0 | 6* |  | * | * | * |
|  | 4 | 0 | 16* |  | * | * | * |
|  | 53 | 39 | 3 | 1.475 | 0.149 | 1.146 | 1.845 |
|  | 6 | 0 | 34* |  | * | * | * |
|  | $7 \quad 2$ | 26 | 15 | 1.099 | 0.431 | 0.279 | 1.792 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 211 | 98 | 1.790 | 0.491 | 1.000 | 2.646 |
| 3 | 30 | 16* |  | * | * * | * |
| 4 | 40 | 17* |  | * * | * * | * |
| 5 | 543 | 4 | 1.599 | 0.181 | 1.255 | 2.137 |
| 6 | 6 - 8 | 62 | 0.957 | 0.379 | 0.699 | 1.531 |
| 7 | 720 | 13 | 1.246 | 0.265 | 0.740 | 1.881 |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ M | ean | StDev | Minimum Maximum |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 2 | 8323 | 1.632 | 0.319 | 1.104 | 2.114 |
| 3 | 3 | 0 18* |  | * * | * * | * |
| 4 | 4 | 0 31* |  | * * | * * | * |
|  | 56 | 4 | 1.511 | 0.154 | 1.176 | 1.892 |
| 6 | 6 | 0 38* |  | * * | * * | * |
| 7 | $7 \quad 26$ | 614 | 1.237 | 0.332 | 0.699 | 1.820 |

Analysis 1. Commercial, Residential, and industrial in rain zones 5 and 7

## Descriptive Statistics: LHARD

| Variable | Rainloc | N | $\mathrm{N}^{*}$ | Mean | Variance | Minimum | Maximum |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| LHARD | 5_TXARA001 | 22 | 0 | 1.5071 | 0.0309 | 1.1461 | 1.8445 |
|  | 5_TXARA002 | 18 | 3 | 1.4559 | 0.0109 | 1.3010 | 1.6821 |
|  | 5_TXARA003 | 7 | 0 | 1.4717 | 0.00794 | 1.3222 | 1.5563 |
|  | 5_TXDAA001 | 6 | 1 | 1.4482 | 0.0178 | 1.2553 | 1.5911 |
|  | 5_TXDAA002 | 18 | 1 | 1.5986 | 0.0187 | 1.3802 | 2.0453 |
|  | 5_TXDAA004 | 17 | 3 | 1.4335 | 0.00894 | 1.3075 | 1.6128 |
|  | 5_TXDAA005 | 6 | 1 | 1.6287 | 0.0217 | 1.4314 | 1.7782 |
|  | 5_TXFWA004 | 19 | 2 | 1.6458 | 0.0440 | 1.3838 | 2.1367 |
|  | 5_TXIRA001 | 20 | 2 | 1.5032 | 0.0248 | 1.1761 | 1.8007 |
|  | 5_TXMEA002 | 7 | 0 | 1.6939 | 0.0233 | 1.4914 | 1.8921 |
|  | 5_TXMEA003 | 6 | 1 | 1.4145 | 0.0201 | 1.2304 | 1.6128 |
|  | 7_OREUA001 | 14 | 2 | 1.4016 | 0.0248 | 1.1461 | 1.7924 |
|  | 7_OREUA003 | 15 | 0 | 1.4529 | 0.0400 | 1.0792 | 1.8195 |
|  | 7_ORGRA003 | 0 | 6 | $*$ | $*$ | $*$ | $*$ |
|  | 7_ORGRA004 | 0 | 6 | $*$ | $*$ | $*$ | $*$ |
|  | 7_ORPOA001 | 12 | 1 | 0.746 | 0.140 | 0.279 | 1.633 |
|  | 7_ORPOA003 | 12 | 2 | 1.1744 | 0.0772 | 0.7404 | 1.8808 |
|  | 7_ORPOA004 | 8 | 5 | 1.3540 | 0.0476 | 1.0000 | 1.6628 |


| 7_ORPOA006 | 11 | 2 | 0.9423 | 0.0534 | 0.6990 | 1.5315 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 7_ORSAA002 | 0 | 6 | $*$ | $*$ | $*$ | $*$ |
| 7_ORSAA003 | 0 | 6 | $*$ | $*$ | $*$ | $*$ |
| 7_ORSAA004 | 0 | 6 | $*$ | $*$ | $*$ | $*$ |

## Xbar-R Chart of LHARD





General Linear Model: LHARD versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed |  | CO, ID, |  |  |  |  |
| EPA_Rain_Zone fixed | 2 | 5,7 |  |  |  |  |
| Analysis of Variance | for LHAR | RD, using | Adjuste | d SS for | Tests |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 0.8799 | 0.5593 | 0.2796 | 4.77 | 0.009 |
| EPA_Rain_Zone | 1 | 5.1893 | 5.2468 | 5.2468 | 89.42 | 0.000 |
|  | 2 | 0.0981 | 0.0981 | 0.0490 | 0.84 | 0.435 |
| Error | 2121 | 12.4389 | 12.4389 | 0.0587 |  |  |
| Total | 217 18 | 18.6061 |  |  |  |  |

Main factors are significant. Interaction is not significant.

## General Linear Model: LHARD versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels | Values |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse | fixed | 3 | CO, ID, | RE |  |  |
| EPA_Rain_Zone | fixed | 2 | 5, 7 |  |  |  |
| Analysis of Variance |  | for LHARD, using |  | Adjuste | SS | Tests |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 0.8799 | 0.5800 | 0.2900 | 4.95 | 0.008 |
| EPA_Rain_Zone | 1 | 5.1893 | 5.1893 | 5.1893 | 88.58 | 0.000 |
| Error | 214 | 12.5369 | 12.5369 | 0.0586 |  |  |
| Total | 217 | 18.6061 |  |  |  |  |

## Main factors are significant.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.13487 | 0.04289 | 3.145 | 0.0057 |
| RE | 0.07044 | 0.03959 | 1.779 | 0.2298 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.06443 | 0.03977 | -1.620 | 0.3201 |

Hardness in rain commercial land use is different than in industrial land use. There are no differences in hardness between commercial and residential land use. There is no difference between residential and industrial land use for hardness concentrations (log)

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of Landuse Landuse $=C O$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.13487 | 0.04289 | 3.145 | 0.0054 |
| RE | 0.07044 | 0.03959 | 1.779 | 0.1790 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.06443 | 0.03977 | -1.620 | 0.2394 |

Same results using Tukey test.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=5$ subtracted from: - - |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.3297 | 0.03503 | -9.412 | 0.0000 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zon |  |  |  |  |
| EPA_Rain_Zone | 5 subtrac | ed from: |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.3297 | 0.03503 | -9.412 | 0.0000 |

Hardness concentrations in EPA rain zones 5 and 7 are significantly different.


Analysis 2. Residential, and industrial in rain zones 2, 5 and 7

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LHARD | 2_KYLOTSR1RE | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR2ID | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR3RE | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR4ID | 0 | 4 | * | * | * | * |
|  | 2 MDAACOMWID | 0 | 3 | * | * | * | * |
|  | 2-MDAACOODRE | 0 | 3 | * | * | * | * |
|  | 2-MDAACORKRE | 0 | 3 | * | * | * | $\star$ |
|  | 2_MDBACOBCID | 0 | 3 | * | * | * | * |
|  | 2_MDBACOSCRE | 0 | 26 | * | * | * | * |
|  | 2 MDBACOTCID | 0 | 3 | * | * | * | * |
|  | 2-MDBACOWCRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYBOID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYFMID | 0 | 3 | * | * | * | * |
|  | 2-MDBCTYHORE | 0 | 3 | * | * | * | * |
|  | 2 MDBCTYHRRE | 0 | 3 | * | * | * | * |
|  | 2_VAARLLP1RE | 8 | 0 | 1.632 | 0.102 | 1.104 | 2.114 |
|  | 2_VAARLTC4ID | 11 | 2 | 1.790 | 0.241 | 1.000 | 2.646 |
|  | 2_VACPTC1ARE | 0 | 8 | * | * | * | * |
|  | 2_VACPTSF2RE | 0 | 3 | * | * | * | * |
|  | 2_VACPTYC1RE | 0 | 7 | * | * | * | * |
|  | 2_VACPTYC3RE | 0 | 14 | * | * | * | * |
|  | 2_VACPTYC5ID | 0 | 14 | * | * | * | * |
|  | 2_VACPTYO1ID | 0 | 3 | * | * | * | * |
|  | 2-VAHATYH2ID | 0 | 19 | * | * | * | * |
|  | 2-VAHATYH3RE | 0 | 17 | * | * | * | * |
|  | 2-VAHATYH4RE | 0 | 17 | * | * | * | * |
|  | 2_VAHATYH5RE | 0 | 17 | * | * | * | * |
|  | 2_VAHCCON1ID | 0 | 2 | * | * | * | * |
|  | 2-VAHCCON2ID | 0 | 3 | * | * | * | * |
|  | 2_VAHCCOR1RE | 0 | 3 | * | * | * | * |
|  | 2_VAHCCOR2RE | 0 | 3 | * | * | * | * |
|  | 2_VANFTMS 6RE | 0 | 3 | * | * | * | * |
|  | 2-VANFTYN2RE | 0 | 29 | * | * | * | * |
|  | 2-VANFTYN3RE | 0 | 27 | * | * | * | * |
|  | 2_VANFTYN5RE | 0 | 27 | * | * | * | * |
|  | 2_VANNTMF1RE | 0 | 1 | * | * | * | * |
|  | 2_VANNTMF4RE | 0 | 3 | * | * | * | * |
|  | 2_VANNTNN1RE | 0 | 10 | * | * | * | * |
|  | 2_VANNTSF4RE | 0 | 2 | * | * | * | * |
|  | 2_VANNTSF6RE | 0 | 2 | * | * | * | * |
|  | 2_VAPMTYP2RE | 0 | 17 | * | * | * | * |
|  | 2_VAPMTYP4RE | 0 | 17 | * | * | * | * |
|  | 2-VAPMTYP5RE | 0 | 17 | * | * | * | * |
|  | 2_VAVBTYIIID | 0 | 3 | * | * | * | * |
|  | 2_VAVBTYR1RE | 0 | 5 | * | $\star$ | * | * |
|  | 2_VAVBTYV1RE | 0 | 27 | * | * | * | * |
|  | 2_VAVBTYV4ID | 0 | 30 | * | * | * | * |
|  | 5_TXARA002RE | 18 | 3 | 1.4559 | 0.0109 | 1.3010 | 1.6821 |
|  | 5-TXARA003RE | 7 | 0 | 1.4717 | 0.00794 | 1.3222 | 1.5563 |
|  | 5_TXDAA001ID | 6 | 1 | 1.4482 | 0.0178 | 1.2553 | 1.5911 |
|  | 5_TXDAA002ID | 18 | 1 | 1.5986 | 0.0187 | 1.3802 | 2.0453 |
|  | 5 TXDAA005RE | 6 | 1 | 1.6287 | 0.0217 | 1.4314 | 1.7782 |
|  | 5_TXFWA004ID | 19 | 2 | 1. 6458 | 0.0440 | 1.3838 | 2.1367 |
|  | 5-TXIRA001RE | 20 | 2 | 1.5032 | 0.0248 | 1.1761 | 1.8007 |
|  | 5_TXMEA002RE | 7 | 0 | 1.6939 | 0.0233 | 1.4914 | 1.8921 |
|  | 5-TXMEA003RE | 6 | 1 | 1.4145 | 0.0201 | 1.2304 | 1.6128 |
|  | 7 OREUA003RE | 15 | 0 | 1.4529 | 0.0400 | 1.0792 | 1.8195 |
|  | 7_ORGRA003RE | 0 | 6 | * | * | * | * |
|  | 7-ORPOA003ID | 12 | 2 | 1.1744 | 0.0772 | 0.7404 | 1.8808 |
|  | 7_ORPOA004ID | 8 | 5 | 1.3540 | 0.0476 | 1.0000 | 1.6628 |
|  | 7_ORPOA006RE | 11 | 2 | 0.9423 | 0.0534 | 0.6990 | 1.5315 |



General Linear Model: LHARD versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels | Values |
| :--- | :--- | ---: | :--- |
| Landuse | fixed | 2 | ID, RE |
| EPA_Rain_Zone | fixed | 3 | $2,5,7$ |

Analysis of Variance for LHARD, using Adjusted SS for Tests

| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Landuse | 1 | 0.2956 | 0.2296 | 0.2296 | 3.80 | 0.053 |
| EPA_Rain_Zone | 2 | 4.1525 | 4.1525 | 2.0762 | 34.33 | 0.000 |
| Error | 168 | 10.1597 | 10.1597 | 0.0605 |  |  |
| Total | 171 | 14.6078 |  |  |  |  |

Significant difference by rain zone but not by land use at 5\%

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
| Landuse $=$ ID subtracted from |  |  |  |  |
|  | Difference | SE Of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.07424 | 0.03810 | -1.948 | 0.0530 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
| Landuse = ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.07424 | 0.03810 | -1.948 | 0.0530 |

No Significant difference by land use

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.1643 | 0.06159 | -2.667 | 0.0252 |
| 7 | -0.4718 | 0.06729 | -7.012 | 0.0000 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.3075 | 0.04338 | -7.090 | 0.0000 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LHARD |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.1643 | 0.06159 | -2.667 | 0.0227 |
| 7 | -0.4718 | 0.06729 | -7.012 | 0.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |


| EPA_Rain_Zone | Of Means | Difference | T-Value | P-Value |
| :--- | ---: | ---: | ---: | ---: |
| 7 | -0.3075 | 0.04338 | -7.090 | 0.0000 |

Significant differences between rain zone 2 and rain zones 5 and 7 . Significant differences between rain zone 5 and 7.



## Oil and Grease

Summary statistics in LOG base 10 scale
Land use: Commercial

| EPA_Rain_Zone | N det/est |  | N ND/NZ Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 2 | 52 | 71 | 0.629 | 0.321 | -0.097 | 1.556 |
| 3 | 3 | 0 | 6* |  | * * | * * | * |
|  | 4 | 14 | 2 | 0.683 | 0.416 | -0.068 | 1.398 |
|  | 5 | 41 | 1 | 0.766 | 0.896 | -0.962 | 2.555 |
|  | 6 | 29 | 5 | 0.620 | 0.487 | -0.374 | 1.778 |
|  | 7 | 37 | 4 | 0.414 | 0.358 | -0.476 | 1.255 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 241 | 68 | 0.531 | 0.417 | -0.415 | 1.342 |
| 3 | 30 | 16* |  | * | * | * |
| 4 | 415 | 2 | 0.443 | 0.402 | -0.291 | 1.146 |
| 5 | 546 | 1 | 0.191 | 1.074 | -2.583 | 2.611 |
| 6 | 655 | 15 | 0.593 | 0.457 | -0.571 | 1.380 |
| 7 | 729 | 4 | 0.569 | 0.314 | -0.205 | 1.204 |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 292 | 239 | 0.500 | 0.336 | -0.742 | 1.505 |
|  | 30 | 0 18* |  | * * | * * | * |
|  | $4 \quad 31$ | 10 | 0.531 | 0.401 | -0.069 | 1.491 |
|  | $5 \quad 69$ | 2 | 0.392 | 0.955 | -1.470 | 3.474 |
|  | $6 \quad 24$ | 414 | 0.474 | 0.573 | -0.732 | 2.176 |
|  | 738 | 8 | 0.252 | 0.445 | -0.804 | 1.491 |

## Descriptive Statistics: LOAG

| Variable | LndRainLoc | N | $\mathrm{N}^{*}$ | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LOAG | 2 KYLOTSR1RE | 0 | 3 | * | * | * | * |
|  | 2-KYLOTSR2ID | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR3RE | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR4ID | 0 | 4 | * | * | * | * |
|  | 2_MDAACOMWID | 3 | 0 | -0.0792 | 0.106 | -0.396 | 0.255 |
|  | 2_MDAACOODRE | 3 | 0 | 0.187 | 0.0909 | -0.155 | 0.415 |
|  | 2 MDAACOPPCO | 22 | 4 | 0.7647 | 0.0919 | 0.1139 | 1.3838 |
|  | 2_MDAACORKRE | 2 | 1 | 0.105 | 0.0453 | -0.0458 | 0.255 |
|  | 2_MDBACOBCID | 2 | 1 | 0.977 | 0.0792 | 0.778 | 1.176 |
|  | 2 mDBACOSCRE | 0 | 26 | * | * | * | * |
|  | 2_MDBACOTCID | 1 | 2 | 0.77815 | * | 0.77815 | 0.77815 |
|  | 2_MDBACOWCRE | 3 | 0 | 0.9662 | 0.00489 | 0.9031 | 1.0414 |
|  | 2_MDBCTYBOID | 1 | 2 | 1.1139 | * | 1.1139 | 1.1139 |
|  | 2_MDBCTYFMID | 3 | 0 | 1.014 | 0.104 | 0.699 | 1.342 |
|  | 2 MDBCTYHORE | 3 | 0 | 0.6667 | 0.00313 | 0.6021 | 0.6990 |
|  | 2_MDBCTYHRRE | 3 | 0 | 0.790 | 0.186 | 0.301 | 1.114 |
|  | 2_MDBCTYKOCO | 3 | 0 | 0.634 | 0.123 | 0.301 | 1.000 |
|  | 2_MDMOCOBCCO | 3 | 0 | 0.234 | 0.0882 | -0.0969 | 0.477 |


| 2_VAARLLP1RE | 8 | 0 | 0.201 | 0.392 | -0.742 | 1.114 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2-VAARLTC4ID | 12 | 1 | 0.535 | 0.302 | -0.415 | 1.322 |
| 2-VACPTC1ARE | 0 | 8 | * | * | * | * |
| 2_VACPTSF2RE | 3 | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VACPTYC1RE | 3 | 4 | 0.525 | 0.0482 | 0.398 | 0.778 |
| 2_VACPTYC3RE | 2 | 12 | 0.548 | 0.0453 | 0.398 | 0.699 |
| 2-VACPTYC4CO | 3 | 11 | 0.498 | 0.0302 | 0.398 | 0.699 |
| 2-VACPTYC5ID | 2 | 12 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VACPTYO1ID | 3 | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAHATYH1CO | 3 | 15 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAHATYH2ID | 3 | 16 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAHATYH3RE | 3 | 14 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2-VAHATYH4RE | 3 | 14 | 0.647 | 0.187 | 0.398 | 1.146 |
| 2_VAHATYH5RE | 3 | 14 | 0.625 | 0.155 | 0.398 | 1.079 |
| 2-VAHCCOC1CO | 2 | 0 | 0.651 | 0.128 | 0.398 | 0.903 |
| 2-VAHCCOC2CO | 3 | 0 | 0.779 | 0.0724 | 0.559 | 1.079 |
| 2_VAHCCON1ID | 2 | 0 | 0.676 | 0.155 | 0.398 | 0.954 |
| 2_VAHCCON2ID | 3 | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAHCCOR1RE | 3 | 0 | 0.583 | 0.103 | 0.398 | 0.954 |
| 2-VAHCCOR2RE | 3 | 0 | 0.498 | 0.0302 | 0.398 | 0.699 |
| 2_VANFTMS5CO | 3 | 0 | 0.498 | 0.0302 | 0.398 | 0.699 |
| 2_VANFTMS6RE | 3 | 0 | 0.767 | 0.409 | 0.398 | 1.505 |
| 2_VANFTMS8CO | 1 | 2 | 0.39794 | * | 0.39794 | 0.39794 |
| 2_VANFTMS 9CO | 1 | 2 | 0.39794 | * | 0.39794 | 0.39794 |
| 2_VANFTYN2RE | 7 | 22 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2-VANFTYN3RE | 5 | 22 | 0.725 | 0.235 | 0.398 | 1.477 |
| 2_VANFTYN4CO | 5 | 22 | 0.630 | 0.268 | 0.398 | 1.556 |
| 2_VANFTYN5RE | 5 | 22 | 0.7868 | 0.0330 | 0.5917 | 1.0414 |
| 2_VANNTMF1RE | , | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VANNTMF4RE | 3 | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VANNTNN1RE | 2 | 8 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VANNTSF4RE | 2 | 0 | 0.651 | 0.128 | 0.398 | 0.903 |
| 2_VANNTSF6RE | 2 | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAPMTYP1CO | 3 | 15 | 0.498 | 0.0302 | 0.398 | 0.699 |
| 2_VAPMTYP2RE | 3 | 14 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAPMTYP4RE | 3 | 14 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2-VAPMTYP5RE | 3 | 14 | 0.498 | 0.0302 | 0.398 | 0.699 |
| 2_VAVBTYIIID | 3 | 0 | 0.599 | 0.0302 | 0.398 | 0.699 |
| 2-VAVBTYR1RE | 4 | 1 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAVBTYV1RE | 4 | 23 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 2_VAVBTYV4ID | 3 | 27 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 4_KATOATWORE | 15 | 0 | 0.572 | 0.211 | -0.0429 | 1.491 |
| 4_KATOBROORE | 16 | 0 | 0.4925 | 0.1225 | -0.0689 | 1.0792 |
| 4_KATOJACKCO | 14 | 2 | 0.683 | 0.173 | -0.0679 | 1.398 |
| 4_KATOSTFEID | 15 | 2 | 0.443 | 0.162 | -0.291 | 1.146 |
| 5-TXARA001CO | 22 | 0 | 0.858 | 0.973 | -0.962 | 2.555 |
| 5-TXARA002RE | 20 | 1 | 0.534 | 1.249 | -1.470 | 2.622 |
| 5-TXARA003RE | 7 | 0 | 0.0255 | 0.350 | -0.855 | 0.699 |
| 5-TXDAA001ID | 7 | 0 | -0.151 | 0.0706 | -0.519 | 0.301 |
| 5-TXDAA002ID | 19 | 0 | -0.0821 | 1.798 | -2.583 | 2.611 |
| 5_TXDAA004CO | 19 | 1 | 0.660 | 0.625 | -0.719 | 2.322 |
| 5-TXDAA005RE | 7 | 0 | -0.2580 | 0.0129 | -0.3010 | 0.000000000 |
| 5-TXFWA004ID | 20 | 1 | 0.571 | 0.735 | -0.981 | 2.265 |
| 5-TXIRA001RE | 21 | 1 | 0.868 | 1.122 | -1.012 | 3.474 |
| 5_TXMEA002RE | 7 | 0 | -0.00695 | 0.318 | -0.732 | 0.699 |
| 5_TXMEA003RE | 7 | 0 | -0.0314 | 0.0623 | -0.2963 | 0.4771 |
| 6_AZMCA001ID | 26 | 1 | 0.7199 | 0.1590 | -0.1160 | 1.3802 |
| 6_AZMCA003ID | 27 | 0 | 0.4549 | 0.2412 | -0.5713 | 1.1461 |
| 6_AZMCA005CO | 26 | 0 | 0.5869 | 0.2482 | -0.3743 | 1.7782 |
| 6_AZMCA006RE | 20 | 0 | 0.403 | 0.365 | -0.732 | 2.176 |
| 6_AZTUA001RE | 2 | 8 | 0.8741 | 0.00168 | 0.8451 | 0.9031 |
| 6_AZTUA002RE | 2 | 6 | 0.77815 | 0.000000000 | 0.77815 | 0.77815 |
| 6_AZTUA003CO | 3 | 5 | 0.903 | 0.0822 | 0.699 | 1.230 |


| 6_AZTUA004ID | 2 | 5 | 0.801 | 0.0208 | 0.699 | 0.903 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 6_CAALAL09ID | 0 | 9 | $*$ | $*$ | $*$ | $*$ |
| 7_OREUA001CO | 14 | 2 | 0.443 | 0.204 | -0.476 | 1.255 |
| 7_OREUA003RE | 14 | 1 | 0.2645 | 0.0742 | -0.2218 | 0.7782 |
| 7_ORGRA003RE | 6 | 0 | -0.247 | 0.168 | -0.804 | 0.362 |
| 7_ORGRA004CO | 5 | 1 | 0.435 | 0.157 | -0.203 | 0.869 |
| 7_ORPOA001CO | 12 | 1 | 0.3800 | 0.0301 | 0.0562 | 0.7559 |
| 7_ORPOA003ID | 12 | 2 | 0.670 | 0.120 | -0.0486 | 1.204 |
| 7_ORPOA004ID | 12 | 1 | 0.5210 | 0.0332 | 0.2041 | 0.8633 |
| 7_ORPOA006RE | 12 | 1 | 0.374 | 0.263 | -0.0969 | 1.491 |
| 7_ORSAA002CO | 6 | 0 | 0.395 | 0.196 | -0.253 | 0.934 |
| 7_ORSAA003ID | 5 | 1 | 0.441 | 0.211 | -0.205 | 0.903 |
| 7_ORSAA004RE | 6 | 0 | 0.478 | 0.130 | -0.0866 | 0.851 |




General Linear Model: LOAG versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID, | RE |  |  |  |
| EPA_Rain_Zone fixed | 5 | 2, 4, | 5, 6, 7 |  |  |  |
| Analysis of Variance for LOAG, using Adjusted SS for Tests |  |  |  |  |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 3.8655 | 3.3711 | 1.6856 | 4.60 | 0.010 |
| EPA_Rain_Zone | 4 | 3.1071 | 2.0783 | 0.5196 | 1.42 | 0.227 |
| Landuse*EPA_Rain_Zone | 8 | 6.2668 | 6.2668 | 0.7833 | 2.14 | 0.031 |
| Error | 5932 | 217.4563 | 217.4563 | 0.3667 |  |  |
| Total | 6072 | 230.6956 |  |  |  |  |

Land use and the interaction land use * EPA rain zone are significant.

| Response Variable LOAG |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of Landuse Landuse $=$ CO subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.1630 | 0.07104 | -2.294 | 0.0664 |
| RE | -0.1945 | 0.06667 | -2.917 | 0.0110 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.03148 | 0.06549 | -0.4807 | 1.000 |

```
Tukey Simultaneous Tests
Response Variable LOAG
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& \begin{tabular}{r} 
Difference \\
Of Means
\end{tabular} & \begin{tabular}{r} 
SE of \\
Difference
\end{tabular} & T-Value & Adjusted \\
P-Value
\end{tabular}
Landuse = ID subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
RE & -0.03148 & 0.06549 & -0.4807 & 0.8804
\end{tabular}
```

Oil and grease in residential is different than in commercial land use. There is not enough evidence to prove a difference between commercial and industrial or between residential and industrial.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of EPA_Rain_ZoneEPA Rain Zone $=2$ subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.0028 | 0.09617 | 0.029 | 1.0000 |
| 5 | -0.0998 | 0.06912 | -1.443 | 1.0000 |
| 6 | 0.0124 | 0.07843 | 0.158 | 1.0000 |
| 7 | -0.1381 | 0.07672 | -1.800 | 0.7242 |
| EPA_Rain_Zone = 4 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.1026 | 0.09701 | -1.057 | 1.000 |
| 6 | 0.0096 | 0.10385 | 0.092 | 1.000 |
| 7 | -0.1409 | 0.10256 | -1.374 | 1.000 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.11214 | 0.07946 | 1.4113 | 1.000 |
| 7 | -0.03832 | 0.07777 | -0.4927 | 1.000 |
| EPA_Rain_Zone = 6 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.1505 | 0.08615 | -1.746 | 0.8125 |

Tukey Simultaneous Tests
Response Variable LOAG
All Pairwise Comparisons among Levels of EPA_Rain_Zone
EPA_Rain_Zone = 2 subtracted from:

|  | Difference | SE of |  | Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.0028 | 0.09617 | 0.029 | 1.0000 |
| 5 | -0.0998 | 0.06912 | -1.443 | 0.5996 |
| 6 | 0.0124 | 0.07843 | 0.158 | 0.9999 |
| 7 | -0.1381 | 0.07672 | -1.800 | 0.3737 |
| EPA_Rain_Zone = 4 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.1026 | 0.09701 | -1.057 | 0.8283 |
| 6 | 0.0096 | 0.10385 | 0.092 | 1.0000 |
| 7 | -0.1409 | 0.10256 | -1.374 | 0.6446 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.11214 | 0.07946 | 1.4113 | 0.6203 |
| 7 | -0.03832 | 0.07777 | -0.4927 | 0.9881 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.1505 | 0.08615 | -1.746 | 0.4054 |

No differences among EPA rain zones for oil and grease.


Residuals fail normality. Tails are larger compared with the normal distribution. A trend was observed in the residuals. Several observations were observed at the same value.


## Total Dissolved Solids

Summary statistics in LOG base 10 scale
Land use: Commercial

| EPA_Rain_Zone | N det/est | N non detected |  |  |  |  |  |  | Mean |  |  |  |  |  | StDev | Minimum Maximum |  |  |
| :---: | :---: | ---: | :---: | :---: | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 82 | 41 | 1.83 | 0.32 | 1.36 | 3.59 |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 6 | 0 | 1.67 | 0.22 | 1.46 | 2.08 |  |  |  |  |  |  |  |  |  |  |  |  |
| 4 | 15 | 1 | 2.20 | 0.28 | 1.81 | 2.94 |  |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 39 | 3 | 1.70 | 0.17 | 1.36 | 2.08 |  |  |  |  |  |  |  |  |  |  |  |  |
| 6 | 20 | 14 | 1.98 | 0.28 | 1.52 | 2.58 |  |  |  |  |  |  |  |  |  |  |  |  |
| 7 | 37 | 4 | 1.62 | 0.40 | 0.60 | 2.09 |  |  |  |  |  |  |  |  |  |  |  |  |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N non detected |  |  |  |  |  | Mean | StDev | Minimum Maximum |
| :---: | :---: | ---: | :---: | :---: | :---: | ---: | :---: | :---: | :---: | :---: |
| 2 | 86 | 23 | 1.81 | 0.50 | 0.65 | 4.05 |  |  |  |  |
| 3 | 16 | 0 | 1.97 | 0.30 | 1.15 | 2.35 |  |  |  |  |
| 4 | 16 | 1 | 2.00 | 0.28 | 1.48 | 2.72 |  |  |  |  |
| 5 | 43 | 4 | 1.88 | 0.20 | 1.43 | 2.38 |  |  |  |  |
| 6 | 56 | 14 | 2.07 | 0.24 | 1.20 | 2.57 |  |  |  |  |
| 7 | 24 | 9 | 1.79 | 0.32 | 0.85 | 2.19 |  |  |  |  |

Land use: Residential

| EPA_Rain_Zone | N det/est | N non detected |  |  |  |  |  |  | Mean |  |  |  |  |  |  | StDev | Minimum Maximum |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 268 | 63 | 1.83 | 0.27 | 1.04 | 3.23 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 17 | 1 | 1.87 | 0.30 | 1.43 | 2.32 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 4 | 31 | 0 | 2.26 | 0.30 | 1.77 | 3.04 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 64 | 7 | 1.85 | 0.18 | 1.52 | 2.28 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 6 | 34 | 4 | 1.98 | 0.22 | 1.62 | 2.45 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 7 | 37 | 3 | 1.64 | 0.38 | 0.48 | 2.24 |  |  |  |  |  |  |  |  |  |  |  |  |  |

## Descriptive Statistics: LTDS

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LTDS | 2_KYLOTSR1RE | 3 | 0 | 2.149 | 0.254 | 1.806 | 2.728 |
|  | 2_KYLOTSR2ID | 3 | 0 | 2.070 | 0.283 | 1.748 | 2.685 |
|  | 2_KYLOTSR3RE | 3 | 0 | 2.469 | 0.0871 | 2.173 | 2.763 |
|  | 2-KYLOTSR4ID | 4 | 0 | 2.4105 | 0.0289 | 2.2148 | 2.5635 |
|  | 2_MDAACOMWID | 1 | 2 | 1.5185 | * | 1.5185 | 1.5185 |
|  | 2_MDAACOODRE | 0 | 3 | * | * | * | * |
|  | 2_MDAACOPPCO | 0 | 26 | * | * | * | * |
|  | 2_MDAACORKRE | 0 | 3 | * | * | * | * |
|  | 2-MDBACOBCID | 0 | 3 | * | * | * | * |
|  | 2_MDBACOSCRE | 0 | 26 | * | * | * | * |
|  | 2_MDBACOTCID | 0 | 3 | * | * | * | * |
|  | 2_MDBACOWCRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYBOID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYFMID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHORE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHRRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYKOCO | 0 | 3 | * | * | * | * |
|  | 2-MDMOCOBCCO | 0 | 3 | * | * | * | * |
|  | 2_VAARLLP1RE | 8 | 0 | 2.394 | 0.195 | 1.968 | 3.230 |
|  | 2_VAARLTC4ID | 12 | 1 | 2.397 | 0.920 | 0.653 | 4.049 |



| 6_AZMCA001ID | 20 | 7 | 1.9987 | 0.0560 | 1.6435 | 2.5717 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6_AZMCA003ID | 24 | 3 | 2.1602 | 0.0191 | 1.8451 | 2.4857 |
| 6 AZMCA005CO | 12 | 14 | 1.8548 | 0.0395 | 1.5185 | 2.2148 |
| 6_AZMCA006RE | 16 | 4 | 1.8932 | 0.0345 | 1.6232 | 2.3909 |
| 6_AZTUA001RE | 10 | 0 | 2.1314 | 0.0350 | 1.9031 | 2.4265 |
| 6_AZTUA002RE | 8 | 0 | 1.9557 | 0.0591 | 1.6532 | 2.4472 |
| 6_AZTUA003CO | 8 | 0 | 2.156 | 0.0933 | 1.748 | 2.583 |
| 6_AZTUA004ID | 6 | 1 | 2.2375 | 0.0233 | 2.0414 | 2.4378 |
| 6 CAALAL09ID | 6 | 3 | 1.785 | 0.139 | 1.204 | 2.204 |
| 7_OREUA001CO | 14 | 2 | 1.8631 | 0.0375 | 1.4914 | 2.0934 |
| 7 OOREUA003RE | 15 | 0 | 1.7557 | 0.0890 | 0.9031 | 2.0414 |
| 7 ORGRA003RE | 6 | 0 | 1.8070 | 0.0336 | 1.4771 | 2.0128 |
| $7{ }^{\text {-ORGRA0 }} 04 \mathrm{CO}$ | 6 | 0 | 1.515 | 0.121 | 0.903 | 1.845 |
| 7 ORPOA001CO | 12 | 1 | 1.368 | 0.269 | 0.602 | 2.079 |
| 7_ORPOA003ID | 11 | 3 | 1.788 | 0.151 | 0.845 | 2.188 |
| 7_ORPOA004ID | 9 | 4 | 1.9331 | 0.0399 | 1.6232 | 2.1523 |
| 7 ORPOA006RE | 11 | 2 | 1.519 | 0.175 | 0.775 | 2.243 |
| 7_ORSAA002CO | 5 | 1 | 1.6984 | 0.00670 | 1.5563 | 1.7634 |
| 7 ORSAA003ID | 4 | 2 | 1.4869 | 0.0130 | 1.3617 | 1.6335 |
| 7 ORSAA004RE | 5 | 1 | 1.321 | 0.228 | 0.477 | 1.643 |

## Xbar-R Chart of LTDS




General Linear Model: LTDS versus Landuse, EPA_Rain_Zone


Not significant by land use
Significant by rain zone
Significant by land use and rain zone interaction.

```
Bonferroni Simultaneous Tests
Response Variable LTDS
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & 0.08739 & 0.03829 & 2.283 & 0.0681 \\
RE & 0.06801 & 0.03647 & 1.865 & 0.1875
\end{tabular}
```

|  | Difference | SE of |  | Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.01937 | 0.03159 | -0.6134 | 1.000 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LTDS |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse $=$ co subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.08739 | 0.03829 | 2.283 | 0.0583 |
| RE | 0.06801 | 0.03647 | 1.865 | 0.1489 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.01937 | 0.03159 | -0.6134 | 0.8128 |

## No significant differences by land use

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTDS |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | 0.0143 | 0.05710 | 0.250 | 1.0000 |
| 4 | 0.3272 | 0.04419 | 7.403 | 0.0000 |
| 5 | -0.0128 | 0.03086 | -0.415 | 1.0000 |
| 6 | 0.1830 | 0.03592 | 5.094 | 0.0000 |
| 7 | -0.1411 | 0.03570 | -3.952 | 0.0013 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.3129 | 0.06813 | 4.593 | 0.0001 |
| 5 | -0.0271 | 0.06034 | -0.449 | 1.0000 |
| 6 | 0.1687 | 0.06308 | 2.675 | 0.1143 |
| 7 | -0.1554 | 0.06295 | -2.468 | 0.2066 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.3400 | 0.04830 | -7.038 | 0.0000 |
| 6 | -0.1442 | 0.05169 | -2.789 | 0.0809 |
| 7 | -0.4683 | 0.05153 | -9.087 | 0.0000 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
| EPA_Rain_Zone | Difference of Means | SE of Difference | T-Value | Adjusted P-Value |


| 6 | 0.1958 | 0.04088 | 4.790 | 0.0000 |
| :---: | :---: | :---: | :---: | :---: |
| 7 | -0.1283 | 0.04068 | -3.153 | 0.0250 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.3241 | 0.04464 | -7.259 | 0.0000 |

TDS in EPA rain zone 2 is significantly different than in rain zones 4,6 , and 7 . No significant differences between EPA rain zone 2 and rain zones 3 and 5 .

TDS in EPA rain zone 3 is significantly different than in rain zones 4, 6, and 7. No significant differences between EPA rain zone 3 and rain zone 5 .

TDS in EPA rain zone 4 is significantly different than in rain zones 5 , and 7 . No significant differences between EPA rain zone 4 and rain zone 6 .

TDS in EPA rain zone 5 is significantly different than in rain zones 6 , and 7 .
TDS in EPA rain zone 6 is significantly different than in rain zone 7 .

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTDS |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | 0.0143 | 0.05710 | 0.250 | 0.9999 |
| 4 | 0.3272 | 0.04419 | 7.403 | 0.0000 |
| 5 | -0.0128 | 0.03086 | -0.415 | 0.9984 |
| 6 | 0.1830 | 0.03592 | 5.094 | 0.0000 |
| 7 | -0.1411 | 0.03570 | -3.952 | 0.0011 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.3129 | 0.06813 | 4.593 | 0.0001 |
| 5 | -0.0271 | 0.06034 | -0.449 | 0.9977 |
| 6 | 0.1687 | 0.06308 | 2.675 | 0.0803 |
| 7 | -0.1554 | 0.06295 | -2.468 | 0.1336 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.3400 | 0.04830 | -7.038 | 0.0000 |
| 6 | -0.1442 | 0.05169 | -2.789 | 0.0591 |
| 7 | -0.4683 | 0.05153 | -9.087 | 0.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |


| 6 | 0.1958 | 0.04088 | 4.790 | 0.0000 |
| :---: | :---: | :---: | :---: | :---: |
| 7 | -0.1283 | 0.04068 | -3.153 | 0.0201 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference SE of |  |  | Adjusted P-Value |
| EPA_Rain_Zone | of Means | Difference | T-Value |  |
| 7 | -0.3241 | 0.04464 | -7.259 | 0.0000 |

TDS in EPA rain zone 2 is significantly different than in rain zones 4,6 , and 7 . No significant differences between EPA rain zone 2 and rain zones 3 and 5 .

TDS in EPA rain zone 3 is significantly different than in rain zone 4 . No significant differences between EPA rain zone 3 and rain zone 5, 6 and 7 .

TDS in EPA rain zone 4 is significantly different than in rain zones 5 , and 7 . No significant differences between EPA rain zone 4 and rain zone 6 .

TDS in EPA rain zone 5 is significantly different than in rain zones 6 , and 7 .

TDS in EPA rain zone 6 is significantly different than in rain zone 7 .


Sites in Kentucky and Arlington Virginia are different than the remaining sites in the database.

## Interaction Plot (fitted means) for LTDS



## Total Suspended Solids

Summary statistics in LOG base 10 scale
Land use: Commercial


Land use: Industrial

| EPA_Rain_Zone | N det/est | N non detected |  | Mean | StDev | Minim | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 210 | 01 | 8 | 1.6434 | 0.4385 | 0.4771 | 2.5185 |
| 3 | $3 \quad 16$ | 16 | 0 | 1.351 | 0.809 | -0.367 | 2.505 |
| 4 | 4 | 16 | 1 | 2.177 | 0.617 | 1.322 | 3.073 |
| 5 | 5 43 | 43 | 4 | 2.1667 | 0.4113 | 1.3802 | 3.3962 |
| 6 | 6 | 57 13 | 13 | 2.458 | 0.5143 | 1.2041 | 3.3664 |
| 7 | 7 24 | 24 | 9 | 2.0825 | 0.3837 | 1.2041 | 3.0334 |

Land use: Residential

| EPA_Rain_Zone $\mathbf{N}$ det/est | N non sampled | Mean |  |  |  |  |  |  | StDev |  |  |  |  | Minimum Maximum |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 309 | 22 | 1.6047 | 0.4328 | 0.4074 | 2.9154 |  |  |  |  |  |  |  |  |  |  |  |
| 3 | 18 | 0 | 1.346 | 0.454 | 0.623 | 2.255 |  |  |  |  |  |  |  |  |  |  |  |
| 4 | 30 | 1 | 2.2706 | 0.5227 | 1.0414 | 3.3913 |  |  |  |  |  |  |  |  |  |  |  |
| 5 | 64 | 7 | 1.8435 | 0.3878 | 0.6021 | 2.7839 |  |  |  |  |  |  |  |  |  |  |  |
| 6 | 33 | 5 | 1.9127 | 0.4257 | 0.4771 | 2.5441 |  |  |  |  |  |  |  |  |  |  |  |
| 7 | 37 | 3 | 1.681 | 0.3785 | 0.8451 | 2.8791 |  |  |  |  |  |  |  |  |  |  |  |

## Descriptive Statistics: LTSS

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| LTSS | 2_KYLOTSR1RE | 3 | 0 | 1.6270 | 0.00411 | 1.5563 | 1.6812 |
|  | 2_KYLOTSR2ID | 3 | 0 | 2.054 | 0.348 | 1.380 | 2.477 |
|  | 2_KYLOTSR3RE | 3 | 0 | 2.179 | 0.167 | 1.708 | 2.435 |
|  | 2_KYLOTSR4ID | 4 | 0 | 1.942 | 0.0652 | 1.623 | 2.241 |
|  | 2_MDAACOMWID | 2 | 1 | 2.361 | 0.0494 | 2.204 | 2.519 |
|  | 2_MDAACOODRE | 3 | 0 | 1.443 | 0.137 | 1.204 | 1.869 |
|  | 2_MDAACOPPCO | 26 | 0 | 1.5368 | 0.0803 | 1.0406 | 2.1276 |
|  | 2_MDAACORKRE | 3 | 0 | 1.625 | 1.077 | 0.513 | 2.568 |
|  | 2_MDBACOBCID | 3 | 0 | 1.764 | 0.0373 | 1.568 | 1.954 |
|  | 2_MDBACOSCRE | 26 | 0 | 1.427 | 0.280 | 0.407 | 2.443 |
|  | 2_MDBACOTCID | 3 | 0 | 2.1936 | 0.00870 | 2.1335 | 2.3010 |
|  | 2_MDBACOWCRE | 3 | 0 | 1.867 | 0.0719 | 1.708 | 2.176 |
|  | 2_MDBCTYBOID | 3 | 0 | 1.704 | 0.0450 | 1.544 | 1.944 |
|  | 2_MDBCTYFMID | 3 | 0 | 1.894 | 0.0537 | 1.690 | 2.146 |
|  | 2_MDBCTYHORE | 3 | 0 | 1.3877 | 0.00590 | 1.3010 | 1.4472 |
|  | 2_MDBCTYHRRE | 3 | 0 | 1.616 | 0.0787 | 1.342 | 1.903 |
|  | 2_MDBCTYKOCO | 3 | 0 | 1.937 | 0.0529 | 1.672 | 2.083 |
| 2_MDMOCOBCCO | 2 | 1 | 1.038 | 0.0742 | 0.845 | 1.230 |  |



| 5_TXMEA002RE | 7 | 0 | 2.115 | 0.111 | 1.708 | 2.784 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 TXMEA003RE | 6 | 1 | 1.7057 | 0.0222 | 1.5441 | 1.8921 |
| 6 AZMCA001ID | 19 | 8 | 2.1062 | 0.1720 | 1.4771 | 2.6998 |
| 6_AZMCA003ID | 24 | 3 | 2.8248 | 0.1505 | 1.7924 | 3.3664 |
| 6_AZMCA005CO | 12 | 14 | 1.781 | 0.241 | 0.903 | 2.316 |
| 6_AZMCA006RE | 15 | 5 | 1.730 | 0.240 | 0.477 | 2.294 |
| 6_AZTUA001RE | 10 | 0 | 2.1178 | 0.0873 | 1.4314 | 2.5185 |
| 6 AZTUA002RE | 8 | 0 | 1.998 | 0.0974 | 1.556 | 2.544 |
| 6_AZTUA003CO | 8 | 0 | 1.997 | 0.141 | 1.462 | 2.708 |
| 6_AZTUA004ID | 7 | 0 | 2.385 | 0.359 | 1.204 | 3.140 |
| 6_CAALAL09ID | 7 | 2 | 2.2286 | 0.0190 | 2.0792 | 2.4472 |
| 7 -OREUA001CO | 14 | 2 | 1.758 | 0.225 | 0.903 | 2.580 |
| 7-OREUA003RE | 15 | 0 | 1.774 | 0.166 | 1.279 | 2.879 |
| 7-ORGRA003RE | 6 | 0 | 1.566 | 0.0616 | 1.176 | 1.914 |
| 7 -ORGRA004CO | 6 | 0 | 1.6843 | 0.0584 | 1.3617 | 2.0531 |
| 7-ORPOA001CO | 12 | 1 | 1.748 | 0.188 | 1.146 | 2.580 |
| 7 ORPOA003ID | 11 | 3 | 2.259 | 0.144 | 1.892 | 3.033 |
| $7{ }^{-}$ORPOA004ID | 9 | 4 | 2.0203 | 0.0659 | 1.5682 | 2.5011 |
| 7-ORPOA006RE | 11 | 2 | 1.7331 | 0.0922 | 1.2553 | 2.1139 |
| 7_ORSAA002CO | 5 | 1 | 1.8444 | 0.0486 | 1.5441 | 2.1492 |
| 7_ORSAA003ID | 4 | 2 | 1.737 | 0.186 | 1.204 | 2.111 |
| 7_ORSAA004RE | 5 | 1 | 1.427 | 0.261 | 0.845 | 1.903 |

## Xbar-R Chart of LTSS



## Individual Value Plot of LTSS vs LndRainLoc



General Linear Model: LTSS versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels | Values |
| :--- | :--- | ---: | :--- |
| Landuse | fixed | 3 | CO, ID, RE |
| EPA_Rain_Zone | fixed | 6 | $2,3,4,5,6,7$ |

Analysis of Variance for LTSS, using Adjusted SS for Tests

| Source | DF | Seq SS | AdjSS | Adj MS | F | P |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Landuse | 2 | 13.6706 | 3.9665 | 1.9833 | 10.20 | 0.000 |
| EPA_Rain_Zone | 5 | 54.4527 | 45.5820 | 9.1164 | 46.88 | 0.000 |
| Landuse*EPA_Rain_Zone | 10 | 18.7052 | 18.7052 | 1.8705 | 9.62 | 0.000 |
| Error | 961 | 186.8854 | 186.8854 | 0.1945 |  |  |
| Total | 978 | 273.7139 |  |  |  |  |

Both main factors and interactions are significant.

```
Bonferroni Simultaneous Tests
Response Variable LTSS
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & 0.0857 & 0.05508 & 1.556 & 0.3600 \\
RE & -0.1177 & 0.05245 & -2.243 & 0.0753
\end{tabular}
```

| Landuse $=$ ID subtracted from: |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  |  |  |  |  |
| Landuse | Difference | SE of |  | Adjusted |
| RE | -0.2034 | 0.04545 | -4.474 | 0.0000 |

Tukey Simultaneous Tests
Response Variable LTSS
All Pairwise Comparisons among Levels of Landuse
Landuse $=C O$ subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.0857 | 0.05508 | 1.556 | 0.2649 |
| RE | -0.1177 | 0.05245 | -2.243 | 0.0642 |

Landuse $=$ ID subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.2034 | 0.04545 | -4.474 | 0.0000 |

No difference in TSS concentrations between commercial and industrial land uses, and commercial and residential. TSS from industrial land uses is different than TSS from residential land use.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTSS |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone | $=2$ subtrac | ted from: |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1060 | 0.08141 | -1.303 | 1.0000 |
| 4 | 0.8089 | 0.06314 | 12.812 | 0.0000 |
| 5 | 0.2241 | 0.04307 | 5.203 | 0.0000 |
| 6 | 0.4657 | 0.05089 | 9.151 | 0.0000 |
| 7 | 0.2256 | 0.05043 | 4.473 | 0.0001 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.9150 | 0.09830 | 9.307 | 0.0000 |
| 5 | 0.3302 | 0.08679 | 3.804 | 0.0023 |
| 6 | 0.5717 | 0.09092 | 6.288 | 0.0000 |
| 7 | 0.3316 | 0.09067 | 3.658 | 0.0040 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.5848 | 0.06994 | -8.361 | 0.0000 |
| 6 | -0.3432 | 0.07500 | -4.576 | 0.0001 |
| 7 | -0.5833 | 0.07470 | -7.809 | 0.0000 |


|  | Difference | SE of |  | Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.241542 | 0.05911 | 4.08597 | 0.0007 |
| 7 | 0.001458 | 0.05873 | 0.02483 | 1.0000 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.2401 | 0.06467 | -3.712 | 0.0033 |

TSS in EPA rain zone 2 is significantly different than in rain zones 4, 5, 6, and 7. No significant differences between EPA rain zone 2 and rain zone 3 .

TSS in EPA rain zone 3 is significantly different than in rain zones $4,5,6$, and 7
TSS in EPA rain zone 4 is significantly different than in rain zones 5,6 , and 7
TSS in EPA rain zone 5 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 5 and rain zone 7 .

TSS in EPA rain zone 6 is significantly different than in rain zone 7.

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTSS |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1060 | 0.08141 | -1.303 | 0.7838 |
| 4 | 0.8089 | 0.06314 | 12.812 | 0.0000 |
| 5 | 0.2241 | 0.04307 | 5.203 | 0.0000 |
| 6 | 0.4657 | 0.05089 | 9.151 | 0.0000 |
| 7 | 0.2256 | 0.05043 | 4.473 | 0.0001 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.9150 | 0.09830 | 9.307 | 0.0000 |
| 5 | 0.3302 | 0.08679 | 3.804 | 0.0020 |
| 6 | 0.5717 | 0.09092 | 6.288 | 0.0000 |
| 7 | 0.3316 | 0.09067 | 3.658 | 0.0035 |
| EPA_Rain_Zone = 4 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.5848 | 0.06994 | -8.361 | 0.0000 |
| 6 | -0.3432 | 0.07500 | -4.576 | 0.0001 |
| 7 | -0.5833 | 0.07470 | -7.809 | 0.0000 |


|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.241542 | 0.05911 | 4.08597 | 0.0006 |
| 7 | 0.001458 | 0.05873 | 0.02483 | 1.0000 |

EPA_Rain_Zone $=6$ subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: | ---: |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.2401 | 0.06467 | -3.712 | 0.0028 |

TSS in EPA rain zone 2 is significantly different than in rain zones $4,5,6$, and 7 . No significant differences between EPA rain zone 2 and rain zone 3 .

TSS in EPA rain zone 3 is significantly different than in rain zones $4,5,6$, and 7
TSS in EPA rain zone 4 is significantly different than in rain zones 5,6 , and 7
TSS in EPA rain zone 5 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 5 and rain zone 7 .

TSS in EPA rain zone 6 is significantly different than in rain zone 7.


The model satisfies normality of residuals. No specific trend was observed in the residuals.

## Interaction Plot (fitted means) for LTSS



## Biochemical Oxygen Demand, 5 day ( $\mathrm{BOD}_{5}$ )

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean |  |  |  |
| :---: | :---: | ---: | :---: | :---: | :---: | ---: |
| 2 | 114 | 9 | 1.10 | 0.29 | 0.47 | 1.98 |
| 3 | 6 | 0 | 1.30 | 0.42 | 0.86 | 1.92 |
| 4 | 15 | 1 | 1.25 | 0.31 | 0.60 | 1.76 |
| 5 | 40 | 2 | 0.80 | 0.20 | 0.40 | 1.23 |
| 6 | 13 | 21 | 1.55 | 0.35 | 0.95 | 2.00 |
| 7 | 37 | 4 | 0.85 | 0.37 | -0.13 | 1.62 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean |  |  |  |
| :---: | :--- | ---: | :--- | :--- | :--- | ---: |
| 2 | 97 | 12 | 0.93 | 0.35 | 0.09 | 2.17 |
| 3 | 16 | 0 | 0.73 | 0.38 | 0.00 | 1.28 |
| 4 | 16 | 1 | 0.72 | 0.25 | 0.30 | 1.15 |
| 5 | 46 | 1 | 0.78 | 0.16 | 0.30 | 1.11 |
| 6 | 21 | 49 | 1.45 | 0.56 | 0.15 | 2.43 |
| 7 | 24 | 9 | 1.45 | 0.44 | 0.60 | 2.20 |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean |  |  |  |
| :---: | :---: | ---: | :---: | :---: | :---: | ---: |
| 2 | 290 | 41 | 1.04 | 0.31 | 0.30 | 2.35 |
| 3 | 16 | 2 | 0.86 | 0.32 | 0.30 | 1.45 |
| 4 | 31 | 0 | 1.17 | 0.35 | 0.30 | 1.84 |
| 5 | 67 | 4 | 0.90 | 0.25 | 0.38 | 1.70 |
| 6 | 15 | 23 | 1.41 | 0.28 | 1.00 | 2.11 |
| 7 | 37 | 3 | 0.70 | 0.32 | 0.13 | 1.61 |

## Descriptive Statistics: LBOD

| Variable |  |  |  |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| LndRainLoc | N | $\mathrm{N}^{*}$ | Mean | Variance | Minimum | Maximum |  |
|  | 2_KYLOTSR1RE | 2 | 1 | 1.224 | 0.286 | 0.845 | 1.602 |
|  | 2_KYLOTSR2ID | 2 | 1 | 1.7794 | 0.0103 | 1.7076 | 1.8513 |
|  | 2_KYLOTSR3RE | 3 | 0 | 1.2721 | 0.0196 | 1.1139 | 1.3802 |
|  | 2_KYLOTSR4ID | 3 | 1 | 1.484 | 0.642 | 0.602 | 2.167 |
|  | 2_MDAACOMWID | 3 | 0 | 0.816 | 0.0842 | 0.602 | 1.146 |
|  | 2_MDAACOODRE | 3 | 0 | 0.9299 | 0.0102 | 0.8451 | 1.0414 |
|  | 2_MDAACOPPCO | 26 | 0 | 1.1532 | 0.0696 | 0.5658 | 1.6217 |
|  | 2_MDAACORKRE | 2 | 1 | 0.929 | 0.213 | 0.602 | 1.255 |
|  | 2_MDBACOBCID | 3 | 0 | 1.0651 | 0.0197 | 0.9031 | 1.1461 |
|  | 2_MDBACOSCRE | 12 | 14 | 1.002 | 0.182 | 0.477 | 1.771 |
|  | 2_MDBACOTCID | 3 | 0 | 0.9105 | 0.0128 | 0.8451 | 1.0414 |
|  | 2_MDBACOWCRE | 3 | 0 | 1.168 | 0.0808 | 0.845 | 1.380 |
|  | 2_MDBCTYBOID | 3 | 0 | 1.292 | 0.0654 | 1.000 | 1.477 |
|  | 2_MDBCTYFMID | 3 | 0 | 1.221 | 0.0684 | 0.954 | 1.477 |
|  | 2_MDBCTYHORE | 3 | 0 | 1.142 | 0.0661 | 0.845 | 1.301 |
|  | 2_MDBCTYHRRE | 3 | 0 | 1.3366 | 0.0249 | 1.1761 | 1.4914 |
|  | 2_MDBCTYKOCO | 3 | 0 | 1.1286 | 0.00910 | 1.0414 | 1.2304 |
|  | 2_MDMOCOBCCO | 3 | 0 | 0.920 | 0.167 | 0.602 | 1.380 |
|  | 2_VAARLLP1RE | 7 | 1 | 0.844 | 0.120 | 0.477 | 1.279 |
|  | 2_VAARLTC4ID | 13 | 0 | 0.9687 | 0.0439 | 0.5052 | 1.2788 |


| 2_VACPTC1ARE | 8 | 0 | 0.940 | 0.106 | 0.477 | 1.301 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2_VACPTSF2RE | 0 | 3 | * | * | * | * |
| 2_VACPTYC1RE | 6 | 1 | 0.730 | 0.0849 | 0.477 | 1.301 |
| 2_VACPTYC3RE | 13 | 1 | 0.9777 | 0.0786 | 0.4771 | 1.3617 |
| 2-VACPTYC4CO | 14 | 0 | 0.8804 | 0.0321 | 0.4808 | 1.1139 |
| 2_VACPTYC5ID | 13 | 1 | 0.4737 | 0.0547 | 0.0931 | 1.0000 |
| 2-VACPTYO1ID | 0 | 3 | * | * | * | * |
| 2_VAHATYH1CO | 18 | 0 | 1.1351 | 0.0798 | 0.4771 | 1.6232 |
| 2-VAHATYH2ID | 19 | 0 | 0.9623 | 0.0655 | 0.6021 | 1.7324 |
| 2-VAHATYH3RE | 17 | 0 | 0.8799 | 0.0625 | 0.3010 | 1.2304 |
| 2_VAHATYH4RE | 17 | 0 | 1.0430 | 0.0626 | 0.6021 | 1.5315 |
| 2_VAHATYH5RE | 17 | 0 | 1.1146 | 0.1622 | 0.6990 | 2.3541 |
| 2_VAHCCOC1CO | 2 | 0 | 1.3551 | 0.0116 | 1.2788 | 1.4314 |
| 2-VAHCCOC2Co | 3 | 0 | 1.3254 | 0.0241 | 1.1461 | 1.4150 |
| 2_VAHCCON1ID | 1 | 1 | 1.1139 | * | 1.1139 | 1.1139 |
| 2_VAHCCON2ID | 2 | 1 | 1.2124 | 0.00879 | 1.1461 | 1.2788 |
| 2_VAHCCOR1RE | 3 |  | 1.3331 | 0.0128 | 1.2041 | 1.4150 |
| 2-VAHCCOR2RE | 3 | 0 | 1.0402 | 0.00157 | 1.0000 | 1.0792 |
| 2_VANFTMS5CO | 0 | 3 | * | * | * | * |
| 2_VANFTMS 6RE | 0 | 3 | * | * | * |  |
| 2_VANFTMS8CO | 0 | 3 | * | * | * |  |
| 2-VANFTMS9Co | 0 | 3 | * | * | * | * |
| 2_VANFTYN2RE | 29 | 0 | 1.1504 | 0.0818 | 0.4771 | 1.8325 |
| 2_VANFTYN3RE | 27 | 0 | 1.1703 | 0.0699 | 0.6990 | 1.7782 |
| 2_VANFTYN4CO | 27 | 0 | 1.1449 | 0.1274 | 0.4671 | 1.9823 |
| 2_VANFTYN5RE | 27 | 0 | 1.0673 | 0.1126 | 0.3010 | 1.5682 |
| 2-VANNTMF1RE | 0 | 1 | * | * | * | * |
| 2-VANNTMF 4RE | 0 | 3 | * | * | * | * |
| 2_VANNTNN1RE | 8 | 2 | 1.069 | 0.154 | 0.602 | 1.820 |
| 2_VANNTSF4RE | 0 | 2 | * | * | * | * |
| 2_VANNTSF6RE | 0 | 2 | * | * | * | * |
| 2_VAPMTYP1CO | 18 | 0 | 1.0331 | 0.0657 | 0.6990 | 1.4914 |
| 2_VAPMTYP2RE | 17 | 0 | 0.9654 | 0.0587 | 0.4771 | 1.4624 |
| 2_VAPMTYP4RE | 17 | 0 | 0.9154 | 0.0213 | 0.6990 | 1.2553 |
| 2_VAPMTYP5RE | 17 | 0 | 0.8928 | 0.1316 | 0.4771 | 1.9494 |
| 2_VAVBTYIIID | 0 | 3 | * | * | * | * |
| 2-VAVBTYR1RE | 0 | 5 | * | * | * | * |
| 2_VAVBTYV1RE | 26 | 1 | 1.1097 | 0.0792 | 0.6021 | 1.8388 |
| 2_VAVBTYV4ID | 29 | 1 | 0.8796 | 0.0448 | 0.4771 | 1.3424 |
| 3_ALMOCREORE | 3 | 0 | 0.760 | 0.368 | 0.301 | 1.447 |
| 3_ALMODAPHCO | 3 | 0 | 1.489 | 0.193 | 1.041 | 1.919 |
| 3_ALMOSARARE | 2 | 1 | 0.9141 | 0.00403 | 0.8692 | 0.9590 |
| 3_ALMOSIIVID | 3 | 0 | 0.8163 | 0.0258 | 0.6335 | 0.9345 |
| 3_ALMOSITVCO | 3 | 0 | 1.105 | 0.129 | 0.863 | 1.519 |
| 3_ALMOSIVIRE | 2 | 1 | 0.9460 | 0.00582 | 0.8921 | 1.0000 |
| 3_ALMOTHEOID | 3 | 0 | 0.8242 | 0.0279 | 0.6532 | 0.9868 |
| 3_GAATAT01ID | 10 | 0 | 0.670 | 0.216 | 0.000000000 | 1.279 |
| 3_GAATAT02RE | 9 | 0 | 0.8641 | 0.0887 | 0.3010 | 1.2304 |
| 4_KATOATWORE | 15 | 0 | 1.0089 | 0.1205 | 0.3010 | 1.5563 |
| 4_KATOBROORE | 16 | 0 | 1.3221 | 0.0771 | 0.9031 | 1.8388 |
| 4-KATOJACKCO | 15 | 1 | 1.2500 | 0.0989 | 0.6021 | 1.7559 |
| 4_KATOSTFEID | 16 | 1 | 0.7156 | 0.0606 | 0.3010 | 1.1461 |
| 5_TXARA001CO | 21 | 1 | 0.7764 | 0.0414 | 0.3979 | 1.0792 |
| 5-TXARA002RE | 19 | 2 | 0.8653 | 0.0899 | 0.3802 | 1.6721 |
| 5-TXARA003RE | 7 | 0 | 0.7098 | 0.0118 | 0.5911 | 0.8921 |
| 5_TXDAA001ID | 7 | 0 | 0.8324 | 0.0319 | 0.5441 | 1.1139 |
| 5_TXDAA002ID | 19 | 0 | 0.6886 | 0.0313 | 0.3010 | 0.9638 |
| 5-TXDAA004CO | 19 | 1 | 0.8254 | 0.0364 | 0.4314 | 1.2304 |
| 5_TXDAA005RE | 7 | 0 | 0.9312 | 0.0242 | 0.8325 | 1.2788 |
| 5_TXFWA004ID | 20 | 1 | 0.8374 | 0.0109 | 0.6628 | 1.1139 |
| 5_TXIRA001RE | 21 | 1 | 1.0430 | 0.0607 | 0.7559 | 1.6990 |
| 5_TXMEA002RE | 7 | 0 | 0.8293 | 0.0123 | 0.5798 | 0.8921 |
| 5_TXMEA003RE | 6 | 1 | 0.7328 | 0.0178 | 0.5315 | 0.9031 |


| 6_AZMCA001ID | 6 | 21 | 1.365 | 0.675 | 0.148 | 2.301 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6_AZMCA003ID | 6 | 21 | 1.462 | 0.279 | 0.929 | 2.431 |
| 6 AZMCA005CO | 6 | 20 | 1.462 | 0.114 | 0.954 | 1.820 |
| 6_AZMCA006RE | 3 | 17 | 1.2479 | 0.00967 | 1.1461 | 1.3424 |
| 6_AZTUA001RE | 6 | 4 | 1.4310 | 0.0583 | 1.2304 | 1.7993 |
| 6_AZTUA002RE | 6 | 2 | 1.469 | 0.141 | 1.000 | 2.114 |
| 6_AZTUA003CO | 7 | 1 | 1.629 | 0.136 | 0.954 | 2.000 |
| 6_AZTUA004ID | 5 | 2 | 1.739 | 0.142 | 1.301 | 2.204 |
| 6 CAALAL09ID | 4 | 5 | 1.1700 | 0.0397 | 0.8751 | 1.3010 |
| 7_OREUA001CO | 14 | 2 | 1.0315 | 0.1038 | 0.6021 | 1.4771 |
| 7_OREUA003RE | 15 | 0 | 0.7638 | 0.0964 | 0.3010 | 1.6128 |
| 7 ORGRA003RE | 6 | 0 | 0.515 | 0.0889 | 0.126 | 0.908 |
| 7_ORGRA004CO | 6 | 0 | 0.7651 | 0.0382 | 0.5051 | 0.9912 |
| 7 ORPOA001CO | 12 | 1 | 0.732 | 0.232 | -0.127 | 1.623 |
| 7_ORPOA003ID | 11 | 3 | 1.7267 | 0.0956 | 1.2304 | 2.2041 |
| 7 ORPOA004ID | 9 | 4 | 1.400 | 0.130 | 0.778 | 1.851 |
| 7 ORPOA006RE | 11 | 2 | 0.8678 | 0.0835 | 0.4771 | 1.4771 |
| 7_ORSAA002CO | 5 | 1 | 0.7268 | 0.0183 | 0.5798 | 0.8976 |
| 7 ORSAA003ID | 4 | 2 | 0.8193 | 0.0324 | 0.6021 | 1.0253 |
| 7_ORSAA004RE | 5 | 1 | 0.3976 | 0.0112 | 0.3222 | 0.5441 |




General Linear Model: LBOD versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed |  | CO, ID, |  |  |  |  |
| EPA Rain Zone fixed | 6 | 2, 3, | 5, 6, |  |  |  |
| Analysis of Variance for LBOD, using Adjusted SS for Tests |  |  |  |  |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 0.6031 | 1.2713 | 0.6357 | 6.35 | 0.002 |
| EPA Rain Zone | 5 | 16.4957 | 15.1699 | 3.0340 | 30.30 | 0.000 |
| Landuse*EPA_Rain_Zone | 10 | 14.3460 | 14.3460 | 1.4346 | 14.33 | 0.000 |
| Error | 882 | 88.3176 | 88.3176 | 0.1001 |  |  |
| Total | 899 | 119.7624 |  |  |  |  |

Main Effects and Interaction are significant.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LBOD |  |  |  |  |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.1341 | 0.04144 | -3.237 | 0.0038 |
| RE | -0.1274 | 0.04009 | -3.178 | 0.0046 |
| Landuse $=$ ID subtracted from: |  |  |  |  |


|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | 0.006708 | 0.03551 | 0.1889 | 1.000 |


| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LBOD |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse $=$ Co subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.1341 | 0.04144 | -3.237 | 0.0035 |
| RE | -0.1274 | 0.04009 | -3.178 | 0.0042 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | 0.006708 | 0.03551 | 0.1889 | 0.9805 |

BOD in commercial land uses is significantly different than in industrial or residential land uses. There is no difference in BOD concentrations between industrial and residential land uses.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of EPA_Rain_ZoneEPA Rain Zone $=2$ subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.0594 | 0.05913 | -1.004 | 1.0000 |
| 4 | 0.0245 | 0.04525 | 0.541 | 1.0000 |
| 5 | -0.1973 | 0.03063 | -6.442 | 0.0000 |
| 6 | 0.4479 | 0.04878 | 9.182 | 0.0000 |
| 7 | -0.0183 | 0.03629 | -0.503 | 1.0000 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.0838 | 0.07100 | 1.181 | 1.0000 |
| 5 | -0.1379 | 0.06270 | -2.200 | 0.4211 |
| 6 | 0.5073 | 0.07330 | 6.921 | 0.0000 |
| 7 | 0.0411 | 0.06565 | 0.626 | 1.0000 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.2218 | 0.04982 | -4.451 | 0.0001 |
| 6 | 0.4234 | 0.06264 | 6.760 | 0.0000 |
| 7 | -0.0427 | 0.05349 | -0.799 | 1.0000 |


|  | Difference | SE of |  | Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.6452 | 0.05304 | 12.164 | 0.0000 |
| 7 | 0.1790 | 0.04185 | 4.278 | 0.0003 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.4662 | 0.05650 | -8.250 | 0.0000 |

BOD5 in EPA rain zone 2 is significantly different than in rain zones 5, and 6 . No significant differences between EPA rain zone 2 and rain zones 3,4 , and 7 .

BOD5 in EPA rain zone 3 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 3 and rain zones 4,5 and 7 .

BOD5 in EPA rain zone 4 is significantly different than in rain zones 5, and 6 . No significant differences between EPA rain zone 4 and rain zone 7 .

BOD5 in EPA rain zone 5 is significantly different than in rain zones 6 , and 7 .
BOD5 in EPA rain zone 6 is significantly different than in rain zone 7 .

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LBOD |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.0594 | 0.05913 | -1.004 | 0.9168 |
| 4 | 0.0245 | 0.04525 | 0.541 | 0.9945 |
| 5 | -0.1973 | 0.03063 | -6.442 | 0.0000 |
| 6 | 0.4479 | 0.04878 | 9.182 | 0.0000 |
| 7 | -0.0183 | 0.03629 | -0.503 | 0.9961 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.0838 | 0.07100 | 1.181 | 0.8462 |
| 5 | -0.1379 | 0.06270 | -2.200 | 0.2377 |
| 6 | 0.5073 | 0.07330 | 6.921 | 0.0000 |
| 7 | 0.0411 | 0.06565 | 0.626 | 0.9891 |
| EPA_Rain_Zone = 4 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.2218 | 0.04982 | -4.451 | 0.0001 |
| 6 | 0.4234 | 0.06264 | 6.760 | 0.0000 |
| 7 | -0.0427 | 0.05349 | -0.799 | 0.9677 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |

$\left.\begin{array}{lrrrr}6 & 0.6452 & 0.05304 & 12.164 & 0.0000 \\ 7 & & 0.1790 & 0.04185 & 4.278\end{array}\right) 0.0003$

BOD5 in EPA rain zone 2 is significantly different than in rain zones 5 , and 6 . No significant differences between EPA rain zone 2 and rain zones 3,4 , and 7 .

BOD5 in EPA rain zone 3 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 3 and rain zones 4, 5 and 7 .

BOD5 in EPA rain zone 4 is significantly different than in rain zones 5 , and 6 . No significant differences between EPA rain zone 4 and rain zone 7 .

BOD5 in EPA rain zone 5 is significantly different than in rain zones 6 , and 7.
BOD5 in EPA rain zone 6 is significantly different than in rain zone 7.


The residuals are normally distributed and did not indicate any specific trend in any of the residuals plots.

## Interaction Plot (fitted means) for LBOD



## Chemical Oxygen Demand (COD)

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ |  |  |  |  |  |  |  | Mean |  |  |  |  | StDev Minimum Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 82 | 41 | 1.84 | 0.32 | 0.90 | 2.80 |  |  |  |  |  |  |  |  |  |
| 3 | 6 | 0 | 1.85 | 0.35 | 1.36 | 2.38 |  |  |  |  |  |  |  |  |  |
| 4 | 0 | $16^{*}$ | ${ }^{*}$ | ${ }^{*}$ | $*$ |  |  |  |  |  |  |  |  |  |  |
| 5 | 41 | 1 | 1.65 | 0.22 | 1.15 | 2.18 |  |  |  |  |  |  |  |  |  |
| 6 | 33 | 1 | 2.29 | 0.24 | 1.89 | 2.76 |  |  |  |  |  |  |  |  |  |
| 7 | 37 | 4 | 1.65 | 0.30 | 0.90 | 2.52 |  |  |  |  |  |  |  |  |  |

Land use: Industrial

| EPA_Rain_Zone $\mathbf{N}$ det/est | N ND/NZ | Mean |  |  |  |  | StDev Minimum Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 2 | 85 | 24 | 1.72 | 0.32 | 0.30 | 2.53 |  |
| 3 | 16 | 0 | 1.52 | 0.36 | 0.60 | 2.06 |  |
| 4 | 0 | $17^{*}$ | ${ }^{*}$ | ${ }^{*}$ | ${ }^{*}$ |  |  |
| 5 | 45 | 2 | 1.59 | 0.29 | 0.98 | 2.40 |  |
| 6 | 56 | 14 | 2.27 | 0.35 | 1.26 | 2.96 |  |
| 7 | 24 | 9 | 1.90 | 0.33 | 1.26 | 2.45 |  |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 265 |  | 66 | 1.79 | 0.29 | 0.70 | 2.79 |
| 3 | 316 |  | 2 | 1.58 | 0.31 | 0.78 | 2.15 |
| 4 | 4 | 0 | 31* | * | * * | * * | * |
| 5 | $5 \quad 69$ |  | 2 | 1.81 | 0.29 | 1.00 | 2.68 |
| 6 | 637 |  | 1 | 2.12 | 0.28 | 1.51 | 2.57 |
| 7 | $7 \quad 37$ |  | 3 | 1.52 | 0.35 | 0.95 | 2.48 |

## Descriptive Statistics: LCOD

| Variable | LndRainLoc | N | $\mathrm{N}^{*}$ | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LCOD | 2_KYLOTSR1RE | 3 | 0 | 1.820 | 0.0899 | 1.477 | 2.033 |
|  | 2_KYLOTSR2ID | 3 | 0 | 2.021 | 0.0591 | 1.792 | 2.276 |
|  | 2_KYLOTSR3RE | 3 | 0 | 2.0211 | 0.00379 | 1.9542 | 2.0755 |
|  | 2_KYLOTSR4ID | 4 | 0 | 1.866 | 0.188 | 1.230 | 2.201 |
|  | 2_MDAACOMWID | 0 | 3 | * | * | * | * |
|  | 2_MDAACOODRE | 0 | 3 | * | * | * | * |
|  | 2_MDAACOPPCO | 0 | 26 | * | * | * |  |
|  | 2_MDAACORKRE | 0 | 3 | * | * | * | * |
|  | 2_MDBACOBCID | 0 | 3 | * | * | * |  |
|  | 2_MDBACOSCRE | 0 | 26 | * | * | * | * |
|  | 2 MDBACOTCID | 0 | 3 | * | * | * | * |
|  | 2-MDBACOWCRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYBOID | 0 | 3 | * | * | * | * |
|  | 2 MDBCTYFMID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHORE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHRRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYKOCO | 0 | 3 | * | * | * | * |
|  | 2-MDMOCOBCCO | 0 | 3 | * | * | * | * |
|  | 2_VAARLLP1RE | 8 | 0 | 1.634 | 0.0917 | 1.146 | 2.204 |
|  | 2_VAARLTC4ID | 12 | 1 | 1.689 | 0.273 | 0.301 | 2.531 |



| 6_AZTUA001RE | 9 | 1 | 2.3530 | 0.0416 | 2.0531 | 2.5682 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 6_AZTUA002RE | 8 | 0 | 2.0413 | 0.0656 | 1.6232 | 2.5185 |
| 6_AZTUA003CO | 8 | 0 | 2.4071 | 0.0733 | 1.9590 | 2.7649 |
| 6_AZTUA004ID | 7 | 0 | 2.4926 | 0.0667 | 2.1761 | 2.9571 |
| 6_CAALAL09ID | 0 | 9 | $*$ | $*$ | $*$ | $\star$ |
| 7_OREUA001CO | 14 | 2 | 1.6907 | 0.0554 | 1.3010 | 2.1139 |
| 7_OREUA003RE | 15 | 0 | 1.5191 | 0.0972 | 0.9638 | 2.1461 |
| 7_ORGRA003RE | 6 | 0 | 1.4900 | 0.0406 | 1.2304 | 1.7404 |
| 7_ORGRA004CO | 6 | 0 | 1.574 | 0.0822 | 1.279 | 2.083 |
| 7_ORPOA001CO | 12 | 1 | 1.643 | 0.177 | 0.903 | 2.519 |
| 7_ORPOA003ID | 11 | 3 | 2.0934 | 0.0832 | 1.5315 | 2.4533 |
| 7_ORPOA004ID | 9 | 4 | 1.8285 | 0.0710 | 1.3424 | 2.1461 |
| 7_ORPOA006RE | 11 | 2 | 1.601 | 0.202 | 0.954 | 2.477 |
| 7_ORSAA002CO | 5 | 1 | 1.6702 | 0.0412 | 1.4472 | 1.9590 |
| 7_ORSAA003ID | 4 | 2 | 1.554 | 0.0746 | 1.255 | 1.839 |
| 7_ORSAA004RE | 5 | 1 | 1.344 | 0.153 | 1.079 | 2.033 |




## General Linear Model: LCOD versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels | Values |
| :--- | :--- | ---: | :--- |
| Landuse | fixed | 3 | CO, ID, RE |
| EPA_Rain_Zone | fixed | 5 | $2,3,5,6,7$ |

Analysis of Variance for LCOD, using Adjusted SS for Tests

| Source | DF | Seq.SS | Adj SS | Adj MS | F | P |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Landuse | 2 | 0.5831 | 0.6005 | 0.3003 | 3.33 | 0.036 |
| EPA_Rain_Zone | 4 | 28.0369 | 25.5423 | 6.3856 | 70.89 | 0.000 |
| Landuse*EPA_Rain_Zone | 8 | 5.2090 | 5.2090 | 0.6511 | 7.23 | 0.000 |
| Error | 834 | 75.1295 | 75.1295 | 0.0901 |  |  |
| Total | 848 | 108.9585 |  |  |  |  |

Main factors and interactions are significant.

```
Bonferroni Simultaneous Tests
Response Variable LCOD
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & -0.05597 & 0.03874 & -1.445 & 0.4468 \\
RE & -0.09694 & 0.03775 & -2.568 & 0.0312
\end{tabular}
```

|  | Difference | SE of |  | Adjusted |
| :---: | :---: | :---: | :---: | :---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.04097 | 0.03237 | -1.265 | 0.6182 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LCOD |  |  |  |  |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.05597 | 0.03874 | -1.445 | 0.3180 |
| RE | -0.09694 | 0.03775 | -2.568 | 0.0276 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.04097 | 0.03237 | -1.265 | 0.4148 |

COD in commercial land use is significantly different than COD in residential areas. There is not enough evidence that indicates a difference between commercial and industrial land uses or between residential and industrial areas.

| Response Variable LCOD |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1301 | 0.05654 | -2.300 | 0.2169 |
| 5 | -0.1008 | 0.02982 | -3.380 | 0.0076 |
| 6 | 0.4446 | 0.03210 | 13.854 | 0.0000 |
| 7 | -0.0914 | 0.03515 | -2.600 | 0.0949 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.02927 | 0.05942 | 0.4926 | 1.0000 |
| 6 | 0.57470 | 0.06060 | 9.4841 | 0.0000 |
| 7 | 0.03866 | 0.06227 | 0.6209 | 1.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.545430 | 0.03693 | 14.7673 | 0.0000 |
| 7 | 0.009388 | 0.03962 | 0.2370 | 1.0000 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |

$\begin{array}{lllll}7 & -0.5360 & 0.04136 & -12.96 & 0.0000\end{array}$
COD in EPA rain zone 2 is significantly different than in rain zones 5 , and 6 . No significant differences between EPA rain zone 2 and rain zones 3 , and 7 .

COD in EPA rain zone 3 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 3 and rain zones 5, and 7 .

COD in EPA rain zone 5 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 5 and rain zone 7 .

COD in EPA rain zone 6 is significantly different than in rain zone 7.

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LCOD |  |  |  |  |
| All Pairwise Comparisons |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1301 | 0.05654 | -2.300 | 0.1447 |
| 5 | -0.1008 | 0.02982 | -3.380 | 0.0065 |
| 6 | 0.4446 | 0.03210 | 13.854 | 0.0000 |
| 7 | -0.0914 | 0.03515 | -2.600 | 0.0703 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.02927 | 0.05942 | 0.4926 | 0.9881 |
| 6 | 0.57470 | 0.06060 | 9.4841 | 0.0000 |
| 7 | 0.03866 | 0.06227 | 0.6209 | 0.9718 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.545430 | 0.03693 | 14.7673 | 0.0000 |
| 7 | 0.009388 | 0.03962 | 0.2370 | 0.9993 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.5360 | 0.04136 | -12.96 | 0.0000 |

COD in EPA rain zone 2 is significantly different than in rain zones 5 , and 6 . No significant differences between EPA rain zone 2 and rain zones 3 , and 7 .

COD in EPA rain zone 3 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 3 and rain zones 5, and 7 .

COD in EPA rain zone 5 is significantly different than in rain zone 6 . No significant differences between EPA rain zone 5 and rain zone 7 .

COD in EPA rain zone 6 is significantly different than in rain zone 7 .


Residuals are normally distributed, no trend was observed in the residual plots.

Interaction Plot (fitted means) for LCOD


## Ammonia ( $\mathbf{N H}_{3}$ )

Summary statistics in LOG base 10 scale
Land use: Commercial

| EPA_Rain_Zone $\mathbf{N}$ det/est | N ND/NZ | Mean |  |  |  |  | StDev |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 81 | 28 | -0.616 | 0.3735 | -1.675 | 0.2041 |
|  | 3 | 0 | 16* |  | * * | * * |  |
|  | 4 | 0 | 17* |  | * * | * * |  |
|  | 5 | 0 | 47* |  | * * | * * |  |
|  | 6 | 51 | 19 | -0.132 | 0.4194 | -1.5229 | 0.716 |
|  | 7 | 18 | 15 | -0.591 | 0.469 | -1.473 | 0.23 |

Land use: Residential

| EPA_Rain_Zone N det/est | N ND/NZ | Mean |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 2 | 262 | 69 | -0.629 | 0.3846 | -1.6214 | 0.1732 |  |
| 3 | 0 | $18^{*}$ | $*$ | $*$ | $*$ |  |  |
| 4 | 0 | $31^{*}$ | $*$ | $*$ | $*$ |  |  |
| 5 | 0 | $71^{*}$ | $*$ | $*$ | $*$ |  |  |
| 6 | 20 | 18 | -0.049 | 0.3025 | -0.7212 | 0.5315 |  |
| 7 | 26 | 14 | -0.962 | 0.771 | -2.662 | 0.748 |  |

## Descriptive Statistics: LNH3

| Variable | LndRainLoc | N | $\mathrm{N}^{*}$ | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LNH3 | 2 KYLOTSR1RE | 3 | 0 | -0.476 | 0.0421 | -0.678 | -0.268 |
|  | 2-KYLOTSR2ID | 3 | 0 | -0.5191 | 0.00777 | -0.6198 | -0.4559 |
|  | 2_KYLOTSR3RE | 3 | 0 | -0.503 | 0.0396 | -0.699 | -0.301 |
|  | 2_KYLOTSR4ID | 4 | 0 | -0.565 | 0.139 | -1.097 | -0.284 |
|  | 2_MDAACOMWID | 0 | 3 | * | * | * | * |
|  | 2_MDAACOODRE | 0 | 3 | * | * | * | * |
|  | 2 MDAACOPPCO | 0 | 26 | $\star$ | * | * | * |
|  | 2-MDAACORKRE | 0 | 3 | * | * | * | * |
|  | 2_MDBACOBCID | 0 | 3 | * | * | * | * |
|  | 2_MDBACOSCRE | 0 | 26 | * | * | * | * |
|  | 2_MDBACOTCID | 0 | 3 | * | * | * | * |
|  | 2-MDBACOWCRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYBOID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYFMID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHORE | 0 | 3 | * | * | * | * |



## Xbar-R Chart of LNH3





## General Linear Model: LNH3 versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID, |  |  |  |  |
| EPA_Rain_Zone fixed | 3 | 2, 6, 7 |  |  |  |  |
| Analysis of Variance for | for LNH3 | 3, using | Adjusted | $S$ for Tes |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 6.6025 | 2.5582 | 1.2791 | 7.18 | 0.001 |
| EPA_Rain_Zone | 2 | 29.6281 | 29.4237 | 14.7118 | 82.60 | 0.000 |
| Landuse*EPA_Rain_Zone | 4 | 2.8313 | 2.8313 | 0.7078 | 3.97 | 0.003 |
| Error | 579 | 103.1278 | 103.1278 | 0.1781 |  |  |
| Total | 587 | 142.1896 |  |  |  |  |

Main factors and interactions are significant.

```
Bonferroni Simultaneous Tests
Response Variable LNH3
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & -0.1302 & 0.06021 & -2.163 & 0.0928 \\
RE & -0.2306 & 0.06099 & -3.782 & 0.0005
\end{tabular}
\begin{tabular}{lrrrr} 
Landuse \(=\) ID subtracted from: & \\
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
RE & -0.1004 & 0.05965 & -1.683 & 0.2786
\end{tabular}
```

Tukey Simultaneous Tests
Response Variable LNH3
All Pairwise Comparisons among Levels of Landuse
Landuse $=$ CO subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.1302 | 0.06021 | -2.163 | 0.0777 |
| RE | -0.2306 | 0.06099 | -3.782 | 0.0005 |

Landuse $=$ ID subtracted from:

|  | Difference | SE of | Adjusted |  |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.1004 | 0.05965 | -1.683 | 0.2116 |

There is a significant difference in ammonia concentrations between commercial and residential land uses. This difference was not observed between commercial and industrial areas nor between residential and industrial areas.

```
Bonferroni Simultaneous Tests
Response Variable LNH3
All Pairwise Comparisons among Levels of EPA_Rain_Zone
EPA_Rain_Zone = 2 subtracted from:
```

|  | Difference <br> Of Means | Difference | T-Value | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| EPA_Rain_Zone |  |  |  |  |
| 6 | 0.5929 | 0.05291 | 11.205 | 0.0000 |
| 7 | -0.2317 | 0.05669 | -4.088 | 0.0001 |

Ammonia in EPA rain zone 2 is significantly different than in rain zones 6 , and 7.
Ammonia in EPA rain zone 6 is significantly different than in rain zone 7.

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.5929 | 0.05291 | 11.205 | 0.0000 |
| 7 | -0.2317 | 0.05669 | -4.088 | 0.0001 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.8246 | 0.06993 | -11.79 | 0.0000 |

Ammonia in EPA rain zone 2 is significantly different than in rain zones 6 , and 7 .
Ammonia in EPA rain zone 6 is significantly different than in rain zone 7.


Residuals are normally distributed. There is no trend in any of the plots of residuals.

## Interaction Plot (fitted means) for LNH3



Nitrite plus Nitrate $\left(\mathbf{N O}_{\mathbf{2}}+\mathbf{N O}_{\mathbf{3}}\right)$
Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ |  | Mean | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 21 | 114 | 9 | -0.297 | 0.424 | -1.7096 | 0.8633 |
| 3 | 3 | 0 | 6 * |  | * * | * * | * |
| 4 | 4 | 0 | 16* |  | * * | * * | * |
| 5 | 5 | 40 | 2 | -0.313 | 0.2076 | -0.8196 | 0.0531 |
| 6 |  | 30 | 4 | -0.004 | 0.3379 | -1.2218 | 0.5911 |
| 7 |  | 37 | 4 | -0.552 | 0.3409 | -1.0969 | 0.415 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ |  | Mean | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 | 94 | 15 | -0.308 | 0.3884 | -1.5229 | 0.3729 |
|  | 3 | 9 | 7 | -0.606 | 0.498 | -1.347 | 0.22 |
|  | 4 | 0 | 17* |  | * * | * * | * |
|  | 5 | 46 | 1 | -0.212 | 0.2213 | -1.1612 | 0.2279 |
|  | 6 | 58 | 12 | 0.2018 | 0.2552 | -0.3372 | 0.6721 |
|  | 7 | 24 | 9 | -0.849 | 0.4197 | -1.661 | -0.1549 |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ |  | Mean | StDev | Minimum M | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 23 | 302 | 29 | -0.307 | 0.3006 | -1.3979 | 0.4624 |
| 3 | 3 | 7 | 11 | -0.068 | 0.659 | -1.102 | 0.5 |
| 4 | 4 | 0 | 31 * |  | * * | * * | * |
|  | 5 | 69 | 2 | -0.177 | 0.26 | -0.9586 | 0.8555 |
|  | 6 | 38 | 0 | 0.0566 | 0.1794 | -0.301 | 0.4713 |
| 7 | 7 | 36 | 4 | -0.248 | 0.4138 | -1.1094 | 0.5441 |

## Descriptive Statistics: LNO2

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LNO2 | 2_KYLOTSR1RE | 0 | 3 | * | * | * | $\star$ |
|  | 2_KYLOTSR2ID | 0 | 3 | * | * | * | $\star$ |
|  | 2 KYLOTSR3RE | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR4ID | 0 | 4 | * | * | * | * |
|  | 2_MDAACOMWID | 3 | 0 | -0.4406 | 0.0173 | -0.5850 | -0.3279 |
|  | 2_MDAACOODRE | 3 | 0 | -0.1157 | 0.0189 | -0.2147 | 0.0414 |
|  | 2 MDAACOPPCO | 26 | 0 | -0.516 | 0.373 | -1.710 | 0.520 |
|  | 2_MDAACORKRE | 3 | 0 | -0.103 | 0.0861 | -0.409 | 0.176 |
|  | 2_MDBACOBCID | 3 | 0 | 0.2635 | 0.0103 | 0.1461 | 0.3222 |
|  | 2_MDBACOSCRE | 26 | 0 | -0.1021 | 0.1018 | -0.6990 | 0.3909 |
|  | 2_MDBACOTCID | 3 | 0 | -0.1204 | 0.0201 | -0.2840 | -0.0315 |
|  | 2-MDBACOWCRE | 3 | 0 | -0.269 | 0.478 | -1.067 | 0.146 |
|  | 2_MDBCTYBOID | 3 | 0 | -0.0220 | 0.0510 | -0.167 | 0.238 |
|  | 2_MDBCTYFMID | 3 | 0 | 0.0348 | 0.0288 | -0.0706 | 0.2304 |
|  | 2_MDBCTYHORE | 3 | 0 | 0.0908 | 0.0298 | -0.1079 | 0.2041 |
|  | 2-MDBCTYHRRE | 3 | 0 | 0.0661 | 0.0356 | -0.149 | 0.204 |



| 7_OREUA001CO | 14 | 2 | -0.5824 | 0.0691 | -1.0402 | -0.0269 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 7_OREUA003RE | 15 | 0 | -0.2205 | 0.0712 | -0.7447 | 0.1461 |
| 7_ORGRA003RE | 6 | 0 | 0.1092 | 0.0518 | -0.0969 | 0.5441 |
| 7_ORGRA004CO | 6 | 0 | -0.427 | 0.0738 | -0.921 | -0.155 |
| 7_ORPOA001CO | 12 | 1 | -0.602 | 0.220 | -1.097 | 0.415 |
| 7_ORPOA003ID | 11 | 3 | -0.860 | 0.171 | -1.607 | -0.301 |
| 7_ORPOA004ID | 9 | 4 | -0.951 | 0.227 | -1.661 | -0.155 |
| 7_ORPOA006RE | 10 | 3 | -0.650 | 0.197 | -1.109 | 0.146 |
| 7_ORSAA002CO | 5 | 1 | -0.493 | 0.0850 | -0.745 | -0.0458 |
| 7_ORSAA003ID | 4 | 2 | -0.587 | 0.0515 | -0.824 | -0.301 |
| 7_ORSAA004RE | 5 | 1 | 0.0437 | 0.0363 | -0.1805 | 0.2553 |




General Linear Model: LNO3 versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels | Values |
| :--- | :--- | ---: | :--- |
| Landuse | fixed | 3 | CO, ID, RE |
| EPA_Rain_Zone | fixed | 4 | $2,5,6,7$ |

Analysis of Variance for LNO3, using Adjusted SS for Tests

| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Landuse | 2 | 0.8956 | 2.0365 | 1.0182 | 9.65 | 0.000 |
| EPA_Rain_Zone | 3 | 24.1930 | 22.5935 | 7.5312 | 71.35 | 0.000 |
| Landuse*EPA_Rain_Zone | 6 | 6.2276 | 6.2276 | 1.0379 | 9.83 | 0.000 |
| Error | 876 | 92.4598 | 92.4598 | 0.1055 |  |  |
| Total | 887 | 123.7760 |  |  |  |  |

Main effects and interactions are significant.

```
Bonferroni Simultaneous Tests
Response Variable LNO3
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & \begin{tabular}{r} 
SE of
\end{tabular} & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & -0.000404 & 0.03497 & -0.01155 & 1.0000 \\
RE & 0.122472 & 0.03309 & 3.70067 & 0.0007
\end{tabular}
```



There is not a significant difference in nitrate concentrations between commercial and industrial land uses. Nitrates in commercial land use areas are significantly different than in residential land use areas. Nitrates in industrial land use areas are significantly different than in residential land use areas.

```
Bonferroni Simultaneous Tests
Response Variable LNO3
All Pairwise Comparisons among Levels of EPA_Rain_Zone
EPA_Rain_Zone = 2 subtracted from:
\begin{tabular}{lrrrr} 
& \begin{tabular}{r} 
Difference \\
of Means
\end{tabular} & \begin{tabular}{rl} 
SE of
\end{tabular} & Adjusted \\
EPA_Rain_Zone & 0.0701 & 0.03138 & 2.234 & 0.1545 \\
5 & 0.3887 & 0.03418 & 11.372 & 0.0000 \\
6 & -0.2455 & 0.03739 & -6.567 & 0.0000
\end{tabular}
EPA_Rain_Zone = 5 subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
EPA_Rain_Zone & of Means & Difference & T-Value & P-Value \\
6 & 0.3186 & 0.04025 & 7.916 & 0.0000 \\
7 & -0.3156 & 0.04301 & -7.339 & 0.0000
\end{tabular}
EPA_Rain_Zone = 6 subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & Adjusted \\
EPA_Rain_Zone & of Means & Difference & T-Value & P-Value \\
7 & -0.6342 & 0.04509 & -14.07 & 0.0000
\end{tabular}
```

Nitrates in EPA rain zone 2 are significantly different than in rain zones 6, and 7. Nitrates in EPA rain zone 2 are not significantly than in rain zone 5 .

Nitrates in EPA rain zone 5 are significantly different than in rain zones 6, and 7.
Nitrates in EPA rain zone 6 are significantly different than in rain zones 7.

```
Tukey Simultaneous Tests
Response Variable LNO3
All Pairwise Comparisons among Levels of EPA Rain_Zone
EPA_Rain_Zone = 2 subtracted from:
\begin{tabular}{lrrrr} 
& \begin{tabular}{r} 
Difference \\
Of Means
\end{tabular} & \begin{tabular}{r} 
SE of
\end{tabular} & & \begin{tabular}{r} 
Adjusted
\end{tabular} \\
EPA_Rain_Zone & \begin{tabular}{rl} 
Difence
\end{tabular} & T-Value & P-Value
\end{tabular}
EPA_Rain_Zone = 5 subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
EPA_Rain_Zone & \begin{tabular}{rl} 
Df Means \\
of Mifference
\end{tabular} & T-Value & P-Value \\
6 & 0.3186 & 0.04025 & 7.916 & 0.0000 \\
7 & -0.3156 & 0.04301 & -7.339 & 0.0000
\end{tabular}
EPA_Rain_Zone = 6 subtracted from:
\begin{tabular}{lrrrr} 
& \begin{tabular}{rl} 
Difference \\
EPA_Rain_Zone & SE of
\end{tabular} & & Adjusted \\
7 & Of Means & Difference & T-Value & P-Value \\
7 & -0.6342 & 0.04509 & -14.07 & 0.0000
\end{tabular}
```

Nitrates in EPA rain zone 2 are significantly different than in rain zones 6, and 7. Nitrates in EPA rain zone 2 are not significantly than in rain zone 5 .

Nitrates in EPA rain zone 5 are significantly different than in rain zones 6, and 7.
Nitrates in EPA rain zone 6 are significantly different than in rain zones 7.



## Total Kjeldahl Nitrogen (TKN)

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 114 | 9 | 0.174 | 0.331 | -1.301 | 0.939 |
| 3 | 6 | 0 | 0.048 | 0.219 | -0.143 | 0.477 |
| 4 | 0 | 16* |  | * * | * * | * 0.477 |
| 5 | 40 | 2 | -0.028 | 0.269 | -0.586 | 0.602 |
| 6 | 34 | 0 | 0.567 | 0.264 | 0.017 | 1.079 |
| 7 | 37 | 4 | -0.043 | 0.471 | -1.696 | 0.919 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  | Minimum | Maximum |  |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 94 | 15 | 0.001 | 0.307 | -0.790 | 1.000 |  |  |
| 3 | 15 | 1 | -0.194 | 0.415 | -1.347 | 0.248 |  |  |
| 4 | 0 | $17^{*}$ |  | $*$ | $*$ |  | $*$ |  |
| 5 | 46 | 1 | -0.074 | 0.294 | -1.000 | 0.699 |  |  |
| 6 | 64 | 6 | 0.519 | 0.411 | -0.602 | 1.204 |  |  |
| 7 | 24 | 9 | 0.192 | 0.487 | -1.843 | 0.771 |  |  |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean | StDev | Minimum | Maximum |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 302 | 29 | 0.125 | 0.299 | -1.000 | 1.556 |  |
| 3 | 17 | 1 | 0.048 | 0.161 | -0.252 | 0.378 |  |
| 4 | 0 | $31^{*}$ |  | $*$ | $*$ | $*$ |  |
| 5 | 69 | 2 | 0.187 | 0.338 | -0.691 | 1.000 |  |
| 6 | 37 | 1 | 0.504 | 0.300 | -0.301 | 1.041 |  |
| 7 | 37 | 3 | -0.167 | 0.579 | -2.314 | 1.065 |  |

## Descriptive Statistics: LTKN

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LTKN | 2_KYLOTSR1RE | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR2ID | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR3RE | 0 | 3 | * | * | * | * |
|  | 2 KYLOTSR4ID | 0 | 4 | * | * | * | * |
|  | 2_MDAACOMWID | 3 | 0 | 0.0889 | 0.114 | -0.222 | 0.447 |
|  | 2_MDAACOODRE | 3 | 0 | -0.180 | 0.505 | -1.000 | 0.230 |
|  | 2_MDAACOPPCO | 26 | 0 | 0.1408 | 0.1412 | -0.6353 | 0.9385 |
|  | 2 MDAACORKRE | 3 | 0 | 0.0698 | 0.255 | -0.222 | 0.653 |
|  | 2_MDBACOBCID | 3 | 0 | 0.143 | 0.669 | -0.789 | 0.740 |
|  | 2_MDBACOSCRE | 26 | 0 | 0.1970 | 0.0886 | -0.4656 | 0.8513 |
|  | 2_MDBACOTCID | 3 | 0 | 0.173 | 0.0699 | 0.000000000 | 0.477 |
|  | 2_MDBACOWCRE | 3 | 0 | 0.4203 | 0.0206 | 0.3010 | 0.5798 |
|  | 2_MDBCTYBOID | 3 | 0 | 0.428 | 0.249 | 0.0792 | 1.000 |
|  | 2_MDBCTYFMID | 3 | 0 | -0.162 | 0.276 | -0.759 | 0.230 |
|  | 2_MDBCTYHORE | 3 | 0 | 0.441 | 0.587 | -0.408 | 1.079 |
|  | 2_MDBCTYHRRE | 2 | 1 | 1.024 | 0.567 | 0.491 | 1.556 |
|  | 2_MDBCTYKOCO | 3 | 0 | 0.360 | 0.0986 | 0.0792 | 0.699 |
|  | 2_MDMOCOBCCO | 3 | 0 | -0.100 | 1.121 | -1.301 | 0.699 |
|  | 2_VAARLLP1RE | 8 | 0 | 0.0813 | 0.0678 | -0.2147 | 0.6021 |
|  | 2_VAARLTC4ID | 12 | 1 | 0.0398 | 0.0767 | -0.4451 | 0.4472 |


| 2_VACPTC1ARE | 8 | 0 | 0.2131 | 0.0243 | 0.0414 | 0.5250 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2_VACPTSF2RE | 0 | 3 | * | * | * | * |
| 2_VACPTYC1RE | 7 | 0 | 0.0479 | 0.0500 | -0.3665 | 0.3579 |
| 2-VACPTYC3RE | 14 | 0 | -0.0127 | 0.0507 | -0.3872 | 0.4014 |
| 2-VACPTYC4CO | 14 | 0 | 0.1206 | 0.0613 | -0.3098 | 0.4298 |
| 2-VACPTYC5ID | 14 | 0 | -0.2955 | 0.0314 | -0.6778 | -0.0362 |
| 2_VACPTYO1ID | 0 | 3 | * | * | * | * |
| 2_VAHATYH1CO | 18 | 0 | 0.2378 | 0.0554 | -0.2007 | 0.6031 |
| 2_VAHATYH2ID | 19 | 0 | 0.0597 | 0.0511 | -0.2676 | 0.5353 |
| 2_VAHATYH3RE | 17 | 0 | -0.1125 | 0.0502 | -0.5850 | 0.2625 |
| 2_VAHATYH4RE | 17 | 0 | 0.1149 | 0.0647 | -0.3188 | 0.5065 |
| 2_VAHATYH5RE | 17 | 0 | 0.2432 | 0.1077 | -0.2757 | 1.0550 |
| 2_VAHCCOC1CO | 2 | 0 | 0.423 | 0.153 | 0.146 | 0.699 |
| 2-VAHCCOC2CO | 3 | 0 | 0.555 | 0.0536 | 0.322 | 0.785 |
| 2_VAHCCON1ID | 2 | 0 | 0.0502 | 0.0184 | -0.0458 | 0.1461 |
| 2_VAHCCON2ID | 3 | 0 | 0.185 | 0.0362 | 0.000000000 | 0.380 |
| 2-VAHCCOR1RE | 3 | 0 | 0.254 | 0.0710 | 0.000000000 | 0.531 |
| 2_VAHCCOR2RE | 3 | 0 | 0.1728 | 0.0293 | 0.000000000 | 0.3424 |
| 2-VANFTMS5CO | 0 | 3 | * |  | * |  |
| 2_VANFTMS 6RE | 0 | 3 | * | * | * |  |
| 2_VANFTMS8CO | 0 | 3 | * | * | * |  |
| 2-VANFTMS9CO | 0 | 3 | * | * | * | * |
| 2_VANFTYN2RE | 29 | 0 | 0.1693 | 0.0589 | -0.2840 | 0.8000 |
| 2-VANFTYN3RE | 27 | 0 | 0.1671 | 0.0461 | -0.2757 | 0.6785 |
| 2_VANFTYN4CO | 27 | 0 | 0.2207 | 0.0879 | -0.3468 | 0.8500 |
| 2_VANFTYN5RE | 27 | 0 | 0.0499 | 0.1049 | -1.0000 | 0.5670 |
| 2_VANNTMF1RE | 0 | 1 | * | * | * | * |
| 2-VANNTMF4RE | 0 | 3 | * | * | * | * |
| 2_VANNTNN1RE | 8 | 2 | 0.2322 | 0.0380 | -0.0132 | 0.5752 |
| 2_VANNTSF4RE | 0 | 2 | * | * | * | * |
| 2_VANNTSF6RE | 0 | 2 | * | 0.04* | * | * |
| 2_VAPMTYP1CO | 18 | 0 | 0.0539 | 0.0448 | -0.3565 | 0.4814 |
| 2_VAPMTYP2RE | 17 | 0 | 0.0386 | 0.0486 | -0.3872 | 0.4150 |
| 2_VAPMTYP4RE | 17 | 0 | -0.0385 | 0.0286 | -0.3098 | 0.3385 |
| 2_VAPMTYP5RE | 17 | 0 | -0.00324 | 0.1173 | -0.4437 | 0.8591 |
| 2-VAVBTYIIID | 0 | 3 | * | * | * | * |
| 2-VAVBTYR1RE | 0 | 5 | * | * | * | * |
| 2_VAVBTYV1RE | 26 | 1 | 0.2966 | 0.0338 | -0.00877 | 0.7574 |
| 2_VAVBTYV4ID | 29 | 1 | -0.00245 | 0.0548 | -0.4815 | 0.6232 |
| 3_ALMOCREORE | 3 | 0 | -0.1197 | 0.0221 | -0.2518 | 0.0414 |
| 3_ALMODAPHCO | 3 | 0 | 0.165 | 0.0739 | -0.0223 | 0.477 |
| 3_ALMOSARARE | 3 | 0 | 0.0680 | 0.0139 | 0.000000000 | 0.2041 |
| 3_ALMOSIIVID | 3 | 0 | -0.0372 | 0.00530 | -0.1024 | 0.0414 |
| 3_ALMOSITVCO | 3 | 0 | -0.0690 | 0.00462 | -0.1427 | -0.00877 |
| 3_ALMOSIVIRE | 3 | 0 | -0.0372 | 0.00530 | -0.1024 | 0.0414 |
| 3_ALMOTHEOID | 3 | 0 | -0.0690 | 0.00462 | -0.1427 | -0.00877 |
| 3_GAATAT01ID | 9 | 1 | -0.288 | 0.273 | -1.347 | 0.248 |
| 3_GAATAT02RE | 8 | 1 | 0.1353 | 0.0234 | -0.0278 | 0.3784 |
| 5-TXARA001CO | 21 | 1 | 0.00765 | 0.0834 | -0.5857 | 0.5635 |
| 5_TXARA002RE | 20 | 1 | 0.1097 | 0.1304 | -0.6905 | 0.7324 |
| 5_TXARA003RE | 7 | 0 | -0.0632 | 0.0101 | -0.1549 | 0.0792 |
| 5_TXDAA001ID | 7 | 0 | -0.0662 | 0.0505 | -0.5229 | 0.1461 |
| 5_TXDAA002ID | 19 | 0 | -0.1466 | 0.1066 | -1.0000 | 0.4624 |
| 5-TXDAA004CO | 19 | 1 | -0.0675 | 0.0613 | -0.4437 | 0.6021 |
| 5_TXDAA005RE | 7 | 0 | 0.1449 | 0.0322 | -0.0969 | 0.3979 |
| 5_TXFWA004ID | 20 | 1 | -0.00875 | 0.0778 | -0.5807 | 0.6990 |
| 5-TXIRA001RE | 21 | 1 | 0.2761 | 0.1655 | -0.6517 | 1.0000 |
| 5_TXMEA002RE | 7 | 0 | 0.4255 | 0.0675 | 0.1461 | 0.8921 |
| 5_TXMEA003RE | 7 | 0 | 0.1962 | 0.0328 | -0.0458 | 0.3802 |
| 6_AZMCA001ID | 26 | 1 | 0.5937 | 0.1516 | -0.2218 | 1.2041 |
| 6_AZMCA003ID | 25 | 2 | 0.6372 | 0.0666 | 0.000000000 | 1.0792 |
| 6_AZMCA005CO | 26 | 0 | 0.6018 | 0.0581 | 0.1139 | 1.0792 |
| 6_AZMCA006RE | 19 | 1 | 0.5089 | 0.0538 | -0.0458 | 1.0414 |


| 6_AZTUA001RE | 10 | 0 | 0.532 | 0.113 | -0.155 | 0.845 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6_AZTUA002RE | 8 | 0 | 0.457 | 0.176 | -0.301 | 0.968 |
| 6_AZTUA003CO | 8 | 0 | 0.452 | 0.102 | 0.0170 | 0.973 |
| 6 AZTUA004ID | 7 | 0 | 0.4765 | 0.0649 | 0.0414 | 0.7340 |
| 6_CAALAL09ID | 6 | 3 | -0.246 | 0.169 | -0.602 | 0.415 |
| 7_OREUA001CO | 14 | 2 | 0.0821 | 0.194 | -0.705 | 0.633 |
| 7-OREUA003RE | 15 | 0 | -0.0675 | 0.193 | -0.827 | 1.064 |
| 7_ORGRA003RE | 6 | 0 | -0.2582 | 0.0490 | -0.6655 | -0.0400 |
| 7_ORGRA004CO | 6 | 0 | -0.276 | 0.0855 | -0.833 | -0.0223 |
| 7_ORPOA001CO | 12 | 1 | 0.0416 | 0.140 | -0.301 | 0.919 |
| 7_ORPOA003ID | 11 | 3 | 0.3577 | 0.0498 | -0.0458 | 0.7709 |
| 7_ORPOA004ID | 9 | 4 | 0.2381 | 0.0558 | -0.0969 | 0.6812 |
| 7_ORPOA006RE | 11 | 2 | 0.0972 | 0.136 | -0.377 | 0.813 |
| 7 ORSAA002CO | 5 | 1 | -0.315 | 0.624 | -1.696 | 0.301 |
| 7-ORSAA003ID | 4 | 2 | -0.366 | 0.978 | -1.843 | 0.204 |
| 7_ORSAA004RE | 5 | 1 | -0.936 | 0.961 | -2.314 | 0.0414 |



## Individual Value Plot of LTKN vs LndRainLoc



General Linear Model: LTKN versus Landuse, EPA_Rain_Zone

| Factor | Type | Levels Values |  |
| :--- | :--- | ---: | :--- |
| Landuse | fixed | 3 | CO, ID, RE |
| EPA_Rain_Zone | fixed | 5 | $2,3,5,6,7$ |

Analysis of Variance for LTKN, using Adjusted SS for Tests

| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Landuse | 2 | 0.1067 | 0.2805 | 0.1403 | 1.20 | 0.300 |
| EPA_Rain_Zone | 4 | 26.9717 | 24.3602 | 6.0901 | 52.29 | 0.000 |
| Landuse*EPA_Rain_Zone | 8 | 5.2809 | 5.2809 | 0.6601 | 5.67 | 0.000 |
| Error | 921 | 107.2734 | 107.2734 | 0.1165 |  |  |
| Total | 935 | 139.6327 |  |  |  |  |

Land use is not significant. EPA rain zone and interaction are significant.

```
Bonferroni Simultaneous Tests
Response Variable LTKN
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & -0.05489 & 0.04388 & -1.251 & 0.6338 \\
RE & -0.00419 & 0.04249 & -0.099 & 1.0000
\end{tabular}
Landuse = ID subtracted from:
```

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | 0.05069 | 0.03657 | 1.386 | 0.4980 |

Tukey Simultaneous Tests
Response Variable LTKN
All Pairwise Comparisons among Levels of Landuse
Landuse $=$ CO subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | -0.05489 | 0.04388 | -1.251 | 0.4232 |
| RE | -0.00419 | 0.04249 | -0.099 | 0.9946 |

Landuse $=$ ID subtracted from:

|  | Difference | SE of |  | Adjusted |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | 0.05069 | 0.03657 | 1.386 | 0.3481 |

## No significant differences among land uses.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTKN |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1325 | 0.06384 | -2.076 | 0.3814 |
| 5 | -0.0716 | 0.03296 | -2.174 | 0.3000 |
| 6 | 0.4300 | 0.03502 | 12.276 | 0.0000 |
| 7 | -0.1056 | 0.03915 | -2.698 | 0.0711 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.06090 | 0.06763 | 0.9005 | 1.0000 |
| 6 | 0.56251 | 0.06866 | 8.1931 | 0.0000 |
| 7 | 0.02693 | 0.07085 | 0.3801 | 1.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.50161 | 0.04153 | 12.0770 | 0.0000 |
| 7 | -0.03397 | 0.04507 | -0.7537 | 1.0000 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 - | -0.5356 | 0.04660 | -11.49 | 0.0000 |

TKN in rain zone 2 is significantly different than TKN in rain zone 6 . No differences were observed between rain zone 2 and zones 3,5 , and 7 .

TKN in rain zone 3 is significantly different than TKN in rain zone 6 . No differences were observed between rain zone 3 and zones 5, and 7 .

TKN in rain zone 5 is significantly different than TKN in rain zone 6 . No differences were observed between rain zone 5 and 7.

TKN in rain zone 6 is significantly different than TKN in rain zone 7.

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTKN |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1325 | 0.06384 | -2.076 | 0.2303 |
| 5 | -0.0716 | 0.03296 | -2.174 | 0.1897 |
| 6 | 0.4300 | 0.03502 | 12.276 | 0.0000 |
| 7 | -0.1056 | 0.03915 | -2.698 | 0.0543 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.06090 | 0.06763 | 0.9005 | 0.8968 |
| 6 | 0.56251 | 0.06866 | 8.1931 | 0.0000 |
| 7 | 0.02693 | 0.07085 | 0.3801 | 0.9956 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.50161 | 0.04153 | 12.0770 | 0.0000 |
| 7 | -0.03397 | 0.04507 | -0.7537 | 0.9436 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.5356 | 0.04660 | -11.49 | 0.0000 |

TKN in rain zone 2 is significantly different than TKN in rain zone 6 . No differences were observed between rain zone 2 and zones 3,5 , and 7 .

TKN in rain zone 3 is significantly different than TKN in rain zone 6 . No differences were observed between rain zone 3 and zones 5, and 7 .

TKN in rain zone 5 is significantly different than TKN in rain zone 6 . No differences were observed between rain zone 5 and 7.

TKN in rain zone 6 is significantly different than TKN in rain zone 7 .


## Interaction Plot (fitted means) for LTKN



## Total Phosphorus (P)

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  |  |  |  |  | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 114 | 9 | -0.624 | 0.279 | -1.222 | 0.243 |  |  |  |  |  |
| 3 | 6 | 0 | -0.722 | 0.330 | -1.187 | -0.337 |  |  |  |  |  |
| 4 | 0 | $16^{*}$ |  | $*$ | $*$ |  | $*$ |  |  |  |  |
| 5 | 40 | 2 | -0.869 | 0.411 | -1.640 | 0.630 |  |  |  |  |  |
| 6 | 34 | 0 | -0.344 | 0.243 | -0.796 | 0.301 |  |  |  |  |  |
| 7 | 37 | 4 | -0.591 | 0.362 | -1.699 | 0.519 |  |  |  |  |  |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  |  |  |  |  | Minimum | Maximum |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: |
| 2 | 96 | 13 | -0.706 | 0.387 | -2.200 | 0.111 |  |  |  |  |  |
| 3 | 15 | 1 | -0.793 | 0.403 | -1.523 | -0.046 |  |  |  |  |  |
| 4 | 0 | $17^{*}$ |  | $*$ | $*$ | $*$ |  |  |  |  |  |
| 5 | 45 | 2 | -0.714 | 0.310 | -1.337 | 0.107 |  |  |  |  |  |
| 6 | 59 | 11 | -0.017 | 0.356 | -0.854 | 0.898 |  |  |  |  |  |
| 7 | 24 | 9 | -0.289 | 0.288 | -1.194 | 0.146 |  |  |  |  |  |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev Minimum |  |  |  |  |  |  | Maximum |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 309 | 22 | -0.503 | 0.281 | -1.301 | 0.277 |  |  |  |  |  |
| 3 | 18 | 0 | -0.724 | 0.409 | -1.187 | 0.444 |  |  |  |  |  |
| 4 | 0 | $31^{*}$ |  | $*$ | $*$ |  | $*$ |  |  |  |  |
| 5 | 69 | 2 | -0.354 | 0.190 | -0.721 | 0.049 |  |  |  |  |  |
| 6 | 38 | 0 | -0.288 | 0.263 | -0.854 | 0.696 |  |  |  |  |  |
| 7 | 37 | 3 | -0.700 | 0.337 | -1.420 | 0.342 |  |  |  |  |  |

## Descriptive Statistics: LTP

| Variable |  |  |  |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| LTADRainLoc | N | $N^{*}$ | Mean | Variance | Minimum | Maximum |  |
| LTP | 2_KYLOTSR1RE | 3 | 0 | -0.275 | 0.131 | -0.602 | 0.114 |
|  | 2_KYLOTSR2ID | 3 | 0 | -0.362 | 0.0626 | -0.585 | -0.0915 |
|  | 2_KYLOTSR3RE | 3 | 0 | -0.750 | 0.111 | -1.097 | -0.432 |
|  | 2_KYLOTSR4ID | 4 | 0 | -0.839 | 0.0760 | -1.155 | -0.523 |
|  | 2_MDAACOMWID | 3 | 0 | -0.176 | 0.116 | -0.553 | 0.111 |
|  | 2_MDAACOODRE | 3 | 0 | -0.301 | 0.138 | -0.523 | 0.127 |
|  | 2_MDAACOPPCO | 26 | 0 | -0.6324 | 0.0941 | -1.1549 | 0.2430 |
|  | 2_MDAACORKRE | 3 | 0 | -0.297 | 0.135 | -0.509 | 0.127 |
|  | 2_MDBACOBCID | 3 | 0 | -0.630 | 0.0436 | -0.824 | -0.409 |
|  | 2_MDBACOSCRE | 26 | 0 | -0.5524 | 0.0746 | -1.1549 | 0.0828 |
|  | 2_MDBACOTCID | 3 | 0 | -0.494 | 0.0310 | -0.620 | -0.292 |
|  | 2_MDBACOWCRE | 3 | 0 | -0.3122 | 0.00315 | -0.3768 | -0.2757 |
|  | 2_MDBCTYBOID | 3 | 0 | -0.3988 | 0.00107 | -0.4318 | -0.3665 |
|  | 2_MDBCTYFMID | 3 | 0 | -0.3982 | 0.000345 | -0.4089 | -0.3768 |
|  | 2_MDBCTYHORE | 3 | 0 | -0.247 | 0.375 | -0.921 | 0.276 |
|  | 2_MDBCTYHRRE | 3 | 0 | -0.194 | 0.119 | -0.409 | 0.204 |
|  | 2_MDBCTYKOCO | 3 | 0 | -0.4426 | 0.0203 | -0.6021 | -0.3279 |
|  | 2_MDMOCOBCCO | 3 | 0 | -0.794 | 0.0483 | -1.046 | -0.638 |
|  | 2_VAARLLP1RE | 8 | 0 | -0.7864 | 0.0564 | -1.0969 | -0.4089 |
|  | 2_VAARLTC4ID | 13 | 0 | -0.6859 | 0.0127 | -0.8861 | -0.5376 |



| 6_AZTUAO01RE | 10 | 0 | -0.146 | 0.121 | -0.456 | 0.695 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 6_AZTUA002RE | 8 | 0 | -0.3468 | 0.0299 | -0.5376 | -0.0809 |
| 6_AZTUA003CO | 8 | 0 | -0.4447 | 0.0355 | -0.6576 | -0.2218 |
| 6_AZTUA004ID | 7 | 0 | -0.0404 | 0.0668 | -0.4437 | 0.3560 |
| 6_CAALAL09ID | 0 | 9 | $\star$ | $\star$ | $*$ | $*$ |
| 7_OREUA001CO | 14 | 2 | -0.4341 | 0.0579 | -0.7696 | -0.1079 |
| 7_OREUA003RE | 15 | 0 | -0.5909 | 0.0987 | -1.0000 | 0.3424 |
| 7_ORGRA003RE | 6 | 0 | -0.6907 | 0.0195 | -0.8861 | -0.5086 |
| 7_ORGRA004CO | 6 | 0 | -0.804 | 0.211 | -1.699 | -0.418 |
| 7_ORPOA001CO | 12 | 1 | -0.630 | 0.200 | -1.222 | 0.519 |
| 7_ORPOA003ID | 11 | 3 | -0.2166 | 0.0208 | -0.4437 | -0.0362 |
| 7_ORPOA004ID | 9 | 4 | -0.1989 | 0.0645 | -0.6383 | 0.1461 |
| 7_ORPOA006RE | 11 | 2 | -0.665 | 0.121 | -1.155 | 0.0792 |
| 7_ORSAA002CO | 5 | 1 | -0.6827 | 0.0113 | -0.8539 | -0.5654 |
| 7_ORSAA003ID | 4 | 2 | -0.688 | 0.138 | -1.194 | -0.374 |
| 7_ORSAA004RE | 5 | 1 | -1.116 | 0.0848 | -1.420 | -0.762 |




## General Linear Model: LTP versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID, RE |  |  |  |  |
| EPA_Rain_Zone fixed | 5 | $2,3,5,6,7$ |  |  |  |  |
| Analysis of Variance | for LTP, | , using A | justed | for Te |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 2.9935 | 1.1228 | 0.5614 | 5.85 | 0.003 |
| EPA Rain Zone | 4 | 20.6147 | 17.8961 | 4.4740 | 46.65 | 0.000 |
| Lan̄${ }^{\text {duse*EPA_Rain_Zone }}$ | 8 | 12.3467 | 12.3467 | 1.5433 | 16.09 | 0.000 |
| Error | 926 | 88.8109 | 88.8109 | 0.0959 |  |  |
| Total | 9401 | 124.7659 |  |  |  |  |

Main factors and interaction are significant in the model

```
Bonferroni Simultaneous Tests
Response Variable LTP
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & 0.1263 & 0.03989 & 3.166 & 0.0048 \\
RE & 0.1163 & 0.03836 & 3.031 & 0.0075
\end{tabular}
```

|  | Difference | SE of | Adjusted |  |
| :--- | ---: | ---: | ---: | ---: |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.01003 | 0.03304 | -0.3036 | 1.000 |


| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTP |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse $=$ co subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.1263 | 0.03989 | 3.166 | 0.0044 |
| RE | 0.1163 | 0.03836 | 3.031 | 0.0069 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.01003 | 0.03304 | -0.3036 | 0.9505 |

Total phosphorus in commercial land use areas is significantly different than total phosphorus in residential or industrial land use areas. There was no difference in total phosphorus concentration between residential and industrial land uses.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTP |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1349 | 0.05760 | -2.342 | 0.1941 |
| 5 | -0.0345 | 0.02994 | -1.151 | 1.0000 |
| 6 | 0.3951 | 0.03183 | 12.413 | 0.0000 |
| 7 | 0.0846 | 0.03548 | 2.385 | 0.1729 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.1004 | 0.06112 | 1.643 | 1.0000 |
| 6 | 0.5300 | 0.06207 | 8.539 | 0.0000 |
| 7 | 0.2195 | 0.06402 | 3.429 | 0.0063 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.4296 | 0.03784 | 11.352 | 0.0000 |
| 7 | 0.1191 | 0.04096 | 2.907 | 0.0373 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |

```
7 -0.3105 0.04236 -7.330 0.0000
```

Total phosphorus in EPA rain zone 2 is different than total phosphorus in EPA rain zone 6 . No difference was observed between EPA rain zone 2 and zones 3, 5, and 7

Total phosphorus in EPA rain zone 3 is different than total phosphorus in EPA rain zones 6 and 7. No difference was observed between EPA rain zone 3 and EPA rain zone 5

Total phosphorus in EPA rain zone 5 is different than total phosphorus in EPA rain zones 6 and 7.
Total phosphorus in EPA rain zone 6 is different than total phosphorus in EPA rain zone 7.

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LTP |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: $\quad-\quad-$ |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1349 | 0.05760 | -2.342 | 0.1319 |
| 5 | -0.0345 | 0.02994 | -1.151 | 0.7792 |
| 6 | 0.3951 | 0.03183 | 12.413 | 0.0000 |
| 7 | 0.0846 | 0.03548 | 2.385 | 0.1194 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.1004 | 0.06112 | 1.643 | 0.4699 |
| 6 | 0.5300 | 0.06207 | 8.539 | 0.0000 |
| 7 | 0.2195 | 0.06402 | 3.429 | 0.0055 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.4296 | 0.03784 | 11.352 | 0.0000 |
| 7 | 0.1191 | 0.04096 | 2.907 | 0.0300 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.3105 | 0.04236 | -7.330 | 0.0000 |

Total phosphorus in EPA rain zone 2 is different than total phosphorus in EPA rain zone 6 . No difference was observed between EPA rain zone 2 and zones 3, 5, and 7

Total phosphorus in EPA rain zone 3 is different than total phosphorus in EPA rain zones 6 and 7. No difference was observed between EPA rain zone 3 and EPA rain zone 5

Total phosphorus in EPA rain zone 5 is different than total phosphorus in EPA rain zones 6 and 7.
Total phosphorus in EPA rain zone 6 is different than total phosphorus in EPA rain zone 7.


Interaction Plot (fitted means) for LTP


## Dissolved Phosphorus (dissolved - P)

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 273 | 30 | -1.185 | 0.403 | -2.168 | -0.444 |
|  | 3 | 60 | -1.168 | 0.438 | -1.602 | -0.602 |
|  | 4 | 0 16* |  | * * | * * | * |
|  | $5 \quad 40$ | 0 | -1.387 | 0.307 | -2.043 | -0.770 |
|  | $6 \quad 26$ | 6 - | -0.500 | 0.354 | -1.301 | 0.204 |
|  | 7 | 734 | -1.546 | 0.317 | -2.000 | -1.051 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  | StDev Minimum |  | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 267 | 42 | -1.103 | 0.360 | -2.000 | -0.347 |
|  | 311 | 5 | -1.460 | 0.213 | -1.602 | -0.991 |
|  | 40 | 17* | * |  | * | * |
|  | $5 \quad 46$ | - 1 | -1.172 | 0.286 | -1.745 | -0.409 |
|  | $6 \quad 52$ | -18 | -0.654 | 0.310 | -1.309 | 0.176 |
| 7 | 7 2 | 31 | -1.150 | 0.415 | -1.444 | -0.857 |

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev Minimum |  | Maximum |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 255 | 76 | -0.869 | 0.370 | -2.088 | 0.029 |  |
| 3 | 14 | 4 | -1.247 | 0.320 | -1.602 | -0.585 |  |
| 4 | 0 | $31^{*}$ |  | $*$ | $*$ | $*$ |  |
| 5 | 69 | 2 | -0.612 | 0.292 | -1.770 | -0.076 |  |
| 6 | 20 | 18 | -0.609 | 0.213 | -0.959 | -0.155 |  |
| 7 | 8 | 32 | -1.691 | 0.349 | -2.060 | -1.046 |  |

## Descriptive Statistics: LDP

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LDP | 2 KYLOTSR1RE | 2 | 1 | -0.734 | 0.101 | -0.959 | -0.509 |
|  | 2_KYLOTSR2ID | 2 | 1 | -0.523 | 0.0620 | -0.699 | -0.347 |
|  | 2_KYLOTSR3RE | 3 | 0 | -1.248 | 0.0927 | -1.523 | -0.921 |
|  | 2_KYLOTSR4ID | 2 | 2 | -0.76955 | 0.000000000 | -0.76955 | -0.76955 |
|  | 2_MDAACOMWID | 0 | 3 | * | * | * | * |
|  | 2_MDAACOODRE | 0 | 3 | * | * | * |  |
|  | 2_MDAACOPPCO | 0 | 26 | * | * | * |  |
|  | 2_MDAACORKRE | 0 | 3 | * | * | * |  |
|  | 2_MDBACOBCID | 0 | 3 | * | * | * |  |
|  | 2_MDBACOSCRE | 0 | 26 | * | * | * | * |
|  | 2_MDBACOTCID | 0 | 3 | * | * | * |  |
|  | 2_MDBACOWCRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYBOID | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYFMID | 0 | 3 | * | * | * |  |
|  | 2_MDBCTYHORE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYHRRE | 0 | 3 | * | * | * | * |
|  | 2_MDBCTYKOCO | 0 | 3 | * | * | * | * |
|  | 2_MDMOCOBCCO | 0 | 3 | * | * * | * | * |
|  | 2_VAARLLP1RE | 8 | 0 | -0.8677 | 0.0654 | -1.2218 | -0.4685 |
|  | 2_VAARLTC4ID | 13 | 0 | -1.0604 | 0.0266 | -1.3979 | -0.8539 |
|  | 2_VACPTC1ARE | 8 | 0 | -0.7347 | 0.0697 | -1.1549 | -0.3665 |


| 2 VACPTSF2RE | 0 | 3 | * | * | * | * |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2_VACPTYC1RE | 7 | 0 | -1.249 | 0.198 | -1.958 | -0.620 |
| 2 VACPTYC3RE | 13 | 1 | -0.9666 | 0.0406 | -1.4776 | -0.6383 |
| 2 VACPTYC4CO | 11 | 3 | -1.205 | 0.202 | -2.044 | -0.585 |
| 2-VACPTYC5ID | 4 | 10 | -2.0000 | 0.000000000 | -2.0000 | -2.0000 |
| 2-VACPTYO1ID | 0 | 3 | * | * | * | * |
| 2_VAHATYH1CO | 17 | 1 | -1.279 | 0.208 | -2.168 | -0.469 |
| 2 VAHATYH2ID | 17 | 2 | -1.0041 | 0.0829 | -1.6641 | -0.5086 |
| 2-VAHATYH3RE | 14 | 3 | -1.1679 | 0.0293 | -1.4466 | -0.7696 |
| 2_VAHATYH4RE | 16 | 1 | -0.9326 | 0.0786 | -1.4864 | -0.4318 |
| 2_VAHATYH5RE | 17 | 0 | -0.6210 | 0.0901 | -1.2100 | 0.0294 |
| 2_VAHCCOC1CO | 0 | 2 | * | * | * | * |
| 2-VAHCCOC2CO | 0 | 3 | * | * | * | * |
| 2_VAHCCON1ID | 0 | 2 | * | * | * | * |
| 2_VAHCCON2ID | 0 | 3 | * | * | * | * |
| 2_VAHCCOR1RE | 0 | 3 | * | * | * | * |
| 2_VAHCCOR2RE | 0 | 3 | * | * | * | * |
| 2-VANFTMS5CO | 0 | 3 | * | * | * | * |
| 2-VANFTMS 6RE | 0 | 3 | * | * | * | * |
| 2_VANFTMS8CO | 0 | 3 | * | * | * | * |
| 2_VANFTMS 9CO | 0 | 3 | * | * | * | * |
| 2_VANFTYN2RE | 29 | 0 | -0.5698 | 0.0803 | -1.1549 | -0.1675 |
| 2_VANFTYN3RE | 27 | 0 | -0.8331 | 0.1009 | -1.4916 | -0.2518 |
| 2_VANFTYN4CO | 27 | 0 | -0.9480 | 0.0990 | -1.5633 | -0.4437 |
| 2_VANFTYN5RE | 27 | 0 | -0.9455 | 0.1574 | -1.8074 | -0.3565 |
| 2_VANNTMF1RE | 0 | 1 | * | * | * | * |
| 2-VANNTMF4RE | 0 | 3 | * | * | * | * |
| 2_VANNTNN1RE | 7 | 3 | -0.698 | 0.129 | -1.155 | -0.0132 |
| 2_VANNTSF4RE | 0 | 2 | * | * | * | * |
| 2_VANNTSF6RE | 0 | 2 | * | * | * | * |
| 2_VAPMTYP1CO | 18 | 0 | -1.4390 | 0.0542 | -1.8681 | -0.9586 |
| 2_VAPMTYP2RE | 17 | 0 | -1.1321 | 0.0380 | -1.4984 | -0.7696 |
| 2_VAPMTYP4RE | 17 | 0 | -1.131 | 0.261 | -2.088 | -0.347 |
| 2_VAPMTYP5RE | 17 | 0 | -0.9125 | 0.1402 | -1.6493 | -0.0757 |
| 2_VAVBTYIIID | 0 | 3 | * | * | * | * |
| 2_VAVBTYR1RE | 0 | 5 | * | * | * | * |
| 2_VAVBTYV1RE | 26 | 1 | -0.6572 | 0.0584 | -1.2218 | -0.1871 |
| 2_VAVBTYV4ID | 29 | 1 | -1.1185 | 0.0902 | -1.7327 | -0.5528 |
| 3_ALMOCREORE | 3 | 0 | -1.471 | 0.0519 | -1.602 | -1.208 |
| 3_ALMODAPHCO | 3 | 0 | -0.813 | 0.0811 | -1.137 | -0.602 |
| 3_ALMOSARARE | 3 | 0 | -1.2252 | 0.00516 | -1.3080 | -1.1805 |
| 3_ALMOSIIVID | 3 | 0 | -1.6021 | 0.000000000 | -1.6021 | -1.6021 |
| 3_ALMOSITVCO | 3 | 0 | -1.5236 | 0.0185 | -1.6021 | -1.3665 |
| 3_ALMOSIVIRE | 3 | 0 | -1.485 | 0.0409 | -1.602 | -1.252 |
| 3_ALMOTHEOID | 3 | 0 | -1.5236 | 0.0185 | -1.6021 | -1.3665 |
| 3_GAATAT01ID | 5 | 5 | -1.337 | 0.0670 | -1.523 | -0.991 |
| 3_GAATAT02RE | 5 | 4 | -0.982 | 0.115 | -1.398 | -0.585 |
| 5_TXARA001CO | 21 | 1 | -1.3512 | 0.1125 | -2.0433 | -0.7696 |
| 5_TXARA002RE | 20 | 1 | -0.6680 | 0.0992 | -1.3979 | -0.0757 |
| 5_TXARA003RE | 7 | 0 | -0.7378 | 0.0128 | -0.8539 | -0.5376 |
| 5_TXDAA001ID | 7 | 0 | -0.755 | 0.102 | -1.222 | -0.409 |
| 5_TXDAA002ID | 19 | 0 | -1.3062 | 0.0486 | -1.7447 | -0.9172 |
| 5_TXDAA004CO | 19 | 1 | -1.4267 | 0.0765 | -2.0000 | -1.0458 |
| 5_TXDAA005RE | 7 | 0 | -0.5263 | 0.0160 | -0.7212 | -0.3372 |
| 5_TXFWA004ID | 20 | 1 | -1.1893 | 0.0329 | -1.6990 | -0.8125 |
| 5_TXIRA001RE | 21 | 1 | -0.6270 | 0.1462 | -1.7696 | -0.1688 |
| 5_TXMEA002RE | 7 | 0 | -0.5001 | 0.0117 | -0.6990 | -0.3565 |
| 5_TXMEA003RE | 7 | 0 | -0.4830 | 0.0534 | -0.8861 | -0.1871 |
| 6_AZMCA001ID | 26 | 1 | -0.5093 | 0.1154 | -1.1549 | 0.1761 |
| 6_AZMCA003ID | 26 | 1 | -0.7985 | 0.0364 | -1.3086 | -0.4685 |
| 6_AZMCA005CO | 26 | 0 | -0.5003 | 0.1250 | -1.3010 | 0.2041 |
| 6_AZMCA006RE | 20 | 0 | -0.6088 | 0.0452 | -0.9586 | -0.1549 |
| 6_AZTUA001RE | 0 | 10 | * | * | * |  |



Xbar-R Chart of LDP



Individual Value Plot of LDP vs LndRainLoc


LndRainLoc

## General Linear Model: LDP versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID, |  |  |  |  |
| EPA_Rain_Zone fixed | 5 | 2, 3, 5 | 6, 7 |  |  |  |
| Analysis of Variance for LDP, using Adjusted SS for Tests |  |  |  |  |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 10.5050 | 0.9326 | 0.4663 | 3.90 | 0.021 |
| EPA Rain Zone | 4 | 28.0949 | 20.2703 | 5.0676 | 42.40 | 0.000 |
| Landuse*EPA_Rain_Zone | 8 | 9.9010 | 9.9010 | 1.2376 | 10.36 | 0.000 |
| Error | 681 | 81.3873 | 81.3873 | 0.1195 |  |  |
| Total | 6951 | 129.8882 |  |  |  |  |

Main factors and interaction are significant in the model

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LDP |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse $=$ co subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.04954 | 0.07029 | 0.7048 | 1.0000 |
| RE | 0.15159 | 0.05581 | 2.7163 | 0.0203 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | 0.1021 | 0.06602 | 1.546 | 0.3678 |


| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LDP |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse = Co subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.04954 | 0.07029 | 0.7048 | 0.7606 |
| RE | 0.15159 | 0.05581 | 2.7163 | 0.0181 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | 0.1021 | 0.06602 | 1.546 | 0.2696 |

Dissolved phosphorus in commercial land use is significantly different than in residential land use areas. No significant difference was observed between commercial and industrial land uses. No significant difference was observed between industrial and residential land uses.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LDP |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.2393 | 0.06929 | -3.454 | 0.0059 |
| 5 | -0.0047 | 0.03529 | -0.133 | 1.0000 |
| 6 | 0.4646 | 0.04315 | 10.767 | 0.0000 |
| 7 | -0.4101 | 0.10310 | -3.978 | 0.0008 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | 0.2346 | 0.07199 | 3.259 | 0.0117 |
| 6 | 0.7040 | 0.07615 | 9.244 | 0.0000 |
| 7 | -0.1708 | 0.12069 | -1.415 | 1.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.4693 | 0.04736 | 9.910 | 0.0000 |
| 7 | -0.4055 | 0.10493 | -3.864 | 0.0012 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.8748 | 0.1078 | -8.113 | 0.0000 |

Dissolved phosphorus concentrations in EPA region 2 are significantly different than in regions 3,6 and 7. No significant differences were observed between EPA regions 2 and 5.

Dissolved phosphorus concentrations in EPA region 3 are significantly different than in regions 5 and 6 . No significant differences were observed between EPA regions 3 and 7.

Dissolved phosphorus concentrations in EPA region 5 are significantly different than in regions 6 and 7.
Dissolved phosphorus concentrations in EPA region 6 are significantly different than in region 7 .

```
Tukey Simultaneous Tests
Response Variable LDP
All Pairwise Comparisons among Levels of EPA_Rain_Zone
EPA_Rain_Zone = 2 subtracted from:
\begin{tabular}{lrrrr} 
& \begin{tabular}{r} 
Difference \\
of Means
\end{tabular} & \begin{tabular}{rl} 
SE of
\end{tabular} & Adjusted
\end{tabular}
EPA_Rain_Zone = 3 subtracted from:
EPA_Rain_Zone \begin{tabular}{rrr} 
Difference & SE of & Adjusted \\
of Means
\end{tabular} Difference \(\quad\) T-Value \begin{tabular}{rl} 
P-Value
\end{tabular}
```

$\left.\begin{array}{lrrrr}5 & 0.2346 & 0.07199 & 3.259 & 0.0099 \\ 6 & 0.7040 & 0.07615 & 9.244 & 0.0000 \\ 7 & & -0.1708 & 0.12069 & -1.415\end{array}\right)$

Dissolved phosphorus concentrations in EPA region 2 are significantly different than in regions 3,6 and 7 . No significant differences were observed between EPA regions 2 and 5.

Dissolved phosphorus concentrations in EPA region 3 are significantly different than in regions 5 and 6 . No significant differences were observed between EPA regions 3 and 7 .

Dissolved phosphorus concentrations in EPA region 5 are significantly different than in regions 6 and 7.
Dissolved phosphorus concentrations in EPA region 6 are significantly different than in region 7 .


The assumption of normality of residuals is valid. No specific trend was observed in the residuals.


## Total Copper ( $\mu \mathrm{g} / \mathrm{L}$ ) (Cu)

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  | Minimum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | Maximum

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean |  | StDev |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | Minimum $\quad$ Maximum

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean | StDev |  | Minimum | Maximum |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 140 | 191 | 1.018 | 0.379 | -0.149 | 2.380 |  |
| 3 | 17 | 1 | 1.079 | 0.463 | 0.699 | 2.009 |  |
| 4 | 31 | 0 | 1.386 | 0.345 | 0.845 | 2.013 |  |
| 5 | 68 | 3 | 0.883 | 0.272 | 0.095 | 1.799 |  |
| 6 | 38 | 0 | 0.972 | 0.496 | 0.146 | 2.255 |  |
| 7 | 37 | 3 | 0.948 | 0.312 | 0.211 | 1.909 |  |

## Descriptive Statistics: LCU

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LCU | 2 KYLOTSR1RE | 3 | 0 | 0.9813 | 0.0234 | 0.8437 | 1.1461 |
|  | 2_KYLOTSR2ID | 3 | 0 | 0.812 | 0.132 | 0.602 | 1.230 |
|  | 2_KYLOTSR3RE | 3 | 0 | 1.293 | 0.0627 | 1.079 | 1.568 |
|  | 2_KYLOTSR4ID | 4 | 0 | 1.3024 | 0.00452 | 1.2553 | 1.3979 |
|  | 2_MDAACOMWID | 3 | 0 | 1.2764 | 0.0126 | 1.1761 | 1.3979 |
|  | 2 MDAACOODRE | 3 | 0 | 1.041 | 0.0625 | 0.845 | 1.322 |
|  | 2_MDAACOPPCO | 26 | 0 | 1.1747 | 0.0491 | 0.7839 | 1.6876 |
|  | 2_MDAACORKRE | 3 | 0 | 1.001 | 0.0493 | 0.845 | 1.255 |
|  | 2_MDBACOBCID | 3 | 0 | 1.2177 | 0.00520 | 1.1761 | 1.3010 |
|  | 2_MDBACOSCRE | 26 | 0 | 1.0590 | 0.1588 | 0.3010 | 1.5641 |
|  | 2 MDBACOTCID | 3 | 0 | 1.2321 | 0.0204 | 1.0792 | 1.3617 |
|  | 2_MDBACOWCRE | 3 | 0 | 1.0698 | 0.0124 | 0.9542 | 1.1761 |
|  | 2_MDBCTYBOID | 3 | 0 | 1.502 | 0.0302 | 1.301 | 1.602 |
|  | 2_MDBCTYFMID | 3 | 0 | 1.541 | 0.0771 | 1.301 | 1.845 |
|  | 2-MDBCTYHORE | 3 | 0 | 1.4184 | 0.0103 | 1.3010 | 1.4771 |
|  | 2 MDBCTYHRRE | 3 | 0 | 2.2086 | 0.0287 | 2.0414 | 2.3802 |
|  | 2_MDBCTYKOCO | 3 | 0 | 1.4184 | 0.0103 | 1.3010 | 1.4771 |
|  | 2_MDMOCOBCCO | 3 | 0 | 1.335 | 0.0494 | 1.146 | 1.580 |
|  | 2_VAARLLP1RE | 8 | 0 | 0.779 | 0.265 | -0.149 | 1.398 |
|  | 2_VAARLTC4ID | 12 | 1 | 0.9564 | 0.0466 | 0.4800 | 1.2041 |
|  | 2_VACPTC1ARE | 0 | 8 | * | * | * | * |


| 2_VACPTSF2RE | 3 | 0 | 0.259 | 0.202 | 0.000000000 | 0.778 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2_VACPTYC1RE | 3 | 4 | 0.6512 | 0.00668 | 0.5798 | 0.7404 |
| 2_VACPTYC3RE | 2 | 12 | 0.8885 | 0.0159 | 0.7993 | 0.9777 |
| 2_VACPTYC4CO | 3 | 11 | 1.030 | 0.107 | 0.732 | 1.380 |
| 2_VACPTYC5ID | 2 | 12 | 0.555 | 0.0747 | 0.362 | 0.748 |
| 2_VACPTYO1ID | 3 | 0 | 1.2041 | 0.0145 | 1.1239 | 1.3424 |
| 2_VAHATYH1CO | 3 | 15 | 1.1283 | 0.0283 | 1.0212 | 1.3222 |
| 2_VAHATYH2ID | 3 | 16 | 1.684 | 0.114 | 1.301 | 1.940 |
| 2_VAHATYH3RE | 3 | 14 | 0.8237 | 0.0296 | 0.6335 | 0.9685 |
| 2_VAHATYH4RE | 3 | 14 | 1.016 | 0.135 | 0.602 | 1.301 |
| 2_VAHATYH5RE | 3 | 14 | 0.7101 | 0.00776 | 0.6232 | 0.7993 |
| 2 VAHCCOC1CO | 2 | 0 | 1.160 | 0.0282 | 1.041 | 1.279 |
| 2_VAHCCOC2CO | 3 | 0 | 1.5386 | 0.00171 | 1.4914 | 1.5682 |
| 2_VAHCCON1ID | 2 | 0 | 1.211 | 0.0574 | 1.041 | 1.380 |
| 2_VAHCCON2ID | 3 | 0 | 1.232 | 0.164 | 0.778 | 1.556 |
| 2_VAHCCOR1RE | 3 | 0 | 0.926 | 0.0402 | 0.699 | 1.079 |
| 2_VAHCCOR2RE | 3 | 0 | 0.9985 | 0.00190 | 0.9542 | 1.0414 |
| 2_VANFTMS5CO | 3 | 0 | 1.4599 | 0.00945 | 1.3802 | 1.5682 |
| 2_VANFTMS 6RE | 3 | 0 | 1.2003 | 0.00484 | 1.1461 | 1.2788 |
| 2_VANFTMS8CO | 3 | 0 | 1.151 | 0.0506 | 0.914 | 1.362 |
| 2-VANFTMS 9CO | 3 | 0 | 1.277 | 0.0444 | 1.041 | 1.447 |
| 2_VANFTYN2RE | 11 | 18 | 1.1917 | 0.0327 | 0.9085 | 1.4624 |
| 2_VANFTYN3RE | 11 | 16 | 1.1920 | 0.0347 | 0.9365 | 1.4771 |
| 2_VANFTYN4CO | 11 | 16 | 1.1798 | 0.0459 | 0.9085 | 1.5315 |
| 2_VANFTYN5RE | 11 | 16 | 1.1569 | 0.0679 | 0.7099 | 1.4624 |
| 2_VANNTMF1RE | 1 | 0 | 1.1461 | * | 1.1461 | 1.1461 |
| 2_VANNTMF4RE | 3 | 0 | 0.6942 | 0.0201 | 0.6021 | 0.8573 |
| 2-VANNTNN1RE | 2 | 8 | 0.5743 | 0.00956 | 0.5051 | 0.6435 |
| 2_VANNTSF4RE | 2 | 0 | 0.6066 | 0.00144 | 0.5798 | 0.6335 |
| 2_VANNTSF6RE | 2 | 0 | 0.7287 | 0.00641 | 0.6721 | 0.7853 |
| 2_VAPMTYP1CO | 3 | 15 | 0.858 | 0.312 | 0.230 | 1.301 |
| 2_VAPMTYP2RE | 3 | 14 | 1.204 | 0.0829 | 0.903 | 1.477 |
| 2_VAPMTYP4RE | 3 | 14 | 0.9261 | 0.0164 | 0.7782 | 1.0000 |
| 2_VAPMTYP5RE | 3 | 14 | 0.752 | 0.0689 | 0.477 | 1.000 |
| 2 VAVBTYIIID | 3 | 0 | 1.438 | 0.0942 | 1.114 | 1.724 |
| 2 VAVBTYR1RE | 3 | 2 | 0.7181 | 0.0111 | 0.6435 | 0.8388 |
| 2_VAVBTYV1RE | 4 | 23 | 0.8506 | 0.00764 | 0.7782 | 0.9638 |
| 2_VAVBTYV4ID | 4 | 26 | 0.639 | 0.0500 | 0.342 | 0.881 |
| 3_ALMOCREORE | 3 | 0 | 0.69897 | 0.000000000 | 0.69897 | 0.69897 |
| 3_ALMODAPHCO | 3 | 0 | 1.0893 | 0.0186 | 0.9584 | 1.2304 |
| 3_ALMOSARARE | 3 | 0 | 0.69897 | 0.000000000 | 0.69897 | 0.69897 |
| 3_ALMOSIIVID | 3 | 0 | 1.279 | 0.0699 | 0.974 | 1.447 |
| 3_ALMOSITVCO | 3 | 0 | 0.799 | 0.0302 | 0.699 | 1.000 |
| 3_ALMOSIVIRE | 3 | 0 | 0.69897 | 0.000000000 | 0.69897 | 0.69897 |
| 3 ALMOTHEOID | 3 | 0 | 1.419 | 0.184 | 0.942 | 1.771 |
| 3_GAATAT01ID | 9 | 1 | 0.933 | 0.252 | 0.301 | 1.756 |
| 3_GAATAT02RE | 8 | 1 | 1.507 | 0.0961 | 1.000 | 2.009 |
| 4_KATOATWORE | 15 | 0 | 1.3209 | 0.1236 | 0.8451 | 1.9243 |
| 4_KATOBROORE | 16 | 0 | 1.4468 | 0.1146 | 0.8451 | 2.0128 |
| 4-KATOJACKCO | 15 | 1 | 2.010 | 0.206 | 0.699 | 2.584 |
| 4_KATOSTFEID | 15 | 2 | 2.064 | 0.356 | 1.000 | 3.134 |
| 5 TXARA001CO | 21 | 1 | 0.7802 | 0.0748 | 0.3010 | 1.3802 |
| 5-TXARA002RE | 20 | 1 | 0.8408 | 0.1362 | 0.0951 | 1.7993 |
| 5_TXARA003RE | 7 | 0 | 0.7522 | 0.0261 | 0.6021 | 0.9542 |
| 5 TXDAA001ID | 7 | 0 | 0.9966 | 0.0398 | 0.6021 | 1.2041 |
| 5-TXDAA002ID | 18 | 1 | 1.2273 | 0.0299 | 0.8451 | 1.4771 |
| 5 TXDAA004CO | 17 | 3 | 1.1499 | 0.1237 | 0.6021 | 1.9138 |
| 5-TXDAA005RE | 6 | 1 | 1.0149 | 0.0131 | 0.9031 | 1.2041 |
| 5_TXFWA004ID | 19 | 2 | 1.3942 | 0.0419 | 1.1139 | 1.9395 |
| 5_TXIRA001RE | 21 | 1 | 0.9402 | 0.0416 | 0.6021 | 1.2788 |
| 5_TXMEA002RE | 7 | 0 | 1.0883 | 0.0231 | 0.9031 | 1.3979 |
| 5 TXMEA003RE | 7 | 0 | 0.6396 | 0.0249 | 0.3010 | 0.7782 |
| 6_AZMCA001ID | 25 | 2 | 1.8095 | 0.1090 | 1.0000 | 2.3010 |


| 6_AZMCA003ID | 25 | 2 | 2.0065 | 0.1989 | 0.3010 | 2.5315 |
| :--- | ---: | :--- | ---: | ---: | ---: | ---: |
| 6_AZMCA005CO | 26 | 0 | 1.2616 | 0.0659 | 0.8451 | 1.7993 |
| 6_AZMCA006RE | 20 | 0 | 1.3454 | 0.1049 | 0.6990 | 2.2553 |
| 6_AZTUA001RE | 10 | 0 | 0.6083 | 0.0739 | 0.1461 | 0.9545 |
| 6_AZTUA002RE | 8 | 0 | 0.4918 | 0.0716 | 0.1461 | 0.8325 |
| 6_AZTUA003CO | 8 | 0 | 0.6083 | 0.0512 | 0.1761 | 0.8325 |
| 6_AZTUA004ID | 7 | 0 | 1.0509 | 0.0388 | 0.7993 | 1.3617 |
| 6_CAALAL09ID | 9 | 0 | 1.6227 | 0.0779 | 1.2041 | 2.1072 |
| 7_OREUA001CO | 14 | 2 | 1.4691 | 0.0940 | 1.1139 | 2.1139 |
| 7_OREUA003RE | 15 | 0 | 0.9659 | 0.1075 | 0.5374 | 1.9085 |
| 7_ORGRA003RE | 6 | 0 | 0.8230 | 0.0420 | 0.4472 | 1.0531 |
| 7_ORGRA004CO | 6 | 0 | 1.0563 | 0.0353 | 0.8129 | 1.2695 |
| 7_ORPOA001CO | 12 | 1 | 1.171 | 0.175 | 0.477 | 1.785 |
| 7_ORPOA003ID | 12 | 2 | 1.6786 | 0.0705 | 1.3222 | 2.0792 |
| 7_ORPOA004ID | 9 | 4 | 1.532 | 0.117 | 1.114 | 2.000 |
| 7_ORPOA006RE | 11 | 2 | 0.975 | 0.153 | 0.211 | 1.690 |
| 7_ORSAA002CO | 5 | 1 | 1.1697 | 0.0140 | 1.0019 | 1.3010 |
| 7_ORSAA003ID | 4 | 2 | 1.2801 | 0.0322 | 1.0414 | 1.4771 |
| 7_ORSAA004RE | 5 | 1 | 0.9829 | 0.0390 | 0.7782 | 1.3010 |



## Individual Value Plot of LCU vs LndRainLoc



## General Linear Model: LCU versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | $3 \mathrm{CO}, \mathrm{ID}, \mathrm{RE}$ |  |  |  |  |  |
| EPA_Rain_Zone fixed | $62,3,4,5,6,7$ |  |  |  |  |  |
| Analysis of Variance for LCU, using Adjusted SS for Tests |  |  |  |  |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 27.1721 | 17.5268 | 8.7634 | 64.52 | 0.000 |
| EPA_Rain_Zone | 5 | 26.3863 | 28.2226 | 5.6445 | 41.56 | 0.000 |
| Landuse*EPA_Rain_Zone | 10 | 13.7000 | 13.7000 | 1.3700 | 10.09 | 0.000 |
| Error | 727 | 98.7367 | 98.7367 | 0.1358 |  |  |
| Total | 7441 | 165.9951 |  |  |  |  |

Main effects and interaction are significant.

```
Bonferroni Simultaneous Tests
Response Variable LCU
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & 0.2403 & 0.04600 & 5.224 & 0.0000 \\
RE & -0.1986 & 0.04333 & -4.583 & 0.0000
\end{tabular}
```

Landuse $=$ ID subtracted from:

| Landuse | of Means | Difference | T-Value | P-Value |
| :---: | :---: | :---: | :---: | :---: |
| RE | -0.4389 | 0.03865 | -11.35 | 0.0000 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LCU |  |  |  |  |
| All Pairwise Comparisons among Levels of Landuse |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.2403 | 0.04600 | 5.224 | 0.0000 |
| RE | -0.1986 | 0.04333 | -4.583 | 0.0000 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.4389 | 0.03865 | -11.35 | 0.0000 |

Significant differences by land use. Copper concentrations are different among the three land uses.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LCU |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.08326 | 0.07089 | -1.175 | 1.0000 |
| 4 | 0.69562 | 0.05582 | 12.463 | 0.0000 |
| 5 | -0.09387 | 0.03973 | -2.362 | 0.2763 |
| 6 | 0.16181 | 0.04106 | 3.940 | 0.0013 |
| 7 | 0.13418 | 0.04512 | 2.974 | 0.0456 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.77889 | 0.08311 | 9.3715 | 0.0000 |
| 5 | -0.01060 | 0.07329 | -0.1446 | 1.0000 |
| 6 | 0.24508 | 0.07402 | 3.3111 | 0.0146 |
| 7 | 0.21745 | 0.07634 | 2.8484 | 0.0678 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.7895 | 0.05883 | -13.42 | 0.0000 |
| 6 | -0.5338 | 0.05974 | -8.94 | 0.0000 |
| 7 | -0.5614 | 0.06260 | -8.97 | 0.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.2557 | 0.04507 | 5.672 | 0.0000 |
| 7 | 0.2280 | 0.04880 | 4.673 | 0.0001 |

```
EPA_Rain_Zone = 6 subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
EPA_Rain_Zone & of Means & Difference & T-Value & P-Value \\
7 & -0.02763 & 0.04989 & -0.5538 & 1.000
\end{tabular}
```

Copper in EPA rain zone 2 is significantly different than in EPA rain zones 4, 6 , and 7 . No differences were observed between rain zone 2 and rain zones 3 and 5 .

Copper in EPA rain zone 3 is significantly different than in EPA rain zones 4, 6, and 7. No differences were observed between rain zone 3 and rain zone 5 .

Copper in EPA rain zone 4 is significantly different than in EPA rain zones 5, 6, and 7.
Copper in EPA rain zone 5 is significantly different than in EPA rain zones 6, and 7. No differences were observed between rain zones 6 and 7 .

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone | $=2$ subtrac | ed from: |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.08326 | 0.07089 | -1.175 | 0.8491 |
| 4 | 0.69562 | 0.05582 | 12.463 | 0.0000 |
| 5 | -0.09387 | 0.03973 | -2.362 | 0.1696 |
| 6 | 0.16181 | 0.04106 | 3.940 | 0.0011 |
| 7 | 0.13418 | 0.04512 | 2.974 | 0.0349 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.77889 | 0.08311 | 9.3715 | 0.0000 |
| 5 | -0.01060 | 0.07329 | -0.1446 | 1.0000 |
| 6 | 0.24508 | 0.07402 | 3.3111 | 0.0120 |
| 7 | 0.21745 | 0.07634 | 2.8484 | 0.0502 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE Of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.7895 | 0.05883 | -13.42 | 0.0000 |
| 6 | -0.5338 | 0.05974 | -8.94 | 0.0000 |
| 7 | -0.5614 | 0.06260 | -8.97 | 0.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.2557 | 0.04507 | 5.672 | 0.0000 |
| 7 | 0.2280 | 0.04880 | 4.673 | 0.0000 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |


| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 7 | -0.02763 | 0.04989 | -0.5538 | 0.9938 |

Copper in EPA rain zone 2 is significantly different than in EPA rain zones 4, 6, and 7. No differences were observed between rain zone 2 and rain zones 3 and 5 .

Copper in EPA rain zone 3 is significantly different than in EPA rain zones 4, 6, and 7. No differences were observed between rain zone 3 and rain zone 5 .

Copper in EPA rain zone 4 is significantly different than in EPA rain zones 5, 6, and 7.
Copper in EPA rain zone 5 is significantly different than in EPA rain zones 6 , and 7 . No differences were observed between rain zones 6 and 7 .

Residual Plots for LCU


Interaction Plot (fitted means) for LCU


## Total Lead ( $\mu \mathrm{g} / \mathrm{L}$ ) ( Pb )

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  | Minimum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | Maximum

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  | Minimum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | Maximum

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev Minimum |  |  | Maximum |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 135 | 196 | 0.836 | 0.596 | -0.341 | 2.566 |  |
| 3 | 15 | 3 | 0.915 | 0.639 | 0.000 | 2.653 |  |
| 4 | 31 | 0 | 1.155 | 0.476 | 0.000 | 2.340 |  |
| 5 | 68 | 3 | 1.073 | 0.357 | 0.224 | 1.949 |  |
| 6 | 30 | 8 | 1.254 | 0.591 | -0.874 | 2.279 |  |
| 7 | 37 | 3 | 0.998 | 0.497 | -0.222 | 2.322 |  |

## Descriptive Statistics: LPB

| Variable | LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| LPB | 2_KYLOTSR1RE | 1 | 2 | 1.7782 | + | 1.7782 | 1.7782 |
|  | 2_KYLOTSR2ID | 0 | 3 | * | * | * | * |
|  | 2_KYLOTSR3RE | 0 | 3 | * | * | * | * |
|  | 2-KYLOTSR4ID | 1 | 3 | 2.1139 | * | 2.1139 | 2.1139 |
|  | 2 MDAACOMWID | 3 | 0 | 0.9748 | 0.0142 | 0.8451 | 1.0792 |
|  | 2_MDAACOODRE | 3 | 0 | 0.436 | 0.345 | 0.0969 | 1.114 |
|  | 2_MDAACOPPCO | 26 | 0 | 0.670 | 0.391 | -0.670 | 1.531 |
|  | 2_MDAACORKRE | 3 | 0 | 0.096910 | 0.000000000 | 0.096910 | 0.096910 |
|  | 2-MDBACOBCID | 3 | 0 | 0.602 | 0.766 | 0.0969 | 1.613 |
|  | 2_MDBACOSCRE | 26 | 0 | 0.2642 | 0.2540 | -0.3010 | 1.5911 |
|  | 2_MDBACOTCID | 3 | 0 | 1.762 | 0.0743 | 1.447 | 1.924 |
|  | 2_MDBACOWCRE | 3 | 0 | 0.226 | 0.832 | -0.301 | 1.279 |
|  | 2_MDBCTYBOID | 3 | 0 | 1.424 | 0.226 | 0.875 | 1.699 |
|  | 2-MDBCTYFMID | 3 | 0 | 1.276 | 0.483 | 0.875 | 2.079 |
|  | 2_MDBCTYHORE | 3 | 0 | 1.150 | 0.226 | 0.875 | 1.699 |


| 2_MDBCTYHRRE | 3 | 0 | 1.076 | 0.121 | 0.875 | 1.477 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2-MDBCTYKOCO | 3 | 0 | 1.522 | 0.137 | 1.117 | 1.845 |
| 2 MDMOCOBCCO | 3 | 0 | 0.520 | 0.133 | 0.178 | 0.903 |
| 2_VAARLLP1RE | 8 | 0 | 0.250 | 0.0997 | -0.341 | 0.778 |
| 2_VAARLTC4ID | 12 | 1 | 0.2128 | 0.0813 | -0.2539 | 0.9031 |
| 2_VACPTC1ARE | 0 | 8 | * | * | * | * |
| 2_VACPTSF2RE | 3 | 0 | 0.65321 | 0.000000000 | 0.65321 | 0.65321 |
| 2_VACPTYC1RE | 3 | 4 | 0.65321 | 0.000000000 | 0.65321 | 0.65321 |
| 2_VACPTYC3RE | 2 | 12 | 1.2075 | 0.0175 | 1.1139 | 1.3010 |
| 2_VACPTYC4CO | 3 | 11 | 1.154 | 0.231 | 0.740 | 1.681 |
| 2_VACPTYC5ID | 2 | 12 | 0.65321 | 0.000000000 | 0.65321 | 0.65321 |
| 2-VACPTYO1ID | 3 | 0 | 1.4696 | 0.0154 | 1.3979 | 1.6128 |
| 2_VAHATYH1CO | 3 | 15 | 1.446 | 0.0495 | 1.279 | 1.699 |
| 2_VAHATYH2ID | 3 | 16 | 1.513 | 0.176 | 1.041 | 1.845 |
| 2_VAHATYH3RE | 3 | 14 | 0.7313 | 0.0237 | 0.5563 | 0.8451 |
| 2_VAHATYH4RE | 3 | 14 | 1.2540 | 0.0204 | 1.1430 | 1.4150 |
| 2-VAHATYH5RE | 3 | 14 | 0.862 | 0.227 | 0.371 | 1.322 |
| 2_VAHCCOC1CO | 2 | 0 | 1.3227 | 0.0170 | 1.2304 | 1.4150 |
| 2_VAHCCOC2CO | 3 | 0 | 1.801 | 0.0331 | 1.623 | 1.987 |
| 2_VAHCCON1ID | 2 | 0 | 1.443 | 0.114 | 1.204 | 1.681 |
| 2_VAHCCON2ID | 3 | 0 | 0.985 | 0.250 | 0.477 | 1.477 |
| 2_VAHCCOR1RE | 3 | 0 | 0.990 | 0.0726 | 0.699 | 1.230 |
| 2_VAHCCOR2RE | 3 | 0 | 1.0428 | 0.0158 | 0.9031 | 1.1461 |
| 2_VANFTMS5CO | 3 | 0 | 1.2461 | 0.0118 | 1.1461 | 1.3617 |
| 2_VANFTMS 6RE | 3 | 0 | 1.2951 | 0.00778 | 1.2041 | 1.3802 |
| 2_VANFTMS8CO | 3 | 0 | 1.440 | 0.0681 | 1.146 | 1.643 |
| 2_VANFTMS 9CO | 3 | 0 | 1.167 | 0.740 | 0.505 | 2.140 |
| 2_VANFTYN2RE | 11 | 18 | 1.022 | 0.185 | 0.321 | 1.792 |
| 2_VANFTYN3RE | 11 | 16 | 1.275 | 0.111 | 0.606 | 1.839 |
| 2_VANFTYN4CO | 11 | 16 | 1.058 | 0.223 | 0.296 | 1.748 |
| 2-VANFTYN5RE | 11 | 16 | 1.250 | 0.322 | 0.193 | 1.959 |
| 2_VANNTMF1RE | 1 | 0 | 1.3802 | * | 1.3802 | 1.3802 |
| 2_VANNTMF4RE | 3 | 0 | 0.65321 | 0.000000000 | 0.65321 | 0.65321 |
| 2_VANNTNN1RE | 2 | 8 | 1.088 | 0.0562 | 0.920 | 1.255 |
| 2_VANNTSF4RE | 2 | 0 | 0.65321 | 0.000000000 | 0.65321 | 0.65321 |
| 2-VANNTSF6RE | 2 | 0 | 1.0051 | 0.0110 | 0.9310 | 1.0792 |
| 2_VAPMTYP1CO | 3 | 15 | 1.382 | 0.0954 | 1.041 | 1.643 |
| 2_VAPMTYP2RE | 3 | 14 | 2.271 | 0.0775 | 2.013 | 2.566 |
| 2_VAPMTYP4RE | 3 | 14 | 0.890 | 0.0376 | 0.778 | 1.114 |
| 2_VAPMTYP5RE | 3 | 14 | 0.878 | 0.0302 | 0.778 | 1.079 |
| 2_VAVBTYIIID | 3 | 0 | 0.973 | 0.307 | 0.653 | 1.613 |
| 2_VAVBTYR1RE | 3 | 2 | 1.0307 | 0.00793 | 0.9368 | 1.1139 |
| 2_VAVBTYV1RE | 4 | 23 | 1.3479 | 0.0214 | 1.1761 | 1.5315 |
| 2_VAVBTYV4ID | 4 | 26 | 0.815 | 0.105 | 0.653 | 1.301 |
| 3_ALMOCREORE | 3 | 0 | 0.39794 | 0.000000000 | 0.39794 | 0.39794 |
| 3_ALMODAPHCO | 3 | 0 | 1.1099 | 0.0195 | 0.9494 | 1.2041 |
| 3_ALMOSARARE | 3 | 0 | 0.8480 | 0.0279 | 0.6691 | 1.0000 |
| 3_ALMOSIIVID | 3 | 0 | 1.3827 | 0.0259 | 1.2788 | 1.5682 |
| 3_ALMOSITVCO | 3 | 0 | 0.922 | 0.0393 | 0.699 | 1.079 |
| 3_ALMOSIVIRE | 3 | 0 | 0.877 | 0.0679 | 0.699 | 1.176 |
| 3_ALMOTHEOID | 3 | 0 | 0.998 | 0.121 | 0.631 | 1.322 |
| 3_GAATAT01ID | 4 | 6 | 0.687 | 0.494 | 0.000000000 | 1.602 |
| 3_GAATAT02RE | 6 | 3 | 1.227 | 0.823 | 0.000000000 | 2.653 |
| 4_KATOATWORE | 15 | 0 | 1.218 | 0.241 | 0.602 | 2.340 |
| 4_KATOBROORE | 16 | 0 | 1.096 | 0.221 | 0.000000000 | 1.699 |
| 4_KATOJACKCO | 15 | 1 | 1.779 | 0.357 | 0.000000000 | 2.340 |
| 4_KATOSTFEID | 15 | 2 | 1.917 | 0.497 | 0.301 | 3.079 |
| 5_TXARA001CO | 21 | 1 | 1.3634 | 0.0945 | 0.7298 | 2.0792 |
| 5_TXARA002RE | 20 | 1 | 0.7964 | 0.0607 | 0.2243 | 1.2041 |
| 5_TXARA003RE | 7 | 0 | 1.075 | 0.0717 | 0.699 | 1.380 |
| 5_TXDAA001ID | 7 | 0 | 1.282 | 0.197 | 0.347 | 1.643 |
| 5_TXDAA002ID | 18 | 1 | 1.1651 | 0.1032 | 0.6021 | 1.6812 |
| 5_TXDAA004CO | 18 | 2 | 1.107 | 0.253 | 0.196 | 2.477 |


| 5_TXDAA005RE | 6 | 1 | 1.599 | 0.0738 | 1.279 | 1.949 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5_TXFWA004ID | 20 | 1 | 1.5859 | 0.1196 | 1.1461 | 2.4314 |
| 5 TXIRA001RE | 21 | 1 | 1.2757 | 0.0611 | 0.7381 | 1.7243 |
| 5-TXMEA002RE | 7 | 0 | 1.1367 | 0.0644 | 0.8451 | 1.6335 |
| 5_TXMEA003RE | 7 | 0 | 0.7399 | 0.0166 | 0.6021 | 0.9542 |
| 6_AZMCA001ID | 27 | 0 | 1.5732 | 0.1349 | 0.9542 | 2.3010 |
| 6_AZMCA003ID | 27 | 0 | 2.3438 | 0.2064 | 0.3010 | 2.7924 |
| 6_AZMCA005CO | 26 | 0 | 0.9853 | 0.0598 | 0.4771 | 1.4314 |
| 6_AZMCA006RE | 20 | 0 | 1.3528 | 0.1251 | 0.7782 | 2.2788 |
| 6_AZTUA001RE | 6 | 4 | 1.460 | 0.212 | 0.699 | 2.114 |
| 6_AZTUA002RE | 4 | 4 | 0.452 | 1.219 | -0.874 | 1.806 |
| 6 AZTUA003CO | 8 | 0 | 1.384 | 0.105 | 1.000 | 1.886 |
| 6_AZTUA004ID | 6 | 1 | 1.735 | 0.0846 | 1.342 | 2.090 |
| 6_CAALAL09ID | 9 | 0 | 1.5948 | 0.0671 | 1.2553 | 2.0792 |
| 7_OREUA001CO | 14 | 2 | 1.439 | 0.230 | 0.477 | 2.362 |
| 7-OREUA003RE | 15 | 0 | 1.320 | 0.195 | 0.699 | 2.322 |
| 7 ORGRA003RE | 6 | 0 | 0.872 | 0.0631 | 0.602 | 1.204 |
| 7_ORGRA004CO | 6 | 0 | 1.158 | 0.130 | 0.863 | 1.792 |
| 7 ORPOA001CO | 12 | 1 | 1.563 | 0.140 | 1.146 | 2.462 |
| 7_ORPOA003ID | 12 | 2 | 1.6963 | 0.0841 | 1.2041 | 2.2304 |
| 7 ORPOA004ID | 9 | 4 | 1.268 | 0.170 | 0.602 | 1.968 |
| 7_ORPOA006RE | 11 | 2 | 0.9724 | 0.0952 | 0.4771 | 1.5798 |
| 7_ORSAA002CO | 5 | 1 | 1.103 | 0.120 | 0.591 | 1.491 |
| 7_ORSAA003ID | 4 | 2 | 1.545 | 0.133 | 1.301 | 2.079 |
| 7_ORSAA004RE | 5 | 1 | 0.237 | 0.0914 | -0.222 | 0.522 |

## Xbar-R Chart of LPB




General Linear Model: LPB versus Landuse, EPA_Rain_Zone

| Factor Type | Levels Values |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID, RE |  |  |  |  |
| EPA_Rain_Zone fixed | 6 | 2, 3, 4, | 5, 6, 7 |  |  |  |
| Analysis of Variance | for LPB, | , using | justed SS | for Tes |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 31.6472 | 12.9341 | 6.4671 | 24.05 | 0.000 |
| EPA_Rain_Zone | 5 | 33.0307 | 33.4042 | 6.6808 | 24.85 | 0.000 |
| Lan̄${ }^{\text {duse*EPA_Rain_Zone }}$ | 10 | 16.0919 | 16.0919 | 1.6092 | 5.99 | 0.000 |
| Error | 7041 | 189.2705 | 189.2705 | 0.2689 |  |  |
| Total | 721 | 270.0403 |  |  |  |  |

Main factors and interaction are significant.

```
Bonferroni Simultaneous Tests
Response Variable LPB
All Pairwise Comparisons among Levels of Landuse
Landuse = CO subtracted from:
\begin{tabular}{lrrrr} 
& Difference & SE of & & Adjusted \\
Landuse & of Means & Difference & T-Value & P-Value \\
ID & 0.1802 & 0.06667 & 2.703 & 0.0211 \\
RE & -0.2163 & 0.06184 & -3.498 & 0.0015
\end{tabular}
```



Copper concentrations are different among commercial, industrial and residential land uses.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LPB |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | 0.05664 | 0.10616 | 0.5336 | 1.0000 |
| 4 | 0.70022 | 0.07912 | 8.8505 | 0.0000 |
| 5 | 0.31270 | 0.05641 | 5.5435 | 0.0000 |
| 6 | 0.49147 | 0.06017 | 8.1684 | 0.0000 |
| 7 | 0.38434 | 0.06421 | 5.9861 | 0.0000 |
| EPA_Rain_Zone = 3 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.6436 | 0.1221 | 5.272 | 0.0000 |
| 5 | 0.2561 | 0.1087 | 2.355 | 0.2823 |
| 6 | 0.4348 | 0.1107 | 3.926 | 0.0014 |
| 7 | 0.3277 | 0.1130 | 2.900 | 0.0577 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.3875 | 0.08256 | -4.694 | 0.0000 |
| 6 | -0.2087 | 0.08517 | -2.451 | 0.2173 |
| 7 | -0.3159 | 0.08807 | -3.587 | 0.0054 |
| EPA_Rain_Zone = 5 subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |


| EPA_Rain_Zone | Of Means | Difference | T-Value | P-Value |
| :--- | ---: | ---: | ---: | ---: |
| 6 | 0.17877 | 0.06462 | 2.766 | 0.0873 |
| 7 | 0.07164 | 0.06840 | 1.047 | 1.0000 |
|  |  |  |  |  |
| EPA_Rain_Zone $=6$ | subtracted from: |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.1071 | 0.07153 | -1.498 | 1.000 |

Copper concentration in EPA rain zone 2 is different than in zones 4, 5, 6, and 7. No differences were observed between rain zones 2 and 3 .

Copper concentration in EPA rain zone 3 is different than in zones 4, 6, and 7. No differences were observed between rain zones 3 and 5 .

Copper concentration in EPA rain zone 4 is different than in zones 5 and 7. No differences were observed between rain zones 4 and 6 .

Copper concentration in EPA rain zone 5 is different than in zone 6 . No differences were observed between rain zones 5 and 7. No differences were observed between zones 6 and 7 .

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LPB |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 3 | 0.05664 | 0.10616 | 0.5336 | 0.9948 |
| 4 | 0.70022 | 0.07912 | 8.8505 | 0.0000 |
| 5 | 0.31270 | 0.05641 | 5.5435 | 0.0000 |
| 6 | 0.49147 | 0.06017 | 8.1684 | 0.0000 |
| 7 | 0.38434 | 0.06421 | 5.9861 | 0.0000 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.6436 | 0.1221 | 5.272 | 0.0000 |
| 5 | 0.2561 | 0.1087 | 2.355 | 0.1725 |
| 6 | 0.4348 | 0.1107 | 3.926 | 0.0012 |
| 7 | 0.3277 | 0.1130 | 2.900 | 0.0433 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.3875 | 0.08256 | -4.694 | 0.0000 |
| 6 | -0.2087 | 0.08517 | -2.451 | 0.1390 |
| 7 | -0.3159 | 0.08807 | -3.587 | 0.0045 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |


| 6 | 0.17877 | 0.06462 | 2.766 | 0.0630 |
| :---: | :---: | :---: | :---: | :---: |
| 7 | 0.07164 | 0.06840 | 1.047 | 0.9018 |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.1071 | 0.07153 | -1.498 | 0.6659 |

Copper concentration in EPA rain zone 2 is different than in zones 4, 5, 6, and 7. No differences were observed between rain zones 2 and 3 .

Copper concentration in EPA rain zone 3 is different than in zones 4, 6, and 7. No differences were observed between rain zones 3 and 5 .

Copper concentration in EPA rain zone 4 is different than in zones 5 and 7. No differences were observed between rain zones 4 and 6 .

Copper concentration in EPA rain zone 5 is different than in zone 6 . No differences were observed between rain zones 5 and 7. No difference was observed between zones 6 and 7

Residual Plots for LPB


## Interaction Plot (fitted means) for LPB



## Total Zinc ( $\mu \mathrm{g} / \mathrm{L}$ ) (Zn)

Summary statistics in LOG base 10 scale Land use: Commercial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev | Minimum | Maximum |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: |
| 2 | 66 | 57 | 2.212 | 0.262 | 1.716 | 2.950 |
| 3 | 6 | 0 | 2.050 | 0.179 | 1.851 | 2.342 |
| 4 | 15 | 1 | 2.552 | 0.367 | 1.663 | 2.969 |
| 5 | 39 | 3 | 1.878 | 0.287 | 1.447 | 2.748 |
| 6 | 34 | 0 | 2.305 | 0.245 | 1.806 | 2.820 |
| 7 | 37 | 4 | 2.060 | 0.325 | 1.591 | 2.964 |

Land use: Industrial

| EPA_Rain_Zone | N det/est | N ND/NZ Mean | StDev |  |  | Minimum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | Maximum

Land use: Residential

| EPA_Rain_Zone | N det/est | N ND/NZ | Mean | StDev | Minimum | Maximum |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 2140 | 191 | 1.702 | 0.426 | 0.047 | 2.726 |
| 3 | 317 | 1 | 1.673 | 0.411 | 0.912 | 2.462 |
| 4 | 431 | 0 | 2.073 | 0.441 | 1.000 | 3.199 |
| 5 | $5 \quad 68$ | 3 | 1.796 | 0.241 | 1.204 | 2.362 |
|  | 638 | 0 | 2.195 | 0.342 | 1.544 | 3.176 |
| 7 | $7 \quad 37$ | 3 | 1.760 | 0.352 | 0.840 | 2.813 |

## Descriptive Statistics: LZN

| Variable |  |  |  |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| LndRainLoc | N | N* | Mean | Variance | Minimum | Maximum |  |
| LZN | 2_KYLOTSR1RE | 3 | 0 | 2.188 | 0.149 | 1.792 | 2.563 |
|  | 2_KYLOTSR2ID | 3 | 0 | 2.357 | 0.101 | 2.152 | 2.723 |
|  | 2_KYLOTSR3RE | 3 | 0 | 2.069 | 0.0444 | 1.851 | 2.272 |
|  | 2_KYLOTSR4ID | 4 | 0 | 2.2135 | 0.0235 | 1.9912 | 2.3243 |
|  | 2_MDAACOMWID | 3 | 0 | 2.2168 | 0.00901 | 2.1139 | 2.3010 |
|  | 2_MDAACOODRE | 3 | 0 | 2.155 | 0.0546 | 1.944 | 2.407 |
|  | 2_MDAACOPPCO | 26 | 0 | 2.1423 | 0.0330 | 1.7747 | 2.5221 |
|  | 2_MDAACORKRE | 3 | 0 | 1.5452 | 0.0211 | 1.3802 | 1.6532 |
|  | 2_MDBACOBCID | 3 | 0 | 2.2719 | 0.00602 | 2.1931 | 2.3483 |
|  | 2_MDBACOSCRE | 26 | 0 | 1.6532 | 0.0584 | 1.2967 | 2.1608 |
|  | 2_MDBACOTCID | 3 | 0 | 2.1928 | 0.0175 | 2.1004 | 2.3444 |
|  | 2_MDBACOWCRE | 3 | 0 | 1.931 | 0.0470 | 1.756 | 2.173 |
|  | 2_MDBCTYBOID | 3 | 0 | 2.506 | 0.0324 | 2.342 | 2.699 |
|  | 2_MDBCTYFMID | 3 | 0 | 2.3858 | 0.0223 | 2.2788 | 2.5563 |
|  | 2_MDBCTYHORE | 3 | 0 | 2.053 | 0.0469 | 1.903 | 2.301 |
|  | 2_MDBCTYHRRE | 3 | 0 | 2.0111 | 0.0211 | 1.9031 | 2.1761 |
|  | 2_MDBCTYKOCO | 3 | 0 | 2.2103 | 0.00351 | 2.1761 | 2.2788 |
|  | 2_MDMOCOBCCO | 3 | 0 | 1.8362 | 0.0162 | 1.7482 | 1.9823 |
|  | 2_VAARLLP1RE | 8 | 0 | 1.231 | 0.392 | 0.0465 | 1.903 |
|  | 2_VAARLTC4ID | 12 | 1 | 1.5903 | 0.0798 | 1.0000 | 1.9542 |


| 2_VACPTC1ARE | 0 | 8 | * | * | * |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2_VACPTSF2RE | 3 | 0 | 0.900 | 0.189 | 0.602 | 1.398 |
| 2_VACPTYC1RE | 3 | 4 | 1.5096 | 0.0173 | 1.3802 | 1.6435 |
| 2_VACPTYC3RE | 2 | 12 | 1.7897 | 0.0109 | 1.7160 | 1.8633 |
| 2-VACPTYC4CO | 3 | 11 | 2.018 | 0.0775 | 1.716 | 2.265 |
| 2_VACPTYC5ID | 2 | 12 | 2.449 | 0.0660 | 2.267 | 2.630 |
| 2_VACPTYO1ID | 3 | 0 | 2.1956 | 0.00161 | 2.1584 | 2.2380 |
| 2_VAHATYH1CO | 3 | 15 | 2.2005 | 0.0138 | 2.0969 | 2.3284 |
| 2_VAHATYH2ID | 3 | 16 | 2.6408 | 0.0176 | 2.4900 | 2.7404 |
| 2_VAHATYH3RE | 3 | 14 | 1.368 | 0.191 | 0.903 | 1.771 |
| 2_VAHATYH4RE | 3 | 14 | 1.7492 | 0.00195 | 1.7160 | 1.7993 |
| 2_VAHATYH5RE | 3 | 14 | 1.080 | 0.542 | 0.230 | 1.519 |
| 2_VAHCCOC1CO | 2 | 0 | 1.988 | 0.0207 | 1.886 | 2.090 |
| 2-VAHCCOC2CO | 3 | 0 | 2.7788 | 0.0251 | 2.6385 | 2.9504 |
| 2_VAHCCON1ID | 2 | 0 | 2.507 | 0.0470 | 2.354 | 2.661 |
| 2_VAHCCON2ID | 3 | 0 | 2.454 | 0.0339 | 2.243 | 2.584 |
| 2_VAHCCOR1RE | 3 | 0 | 1.7427 | 0.0110 | 1.6232 | 1.8195 |
| 2_VAHCCOR2RE | 3 | 0 | 2.182 | 0.230 | 1.820 | 2.726 |
| 2_VANFTMS5CO | 3 | 0 | 2.371 | 0.0631 | 2.130 | 2.631 |
| 2_VANFTMS6RE | 3 | 0 | 1.9543 | 0.00817 | 1.8692 | 2.0492 |
| 2_VANFTMS8CO | 3 | 0 | 2.1748 | 0.0124 | 2.0899 | 2.3010 |
| 2_VANFTMS 9 CO | 3 | 0 | 2.450 | 0.195 | 2.053 | 2.925 |
| 2_VANFTYN2RE | 11 | 18 | 1.742 | 0.162 | 0.699 | 2.314 |
| 2_VANFTYN3RE | 11 | 16 | 1.876 | 0.179 | 0.740 | 2.292 |
| 2_VANFTYN4CO | 11 | 16 | 2.3598 | 0.0440 | 1.8808 | 2.6232 |
| 2_VANFTYN5RE | 11 | 16 | 1.9504 | 0.0638 | 1.3617 | 2.3181 |
| 2_VANNTMF1RE | 1 | 0 | 1.8325 | * | 1.8325 | 1.8325 |
| 2_VANNTMF4RE | 3 | 0 | 1.609 | 0.0511 | 1.431 | 1.863 |
| 2_VANNTNN1RE | 2 | - | 1.6048 | 0.00470 | 1.5563 | 1.6532 |
| 2_VANNTSF4RE | 2 | 0 | 1.399 | 0.0287 | 1.279 | 1.519 |
| 2_VANNTSF6RE | 2 | 0 | 1.6324 | 0.00184 | 1.6021 | 1.6628 |
| 2_VAPMTYP1CO | 3 | 15 | 2.083 | 0.0503 | 1.833 | 2.265 |
| 2_VAPMTYP2RE | 3 | 14 | 2.071 | 0.0392 | 1.869 | 2.265 |
| 2_VAPMTYP4RE | 3 | 14 | 1.558 | 0.0525 | 1.301 | 1.740 |
| 2_VAPMTYP5RE | 3 | 14 | 1.179 | 0.688 | 0.221 | 1.663 |
| 2_VAVBTYIIID | 3 | 0 | 2.466 | 0.0430 | 2.228 | 2.609 |
| 2_VAVBTYR1RE | 3 | 2 | 1.384 | 0.182 | 0.903 | 1.716 |
| 2_VAVBTYV1RE | 4 | 23 | 1.6995 | 0.00442 | 1.6335 | 1.7709 |
| 2_VAVBTYV4ID | 4 | 26 | 2.159 | 0.0548 | 1.898 | 2.467 |
| 3_ALMOCREORE | 3 | 0 | 1.579 | 0.636 | 0.912 | 2.462 |
| 3_ALMODAPHCO | 3 | 0 | 2.149 | 0.0368 | 1.959 | 2.342 |
| 3_ALMOSARARE | 3 | 0 | 1.6884 | 0.0133 | 1.6021 | 1.8195 |
| 3_ALMOSIIVID | 3 | 0 | 2.319 | 0.0771 | 2.079 | 2.623 |
| 3_ALMOSITVCO | 3 | 0 | 1.9498 | 0.0137 | 1.8513 | 2.0792 |
| 3_ALMOSIVIRE | 3 | 0 | 1.795 | 0.0391 | 1.568 | 1.934 |
| 3_ALMOTHEOID | 3 | 0 | 2.278 | 0.124 | 1.892 | 2.580 |
| 3_GAATAT01ID | 9 | 1 | 1.8625 | 0.0591 | 1.3979 | 2.2304 |
| 3_GAATAT02RE | 8 | 1 | 1.657 | 0.179 | 1.000 | 2.104 |
| 4_KATOATWORE | 15 | 0 | 2.087 | 0.289 | 1.000 | 3.199 |
| 4_KATOBROORE | 16 | 0 | 2.0592 | 0.1192 | 1.4771 | 2.7853 |
| 4_KATOJACKCO | 15 | 1 | 2.5519 | 0.1345 | 1.6628 | 2.9685 |
| 4_KATOSTFEID | 15 | 2 | 2.5707 | 0.1254 | 2.0792 | 3.2014 |
| 5_TXARA001CO | 21 | 1 | 1.7920 | 0.0514 | 1.4472 | 2.4624 |
| 5_TXARA002RE | 20 | 1 | 1.7226 | 0.0637 | 1.2041 | 2.1139 |
| 5_TXARA003RE | 7 | 0 | 1.6831 | 0.0447 | 1.3010 | 1.9542 |
| 5-TXDAA001ID | 7 | 0 | 2.0214 | 0.0204 | 1.8451 | 2.2304 |
| 5-TXDAA002ID | 18 | 1 | 1.8553 | 0.0370 | 1.4771 | 2.1732 |
| 5_TXDAA004CO | 18 | 2 | 1.9784 | 0.1037 | 1.6021 | 2.7482 |
| 5-TXDAA005RE | 6 | 1 | 1.8889 | 0.0444 | 1.6021 | 2.1461 |
| 5_TXFWA004ID | 20 | 1 | 2.4855 | 0.1225 | 1.4472 | 3.1461 |
| 5-TXIRA001RE | 21 | 1 | 1.9164 | 0.0303 | 1.4771 | 2.1761 |
| 5-TXMEA002RE | 7 | 0 | 1.915 | 0.0730 | 1.477 | 2.362 |
| 5_TXMEA003RE | 7 | 0 | 1.5558 | 0.0131 | 1.4771 | 1.7782 |


| 6_AZMCA001ID | 27 | 0 | 2.5259 | 0.1205 | 1.9031 | 3.0000 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| 6_AZMCA003ID | 27 | 0 | 2.7165 | 0.0785 | 1.8501 | 3.1461 |
| 6_AZMCA005CO | 26 | 0 | 2.3006 | 0.0643 | 1.8451 | 2.8195 |
| 6_AZMCA006RE | 20 | 0 | 2.2649 | 0.1261 | 1.6021 | 3.1761 |
| 6_AZTUA001RE | 10 | 0 | 2.2332 | 0.0702 | 1.8451 | 2.6335 |
| 6_AZTUA002RE | 8 | 0 | 1.974 | 0.113 | 1.544 | 2.556 |
| 6_AZTUA003CO | 8 | 0 | 2.3194 | 0.0519 | 1.8062 | 2.5441 |
| 6_AZTUA004ID | 7 | 0 | 2.6825 | 0.0302 | 2.5185 | 3.0414 |
| 6_CAALAL09ID | 9 | 0 | 2.5108 | 0.0354 | 2.2553 | 2.8451 |
| 7_OREUA001CO | 14 | 2 | 2.0880 | 0.0693 | 1.7324 | 2.6232 |
| 7_OREUA003RE | 15 | 0 | 1.8384 | 0.1033 | 1.4771 | 2.8129 |
| 7_ORGRA003RE | 6 | 0 | 1.6522 | 0.0216 | 1.4624 | 1.8865 |
| 7_ORGRA004CO | 6 | 0 | 1.949 | 0.0690 | 1.591 | 2.312 |
| 7_ORPOA001CO | 12 | 1 | 2.192 | 0.170 | 1.613 | 2.964 |
| 7_ORPOA003ID | 12 | 2 | 2.747 | 0.165 | 2.288 | 3.908 |
| 7_ORPOA004ID | 9 | 4 | 2.404 | 0.130 | 1.672 | 2.799 |
| 7_ORPOA006RE | 11 | 2 | 1.9330 | 0.0652 | 1.6128 | 2.4914 |
| 7_ORSAA002CO | 5 | 1 | 1.7955 | 0.00895 | 1.6532 | 1.9085 |
| 7_ORSAA003ID | 4 | 2 | 1.8372 | 0.0131 | 1.6990 | 1.9542 |
| 7_ORSAA004RE | 5 | 1 | 1.273 | 0.143 | 0.840 | 1.699 |

## Xbar-R Chart of LZN




General Linear Model: LZN versus Landuse, EPA_Rain_Zone

| Factor Type | Levels | Values |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Landuse fixed | 3 | CO, ID |  |  |  |  |
| EPA_Rain_Zone fixed | 6 | 2, 3, | 5, 6, |  |  |  |
| Analysis of Variance | for LZN, | , using | justed | for Te |  |  |
| Source | DF | Seq SS | Adj SS | Adj MS | F | P |
| Landuse | 2 | 41.3490 | 21.4483 | 10.7241 | 85.53 | 0.000 |
| EPA_Rain_Zone | 5 | 21.5544 | 19.7393 | 3.9479 | 31.49 | 0.000 |
| Landuse*EPA_Rain_Zone | 10 | 5.6755 | 5.6755 | 0.5676 | 4.53 | 0.000 |
| Error | 733 | 91.9066 | 91.9066 | 0.1254 |  |  |
| Total | 7501 | 160.4855 |  |  |  |  |

Main factors and interaction are significant.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LZN |  |  |  |  |
| All Pairwise Comparisons among Landuse $=$ CO subtracted from: |  |  |  |  |
|  |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.1634 | 0.04412 | 3.703 | 0.0007 |
| RE | -0.3098 | 0.04160 | -7.446 | 0.0000 |

Landuse $=$ ID subtracted from:

Difference SE of Adjusted

| Landuse | of Means | Difference | T-Value | P-Value |
| :---: | :---: | :---: | :---: | :---: |
| RE | -0.4732 | 0.03708 | -12.76 | 0.0000 |
| Tukey Simultaneous Tests |  |  |  |  |
| Response Variable LZN |  |  |  |  |
| All Pairwise Comparisons among Levels of LanduseLanduse $=$ CO subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| ID | 0.1634 | 0.04412 | 3.703 | 0.0006 |
| RE | -0.3098 | 0.04160 | -7.446 | 0.0000 |
| Landuse $=$ ID subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| Landuse | of Means | Difference | T-Value | P-Value |
| RE | -0.4732 | 0.03708 | -12.76 | 0.0000 |

Zinc concentration is different in commercial compared with residential and industrial. Zinc concentrations in residential are different than in industrial.

| Bonferroni Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LZN |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1102 | 0.06811 | -1.618 | 1.0000 |
| 4 | 0.3684 | 0.05363 | 6.869 | 0.0000 |
| 5 | -0.0850 | 0.03796 | -2.240 | 0.3813 |
| 6 | 0.3412 | 0.03930 | 8.680 | 0.0000 |
| 7 | 0.0691 | 0.04335 | 1.594 | 1.0000 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.47863 | 0.07986 | 5.9935 | 0.0000 |
| 5 | 0.02521 | 0.07030 | 0.3586 | 1.0000 |
| 6 | 0.45139 | 0.07103 | 6.3547 | 0.0000 |
| 7 | 0.17934 | 0.07335 | 2.4449 | 0.2209 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.4534 | 0.05638 | -8.042 | 0.0000 |
| 6 | -0.0272 | 0.05729 | -0.475 | 1.0000 |
| 7 | -0.2993 | 0.06014 | -4.976 | 0.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 6 | 0.4262 | 0.04298 | 9.916 | 0.0000 |


| 7 | 0.1541 | 0.04671 | 3.300 | 0.0152 |
| :---: | :---: | :---: | :---: | :---: |
| EPA_Rain_Zone $=6$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 7 | -0.2720 | 0.04781 | -5.690 | 0.0000 |

Zinc concentration in EPA rain zone 2 is different than in rain zones 4 and 6 . There were no differences between rain zone 2 and rain zones 3,5 and 7 .

Zinc concentration in EPA rain zone 3 is different than in rain zones 4 and 6 . There was no differences between rain zone 3 and rain zones 5 and 7 .

Zinc concentration in EPA rain zone 4 is different than in rain zones 5 and 7. There was no differences between rain zone 4 and rain zone 6 .

Zinc concentration in EPA rain zone 5 is different than in rain zones 6 and 7.
Zinc concentration in EPA rain zone 6 is different than in rain zone 7.

| Tukey Simultaneous Tests |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Response Variable LZN |  |  |  |  |
| All Pairwise Comparisons among Levels of EPA_Rain_Zone |  |  |  |  |
| EPA_Rain_Zone $=2$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA Rain Zone | of Means | Difference | T-Value | P-Value |
| 3 | -0.1102 | 0.06811 | -1.618 | 0.5864 |
| 4 | 0.3684 | 0.05363 | 6.869 | 0.0000 |
| 5 | -0.0850 | 0.03796 | -2.240 | 0.2196 |
| 6 | 0.3412 | 0.03930 | 8.680 | 0.0000 |
| 7 | 0.0691 | 0.04335 | 1.594 | 0.6024 |
| EPA_Rain_Zone $=3$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 4 | 0.47863 | 0.07986 | 5.9935 | 0.0000 |
| 5 | 0.02521 | 0.07030 | 0.3586 | 0.9992 |
| 6 | 0.45139 | 0.07103 | 6.3547 | 0.0000 |
| 7 | 0.17934 | 0.07335 | 2.4449 | 0.1410 |
| EPA_Rain_Zone $=4$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |
| EPA_Rain_Zone | of Means | Difference | T-Value | P-Value |
| 5 | -0.4534 | 0.05638 | -8.042 | 0.0000 |
| 6 | -0.0272 | 0.05729 | -0.475 | 0.9970 |
| 7 | -0.2993 | 0.06014 | -4.976 | 0.0000 |
| EPA_Rain_Zone $=5$ subtracted from: |  |  |  |  |
|  | Difference | SE of |  | Adjusted |



Zinc concentration in EPA rain zone 2 is different than in rain zones 4 and 6 . There was no differences between rain zone 2 and rain zones 3,5 and 7 .

Zinc concentration in EPA rain zone 3 is different than in rain zones 4 and 6 . There was no differences between rain zone 3 and rain zones 5 and 7 .

Zinc concentration in EPA rain zone 4 is different than in rain zones 5 and 7. There was no differences between rain zone 4 and rain zone 6 .

Zinc concentration in EPA rain zone 5 is different than in rain zones 6 and 7.
Zinc concentration in EPA rain zone 6 is different than in rain zone 7


## Interaction Plot (fitted means) for LZN




[^0]:    * Ammonia in Open Space was found in 22 events. Only 3 events had values above the detection limit

